

Hydrologic Benchmark Network Stations in the Midwestern U.S. 1963-95 (USGS Circular 1173-B)

Abstract and Map	List of all HBN	Introduction to	Analytical		
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South Hogan Creek near Dillsboro, Indiana (Station 03276700)

This report details one of the approximately 50 stations in the Hydrologic Benchmark Network (HBN) described in the four-volume U.S. Geological Survey Circular 1173. The suggested citation for the information on this page is:

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All of the tables and figures are numbered as they appear in each circular. Use the navigation bar above to view the abstract, introduction and methods for the entire circular, as well as a map and list of all of the HBN sites. Use the table of contents below to view the information on this particular station.

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Site Characteristics and Land Use

The South Hogan Creek HBN Basin is in the Central Lowlands physiographic province in southeastern Indiana (Figure 6. Map of study area in the South Hogan Creek Basin and photograph of the Whitaker Creek tributary). The 99-km² basin ranges in elevation from 174 to 303 m and drains the remnants of a broad undulating plateau

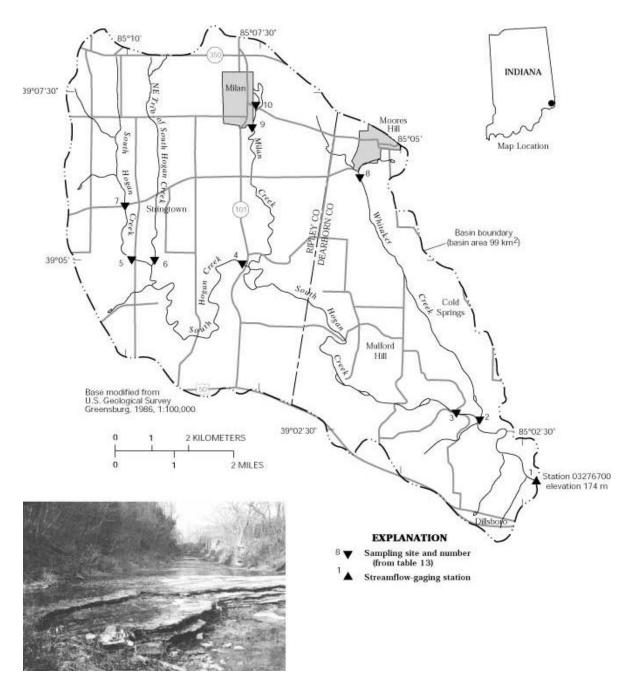


Figure 6. Map of study area in the South Hogan Creek Basin and photograph of the Whitaker Creek tributary

dissected by deeply entrenched streams forming rugged V-shaped valleys. The USGS gaging station is located 1.9 km northeast of the town of Dillsboro, Ind., at latitude 39×01'47" and longitude 85×02'17". South Hogan Creek is a southeast-flowing tributary of the Ohio River and has a channel length of about 29 km upstream from the gaging station and an average stream gradient of 4.2 m/km. The main channel is perennial, and mean monthly discharge varies from 0.23 m³/s in September to 2.8 m³/s in March; there are occasional periods of no flow during the late summer and fall (July through November). Streamflow is highly variable during storms as a result of rapid runoff from the deeply incised stream channel and rocky streambed. Average annual runoff from the basin was 39 cm from 1961 through 1993 (Stewart and others, 1994). The climate is characterized by cold winters and hot summers (McWilliams, 1985). Average daily air temperatures range from -1.9×C in January to 22.9×C in July. Average annual precipitation is 100 cm, of which about 50 percent falls between March and July. Average annual snowfall is 42 cm, and, on average, 11 days of the year have at least 2 cm of snow on the ground (McWilliams, 1985).

The South Hogan Creek Basin is in the Eastern Broadleaf Forest (Continental) ecoregion, and forest vegetation, which covers about 30 percent of the basin, is dominated by the oak-hickory association (Bailey and others, 1994). The predominant tree species are American beech, white ash, sugar maple, white oak, chinquapin oak, red oak, shagbark hickory, blue ash, tuliptree, Ohio buckeye, and black walnut (Homoya and others, 1985). Common understory species include flowering dogwood, sassafras, and hophornbeam. Soils in the basin are classified as Alfisols (Fragiudalfs, Ochraqualfs, and Hapludalfs) and are mapped in the Cobbsfork-Avonburg and Cincinnati- Rossmoyne-Hickory soil associations (McWilliams, 1985). The Cobbsfork-Avonburg soils are on the almost level to gently sloping upland plateau surface and are deep, poorly drained soils developed in loess and in silty glacial drift. A typical soil profile has a surface layer of dark grayish-brown silt loam as much as 25 cm thick overlying a subsoil of light-gray to yellowish-brown mottled silt loam and silty clay loam that extents to a depth of 200 cm. The Cincinnati-Rossmoyne-Hickory association includes well-drained soils on hillslopes that have developed in loess and in silty glacial drift. These soils range in thickness from 200 to 300 cm and often have a fragipan in the lower part of the subsoil. Soils in both associations have acidic surface layers (pH 5.1 to 6.0) that are sometimes limed to raise the soil pH for certain crops (McWilliams, 1985). The clay fraction of soils is dominated by smectite, which probably is derived from the underlying Peoria loess (Burras and others, 1996).

The basin is underlain by flat-lying marine sedimentary rocks of Ordovician age consisting of the Dillsboro Formation and Saluda Formation (Gray and others, 1972; Renn and Arihood, 1991). The Dillsboro Formation consists of argillaceous, rubbly limestone and calcareous shale, and the overlying Saluda Formation is a gray, silty dolomitic limestone. The Saluda Formation is more resistant to weathering than the Dillsboro Formation and tends to be the ridge-forming rock unit, whereas the Dillsboro Formation is the valley-forming rock unit (Renn and Arihood, 1991). In the upland areas, the bed- rock is covered by unconsolidated glacial till of late Wisconsinan age, which has an average thickness of 6 m. The till consists of clay to gravel-size deposits and, in parts of the basin, is mantled by a thin (30- to 100-cm) layer of Peoria loess (McWilliams, 1985).

The South Hogan Creek Basin drains parts of Ripley and Dearborn Counties in southern Indiana. Land ownership in the basin is 99 percent private and 1 percent county and State (Irvin Harmeyer, U.S. Department of Agriculture Natural Resources Conservation Service, written commun., 1996). More than 80 km of paved and gravel roads provide access to most areas of the basin. Access to waterways is limited to rights-of-way at public road crossings.

Current land use in the basin is about 40 percent row crops, 30 percent pasture, and 30 percent forest (Irvin Harmeyer, written commun., 1996). Agricultural lands tend to be located on the upland plateau surface, and forested areas are concentrated along the drainages. The major row crops are corn and soybeans, which are tilled conventionally (Renn and Arihood, 1991). In addition to agricultural land, the basin contains the communities of Milan (population 1,530), Moores Hill (population 650), and part of Dillsboro (population 1,200). The town of Milan has some industrial development and a wastewater-treatment facility that discharges treated water into the head-waters of Milan Creek, Moores Hill has a wastewatertreatment facility that discharges into the headwaters of Whitaker Creek. Neither treatment facility was in operation when South Hogan Creek was selected as an HBN station (Cheryl Silcox, U.S. Geological Survey, written commun., 1997). Most residents in the basin obtain domestic water from reservoirs because the bedrock and glacial deposits are poor sources of ground water (Renn and Arihood, 1991). Other human-related activities in the basin include road salting, channel-bottom excavation for fossil exploration, several livestock yards, and an active freight-hauling railroad paralleling almost the entire length of the main channel of South Hogan Creek (Cheryl Silcox, written commun., 1997). The only change in land use since the establishment of the HBN site has been a slight increase in urban development (Irvin Harmeyer, written commun., 1996).

Historical Water-Quality Data and Time-Series Trends

The data set for the South Hogan Creek HBN Station analyzed for this report includes 189 water-quality samples that were collected from October 1968 through May 1993. Samples were collected monthly from 1969 through 1982 and quarterly from 1983 through 1993. Water-quality samples in the early part of the period of record probably were analyzed at a USGS laboratory in Columbus, Ohio, that was operated until 1973 (Durum, 1978). After establishment of the central laboratory system, samples were analyzed at a laboratory in Atlanta, Ga., from 1973 through 1985 and at the NWQL in Arvada, Colo., from 1986 through 1993. Daily discharge records are available for South Hogan Creek (station 03276700) from July 1961 through September 1993, and sediment discharge records are available from 1969 through 1993.

Calculated ion balances for 186 samples with complete major-ion analyses are shown in Figures 7a and 7b. Temporal variation of discharge, field pH, major dissolved constituents, and ion balance at South Hogan Creek, Indiana. Ion balances ranged from -8.2 to 12 percent, and more than 95 percent of samples had values within the ± 5 percent range, indicating that the analytical measurements were of high quality. The average charge balance for all samples was 0.4 percent, indicating that unmeasured constituents, such as organic anions, do not seem

to contribute substantially to the ion balance of stream water at this site. No unusual patterns were evident in the time-series plots to indicate any method-related biases in the water-quality data for this HBN station (fig. 7).

Median concentrations and ranges of major constituents in stream water at the South Hogan Creek Station and VWM concentrations in wet-only precipitation measured at the Oxford NADP Station are listed in table 10. Precipitation chemistry at the NADP station, which is located about 63 km northeast of the HBN station, was dilute and acidic with a VWM pH of 4.3 for 12 years of record. The predominant cations were hydrogen, which accounted for 63 percent of the total cation charge, and ammonium, which accounted for 20 percent. Sulfate and nitrate were the predominant anions and contributed almost 95 percent of the total anion charge. The low pH and predominance of strong acid anions indicate that precipitation at the nearby NADP station probably is affected by anthropogenic emissions of sulfur and nitrogen compounds, which cause acid rain.

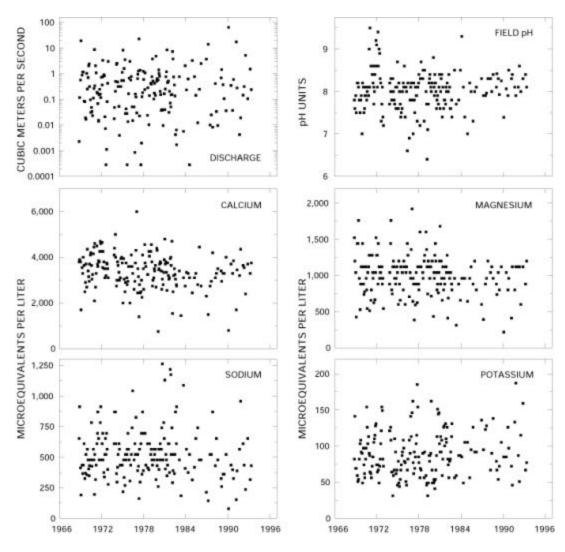


Figure 7a. Temporal variation of discharge, field pH, major dissolved constituents, and ion balance at South Hogan Creek, Indiana

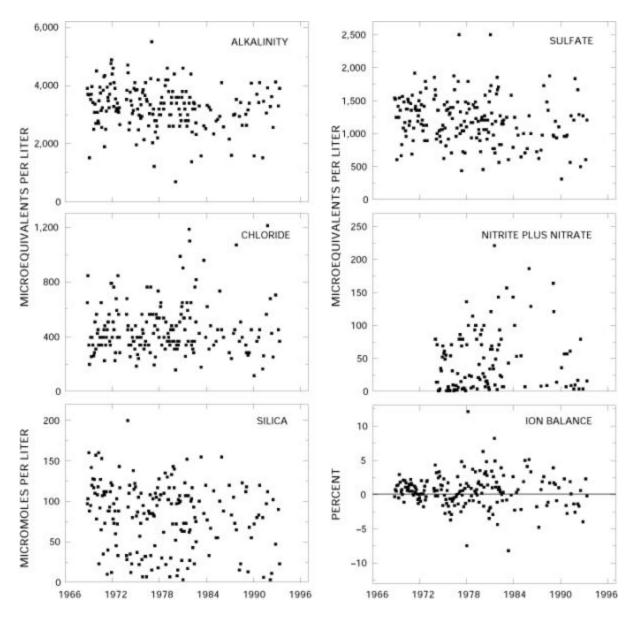


Figure 7b. Temporal variation of discharge, field pH, major dissolved constituents, and ion balance at South Hogan Creek, Indiana - Continued

Table 10. Minimum, first quartile, median, third quartile, and maximum values of physical properties and major dissolved constituents measured in water-quality samples from South Hogan Creek, Indiana, October 1968 through May 1993, and volume-weighted mean concentrations in wet precipitation collected at the Oxford Station, Ohio

[Concentrations in units of microequivalents per liter, discharge in cubic meters per second, specific conductance in microsiemens per centimeter at 25 degrees Celsius, pH in standard units, and silica in micromoles per liter; n, number of stream samples; VWM, volume-weighted mean; inst., instantaneous; spec. cond., specific conductance; <, less than; -, not reported]

D		Precipitation					
Parameter	Minimum	First quartile	Median	Third quartile	Maximum	n	VWMa
Inst. discharge	0.0003	0.045	0.25	0.76	64	187	
Spec. cond., field	120	420	470	520	720	188	28
pH, field	6.4	7.8	8.1	8.3	9.5	187	4.3 ^b
Calcium	750	3,000	3,30	3,800	6,000	189	6.9
Magnesium	220	850	1,070	1,150	1,920	189	2.0
Sodium	78	400	480	610	1,260	188	3.2
Potassium	31	67	85	110	190	188	.5
Ammonium	<.7	1.4	2.1	2.9	11	57	16
Alkalinity, laboratory	700	2,800	3,320	3,760	5,500	188	
Sulfate	310	960	1,200	1,410	2,500	189	57
Chloride	120	340	390	540	1,210	188	4.2
Nitrite plus nitrate	<.7	7.1	24	71	220	130	24 ^c
Silica	.7	55	88	110	200	189	

^a Data are volume-weighted mean concentrations for 1984—95.

^b Laboratory pH.

^c Nitrate only.

Stream water in South Hogan Creek is concentrated and strongly buffered; specific conductance ranged from 120 to 720 µmS/cm, and alkalinity was between 700 and 5,500 meg/L (table 10). The major ions in stream water were calcium and alkalinity, which accounted for almost 70 percent of the total ionic charge. The high concentrations of these ions in stream water are attributed to the dissolution of carbonate minerals in the limestone bedrock and glacial till. The median chloride concentration in stream water was 390 meg/L compared to the VWM concentration of 4.2 meg/L in precipitation. This large difference in concentration indicates that most stream-water chloride is derived from sources in the basin. Although some chloride may be derived from bedrock sources, most is probably derived from human activities. A number of point and nonpoint sources of chloride exist in the basin, including discharge from two wastewater-treatment plants and road salt and fertilizer applications. Sulfate concentrations also were considerably higher in stream water than in precipitation; the median stream-water concentration was 1,200 meg/L compared to 57 meg/L in precipitation. The main source of stream-water sulfate probably is weathering of calcium sulfate minerals, such as gypsum and anhydrite in the marine sedimentary rocks. The median nitrate concentration of stream water was similar to precipitation, and stream-water ammonium concentrations were considerably lower. Although these stream-water nitrogen concentrations are low compared to concentrations considered indicative of human-related pollution (Kross, 1990), the nitrate concentrations were substantially higher than at most other stations in the HBN, indicating human activities have some effect on nutrient concentrations in stream water at this station.

The solute composition of stream water was further evaluated by analyzing correlations between solutes and stream discharge (table 11). Most solutes were weakly correlated with discharge, except for sodium (rho = -0.733), chloride (rho = -0.669), and nitrogen (rho = 0.703). For the solutes, the strongest correlations were found among the weatheringderived solutes, particularly calcium, magnesium, and alkalinity. The strong inverse correlations of sodium and chloride with discharge simply may be due to dilution of wastewater discharge during periods of increased discharge. Alternatively, these correlations may indicate that human activities have resulted in the contamination of the local groundwater system. Although nitrate also is derived primarily from human-related sources, it had a positive correlation with discharge. This indicates that nitrate primarily is transported during periods of high flow, probably due to surface runoff from agricultural lands and livestock yards. The strong correlations among calcium, magnesium, and alkalinity reflect the weathering of carbonate minerals, but the lack of correlation with discharge was somewhat unexpected. A more detailed examination of the data revealed that the concentrations of these constituents were, in fact, related to discharge, but the relation was not monotonic. Concentrations had the expected inverse correlation with discharge, but only for discharges greater than 0.03 m³/s. At flows less than 0.03 m³/s, stream-water concentrations were fairly constant as a function of discharge, perhaps reflecting a uniform source of streamflow from the bedrock aquifer.

Table 11. Spearman rank correlation coefficients (rho values) showing the relation among discharge, pH, and major dissolved constituents, South Hogan Creek, Indiana, 1969—93

[Q, discharge; Ca, calcium; Mg, magnesium; Na, sodium; K, potassium; Alk, alkalinity; SO₄, sulfate; Cl, chloride; N, nitrite plus nitrate; Si, silica]

	Q	рН	Ca	Mg	Na	K	Alk	SO ₄	C1	N
pН	0.345									
Ca	250	0.303								
Mg	333	.252	0.866							
Na	733	206	.416	0.555						
K	550	320	.033	052	0.487					
Alk	264	.250	.914	.846	.388	-0.010				
SO ₄	429	.224	.728	.734	.617	.145	0.628			
Cl	669	164	.406	.393	.871	.639	.354	0.569		
N	.703	.122	020	230	426	316	126	216	-0.292	
Si	.227	139	286	521	287	.221	358	399	126	0.478

Results of the seasonal Kendall test for trends in discharge and major dissolved constituents are listed in table 12. No statistically significant trends were observed at the South Hogan Creek Station at the 0.01 probability level. Upward trends in flow-adjusted pH and sulfate concentrations were detected at a slightly less significant level of p \pm 0.02. The lack of trends in most dissolved constituents, particularly sodium, chloride, and nitrate, indicates that land-use activities in the basin, such as road salting, agricultural runoff, and wastewater discharge, have remained fairly constant during the period of record. The source of trends in flow-adjusted pH and sulfate concentrations are less clear. Atmospheric deposition of sulfate has decreased substantially in the industrialized Midwest since the 1970's (Husar and others, 1991); however, changes in precipitation chemistry probably would not have much of an effect on stream chemistry in this basin because of the large internal source of sulfate from weathering. The trends in sulfate and pH also might be caused by changes in land use; however, there are few human-related activities that would impact pH and sulfate concentrations without also affecting the concentrations of the other dissolved constituents, particularly chloride, sodium, and nitrate.

Table 12. Results of the seasonal Kendall test for trends in discharge and unadjusted and flow-adjusted pH and major dissolved constituents, South Hogan Creek, Indiana, October 1968 through May 1993

[Trends in units of microequivalents per liter per year, except discharge in cubic meters per second per year, pH in standard units per year, and silica in micromoles per liter per year; inst., instantaneous; <, less than; --, not calculated]

Parameter	Unadju	sted	Flow adjusted		
T drameter	Trend	p-value	Trend	p-value	
Discharge	-0.001	0.294			
pН	.02	.025	0.02	0.015	
Calcium	-4.4	.446	-3.1	.649	
Magnesium	<.1	.225	-1.8	.221	
Sodium	.1	.698	-1.4	.494	
Potassium	.6	.064	.6	.113	
Alkalinity	-3.3	.688	1.0	.887	
Sulfate	-7.7	.111	-7.6	.020	
Chloride	1.8	.331	2.6	.139	
Nitrite plus nitrate	.3ª	.763			
Silica	6	.108	8	.053	

^a Trend calculated for 1974—93 using a trend test for censored data.

Synoptic Water-Quality Data

Results of the surface-water synoptic sampling in the South Hogan Creek Basin on November 5 and 6, 1991, are listed in table 13, and the locations of the sampling sites are shown in figure 6. During the sampling period, discharge at the gaging station was about 0.01 m³/s compared to the median discharge of 0.15 m³/s for November (Lawrence, 1987), indicating that the basin was sampled during low-flow conditions for that time of year. Because of the low-flow conditions, solute concentrations measured at site 1 were greater than the third-quartile concentrations reported for the HBN station during the entire period of record (table 10). The tributary streams were similar in composition to stream water collected at the gaging station (site 1); calcium and magnesium were the predominant cations, and bicarbonate and sulfate were the predominant anions. Ion balances for the synoptic samples were around zero (range -2.4 to 2.4 percent), indicating that unmeasured constituents, such as organic anions, did not seem to be an important component of stream water during the sampling period.

Considerable spatial variability in stream chemistry was measured in the South Hogan Creek Basin, particularly for sodium, chloride, and nitrate, which seems to be related to wastewater discharge from treatment facilities in the towns of Milan and Moores Hill. For example, chloride concentrations in tributaries sampled downstream from the towns of Moores Hill (site 8) and Milan (site 9) were 1,940 and 2,510 meg/L (table 13), respectively, compared to the average concentration of 580 meg/L at background sites in the basin (sites 3–7, 10). Sodium concentrations were 3,740 and 3,000 meg/L at sites 8 and 9, respectively, compared to the average background concentration of 650 meg/L. The most substantial difference in concentration was measured for nitrate, which had a concentration of 590 meg/L at site 8 and 1.710 meg/L at site 9 compared to the average background concentration of less than 10 meq/L. The synoptic samples also demonstrate that the water quality downstream at the HBN station was impacted by the wastewater-treatment facilities based on the elevated chloride concentration measured at site 1. By contrast, the nitrate concentration measured at site 1 was similar to background concentrations, implying that, at least during low-flow conditions, the biota were capable of consuming excess nitrate upstream from the HBN station. In contrast to sodium, chloride, and nitrate, the concentration patterns of the weathering-derived constituents were more uniform across the basin. For example, calcium concentrations ranged from 1,850 to 4,600 meg/L, and magnesium ranged from 620 to 2,000 meg/L. This pattern not only reflects the widespread presence of carbonate minerals in the glacial till and bedrock, but indicates that the towns of Milan and Moores Hill did not greatly affect the concentrations of these constituents in surface water. Sulfate concentrations were much lower in tributaries in the western one-half of the basin compared to the eastern one-half. For example, the average sulfate concentration at sites 4, 5, 6, and 7 was 400 meg/L compared to the average concentration of 1,740 meg/L at sites 2, 3, 8, 9, and 10. One explanation for this spatial pattern is a difference in bedrock mineralogy between these two areas of the basin. The streams in the northwestern part of the basin drain areas covered by glacial till, whereas drainages in the southeastern part of the basin are incised into the Ordovician bedrock, which, in places, contains evaporite beds of gypsum and anhydrite.

Table 13. Physical properties and major dissolved constituents from surface-water sampling sites in the South Hogan Creek Basin, Indiana, collected November 5—6, 1991

[Site locations shown in fig. 6; Q, discharge in cubic meters per second; SC, specific conductance in microsiemens per centimeters at 25 degrees Celsius; Ca, calcium; Mg, magnesium; Na, sodium; K, potassium; Cl, chloride; NO_3 , nitrate; SO_4 , sulfate; Alk, alkalinity; Si, silica; concentrations in microequivalents per liter, except silica in micromoles per liter; WTP, wastewater-treatment plant; –, not reported; criteria used in selection of sampling sites: $BG = bedrock \ geology$, $MT = major \ tributary$, $LU = land \ use$]

Site	Station number	Q	SC	pН	Ca	Mg	Na	K	Alk	SO ₄	Cl	NO ₃	Si	Criteria	Remarks
1	3276700	0.0096	720	8.2	4,650	1,200	1,090	160	4,000	2,040	1,320	0.7	110		Gaging station
2	390241085031600		670	8.5	4,600	1,280	830	62	3,540	2,710	850	.4	120	МТ	Downstream from Moores Hill
3	390246085034400	.0079	650	8.1	3,800	1,040	1,130	200	3,700	1,440	730	.6	87	MT, BG	Limestone bedrock
4	390456085074700		590	7.6	3,900	1,200	370	250	4,700	400	560	3.4	170	MT, BG	Glacial till
5	390500085095100		660	8.2	3,500	2,000	780	160	5,300	500	730	3.5	160	BG	Glacial till
6	390502085092800		320	7.8	1,850	620	130	200	2,480	180	160	.6	63	BG	Glacial till
7	390550085100300		640	7.9	3,350	2,000	870	460	5,620	520	510	3.4	230	BG	Glacial till
8	390620085053700	.0017	100	8.0	4,100	1,680	3,740	240	5,680	1,650	1,940	590	270	LU	Downstream from WTP
9	390706085073700	.0059	830	7.4	2,650	960	3,000	440	1,520	1,480	2,510	1,710	120	LU	Downstream from WTP
10	390715085073500	.0006	560	7.7	3,300	1,120	610	150	2,780	1,400	760	29	110	LU	Upstream from WTP

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Renn, D.E., and Arihood, L.D., 1991, Sedimentation in Versailles Lake, Ripley County, southeastern Indiana, 1956–88: U.S. Geological Survey Water-Resources Investigations Report 92–4011, 63 p.

Stewart, J.A., Keeton, C.R., Benedict, B.L., and Hammil, L.E., 1994, Water resources data, Indiana, water year 1993: U.S. Geological Survey Water-Data Report IN–93–1, 344 p.

Appendix A. List of Map References

- a. U.S. Geological Survey topographic maps:
 - Dillsboro, Indiana (1:24,000), 1980, gaging station on this quadrangle
 - Milan, Indiana (1:24,000), 1980
 - Pierceville, Indiana (1:24,000), 1980
 - Greensburg, Indiana (1:100,000), 1986

b. Geologic maps:

• Gray, H.H., Forsyth, J.L., Schneider, A.F., and Gooding, A.M., 1972, Regional geologic map no. 7, Cincinnati sheet, part B: Bloomington, Indiana Geological Survey, 1 sheet, scale 1:250,000.

c. Soil surveys:

- McWilliams, K.M., 1985, Soil survey of Ripley County and part of Jennings County, Indiana: U.S. Department of Agriculture Soil Conservation Service, 125 p.
- Nickell, A.K., 1981, Soil survey of Dearborn and Ohio Counties, Indiana: U.S. Department of Agriculture Soil Conservation Service.

d. Miscellaneous maps:

• U.S. Department of Agriculture, 1986, Important farmland, Ripley County, Indiana: U.S. Department of Agriculture Soil Conservation Service, scale 1:50,000.

Appendix B. NWIS Site-Identification Numbers

Table B–1. NWIS site-identification numbers and site names for water-quality sampling sites.

Site	Identification Number	Site Name
1	03276700	SOUTH HOGAN CR NR DILLSBORO IN
2	390241085031600	WHITAKER CR ABOVE DILLSBORO IN
3	390246085034400	S HOGAN CR AB CHANCE BRANCH NR DILLSBORO IN
4	390456085074700	S HOGAN CR AT HWY 101 NR MILAN IN
5	390500085095100	S HOGAN CR AT 525 EAST RD NR STRINGTOWN IN
6	390502085092800	NORTHEAST TRIB OF S HOGAN CR NR STRINGTOWN IN
7	390550085100300	S HOGAN CR AB HOG FARM NR STRINGTOWN IN
8	390620085053700	WHITAKER CR BLW MOORES HILL IN
9	390706085073700	MILAN CR BELOW WWTP AT MILAN IN
10	390715085073500	MILAN CR AB WWTP AT MILAN IN

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