



Biological Assessment of Environmental Flows for Oklahoma

By William L. Fisher, Titus S. Seilheimer, and Jason M. Taylor

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Conversion Factors

Inch/Pound to SI

Multiply	By	To obtain
Length		
inch (in.)	25.4	millimeter (mm)
foot (ft)	0.3048	meter (m)
mile (mi)	1.609	kilometer (km)
Area		
square mile (mi ²)	2.590	square kilometer (km ²)
Flow rate		
cubic foot per second (ft ³ /s)	0.02832	cubic meter per second (m ³ /s)
cubic foot per second per square mile [(ft ³ /s)/mi ²]	0.01093	cubic meter per second per square kilometer [(m ³ /s)/km ²]

Temperature in degrees Celsius (°C) may be converted to degrees Fahrenheit (°F) as follows:

$$^{\circ}\text{F}=(1.8\times^{\circ}\text{C})+32$$

Temperature in degrees Fahrenheit (°F) may be converted to degrees Celsius (°C) as follows:

$$^{\circ}\text{C}=(^{\circ}\text{F}-32)/1.8$$

Vertical coordinate information is referenced to the North American Vertical Datum of 1988 (NAVD 88).

Horizontal coordinate information is referenced to the North American Datum of 1983 (NAD 83).

Altitude, as used in this report, refers to distance above the vertical datum.

*Transmissivity: The standard unit for transmissivity is cubic foot per day per square foot times foot of aquifer thickness [(ft³/d)/ft²ft]. In this report, the mathematically reduced form, foot squared per day (ft²/d), is used for convenience.

Specific conductance is given in microsiemens per centimeter at 25 degrees Celsius (μS/cm at 25 °C).

Concentrations of chemical constituents in water are given either in milligrams per liter (mg/L) or micrograms per liter (μg/L).

Abbreviations

BCD	Bray-Curtis dissimilarity
HIP	hydroecological integrity assessment process
HIT	hydrologic index tool
IHA	indicators of hydrologic alteration
NMS	nonmetric multidimensional scaling
RVA	range of variability
USGS	U.S. Geological Survey

Biological Assessment of Environmental Flows for Oklahoma

By William L. Fisher,¹ Titus S. Seilheimer,² and Jason M. Taylor²

Abstract

Large-scale patterns in fish assemblage structure and functional groups are influenced by alterations in streamflow regime. In this study, we defined an objective threshold for alteration for Oklahoma streams using a combination of the expected range of 27 flow indices and a discriminant analysis to predict flow regime group. We found that fish functional groups in reference flow conditions had species that were more intolerant to flow alterations and preferences for stream habitat and faster flowing water. In contrast, altered sites had more tolerant species that preferred lentic habitat and slower water velocity. Ordination graphs of the presence and functional groups of species revealed an underlying geographical pattern roughly conforming to ecoregions, although there was separation between reference and altered sites within the larger geographical framework.

Additionally, we found that reservoir construction and operation significantly altered fish assemblages in two different systems, Bird Creek in central Oklahoma and the Kiamichi River in southeastern Oklahoma. The Bird Creek flow regime shifted from a historically intermittent stream to one with stable perennial flows, and changes in fish assemblage structure covaried with changes in all five components of the flow regime. In contrast, the Kiamichi River flow regime did not change significantly for most flow components despite shifts in fish assemblage structure; however, most of the species associated with shifts in assemblage structure in the Kiamichi River system were characteristic of lentic environments and were likely related more to proximity of reservoirs in the drainage system than changes in flow. The spatial patterns in fish assemblage response to flow alteration, combined with different temporal responses of hydrology and fish assemblage structure at sites downstream of reservoirs, indicate that interactions between flow regime and aquatic biota vary depending on ecological setting. This supports the notion that regional variation in natural flow regimes could affect the development of flow recommendations.

Introduction

Oklahoma has a wide range of climatic and environmental conditions that create variation in stream hydrology and water availability across the State. Climate ranges from humid subtropical in the east to semiarid in the west, and these climatic differences influence environmental conditions and the flow regime of streams and rivers throughout the State (Turton

¹ U.S. Geological Survey.

² Cornell University.

and others, 2009). Environmental flows are defined as the quantity, timing, and quality of water flows required to sustain freshwater and estuarine ecosystems and the human livelihood and well-being that depend on these ecosystems (Poff and others, 2010). A better understanding of environmental flows and how flow regimes influence aquatic ecosystems in Oklahoma can help assess the impacts of flow-related changes on aquatic organisms, such as fish assemblages. Such an assessment could inform water management in Oklahoma under current and future conditions.

The hydrologic flow regime has a direct influence on the structure and function of aquatic ecosystems and is the primary driver of ecological integrity of a river ecosystem. Alteration of the hydrologic regime of rivers from impoundments and flow diversions modifies the structure and function of river ecosystems (Poff and others, 1997; Rosenberg and others, 2000; Postel and Richter, 2003; Poff and others, 2007). Hydrologic alterations—such as flow stabilization, prolonged low flows, loss of seasonal flow peaks, rapid changes in river stage, and low or high water temperatures downstream—disrupt life cycles of aquatic plants, invertebrates, and fishes, resulting in a reduction in species diversity and modifying reproduction and growth rates that oftentimes lead to local extinctions of native species and the invasion and establishment of exotic species (Poff and others, 1997). Large water diversions deplete streamflows, sometimes to damaging levels, affecting aquatic and floodplain habitats, aquatic biodiversity, sport and commercial fisheries, natural floodplain fertility, and natural flood control (Postel and Richter, 2003). The development of water resources to meet the demands of urban population centers is growing and threatens the ecological integrity of many freshwater ecosystems (Fitzhugh and Richter, 2004).

Alterations in the flow regime of streams can be reflected in the composition of the aquatic biota (Poff and Zimmerman, 2010), which makes flow regime a useful indicator of ecological integrity. Procedures and tools for measuring flow regime and alteration of flow are available through the software Indicators of Hydrologic Alteration (IHA; Richter and others, 1996; Matthews and Richter, 2007) and Hydroecological Integrity Assessment Process (HIP; Olden and Poff, 2003; Henriksen and others, 2006). A classification of unaltered streams using the HIP approach was used for Oklahoma streams and indicated that there were four general types of flow regime, describing a wide range of flow conditions: perennial runoff, perennial flashy, stable groundwater, and intermittent (Turton and others, 2009). These baseline conditions provide a reference to determine the level of alteration that has occurred in recent times and to relate to changes in biotic communities.

Individual fish species and fish assemblages respond differently to changes in flow regime. Poff and Allan (1995) found that the functional organization of fish assemblages was influenced by the degree of flow stability in a stream, whereas more variable flow regimes resulted in shifts to more generalist strategies (for example, habitat use, feeding) by the fish. Hypotheses of how ecological communities respond to hydrologic alteration provide a useful framework for assessing changes in fish assemblages to alterations in flow. In a review of the impacts of altered flow on aquatic biodiversity, fish abundance and diversity were negatively influenced by both elevated and decreased flow magnitude (Poff and Zimmerman 2010). The authors called for better-defined thresholds for flow alteration-ecological response relationships. This study addresses this need for rivers in Oklahoma.

Improving our understanding of the response of aquatic organisms to flow alterations is an important step in the development of water resource management that incorporates protection of Oklahoma's aquatic resources. Our objectives were to (1) quantify the thresholds of alteration for important flow indices, (2) test the configuration of functional fish groups in relation to flow

alteration, and (3) identify patterns in fish assemblages associated with flow regime in all sites. Finally, we provide a detailed example of the impacts of impoundments on downstream flow regimes and resulting changes in fish assemblages at two sites in Oklahoma, the Kiamichi River and Bird Creek.

Methods for Collecting Data

Streamgage Selection and Streamflow

We chose 168 streams with gages draining areas smaller than 2,500 square miles in or adjacent to Oklahoma (Esralew, 2010) for our original analysis of reference streams. From this larger set of gaged streams, we chose a subset of 88 streams with a known reference flow regime and that were previously classified into streamflow groups (Turton and others, 2009). These streams and their previous classifications provided a benchmark to determine if their flow regime had been altered. To minimize the effect of climatic variation on localized streamflow, we selected streamgages that monitored flow for a concurrent period. A 15-year minimum period of flow data and at least 50 percent concurrent period is recommended to correctly characterize flows (Kennard and others, 2010). We selected a 20 concurrent year period (1988–2007), which reduced the number of potential streamgages from 88 to 41. Streamgage information (daily mean and peak flows) was obtained from the U.S. Geological Survey (USGS) National Water Information System (<http://waterdata.usgs.gov/nwis/>) for full water years (October–August) from 1988 through 2007. This recent period was selected because it was more likely to coincide with available fish data. We used the Hydrologic Index Tool (HIT, version 1.48; U.S. Geological Survey, undated) to calculate 171 flow indices for each study stream (Turton and others, 2009). These flow indices describe the entire flow regime, including magnitude of average, low, and high flows, frequency of low and high flow, duration of low and high flow, timing of flow events, and rate of change (Olden and Poff, 2003).

Flow Alteration

We used two statistical techniques to quantify alteration in Oklahoma streams. In the first, we fit distributions and identified the 5th and 95th percentiles (Arthington, 2006) for the reference flow period of the four flow groups reported by Turton and others (2009) for 88 gaged streams. The group notation adopted by Turton and others (2009) was retained (41, 42, 43, and 44), which corresponded to general flow regime descriptions of perennial runoff (41), perennial flashy (41), stable groundwater (43), and intermittent (44). The 27 flow indices we used (table 1) were identified as the best combination of variables from the available 171 flow indices to categorize the regions flow types. Six types of distribution—normal, lognormal, exponential, Weibull 2 parameter, Weibull 3 parameter, and Johnson Su—were fit in JMP (version 7; SAS Institute Inc., undated). The best distribution for each flow group was determined by a combination of *p*-value and visual inspection of the fit curve. The distributions were least reliable for group 43, which had only 6 member streams but had a distinctive flow regime compared with the other groups, so the distributions for group 43 were retained.

The second method used to quantify alteration was discriminant analysis. Discriminant analysis is a technique that assigns variables to pre-existing classes or groups using multiple explanatory variables (for example, flow indices; Gotelli and Ellison, 2004). We developed an algorithm to predict group classification using the same 27 flow indices as in the first method

(table 1). The algorithm was then used to assign the streams of interest in this study to one of four flow groups. We noted when a stream changed membership in recent period because it indicated a large change in flow regime. It was also possible that a stream could not be classified into one of the four groups, which would indicate that the flow regime has been altered and does not reflect the four Oklahoma flow regimes identified by Turton and others, (2009). The current streams were assigned to reference or altered classes based on these two techniques.

Fish Data

The primary source of fish species data was from a database of Oklahoma fishes (E. Tejan, Enogex, LCC, written commun. , 2009), which contained records from the Oklahoma Conservation Commission, the Oklahoma Department of Environmental Quality, the U.S. Forest Service, and published records. Additional fish data were provided by the Kansas Department of Wildlife and Parks, the Missouri Resource Assessment Partnership, and The Nature Conservancy of Arkansas. A 15-kilometer (km) buffer (that is, radius; Poff and Allan, 1995) from each streamgage location was used to select nearest fish sampling locations in close proximity to USGS streamgages using ArcGIS (version 8.3; Environmental Research Systems Institute, Inc., undated). Only sites on the main stem of the gaged river were selected (Poff and Allan, 1995). There were 28 streamgage locations with sufficient fish data within the 15-km buffer (table 2; fig. 1). The selected streams fall within two major river basins, the Arkansas River to the north and the Red River to the south. The streams can also be divided between the wetter eastern portion of Oklahoma and drier western portion by running a vertical line to the east of streamgage 10 and west of streamgage 18 (fig. 1). A total of 150 species were found in the waters near the 28 streamgage sites (appendix 1).

Response of Fish Assemblages to Flow Alteration

Fish data were collected by different sources and with different methods, therefore we limited analysis to fish presence and the proportion of fish functional groups at each site. The coarse-scale nature of the data allowed us to identify large-scale patterns that affect Oklahoma at a regional level (Poff and Allan, 1995). Functional groupings reflect preferences for a narrow range of habitat conditions or tolerance to alteration in their environment. Fish species were assigned to groups in five functional categories: habitat type (creek, river, creek/river, lentic, general), flow velocity, trophic guild, reproductive guild, and tolerance (tolerance to water quality degradation, tolerance to habitat degradation, sensitive benthic species, conservation status). Species were assigned to functional groups based on published classes, primarily developed specifically for Oklahoma species (Jester and others, 1992), but also included additional sources for current speed (Frimpong and Angermier, 2009) and conservation status (NatureServe, 2009). Functional groups that require additional explanation are lithophilic species, reproductive guilds, and those based on Oklahoma State conservation status. Lithophilic species are species that spawn on gravel or cobble. Reproductive guilds were based on Goldstein and Simon's (1998) classification and were as follows: A_1, nonguarders, open substratum spawners; A_2, nonguarders, brood hidiers; B_1, guarders, substratum choosers; B_2, guarders, nest spawners; and C_2, internal live bearers. Conservation status of species in Oklahoma is as follows: S1, critically imperiled; S2, imperiled; S3, vulnerable; S4, apparently secure; S5, secure; exotic (NatureServe, 2009). Species in multiple categories were assigned to the dominant classification (Goldstein and Simon, 1999).

We analyzed patterns in fish assemblage structure among sites representing a range of flow alteration using ordination analysis followed by comparison of functional groups among alteration classes. Differences in assemblage structure among stream sites were quantified using nonmetric multidimensional scaling (NMS) ordinations (Minchin, 1987; Clarke, 1993) on the presence or absence of fish species with Jaccard's similarity as the distance measure. Jaccard's similarity compares homogeneity in groups, with greater distance between points indicating less similarity (McCune and Grace, 2002). Principal components analysis of the correlations was used to identify patterns in the proportions of functional groups among stream sites using the broken stick method (that is actual eigenvalues higher than eigenvalues from random variation) to identify significant components (McCune and Grace, 2002). Rank correlations between flow variables and ordination axes were calculated to determine relationships between fish assemblage structure and flow components. Differences in proportions of functional groups between altered and reference sites were tested with analysis of variance (ANOVA) in JMP. Multivariate analysis of species presence and functional groups were done in PC-ORD (version 4; Peck, 2010).

Temporal Response of Fish Assemblage Structure to Flow Alteration

We analyzed trends in hydrologic parameters associated with different periods related to reservoir construction within the Kiamichi River and Bird Creek watersheds in Oklahoma (fig. 2). Daily discharge records were obtained from USGS streamgages located downstream of reservoirs on each stream. Discharge in the Kiamichi River is influenced by Jackfork Creek, a tributary that was impounded in 1982 to form Sardis Lake. Changes in Kiamichi River flows were assessed using the combined records of USGS streamgages 07336200 and 07336500, both located near Antlers, Okla., downstream of Sardis Lake and upstream of Hugo Lake. Daily flow records from two gages were used to incorporate a period of record long enough to estimate pre-dam conditions (more than 20 years) (Kennard and others, 2010). The combined record spanned 81 years (1926–2007) and included daily flow data that corresponded with available fish data (1977–1996). To assess changes in hydrology within Bird Creek, we used USGS streamgage 07177500 located downstream of Birch and Skiatook Lakes near Sperry, Okla. The gage record spanned 49 years (1938–2007) and included daily flow data that corresponded with fish data (1978–1996) collected on Bird Creek. Discharge in Bird Creek is influenced by a tributary that was impounded in 1977 to form Birch Lake and a second reservoir on the main stem, Skiatook Lake, which was impounded in 1984.

We used a dataset consisting of fish collections from about 40 different sampling events at the same location on each stream to assess temporal changes in fish assemblage structure between 1977 and 1996. Fish sampling was conducted by the Oklahoma Department of Environmental Quality using a standardized sampling effort of twenty 10-meter (m) seine hauls made with a 3-m \times 1.5-m (4.7-millimeter mesh) lead weighted seine. Sampling effort was approximately 1 hour per sampling event. We limited data for analysis to collections that were made in early summer (June and July) because this period was sampled every year, whereas late summer and early fall sampling effort varied among years. When multiple early summer sampling events occurred within one year, we kept both events to include potential effects of within-year variation in the analysis. We excluded data for the year preceding and the year of impoundment to remove any potential effects of dam construction not related to flow alteration from the dataset. This resulted in a total of 26 (Kiamichi River) and 23 (Bird Creek) sampling events, representing 17 different summers that we retained for analysis.

We used the IHA software (version 7.1; ConserveOnline, undated) to assess potential differences in hydrologic conditions between pre- and post-impoundment periods for each stream. The pre-impoundment period was between 1926 and 1980 for the Kiamichi River, after which flows were potentially altered by Sardis Lake reservoir operations (1983–2007). Bird Creek had three periods, including the pre-impoundment (1958–1976), post-Birch Lake impoundment (1978–1982), and post-Skiatook Lake impoundment (1985–2007) periods.

The IHA method is based on 33 biologically relevant hydrologic parameters divided among five fundamental characteristics of hydrologic regimes that represent temporal changes in water conditions (for example, water levels, discharge rates; Richter and others, 1996). These characteristics include the magnitude, timing of occurrence, frequency of occurrence, duration of time, and the rate of change in specific water conditions. We used the nonparametric statistics option to compare median and coefficient of variation values between pre- and post-impoundment periods using a permutation procedure (1,000 permutations) to identify significant differences between periods. We also used the range of variability (RVA) approach to assess changes in IHA parameters between periods (Richter and others, 1997). The RVA approach divides the data from the period before the event for each parameter into three different categories based on the 17th percentiles from the median yielding three categories of equal size. We used RVA to compare medians between pre- and post-impoundment periods in relation to RVA boundaries calculated for the pre-impoundment period.

Patterns in fish assemblage structure between sampling events were quantified using NMS ordinations (Minchin, 1987; Clarke, 1993) on $\log_{10}(x)$ -transformed abundances of fish species with Bray-Curtis dissimilarity (BCD) as the distance measure. Prior to analysis, fish species occurring in less than two sampling events (< 5 percent) were excluded from the dataset to reduce noise (McCune and Grace, 2002). We used rotational vector fitting to relate hydrologic parameters from IHA to gradients in fish assemblage structure quantified by the NMS ordination (Faith and Norris, 1989). Vector fitting finds the direction of the maximum correlation for each environmental variable and assesses significance of the fitted environmental vectors using a permutation procedure (1,000 permutations) (Oksanen and others, 2010). Some hydrologic variables were $\log_{10}(x)$ -transformed to improve assumptions of normality (McCune and Grace, 2002). Ordination plots were rotated to have the strongest correlation with environmental flow vectors on the horizontal x-axis (axis 1) of the plots. We also examined the response of downstream fish assemblages to impoundment by comparing centroids and dispersion of sampling events within pre- and post-impoundment groups within ordination space using permutational multivariate analysis of variance (PERMANOVA) and permutational analysis of multivariate dispersion (PERMDISP; Anderson 2001, 2006). We assigned sampling events to pre-impoundment and post-impoundment periods for both the Kiamichi River and Bird Creek. An additional sampling period, transitional, was added between the pre-impoundment and post-impoundment periods to assess differences in fish assemblage structure among these three periods.

We used Indicator Species Analysis (ISA; Dufrêne and Legendre, 1997) to test for affinities of different species to our period grouping. ISA measures association between species and predetermined groups by calculating an indicator value (IndVal) based on species abundances weighted by their occurrence within each group. We estimated the probability of achieving an equal or larger IndVal value among groups (p) using a bootstrap method based on 999 random permutations of the original data. All ordination analysis was conducted in R statistical language, version 2.11.2 (R Core Development Team, 2010) using the vegan package

for NMS, PERMANOVA, and PERMDISP (R Core Development Team, 2010; Oksanen and others, 2011). Indicator species analysis was performed using PC-ORD (version 5.0; McCune and Grace, 2002).

Quantifying Flow Alteration

The 20-year period of flow data for 14 streams corresponded to the reference period used in the earlier classification; these data are referred to as “reference” in table 2 (Turton and others, 2009). Streams were considered to have an altered flow regime when upstream anthropogenic regulation of flows, such as impoundments or water withdrawal, were present and when the period of the streamflow record we analyzed was not in the least-altered period (Esralew, 2010). The expected range (5th and 95th percentile; table 3) of flow indices differed greatly between flow regime groups. We compared index values for streams with the expected range and counted the number exceeding the expected range of threshold and average percent exceeded (table 4). The reference streams had lower numbers and percentages of flow indices exceeding the expected values. The mean percent exceeding was higher in the altered streams than the reference streams.

The discriminant analysis using 27 flow metrics correctly classified 86 of the 88 streams into the four flow groups. The discriminant analysis algorithm was then used to predict flow regime groups for the 28 study streams. The reference streams were all classified correctly, while only 5 of 14 altered streams were the same group as before discriminant analysis (table 4). Three streams shifted from group 42 to 43, while one stream in group 42 did not classify into any of the defined four groups. Five streams in group 44 changed to 41, 42, and 43. The least common flow regime group in Turton and others (2009) was 43 (stable groundwater; six streams), but it was the most common group that altered streams fell under because of elevated flows, which indicates there was an increase in flow and stability.

Fish Assemblage Structure

Nonmetric multidimensional scaling (NMS) ordination of 28 fish sampling sites identified two axes (2D) that explained 85.8 percent of the variation in original distances among fish sampling sites (2D stress = 13.5). NMS axis 1 (x axis) represented changes in species composition along a gradient of decreasing variability in daily flows (MA04), variability in May flows (MA28), frequency of low flow spells (FL03), high flood pulse count (FH04), variability in annual maximums of 90-day means of daily discharge (DH10), constancy (TA01), and fall rate (RA03). NMS axis 2 (y axis) represented changes in species composition related to increasing variability in annual maximums of 3-day means of daily discharge (DH07), DH10, and RA03 (table 5). These changes in flow environments appeared to represent geographic variation in flow more than flow alteration, as sites were primarily ordinated based on location rather than alteration, although eastern Oklahoma sites show some difference between altered and reference samples (fig. 3). Streams grouped by basin were clearly separated from each other, particularly the eastern basins from the western streams. Altered sites within the two eastern basins tended to shift away from reference sites along a trajectory correlated with increasing MA04, MA28, FL03, FH04, and TA01 (fig. 3; table 5). In contrast, there was considerable overlap between reference and streams with altered hydrology among western streams (fig. 3). In general, westerns stream represented fish assemblages associated with higher DH10 and RA03 (fig. 3; table 5).

Principal components analysis of fish functional group proportions at sites identified two significant principal components (fig. 4). These two axes explained 63.2 percent of the variation in the data. There were nine indices positively correlated with PC1 (MA04, MA28, specific mean annual maximum flows (MH20), FL03, flood frequency (FH05), number of zero-flow days (DL18), DH10, TA01, RA03, and five indices negatively correlated with PC2 [mean daily flow (MA01), mean minimum January flow (ML01), mean maximum April flow (MH04), and no daily rises (RA05)] (table 5). There were similar relationships in the PCA plot compared with those in the NMS plot (fig. 3). There was reasonable separation among eastern Red River, Arkansas River, western stream sites, and there was some separation of reference streams versus those with altered hydrology along a trajectory associated with small streams on the steams loading the highest (that is more distinct functional group configuration) on the second principle component compared to larger reference streams with negative values on the second axis in eastern streams (fig. 4). There was no evidence of differing functional structure within fish assemblages between reference and altered streams among the western basins (fig. 4).

We compared the proportion of functional groups between altered and reference sites assigned with discriminant analysis (table 6). There were higher proportions of species in reference locations that were intolerant to water quality and habitat degradation, more sensitive benthic species and lithophilic spawners, and species associated with fast flowing water. Altered sites had more tolerant species that preferred lentic conditions and slower flowing waters. Generalist broadcast spawners (A_1) were more common in altered locations, while brood hidiers were more common in reference sites (A_2). We identified a geographical trend of more intolerant species in streams with fast flowing water and imperiled species in the eastern portion of the State compared with the more generalist groups in the western portion (table 6).

Influence of Impoundment on Downstream Hydrologic Conditions

Kiamichi River

The impoundment of Jackfork Creek by Sardis Lake dam had limited effects on downstream hydrology of the Kiamichi River. Median December flows and variation in January flows were significantly higher after impoundment (table 7). There were no differences in magnitude and duration of minimum flows between the periods, but median 3- and 7-day maximum flows significantly decreased between the two periods (table 7). We did not observe any significant differences in the timing of annual extreme events between the pre- and post-impoundment periods. High flow pulses and flow reversals were more frequent after impoundment (table 7).

Bird Creek

Flow alteration due to the impoundment of Birch Creek, a tributary of Bird Creek, had minimal effects on the overall hydrologic conditions in Bird Creek. Only three parameters, median monthly flows for May and June and the number of flow reversals in Bird Creek, were significantly higher after the construction of Birch Lake (table 8). In contrast, several hydrologic parameters, representing all five components of the flow regime, were significantly altered by the impoundment of Bird Creek and construction of Skiatook Lake. Reservoir release patterns resulted in significantly higher median monthly flows for all 12 months with the months of May and June exhibiting more variation in median flow. The magnitude and duration of annual extreme conditions, particularly low flows, was significantly altered. Median annual minimum

flows increased dramatically and the base flow index increased by an order of magnitude between the pre-impoundment and post-Skiatook-impoundment periods. Additionally, the median annual 30-day maximum flow was higher post impoundment. There were no significant reservoir effects on the timing of minimum flows, but the annual 1-day maximum flow occurred later in the year after the impoundment of Skiatook Lake. Low flow pulses never occurred and duration of high flow pulses increased significantly from 4 to 6 days after impoundment of Bird Creek by Skiatook Lake. The rate and frequency of changes in discharge were significantly altered by reservoir operations with faster fall rates and more frequent and variable flow reversals per year (table 8).

Influence of Impoundment on Downstream Fish Assemblage Structure

Kiamichi River

Twenty-six sampling events ordinated in fish species multidimensional space identified three axes (3D) that explained 76.5 percent of the variation in original distances among sampling events (3D stress = 15.8). Despite only explaining 17 percent of the variation, NMS axis 1 (x axis) had strong associations with median flows for September, magnitude and duration of low flows (7-day minimum), frequency of flow reversals, and daily discharge for the sampling date (fig. 5A, table 9). Frequency of flow reversals was the only hydrologic parameter related to fish community structure that was altered after the impoundment of Sardis Lake (fig. 6C, table 7). However, there were positive trends in September median flows, 7-day minimum flows, and number of flow reversals across the period of record associated with available fish data (fig. 6B, D, and F). Additionally, there were significant differences in fish species composition (F) between pre- and post-impoundment periods using both the two-group (PERMANOVA $F_{1,24} = 5.52$, $p < 0.002$) and three-group (PERMANOVA $F_{1,24} = 6.96$, $p < 0.001$) structure. There were no differences in multivariate dispersion among the pre- and post-dam construction periods (PERMDISP $F_{1,24} = 0.07$, $p = 0.792$) or among the pre-dam, transitional, or post-dam construction periods (PERMDISP $F_{2,23} = 0.09$, $p = 0.911$). Brook silversides (*Labidesthes sicculus*) was a significant indicator species for pre-impoundment conditions, whereas increased abundance of blacktail shiners (*Cyprinella venusta*) was positively associated with the post-impoundment period (table 10). Among bass species, spotted bass (*Micropterus punctulatus*) were more abundant and frequent in pre-impoundment collections while largemouth bass (*Micropterus salmoides*) were significant indicators of post-impoundment conditions, particularly during the transitional period (table 10).

Bird Creek

Nonmetric multidimensional scaling (NMS) ordination of 23 sampling events in fish species space identified three axes that explained 83.3 percent of the variation in original distances among fish sampling events (3D stress = 13.07). More than half of the explained variation (46.6 percent) was attributed to NMS axis 1 (x axis), which had strong associations with increased median flows during typical low flow months (July, August, September, and October), December flows, minimum flows (1st, 3d, 7th, 30th, and 90th day), and number of flow reversals, as well as decreased number and duration of low flow events (fig. 5B, table 9). All these hydrologic parameters responded dramatically to flow alteration associated with water management practices at Skiatook Lake (fig. 7, table 8). Pre- and post-impoundment periods were clearly separated on NMS axis 1 (fig. 5B), and some separation between the transitional

and post-impoundment periods occurred on NMS axis 2 (y axis; fig. 5B). Pre- and post-impoundment fish assemblages differed significantly in their species composition using both the two group (PERMANOVA $F_{1,21} = 10.12$, $p < 0.001$) and three group (PERMANOVA $F_{1,22} = 9.21$, $p < 0.001$) structure. There was a marginal difference in multivariate dispersion among the pre- and post-dam construction periods (PERMDISP $F_{1,21} = 3.24$, $p = 0.094$) but no difference among pre-dam, transitional, or post-dam construction periods (PERMDISP $F_{2,20} = 2.56$, $p = 0.103$). No indicator species were identified for the pre-dam construction period, but there were 13 species that had affinities with the period after dam construction (table 10). Several cyprinid minnows, including central stonerollers (*Campostoma anomalum*), ghost shiners (*Notropis buchanni*), sand shiners (*Notropis stramineus*), and suckermouth minnows (*Phenacobius mirabilis*), were strong indicators of post-dam construction fish assemblages (table 10). Western mosquitofish (*Gambusia affinis*) and inland silversides (*Menidia beryllina*) showed affinities for post-dam construction conditions. Two centrarchids, largemouth bass (*Micropterus salmoides*) and white crappie (*Pomoxis annularis*), were significant indicator species for the post-dam construction period. The slenderhead darter (*Percina phoxocephala*) was also associated with post-dam construction conditions in Bird Creek (table 10).

Effects of Flow Alteration on Fish Assemblages

Development of flow-ecology relationships is critical to setting management goals for protecting freshwater biodiversity in altered ecosystems (Poff and others, 2010; Poff and Zimmerman, 2010). Analyses of spatial datasets that represent hydrologic alteration gradients as well as temporal biological datasets spanning pre-and post-hydrologic alteration events aid in this process by identifying components of altered hydrologic regimes and other factors that influence biological change (Propst and Gido, 2004; Armstrong and others, 2010; Gido and others, 2010; Poff and Zimmerman, 2010; Taylor, 2010). Results from such studies can be used by water resource managers to help restore critical components of the flow regime while continuing to meet anthropogenic needs for water (Richter and others, 2003). The study of this report focused on relationships between flow alteration and fish assemblage structure and function, relying on a fish dataset with sites that represented different degrees of flow alteration, as well as two temporal fish datasets representing pre- and post-river impoundment environments. At the broad scale, fish assemblage and functional structure were associated with several flow metrics (Poff and Allan, 1995) but had stronger associations with regional factors, evidence that regional differences in flow regimes potentially influence fish assemblage structure (Gehrke and Harris, 2001). Likewise, patterns in functional group membership were associated with reference versus altered flows, but similar patterns were observed from east to west. Thus, the development of flow-ecology relationships needs to be stratified across these ecosystem types.

Although not presented in this report, we classified altered sites into three levels of alteration (low, moderate, high) based on the number and percentage exceeding the expected range of values. There were no clear relationships between increasing flow alteration and fish functional groups. A study of this type of relationship between degree of flow alteration and fish ecology could prove useful, but would be better served by a planned study with active data collection to control the level and location of fish sampling relative to alteration. The differences in fish assemblages and flow regime between the eastern and western portions of Oklahoma would be best served by separate studies of these areas.

A previous study of Oklahoma fishes found that they are more sensitive to habitat degradation than water quality degradation (Jester and others, 1992), so efforts to identify specific instream habitat changes directly related to flow regime alteration in Oklahoma could enhance the results in this report. Fish species also can be used as indicators of flow alteration, as has been done with water quality in Great Lakes wetlands (Seilheimer and Chow-Fraser, 2007) and in this report with the ISA in Bird Creek. This method would take advantage of the greater distribution of fish collection data, as well as the relative ease of collecting fish samples compared with the 15 or more years of flow data needed to assess flow regime.

One hydrologic change that is clearly shown in our analysis of flow alteration is the shift to higher, more stable flows that are reflected by Turton and others (2009), as classified in group 43. Conditions in group 43 only occurred in a small group of only 6 streams in northeastern Oklahoma from the initial total of 88. We observed perennial flashy (42) and intermittent (44) streams undergoing significant changes to flow regime to group 43 in current flows. The impacts of this change are evident in the analysis of Bird Creek, which changed from an intermittent to a stable groundwater classification after the impoundment of Skiatook Lake. Interestingly, following the impoundment of Sardis Lake, the Kiamichi River did not shift from a perennial flashy stream to one with stable flows characteristic of group 43. An analysis of fish assemblage structure across a 17-year period associated with reservoir impoundment in these two river systems indicated that there were significant shifts in species composition for both sites. However, the degree to which reservoir impoundment altered downstream flow regimes within the study reaches was not consistent between the two river systems. The construction of Skiatook Lake within the Bird Creek river system significantly altered aspects of all five components of flow (magnitude, timing of occurrence, frequency of occurrence, duration of time, and rate of change in flow conditions), resulting in a change from a naturally intermittent stream flow regime to more stable flows, characteristic of streams in northeastern Oklahoma (Turton and others, 2009). In contrast, the Kiamichi River hydrologic regime downstream of Sardis Lake was relatively unaltered with modest declines in maximum flows and slight increases in number of reversals. This system retained the characteristics of a perennial flashy stream identified by Turton and others (2009). Despite this lack of identifiable change, there were still significant changes in fish assemblage structure within both systems across the 17-year period associated with reservoir construction, but species associated with these changes differed between the two systems. Several flow-dependent species as well as reservoir species, including largemouth bass, increased within the period of record associated with flow alteration in Bird Creek, a historically intermittent stream, whereas declines in two riverine-adapted fish coupled with an increase in largemouth bass were associated with the impoundment of Sardis Lake within the Kiamichi River basin. These contrasting responses of flow regimes and fish assemblages to reservoir impoundment within these the two river systems highlight the fact that linkages between reservoir impoundments, hydrologic alteration, and biological attributes are dependent on the ecological setting and associated natural hydrologic regimes of study systems.

Bird Creek lies in north-central Oklahoma and exhibits a historical flow regime with high flow variability and extended periods of extreme low flows, typical of streams draining the Great Plains region of North America (Poff and Ward, 1989; Dodds and others, 2004). Fishes of prairie streams have evolved behavioral and physiological mechanisms as well as life history traits to cope with variable and sometimes harsh conditions associated with dynamic hydrology (Matthews, 1987, 1988; Labbe and Fausch, 2000; Matthews and Marsh-Matthews, 2003). It is not surprising that there were no significant indicator species associated with Bird Creek fish

assemblages during the pre-impoundment period because fish assemblage structure likely varies with stream flow across time in intermittent streams (Matthews and Marsh-Matthews, 2003). Higher flow conditions and loss of low- to no-flow periods associated with post-impoundment habitats corresponded with increases in frequency and abundance of fluvial-dependent species, including central stonerollers, sand shiners, suckermouth minnows, and slenderhead darters. Central stonerollers can occupy a wide variety of stream conditions ranging from intermittent headwater streams to downstream perennial reaches and adjust life history strategies to persist along fluvial gradients experienced in prairie streams (Spranza and Stanley, 2000). Fluvial habitats with perennial flow provide more predictable habitat conditions, which may confer more stable fish populations. Sand shiners are broadcast spawners that prefer flowing water over sand substrates, particularly in the presence of other *Notropis* sp. (Mueller and Pyron, 2011). Conditions downstream of reservoirs may provide other benefits to sand shiners, such as decreased turbidity, which has been shown to negatively impact feeding on terrestrial and aquatic insects by this species (Bonner and Wilde, 2002). Suckermouth minnows and slenderhead darters prefer fast flowing waters and require high-velocity areas during spring for spawning (Brewer and others, 2006). These two species seem to persist downstream of dams but are often extirpated upstream of reservoirs; however, spawning habitats may be impacted by flood control reservoirs due to impacts on the timing and magnitude of flows (Quist and others, 2005; Brewer and others, 2006). We also observed an increase in western mosquitofish, a species not generally associated with higher flows. Mosquitofish exhibit an opportunistic life history strategy characterized by small body size with early maturation, low fecundity per spawning event, and low juvenile survivorship (Winemiller and Rose, 1992; Hoeinghaus and others, 2007). This strategy is well adapted to fish occupying habitats that experience frequent and intense disturbance and likely confers an ability to monopolize and maintain higher abundances in highly variable shallow stream margin habitats associated with substantially more flow reversals in post-impoundment conditions.

Southeastern Oklahoma streams lie within the western extension of the Ozark Mountains and support perennial flow with a certain degree of flashiness. While impoundment of Sardis Lake did not have strong effects on long-term patterns in stream discharge, general increases in September median flows, low-flow duration, and number of reversals for 1976–1996 were associated with shifts in fish assemblage structure during the study period. However, these increases explained a small proportion of the variation in fish assemblage structure, and there were few significant indicator species for the pre- or post-impoundment periods.

Patterns in indicator species analysis suggest that factors associated with reservoir impoundment other than flow may have influenced changes in fish assemblage structure. We observed an increase in frequency and abundance in largemouth bass coupled with a decrease in spotted bass. The flow analysis results do not support altered hydrology as a primary mechanism for this shift, and this result may be due simply to increased largemouth bass abundance associated with stocking in Sardis Lake as well as Hugo Lake downstream. Additionally, reservoir-like conditions can occur during high flow periods within the stream reach due to backwater from Lake Hugo. This potential mechanism was not limited to the Kiamichi River site as we also observed higher frequency and abundance of inland silversides and largemouth bass in post-impoundment collections downstream of Skiatook Lake in Bird Creek. Shifts in largemouth bass frequency and abundance downstream of reservoirs may be due to lentic species having more generalist life history traits which provide certain advantages in flow-altered environments (Arthington and others, 2006). However, increases in largemouth bass abundance

across two systems that differ in flow alteration suggest that their abundance may be more related to dispersal from introductions within the reservoir rather than flow requirements. Previous studies have demonstrated that upstream as well as regional patterns in fish assemblage structure are influenced by proximity to reservoir assemblages (Falke and Gido, 2006; Gido and others, 2010).

Quist and others (2005) observed introduced game fish species both upstream and downstream after construction of a reservoir. This may have been a mechanism for increased largemouth bass and forage fish frequency and abundance downstream of reservoirs in the current study. The distribution of inland silversides was historically limited to Atlantic and Gulf Coast waters and adjacent coastal portions of rivers. Inland silversides has been widely introduced in reservoirs of Midwestern States as forage for game fish and is responsible for large declines in brook silverside populations in Oklahoma (Gomez and Lindsay, 1972; McCormas and Drenner, 1982). While native to Oklahoma streams, largemouth bass populations were augmented through stocking of Skiatook reservoir immediately after impoundment. However, spotted bass populations increased in the early 1990s and displaced largemouth bass from much of their habitat (Long and Fisher, 2005). Factors influencing black bass habitat use within the reservoir (stocking, displacement) potentially influenced temporal patterns of frequency and occurrence of largemouth bass in downstream habitats of both systems.

Several studies have shown changes in fish assemblage structure associated with water resource development and reservoir operations (Bain and others, 1988; Bonner and Wilde, 2002; Marchetti and Moyle, 2001; Perkin and Bonner, 2011). Our analysis examines a temporal sequence of fish assemblage structure in response to reservoir construction in two different stream systems representing different ecological settings. However, other factors related to reservoirs, including altered thermal regimes due to hypolimnetic releases (Olden and Naiman, 2010) and proximity to reservoir habitats (Falke and Gido, 2006; Gido and others, 2010), may also influence fish assemblage structure in lotic environments. Additionally, while the biological datasets used in this study were reasonably long, these are temporal datasets of limited length. For example, in our Kiamichi River analysis, there was little change in long-term hydrologic patterns. Yet, within the period of fish sampling there were trends in several hydrologic variables that were correlated with fish assemblage structure.

The Kiamichi River is a well-studied system, and recent studies examining longer time trends have made some alarming observations. Drought conditions after the period of record that this study assessed, combined with flow management have resulted in serious dewatering of fluvial habitats and an associated decline in mussel densities with a shift in assemblage structure from thermally sensitive to tolerant species (Galbraith and others, 2010). This shift in mussel assemblages has important consequences for ecosystem function (Spooner and Vaughn, 2008; Vaughn, 2010). It is possible that fluvial fish species have also declined during this period. For this reason it is important that water management goals be developed by considering a range of variability approach that identifies natural flow regimes for dry, moderate, and wet periods to maintain key components of the natural flow regime (Richter and others, 2003). The results of this study indicate that interactions between flow regime and aquatic biota vary depending on ecological setting.

This study was limited by the low number of gaged locations with associated fish collections. Of these locations with streamflow and fish information, only a subset had a known reference period and could be used. We were unsuccessful at drawing a conclusion associated with a gradient of flow alteration, which may have been due to the types of streams in each class

or from other factors impacting species other than flow alteration. The streams considered altered were from different ecoregions; thus, there may be biogeographical constraints on the biota that could confound the analysis with regards to altered flow regime.

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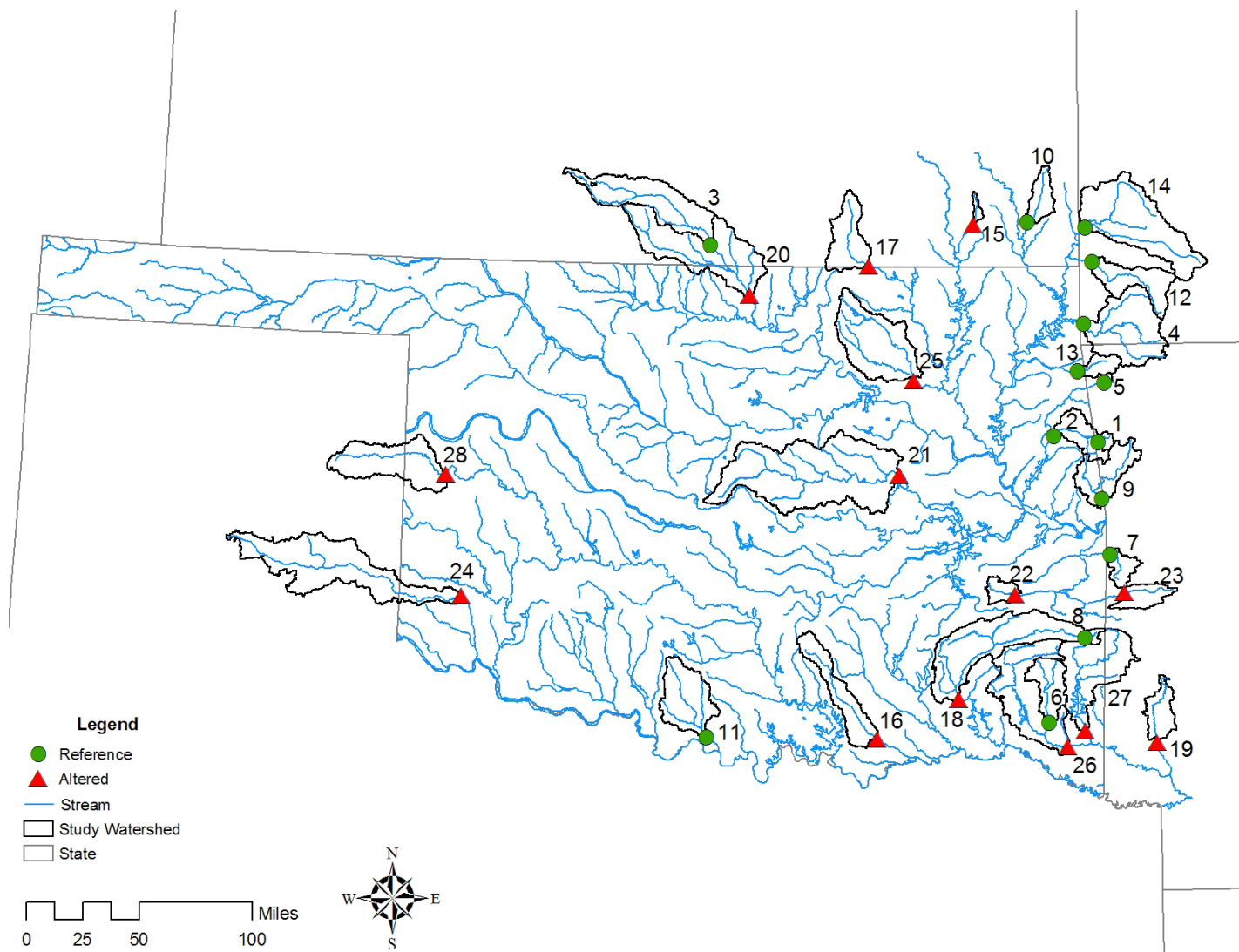
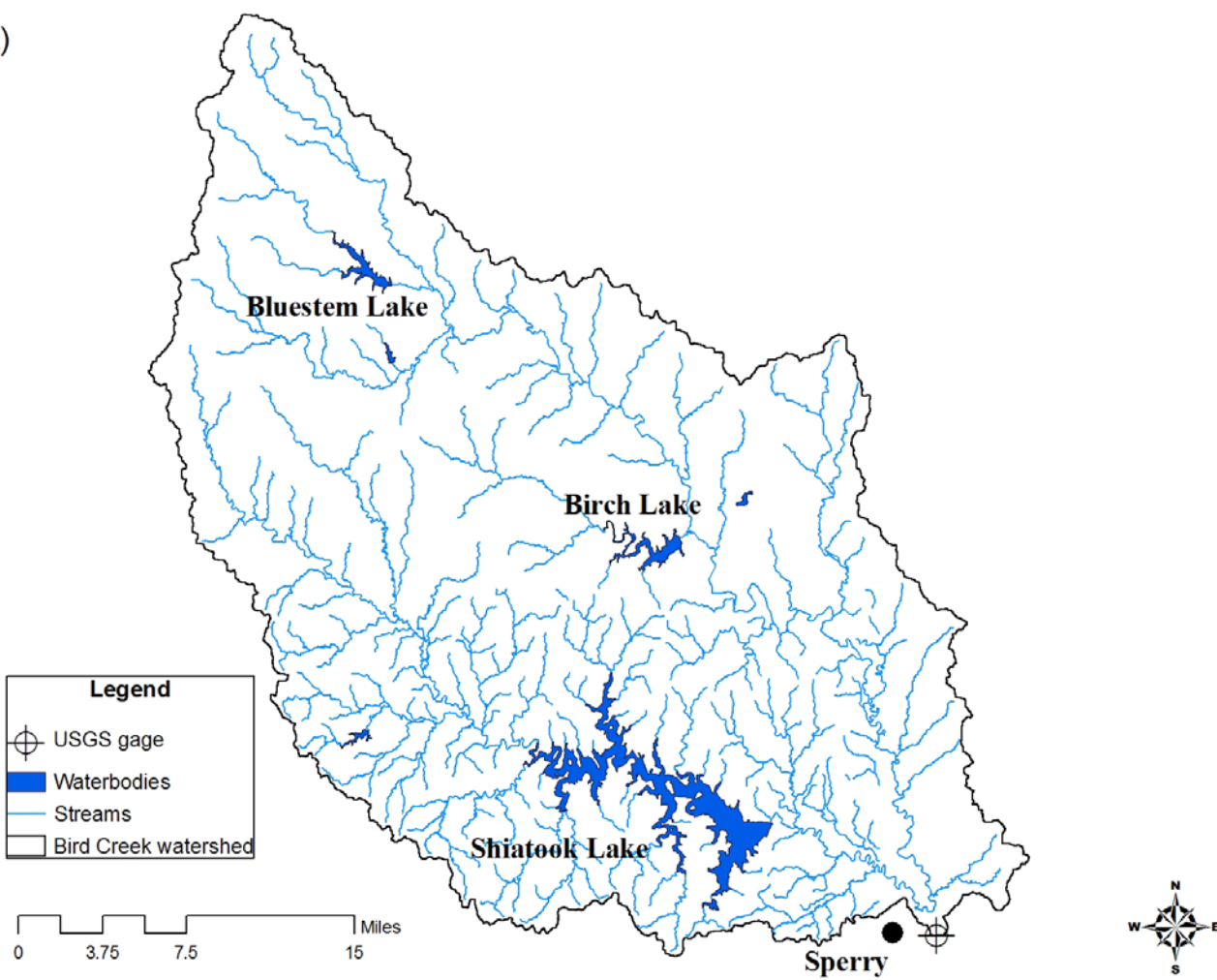


Figure 1. Map showing streamgage location and contributing watersheds for reference and altered sites in Oklahoma. Numbers next to streamgage symbols correspond to streamgage names in table 2. Digital data from U.S. Geological Survey, U.S. Environmental Protection Agency, and Oklahoma Water Resource Board, various scales. Gage locations and gage watersheds developed by authors. Projection is Geographic Coordinate System (GCS) North American 1927. Horizontal datum is North American Datum of 1927, Albers conical projection. Map created by Titus Seilheimer.

A)



B)

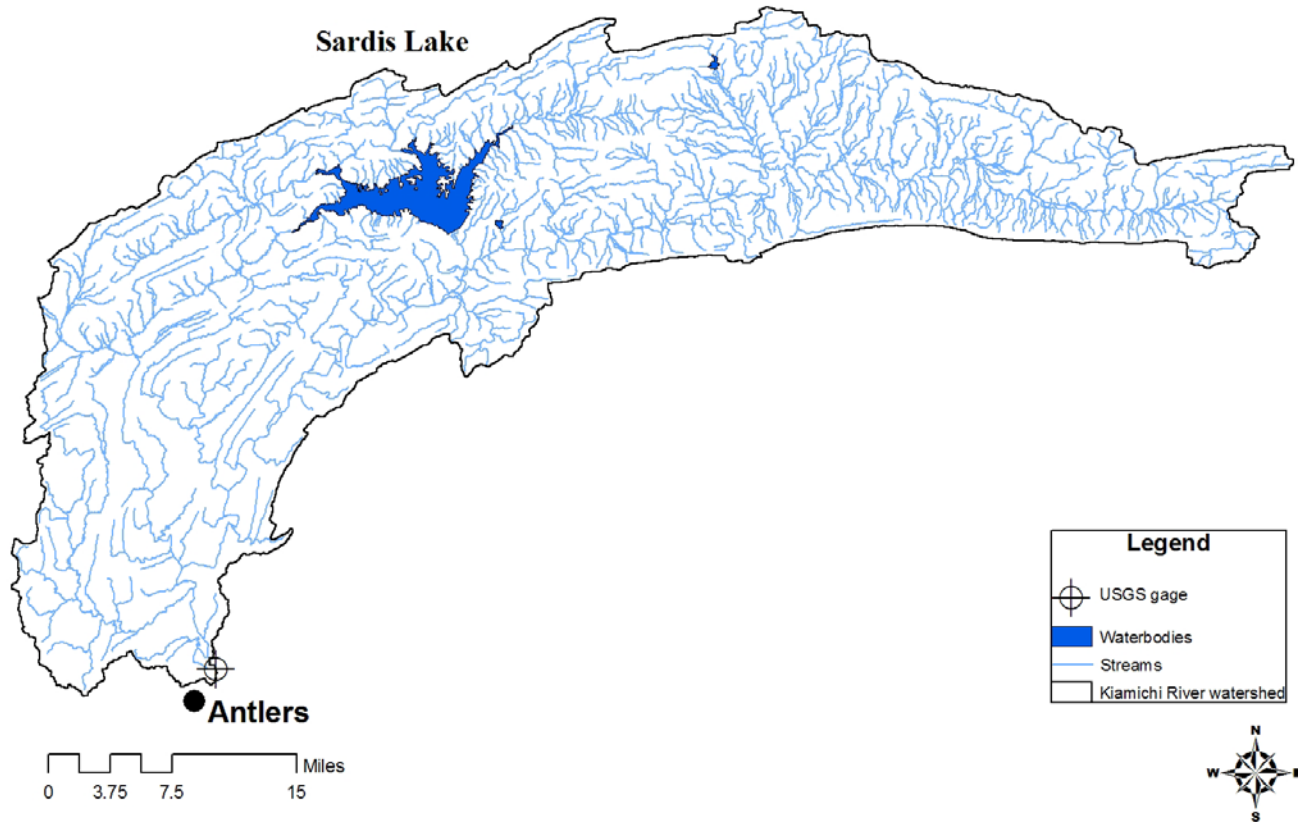


Figure 2. Map of the watershed on A, Bird Creek and B, Kiamichi River, Oklahoma. Sampling site on Bird Creek is in relation to U.S. Geological Survey (USGS) streamgage 07177500 near Sperry, Okla., Birch Lake, and Skiatook Lake; sampling site on Kiamichi River is in relation to USGS streamgage 07336200 near Antler, Okla., and Sardis Lake. Stream data from U.S. Environmental Protection Agency and water body data from the Oklahoma Water Resources Board, various scales. Gage locations and gage watersheds developed by authors. Projection is Geographic Coordinate System (GCS) North American 1927. Horizontal datum is North American Datum of 1927, Albers conical projection. Map created by Titus Seilheimer.

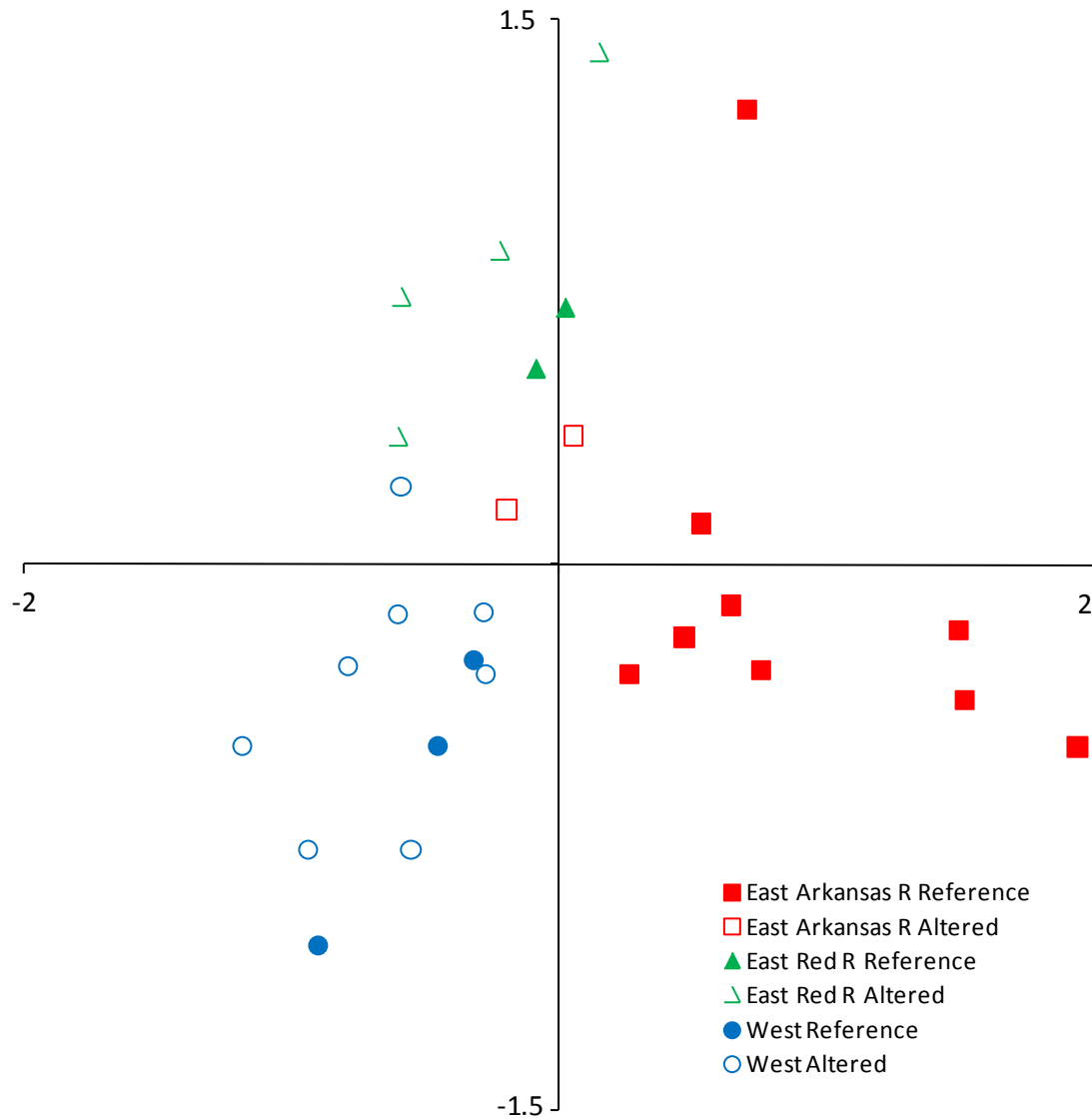


Figure 3. Ordination biplot from nonmetric multidimensional scaling (NMS; Jaccard) of species presence for fishes associated with 28 streamgages on reference and altered streams in Oklahoma. The x-axis is NMS 1, and the y-axis is NMS 2.

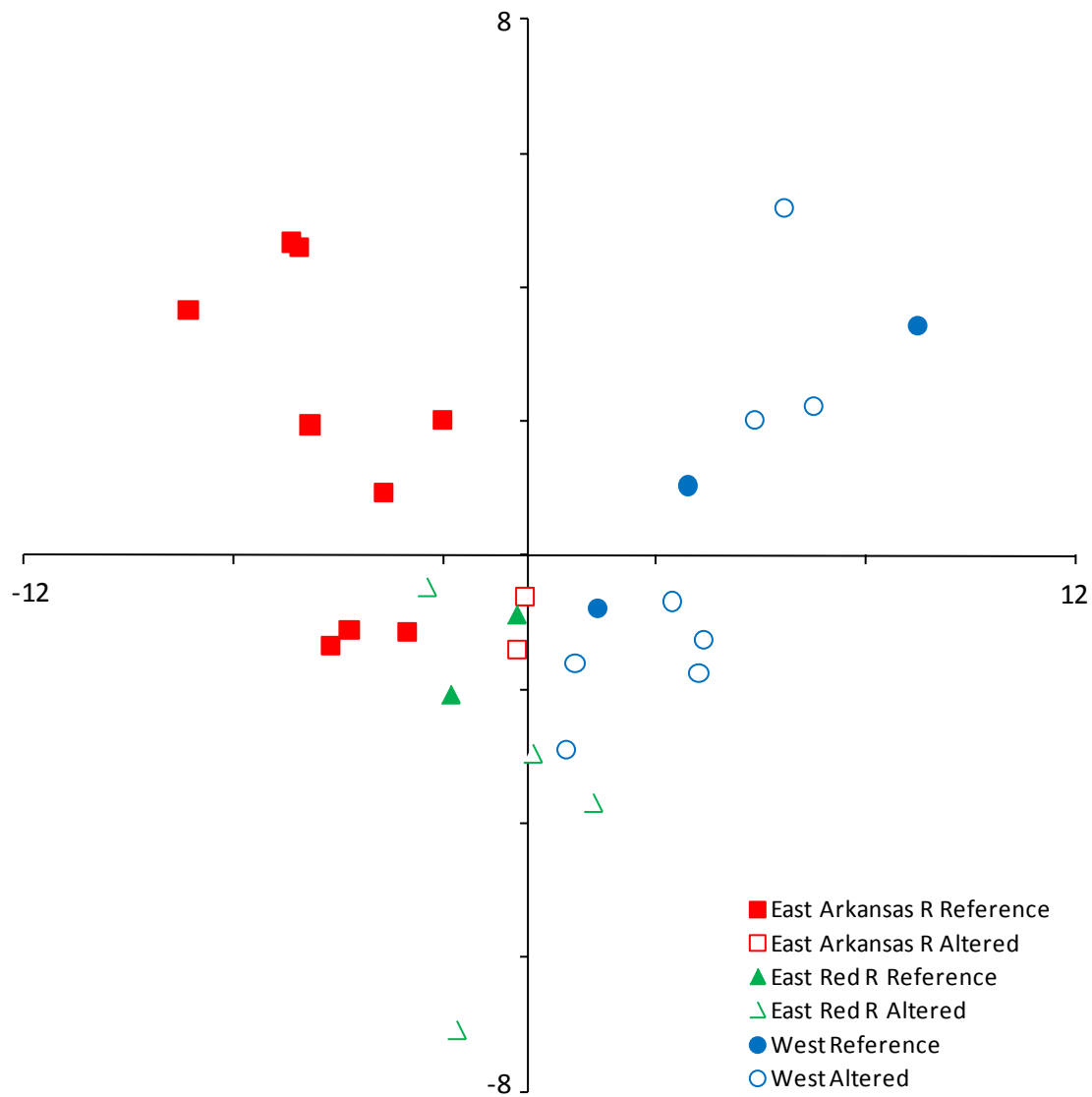


Figure 4. Ordination biplot from principal components (PC) analysis of proportion of functional groups for 28 streamgages on reference and altered streams in Oklahoma. The x-axis is PC1, and the y-axis is PC2.

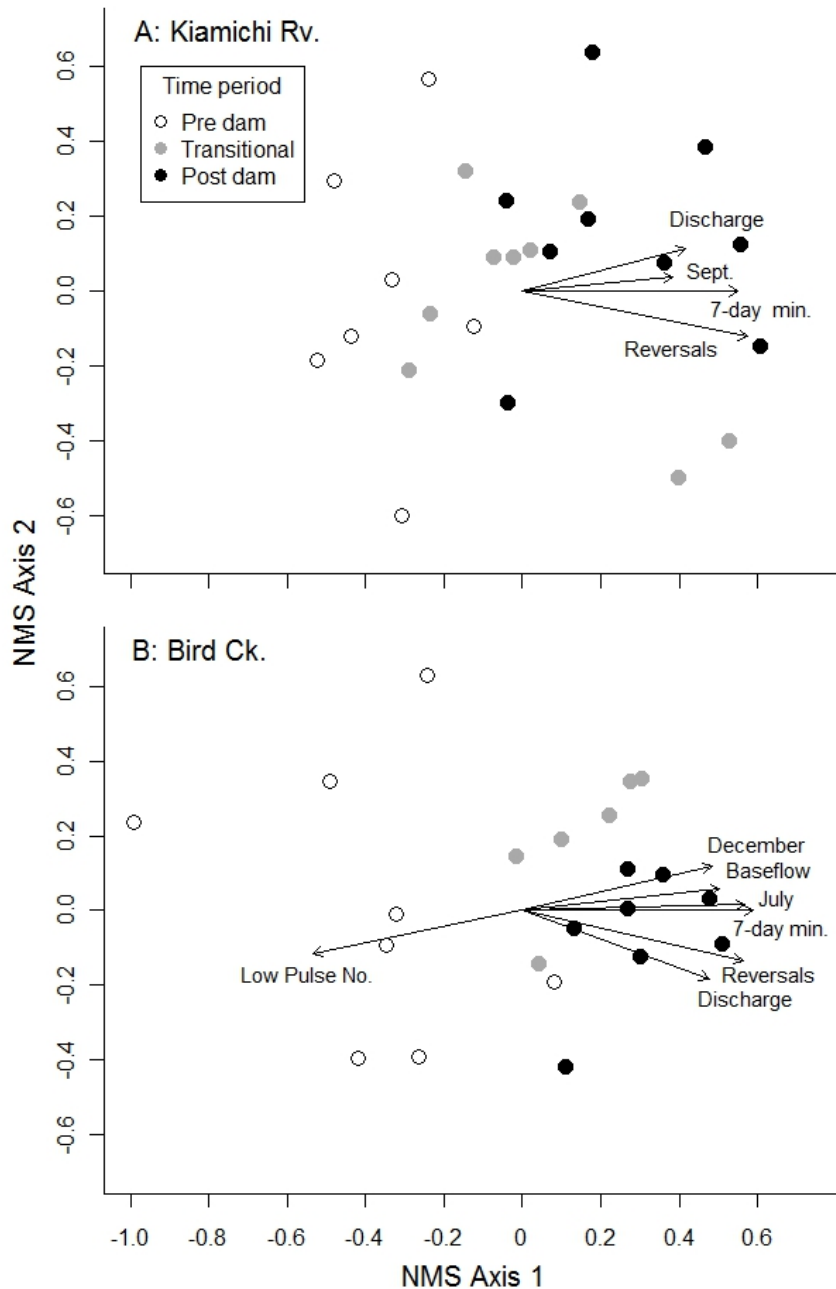


Figure 5. Charts showing nonmetric multidimensional scaling (NMS) ordination based on fish species composition for sampling events conducted during the early summer over 18 years at one location on A, the Kiamichi River (stress = 15.08) and B, Bird Creek (stress = 13.07), Oklahoma. The collection sites lie downstream of Sardis Lake on the Kiamichi River and downstream of Birch and Skiatook Lakes on Bird Creek, which were formed by impoundments in 1977 and 1982, respectively. NMS axes 1 and 2 represent ecological distances between sites based on differences in their species composition. Open symbols indicate pre-dam sampling events, whereas solid symbols represent post-dam sampling events divided into transitional (grey), and post-dam construction (black) periods for the two river systems. Please refer to text for years associated with each period. Arrows indicate direction and magnitude of correlations between nonmetric multidimensional scaling scores and hydrologic parameters calculated with indicators of hydrologic alteration analysis (table 9).

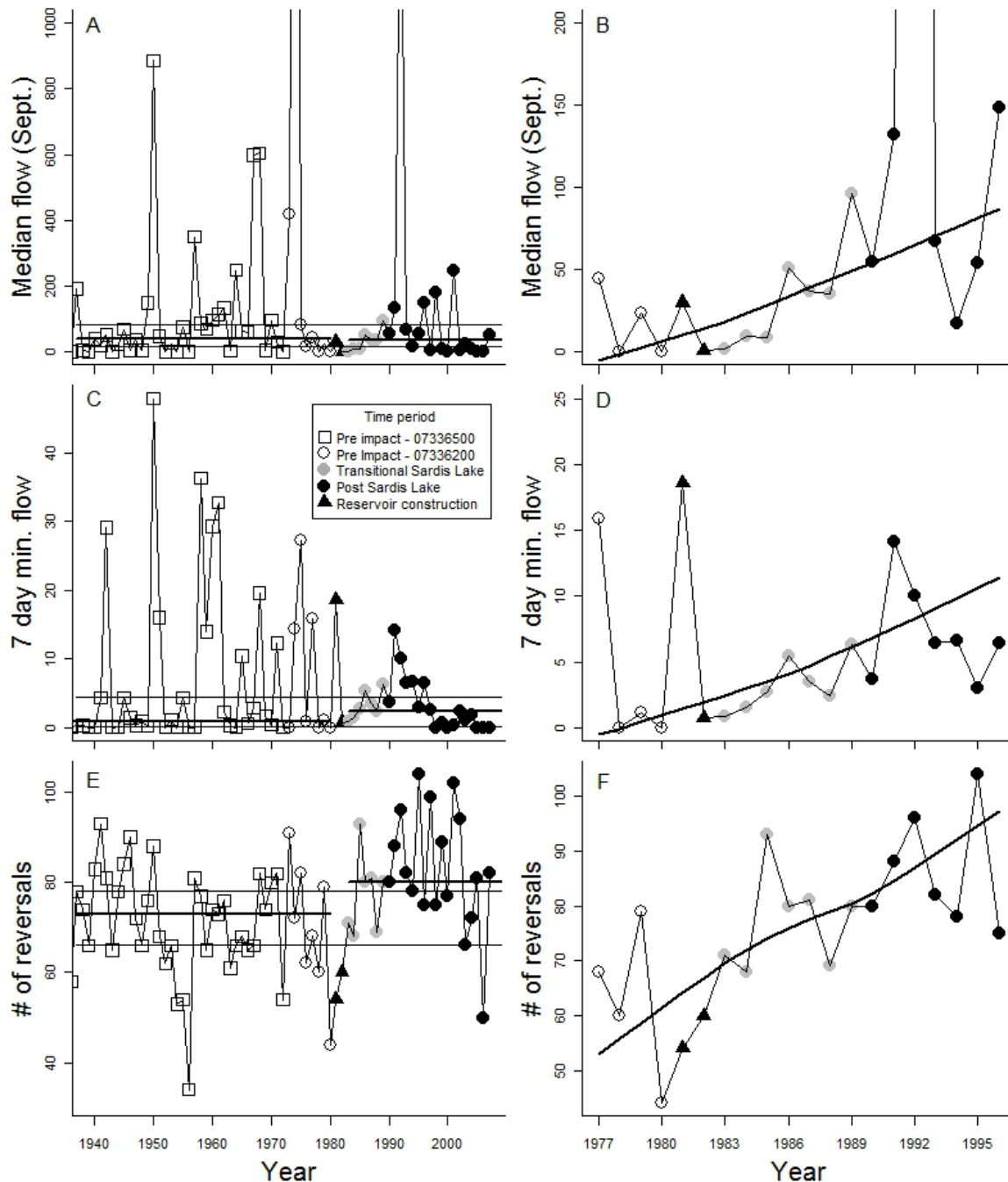


Figure 6. Chart showing annual variation in three hydrologic parameters that were significantly correlated with differences in fish assemblage structure (fig. 7A, table 9) in the Kiamichi River near Antlers, Oklahoma. Parameters include A and B, median flows for September; C and D, annual 7-day minimum (min.) flow; and E and F, number (#) of flow reversals per year. Each parameter includes plots of period of record (A, C, E) followed by record associated with fish data (B, D, F). Thin horizontal lines represent range of variation for the period before the construction of the impoundment, and thick horizontal lines represent median values for the periods before the construction of the impoundment and after the construction of the Sardis Lake impoundment. Smoothers (B, D, F) are based on locally weighted regression (LOWESS). All flow data are presented in cubic feet per second.

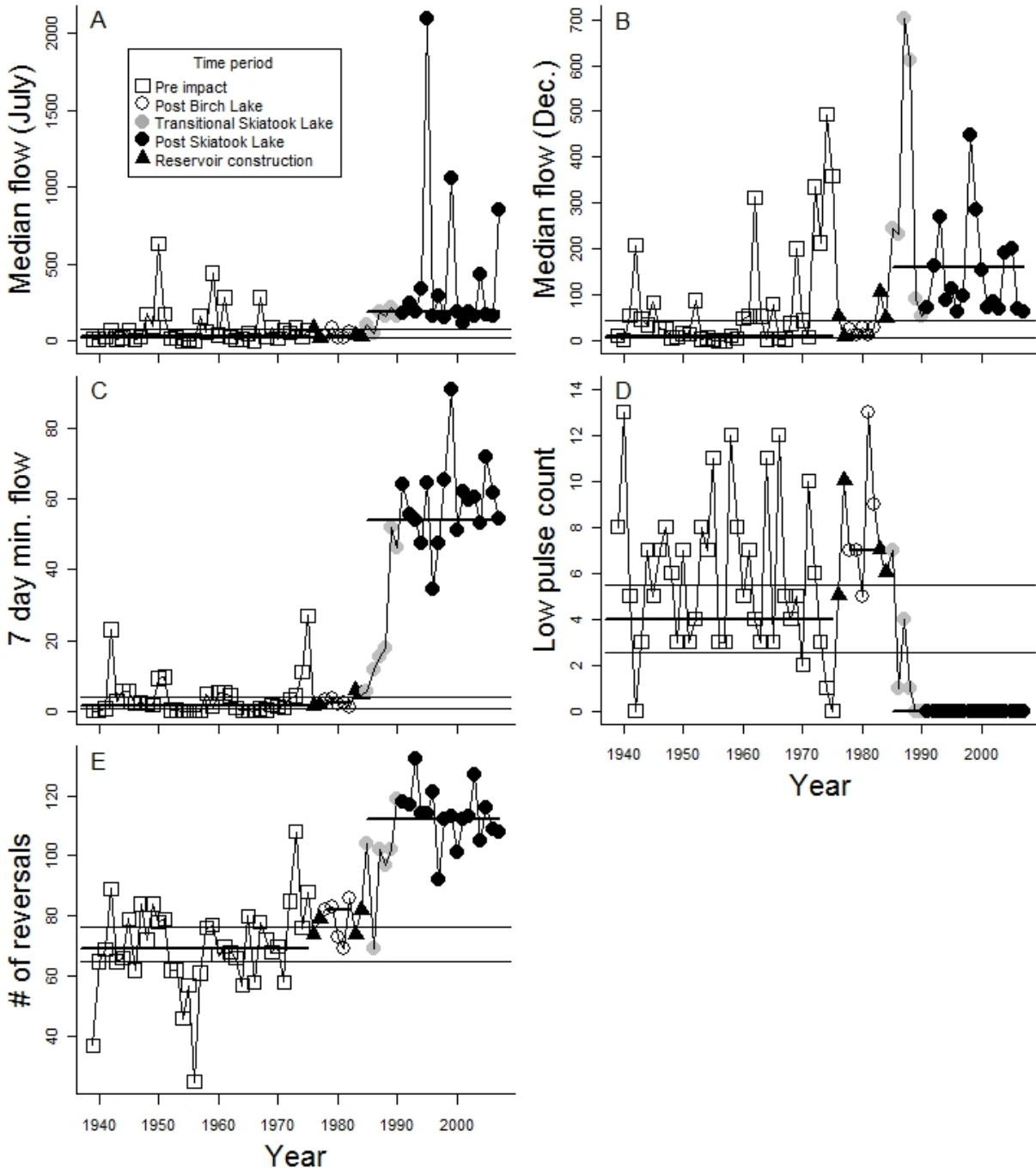


Figure 7. Charts showing annual variation in five hydrologic parameters that were significantly correlated with differences in fish assemblage structure (fig. 7B, table 9) in Bird Creek near Sperry, Oklahoma. Parameters include A, median flow for July; B, median flow for December; C, annual 7-day minimum (min.) flow; D, number of low pulses per year; and E, number (#) of flow reversals per year. Thin horizontal lines represent range of variation for period before the construction of the impoundment, and thick horizontal lines represent median values for the periods before construction of the impoundment and after construction of the Birch Lake and Skiatook Lake impoundments. All flow data are presented in cubic feet per second.

Table 1. Flow indices used in a biological assessment of flows of 28 streams in Oklahoma.[ft³/s, cubic feet per second; ft³/s/mi², cubic feet per second per square mile; ft³/s/d, cubic feet per second per day; %, percent]

Code	Hydrologic index	Units	Definition
Magnitude			
MA01	Mean daily flows	ft ³ /s	Mean daily flows.
MA04	Variability in daily flows 1	%	Coefficient of variation of the logs in daily flows corresponding to the {5th, 10th, 15th, ..., 85th, 90th, 95th} percentiles.
MA28	Variability in May flows	%	Coefficient of variation in monthly flows for May.
ML01	Mean minimum January flows	ft ³ /s	Mean minimum monthly flow for January.
ML09	Mean minimum September flows	ft ³ /s	Mean minimum monthly flow for September.
MH04	Mean maximum April flows	ft ³ /s	Mean of the maximum monthly flows for April.
MH14	Median of annual maximum flows	Dimensionless	Median of the highest annual daily flow divided by the median annual daily flow averaged across all years.
MH20	Specific mean annual maximum flows	ft ³ /s /mi ²	Mean annual maximum flows divided by catchment area.
Frequency			
FL03	Frequency of low flow spells	Events per year	Total number of low flow spells (threshold equal to 5 % of mean daily flow) divided by the record length in years.
FH01	High flood pulse count 2	Events per year	Mean number of high pulse events, where the 75th percentile is the high pulse threshold.
FH04	High flood pulse count 1	d/yr	Mean number of days per year above the upper threshold (defined as 7 times median daily flow), and the value is represented as an average instead of a tabulated count.
FH05	Flood frequency 1	Events per year	Mean number of high flow events per year using an upper threshold of 1 times median flow over all years.
Duration			
DL03	Annual minima of 7-day means of daily discharge	ft ³ /second	Magnitude of minimum annual flow of 7-day mean daily discharge.
DL05	Annual minima of 90-day means of daily discharge	ft ³ /second	Magnitude of minimum annual flow of 90-day mean daily discharge.
DL18	Number of zero-flow days	d/yr	Mean annual number of days having zero daily flow.
DH07	Variability in annual maxima of 3-day means of daily discharge	%	Coefficient of variation in the 3-day moving average flows.
DH10	Variability in annual maxima of 90-day means of daily discharge	%	Coefficient of variation in the 90-day moving average flows.
DH15	High flow pulse duration	d/yr	Mean duration of FH1 (high flood pulse count 1).
DH21	High flow duration 1	Days	Average duration of flow events with flows above a threshold equal to the 25th percentile value for the entire set of flows.

Code	Hydrologic index	Units	Definition
DH23	Flood duration 1	Days	Mean annual number of days that flows remain above the flood threshold averaged across all years.
Timing			
TA01	Constancy	Dimensionless	See Colwell (1974).
TL01	Julian date of annual minimum	Julian days	The mean Julian date of the 1-day annual minimum flow over all years.
TH01	Julian date of annual maximum	Julian days	The mean Julian date of the 1-day annual maximum flow over all years.
Rate of change			
RA03	Fall rate	ft ³ /s /d	Mean rate of negative changes in flow from one day to the next.
RA05	No day rises	Dimensionless	Ratio of days where the flow is higher than the previous day.
RA07	Change of flow	ft ³ /s/d	Median of difference between natural logarithm of flows between two consecutive days with decreasing flow.

Table 2. Streamgages used in a biologic assessment of the flows of 28 streams in Oklahoma,
[Streamgage number corresponds to numbers in figure 1. HIP, hydroecological integrity assessment process; no., number;
mi², square miles. HIP classifications are from Turton and others (2009)]

Gage no.	Streamgage code	Gage name	HIP group	Alteration status	Latitude	Longitude	Basin size (mi ²)
1	07196900	Baron Fork at Dutch Mills, Ark.	42	Reference	35.88	-94.49	42.2
2	07197000	Baron Fork at Eldon, Okla.	41	Reference	35.92	-94.84	319.7
3	07151500	Chickaskia River near Corbin, Ks.	41	Reference	37.13	-97.60	833.6
4	07189000	Elk River near Tiff City, Mo.	43	Reference	36.63	-94.59	872.7
5	07195800	Flint Creek at Springtown, Ark.	41	Reference	36.26	-94.43	15.1
6	07337900	Glover River near Glover, Okla.	42	Reference	34.10	-94.90	328.6
7	07249400	James Fork near Hackett, Ark.	42	Reference	35.16	-94.41	150.5
8	07335700	Kiamichi River near Big Cedar, Okla.	42	Reference	34.64	-94.61	40.7
9	07249985	Lee Creek near Short, Okla.	42	Reference	35.52	-94.46	445.3
10	07184000	Lightning Creek near McCune, Ks.	44	Reference	37.28	-95.03	201
11	07315700	Mud Creek near Courtney, Okla.	44	Reference	34.00	-97.57	589.3
12	07187000	Shoal Creek above Joplin, Mo.	43	Reference	37.02	-94.52	438.5
13	07191220	Spavinaw Creek near Sycamore, Okla.	41	Reference	36.33	-94.64	135
14	07186000	Spring River near Waco, Mo.	43	Reference	37.25	-94.57	1,188.1
15	07170700	Big Hill Creek near Cherryvale, Ks.	44	Altered	37.27	-95.47	37.8
16	07332500	Blue River near Blue, Okla.	41	Altered	34.00	-96.24	489.8
17	07172000	Caney River near Elgin, Ks.	42	Altered	37.00	-96.32	439.6
18	07336200	Kiamichi River near Antlers, Okla.	42	Altered	34.25	-95.61	1,158.3
19	07341200	Saline River near Lockesburg, Ark.	42	Altered	33.96	-94.06	259.3
20	07152000	Chikaskia River near Blackwell, Okla.	42	Altered	36.81	-97.28	1,921.6
21	07243500	Deep Fork near Beggs, Okla.	42	Altered	35.67	-96.07	2,056.2
22	07247500	Fourche Maline near Red Oak, Okla.	44	Altered	34.91	-95.16	123.5
23	07247000	Poteau River at Cauthron, Ark.	44	Altered	34.92	-94.30	208.8
24	07300500	Salt Fork Red River at Mangum, Okla.	44	Altered	34.86	-99.51	1,380.4
25	07177500	Bird Creek near Sperry, Okla.	44	Altered	36.28	-95.95	930.5
26	07338500	Little River near Idabel, Okla.	42	Altered	33.94	-94.76	1,260
27	07339000	Mountain Fork near Eagletown, Okla.	42	Altered	34.04	-94.62	820.5
28	07316500	Washita River near Cheyenne, Okla.	44	Altered	35.63	-99.67	782.3

Table 3. Distribution of 5th and 95th percentiles for 27 hydrologic indices for the four hydroecological integrity assessment process groups at 28 streams in Oklahoma.

[Distribution types: N, normal; LN, lognormal; EX, exponential; W2, Weibull 2 parameter; W3, Weibull 3 parameter; JS, Johnson Su. 5Q, 5th percentile; 95Q, 95th percentile; 41–44, group designations from Turton and others (2009). Hydrologic indices are described in table 1]

	41			42		
	5Q	95Q	Distribution	5Q	95Q	Distribution
MA01	3.55	410.01	W2	52.11	1,741.29	W2
MA04	52.19	151.38	N	130.35	223.78	LN
MA28	51.73	145.58	N	104.49	173.71	W2
ML01	2.12	112.6	W2	11.16	412.55	LN
ML09	0.55	53.61	W2	0.72	34.6	W2
MH04	56.35	3,291.1	EX	657.43	16,780.43	W2
MH14	19.57	94.54	N	49.44	271.9	LN
MH20	2.66	44.01	W2	7.58	57.67	N
FL03	0	4.84	LN	3.8	10.27	JS
FH01	4.8	14.65	N	8.09	13.61	W2
FH04	2.9	31.78	N	29.79	77.25	N
FH05	4.47	14.1	N	5.74	10.84	N
DL03	0.21	40.19	W2	0.18	17.18	W2
DL05	1.3	101.41	W2	7.16	273.01	LN
DL18	0	25.97	LN	0	37.11	W2
DH02	61.73	7,328.48	W2	1,129.89	28,740.19	W2
DH07	41.84	108.1	N	45.53	88.29	JS
DH10	34.17	94.76	LN	37.31	78.07	LN
DH15	4.59	12.97	LN	6.38	11.59	JS
DH21	31.56	118.72	N	59.12	148.17	JS
DH23	1.19	2.69	LN	1.33	6.82	JS
TA01	0.22	0.57	N	0.19	0.37	N
TL01	233.62	272.37	N	241.76	276.33	LN
TH01	36.94	220.7	N	33.14	175.88	N
RA03	1.43	172.29	W2	31.05	568.34	W2
RA05	0.18	0.31	N	0.17	0.28	N
RA07	0.05	0.14	LN	0.11	0.2	N

	43			44		
	5Q	95Q	Distribution	5Q	95Q	Distribution
MA01	315.56	1,898.69	LN	12.49	294.59	W2
MA04	81	142.91	N	131.15	280.78	N
MA28	51.29	108.6	N	153.49	266.07	N
ML01	151.89	507.03	LN	0.85	27.94	W2
ML09	66.34	213.08	LN	0.03	8.08	W2
MH04	939.04	18,286.59	W2	103.99	4,592.21	W2
MH14	16.86	56.82	N	86.83	1,087.91	W2
MH20	12.64	29.38	LN	2.89	60.49	W2
FL03	0.01	2.13	W2	5.46	12.14	W3
FH01	5.77	10.49	LN	9.66	15.48	N
FH04	7.54	34.7	N	20.32	105.14	N
FH05	4.86	7.36	LN	6.31	15.36	LN
DL03	50.16	169.4	LN	0	5.14	W3
DL05	82.02	351.56	N	0.98	32.98	W2
DL18	0	0		6.09	140.62	W2
DH02	2,149.5	34,131.3	W2	447.79	7,895.62	W2
DH07	61.64	88.78	W2	46.25	138.96	LN
DH10	52.1	64.36	N	52.67	112.7	LN
DH15	6.71	13.78	N	3.92	8.5	N
DH21	77.76	132.91	N	38.44	122.13	N
DH23	1.33	4.07	N	1.3	3.66	LN
TA01	0.48	0.59	N	0.2	0.38	LN
TL01	258.93	280.4	N	236.26	271.14	W3
TH01	82.6	146.14	N	88.46	207.1	N
RA03	45.6	470.91	W2	28.64	206.04	LN
RA05	0.21	0.26	N	0.16	0.28	LN
RA07	0.04	0.08	N	0.14	0.37	LN

Table 4. Observed and predicted hydroecological integrity assessment process grouping with discriminant analysis for 28 Oklahoma streamgages.

[Misclassified gages are in bold. HIP, hydroecological integrity assessment process; No., no., number; USGS, U.S. Geological Survey]

USGS streamgage no.	Actual HIP	Predicted HIP	No. of indices exceeded	Mean percent indices exceeded	Stream type
07196900	42	42	5	11.2	Reference
07197000	41	41	2	25.4	Reference
07151500	41	41	0	0.0	Reference
07189000	43	43	2	11.9	Reference
07195800	41	41	0	0.0	Reference
07337900	42	42	3	15.4	Reference
07249400	42	42	0	0.0	Reference
07335700	42	42	6	23.4	Reference
07249985	42	42	3	3.4	Reference
07184000	44	44	0	0.0	Reference
07315700	44	44	0	0.0	Reference
07187000	43	43	5	24.3	Reference
07191220	41	41	1	2.8	Reference
07186000	43	43	4	4.5	Reference
07170700	44	44	6	33.1	Altered
07332500	41	41	3	14.3	Altered
07172000	42	42	2	7.0	Altered
07336200	42	42	0	0.0	Altered
07341200	42	42	7	8.3	Altered
07152000	42	43	6	70.1	Altered
07243500	42	43	9	32.7	Altered
07247500	44	42	5	9.3	Altered
07247000	44	42	6	56.6	Altered
07300500	44	42	9	24.4	Altered
07177500	44	43	16	243.4	Altered
07338500	42	43	13	42.3	Altered
07339000	42	No class	17	181.3	Altered
07316500	44	41	12	49.8	Altered

Table 5. Correlations between first and second nonmetric multidimensional scaling (NMS) axes of species presence and functional groups with flow variables.

[Correlations in bold typeface have an absolute r value greater than 0.5. Hydrologic indices (variables) are described in table 1. NMS, nonmetric multidimensional scaling]

Variable	Species		Functional group	
	NMS1	NMS2	PC1	PC2
MA01	-0.28	0.32	-0.03	-0.63
MA04	-0.63	-0.04	0.65	-0.16
MA28	-0.53	0.03	0.57	-0.16
ML01	-0.04	0.34	-0.23	-0.58
ML09	0.14	-0.13	-0.23	-0.15
MH04	-0.34	0.33	0.04	-0.67
MH14	-0.40	-0.24	0.55	0.01
MH20	0.25	0.03	-0.17	0.09
FL03	-0.56	-0.06	0.63	0.01
FH01	-0.44	0.27	0.34	-0.29
FH04	-0.61	0.02	0.58	-0.18
FH05	-0.46	0.03	0.45	-0.06
DL03	0.20	-0.16	-0.27	-0.09
DL05	-0.07	0.15	-0.17	-0.49
DL18	-0.41	-0.40	0.59	0.27
DH07	-0.14	-0.70	0.39	0.42
DH10	-0.52	-0.64	0.74	0.32
DH15	0.48	-0.14	-0.41	0.17
DH21	0.05	-0.31	-0.02	0.05
DH23	-0.39	-0.28	0.31	0.03
TA01	-0.52	-0.05	0.59	0.01
TL01	0.15	-0.24	-0.20	-0.10
TH01	-0.12	-0.12	0.00	-0.08
RA03	-0.52	-0.54	0.63	0.08
RA05	-0.41	0.35	0.12	-0.71
RA07	-0.17	0.08	0.03	-0.13

Table 6. Average proportions of species in functional groups for alteration and geographic location of streamgages on the Kiamichi River and Bird Creek, Oklahoma.

[Values in bold typeface are higher and significantly different between reference and altered periods]

	Alteration			Location		
	Reference	Altered	p-value	Eastern	Western	p-value
Tolerance of water quality change						
Intolerant	0.26	0.08	<0.01	0.25	0.04	<0.01
Intermediate	0.39	0.40	0.72	0.44	0.33	0.01
Tolerant	0.36	0.52	0.03	0.31	0.64	<0.01
Tolerance of habitat change						
Intolerant	0.38	0.24	0.02	0.42	0.14	<0.01
Intermediate	0.33	0.37	0.13	0.34	0.36	0.61
Tolerant	0.29	0.40	0.06	0.24	0.50	<0.01
Sensitive benthic species						
No	0.73	0.82	0.01	0.73	0.85	<0.01
Yes	0.27	0.18	0.01	0.27	0.15	<0.01
Lithophilic species						
No	0.54	0.79	<0.01	0.56	0.84	<0.01
Yes	0.46	0.21	<0.01	0.44	0.16	<0.01
Habitat type						
Creek	0.31	0.16	<0.01	0.30	0.13	<0.01
River	0.09	0.10	0.49	0.11	0.07	0.02
River or creek	0.21	0.18	0.13	0.22	0.15	<0.01
Lentic	0.06	0.15	<0.01	0.06	0.16	<0.01
General	0.34	0.41	0.12	0.31	0.49	<0.01
Current speed						
Slow	0.45	0.57	<0.01	0.45	0.60	<0.01
Slow to moderate	0.19	0.17	0.14	0.17	0.18	0.72
Moderate	0.07	0.04	0.03	0.07	0.02	<0.01
Moderate to fast	0.22	0.16	0.03	0.22	0.13	<0.01
Fast	0.02	0.00	0.01	0.01	0.00	0.11
General	0.07	0.06	0.78	0.07	0.06	0.75
Trophic group						
Generalist	0.07	0.06	0.25	0.06	0.08	0.15
Herbivore	0.05	0.03	0.02	0.05	0.03	0.07
Invertivore	0.64	0.59	0.21	0.69	0.50	<0.01
Omnivore	0.09	0.15	0.08	0.07	0.21	<0.01
Piscivore	0.14	0.18	0.02	0.14	0.19	<0.01
Reproductive guild						
A_1	0.36	0.46	<0.01	0.37	0.47	<0.01
A_2	0.23	0.14	0.01	0.23	0.10	<0.01
B_1	0.01	0.02	0.09	0.01	0.01	0.29
B_2	0.38	0.36	0.32	0.36	0.39	0.29
C_2	0.02	0.03	0.65	0.02	0.03	0.05
Conservation status						
Exotic	0.01	0.02	0.07	0.01	0.04	<0.01
S1	0.02	0.02	0.75	0.02	0.01	0.42
S2	0.02	0.03	0.38	0.03	0.00	0.01
S3	0.09	0.09	0.76	0.10	0.08	0.26
S4	0.23	0.22	0.82	0.23	0.22	0.68
S5	0.64	0.61	0.37	0.61	0.65	0.26
Richness						
Family richness	9.8	11.3	0.08	10.4	10.7	0.73
Species richness	37.4	43.9	0.28	43.3	36.5	0.27

Table 7. Differences between periods in median and coefficient of dispersion values for hydrologic parameters for the Kiamichi River, Oklahoma.

[Results are based on indicators of hydrologic analysis between pre- and post-impoundment periods. Values in bold typeface are significantly different between pre- and post-impoundment periods. CD, coefficient of dispersion; L–RVA, boundary for lower range of natural variability; H–RVA, boundary for upper range of natural variability; $p(\text{Med.})$, proportion of randomized trials where difference between pre and post impoundment median values were greater than observed values; $p(\text{CD})$, proportion of randomized trials where difference between pre and post impoundment CD values were greater than observed values]

	Pre-impoundment				Post Sardis Lake impoundment			
	Median	CD	L-RVA	H-RVA	Median	CD	$p(\text{Med.})$	$p(\text{CD})$
Magnitude of monthly water conditions								
October	90	2.9	14	172	31	5.4	0.38	0.22
November	178.5	2.9	96	482.1	339	5.4	0.19	0.11
December	356	1.6	246.4	676	1,100	2.0	<0.01	0.54
January	641	1.3	401.4	943	553	3.5	0.62	<0.01
February	986	1	555.4	1,376	872.5	1.6	0.91	0.16
March	1,100	1	793	1,520	1,420	1	0.26	0.79
April	1,055	1.5	845.2	1,610	1,077	2.2	0.87	0.08
May	916	1.3	593	1,368	1,350	1.6	0.10	0.46
June	272	1.8	155.7	507.1	322.5	3.1	0.32	0.09
July	53	2.8	26.5	127.6	46	2.1	0.71	0.59
August	27	2.9	6.2	57.5	12	5.1	0.24	0.16
September	40	2.7	14.7	84.1	35.5	2.2	0.80	0.68
Magnitude and duration of annual extreme conditions								
1-day minimum flow	0.6	11.7	0.1	3.9	1.5	2.8	0.13	0.22
3-day minimum flow	0.6	12.5	0.1	4.2	1.9	2.4	0.10	0.23
7-day minimum flow	1	11.9	0.2	4.5	2.4	2.2	0.13	0.23
30-day minimum flow	5.5	5	2.9	18.7	6.5	1.2	0.72	0.23
90-day minimum flow	91.4	2.6	46.2	150.3	77.8	1.1	0.58	0.31
1-day maximum flow	31,100	0.6	24,800	38,960	23,100	0.6	0.07	0.69
3-day maximum flow	26,070	0.7	21,580	34,460	18,630	0.7	0.04	0.98
7-day maximum flow	14,880	0.7	13,700	21,410	11,160	0.7	0.04	0.99
30-day maximum flow	6,630	0.7	4,974	8,173	5,299	0.7	0.17	0.86
90-day maximum flow	3,494	0.7	3,178	4,703	3,419	0.5	0.76	0.32
# of zero-flow days	0	0	0	0	0	0		
Timing of annual extreme water conditions								
Julian date of 1-day max	253	0.1	232	272.5	258	0.1	0.82	0.60
Julian date of 1-day min	124	0.5	106.5	154.4	136	0.5	0.41	0.94
Frequency and duration of high and low pulses								
# of low pulses per year	4	0.8	4	5	4	0.8	0.71	0.85
Duration of low pulses	9	1.2	8	15.3	10.5	1.3	0.48	0.81
# of high pulses per year	12	0.4	11	14.5	14	0.3	0.05	0.25
Duration of high pulses	5	0.4	4	5.5	4.5	0.7	0.29	0.05
Rate and frequency of water condition changes								
Rise rates	235	1.4	124.2	343.8	201	1.4	0.68	0.91
Fall rates	-48	-1.1	-68.1	-32.5	-45	-1.6	0.91	0.08
# of reversals	73	0.2	66	78	80	0.2	0.04	0.99

Table 8. Differences between periods in median and coefficient of dispersion values for hydrologic parameters for Bird Creek, Oklahoma. [Results are based on indicators of hydrologic analysis between pre- and post-impoundment periods. Values in bold typeface are significantly different between pre and post impoundment periods. CD, coefficient of dispersion; L–RVA, boundary for lower range of natural variability; H–RVA, boundary for upper range of natural variability; $p(\text{Med.})$, proportion of randomized trials where difference between pre and post impoundment median values were greater than observed values; $p(\text{CD})$, proportion of randomized trials where difference between pre and post impoundment CD values were greater than observed values]

	Pre-impoundment				Post Birch Lake impoundment				Post Skiatook Lake impoundment			
	Median	CD	L–RVA	H–RVA	Median	CD	$p(\text{Med.})$	$p(\text{CD})$	Median	CD	$p(\text{Med.})$	$p(\text{CD})$
Magnitude of monthly water conditions												
October	12	6.8	3.6	64.2	8.7	1.3	0.81	0.34	154	0.5	<0.01	0.18
November	19	3.9	6.2	45.5	18.5	5.7	0.98	0.68	79	6.2	<0.01	0.49
December	25	3.0	7.6	53	25	0.6	0.98	0.52	114	1.5	<0.01	0.23
January	37	2.3	15.6	76.8	19	0.9	0.48	0.37	88	2.4	<0.01	0.92
February	36	2.9	15.1	90.3	87	2.1	0.17	0.77	107	1.5	<0.01	0.28
March	60	3.4	36.0	147.4	93	2.5	0.52	0.78	220	6.0	0.01	0.25
April	137	1.9	47.6	251	204	1.4	0.69	0.76	249	3.8	0.01	0.20
May	153	1.5	98.6	244	446	4.4	0.02	0.35	315	5.2	<0.01	0.03
June	91.5	1.6	39.8	153.9	181.5	1.5	0.06	0.95	273	5.6	<0.01	0.07
July	27	2.8	15.1	69	27	1.9	0.92	0.70	183	0.7	<0.01	0.19
August	10	2.2	5.3	19	15	1.2	0.48	0.70	159	0.3	<0.01	0.25
September	10	7.0	6.5	42.2	9.4	0.9	0.88	0.35	159.5	0.1	<0.01	0.31
Magnitude and duration of annual extreme conditions												
1-day minimum flow	1.2	3.4	0.2	2.1	1.8	0.6	0.58	0.46	49	0.4	<0.01	0.21
3-day minimum flow	1.2	3.5	0.3	2.2	2	0.6	0.48	0.41	50.7	0.3	<0.01	0.19
7-day minimum flow	1.6	3.0	0.5	3.8	2.3	0.8	0.60	0.38	54.1	0.3	<0.01	0.12
30-day minimum flow	3.6	2.8	1.3	8.6	5.5	0.5	0.46	0.45	64.6	0.3	<0.01	0.09
90-day minimum flow	17	4.3	10.7	48.6	21.2	0.7	0.79	0.27	132.3	0.8	<0.01	0.13
1-day maximum flow	13,800	0.7	10,760	17,330	9,410	0.6	0.20	0.95	12,900	0.7	0.77	0.70
3-day maximum flow	9,853	0.8	7,404	12,440	6,540	0.6	0.27	0.84	8,130	0.9	0.62	0.56
7-day maximum flow	5,645	1.0	3,754	6,864	3,942	0.9	0.31	0.88	6,613	0.7	0.43	0.40
30-day maximum flow	2,169	0.8	1,479	2,607	1,954	1.2	0.77	0.65	3,115	0.8	0.04	0.95
90-day maximum flow	1,136	1.0	791.2	1,340	1,215	0.8	0.79	0.82	1,429	1.4	0.10	0.25
# of zero-flow days	0	0	0	0	0	0.0			0	0.0		
Base flow index	0	2.5	0	0	0	3.0	0.34	0.77	0.06	1.7	<0.01	0.59
Timing of annual extreme water conditions												
Julian date of 1-day max	257	0.1	246.5	275	283	0.1	0.11	0.63	311	0.2	<0.01	0.40
Julian date of 1-day min	163	0.4	133.1	226.4	139	0.2	0.63	0.12	128	0.3	0.12	0.44

	Pre-impoundment				Post Birch Lake impoundment				Post Skiatook Lake impoundment			
	Median	CD	L-RVA	H-RVA	Median	CD	p(Med.)	p(CD)	Median	CD	p(Med.)	p(CD)
Frequency and duration of high and low pulses												
# of low pulses per year	4	1.3	2.5	5.5	4	1.5	0.73	0.67	0	0	0.21	0.02
Duration of low pulses	11.5	1	9	16.8	5	0.6	0.24	0.43	5.5		0.12	
# of high pulses per year	13	0.4	11	14	10	0.4	0.14	0.93	14	0.4	0.09	0.75
Duration of high pulses	4	0.5	3.0	5	5.5	1.0	0.09	0.24	6	0.7	<0.01	0.34
Rate and frequency of water condition changes												
Rise rates	25	2.4	11.1	46.7	8	1.3	0.30	0.70	22	2.7	0.75	0.82
Fall rates	-9	-1.1	-11	-5	-9	-0.8	0.80	0.70	-16.5	-1.2	<0.01	0.83
# of reversals	69	0.2	65	76	82	0.2	0.05	0.53	112	0.1	<0.01	0.03

Table 9. Correlation coefficients and p-values for best fit environmental vectors describing relationships between environmental flow variables and fish assemblage structure in the Kiamichi River and Bird Creek, Oklahoma.

[Values are based on rotational vector fitting between variables and axes 1 and 2 NMS scores for each site. Bold values represent significant relationships. na, not available]

Hydrologic parameter	Kiamichi River		Bird Creek	
	<i>r</i>	<i>P</i>	<i>r</i>	<i>P</i>
Magnitude of monthly water conditions				
October	0.37	0.182	0.72	0.001
November	0.23	0.526	0.44	0.118
December	0.14	0.801	0.66	0.002
January	0.13	0.809	0.59	0.011
February	0.25	0.505	0.25	0.523
March	0.04	0.986	0.28	0.412
April	0.08	0.918	0.43	0.117
May	0.24	0.525	0.11	0.887
June	0.41	0.126	0.53	0.045
July	0.37	0.202	0.75	0.002
August	0.49	0.051	0.71	0.001
September	0.51	0.024	0.51	0.033
Magnitude and duration of annual extreme conditions				
1-day minimum flow	0.73	0.001	0.76	0.001
3-day minimum flow	0.72	0.001	0.78	0.001
7-day minimum flow	0.73	0.001	0.78	0.001
30-day minimum flow	0.56	0.013	0.73	0.001
90-day minimum flow	0.34	0.244	0.74	0.002
1-day maximum flow	0.26	0.447	0.44	0.126
3-day maximum flow	0.19	0.678	0.47	0.084
7-day maximum flow	0.21	0.624	0.56	0.022
30-day maximum flow	0.24	0.538	0.44	0.092
90-day maximum flow	0.26	0.447	0.39	0.171
# of zero-flow days	0.48	0.058	na	na
Base flow index	0.42	0.117	0.67	0.002
Timing of annual extreme conditions				
Julian date of 1-day max	0.41	0.117	0.2	0.684
Julian date of 1-day min	0.39	0.118	0.14	0.832
Frequency and duration of high and low pulses				
# of low pulses per year	0.3	0.347	0.73	0.001
Duration of low pulses	0.19	0.641	0.35	0.264
# of high pulses per year	0.3	0.317	0.01	0.999
Duration of high pulses	0.4	0.132	0.17	0.739
Rate and frequency of water condition changes				
Rise rates	0.24	0.505	0.079	0.939
Fall rates	0.44	0.079	0.33	0.292
# of reversals	0.57	0.009	0.77	0.001
Discharge at time of sampling	0.78	0.001	0.67	0.005

Table 10. Fish species that were significant indicators of pre- or post-impoundment periods on the Kiamichi River and Bird Creek, Oklahoma.

[Periods are divided into pre and post for the categorical variable “Dam.” Additionally, periods were divided into pre (pr), transitional (tran), and post for the categorical variable “Time.” Significance (*p*) is based on a Monte Carlo randomization test where the proportion of randomized trials with indicator value (indval) equal to or exceeding the observed indicator value]

Common Name	Variable “Dam”			Variable “Time”		
	Period	Indval	<i>p</i>	Period	Indval	<i>p</i>
Kiamichi River						
Blacktail shiner	—	—	—	post	51.6	0.013
Brook silverside	pre	74.6	0.002	pre	59.5	0.006
Spotted bass	pre	54.3	0.037	—	—	—
Largemouth bass	post	61.1	0.045	tran	49.2	0.003
Bird Creek						
Central stoneroller	post	84.2	0.001	tran	56.6	0.007
Ghost shiner	post	84	0.001	—	—	—
Sand shiner	post	42.9	0.05	tran	46.9	0.041
Suckermouth minnow	post	71.2	0.005	—	—	—
Bullhead minnow	post	67.5	0.007	post	67.2	0.001
Inland silverside	—	—	—	post	49	0.031
Western mosquitofish	post	62.5	0.039	post	47	0.042
Largemouth bass	post	56.9	0.026	—	—	—
White crappie	—	—	—	post	57.7	0.012
Slenderhead darter	post	50	0.024	—	—	—

Appendix 1.—Family, Scientific Name, Common Name, and Functional Group Membership for Species Observed in This Study

Family	Scientific name	Common name
Aphredoderidae	<i>Aphredoderus sayanus</i>	Pirate perch
Aphredoderidae	<i>Labidesthes sicculus</i>	Brook silverside
Aphredoderidae	<i>Menidia beryllina</i>	Inland silverside
Catostomidae	<i>Carpiodes carpio</i>	River carpsucker
Catostomidae	<i>Carpiodes cyprinus</i>	Quillback
Catostomidae	<i>Catostomus commersonii</i>	White sucker
Catostomidae	<i>Cycleptus elongatus</i>	Blue sucker
Catostomidae	<i>Erimyzon sucetta</i>	Lake chubsucker
Catostomidae	<i>Hypentelium nigricans</i>	Northern hog sucker
Catostomidae	<i>Ictiobus bubalus</i>	Smallmouth buffalo
Catostomidae	<i>Ictiobus cyprinellus</i>	Largemouth buffalo
Catostomidae	<i>Ictiobus niger</i>	Black buffalo
Catostomidae	<i>Minytrema melanops</i>	Spotted sucker
Catostomidae	<i>Moxostoma carinatum</i>	River redhorse
Catostomidae	<i>Moxostoma duquesnei</i>	Black redhorse
Catostomidae	<i>Moxostoma erythrurum</i>	Golden redhorse
Catostomidae	<i>Moxostoma macrolepidotum</i>	Shorthead redhorse
Centrarchidae	<i>Ambloplites ariommus</i>	Shadow bass
Centrarchidae	<i>Ambloplites rupestris</i>	Rock bass
Centrarchidae	<i>Centrarchus macropterus</i>	Flier
Centrarchidae	<i>Lepomis cyanellus</i>	Green sunfish
Centrarchidae	<i>Lepomis gulosus</i>	Warmouth sunfish
Centrarchidae	<i>Lepomis humilis</i>	Orangespotted sunfish
Centrarchidae	<i>Lepomis macrochirus</i>	Bluegill sunfish
Centrarchidae	<i>Lepomis marginatus</i>	Dollar sunfish
Centrarchidae	<i>Lepomis megalotis</i>	Longear sunfish
Centrarchidae	<i>Lepomis microlophus</i>	Redear sunfish
Centrarchidae	<i>Lepomis miniatus</i>	Redspotted sunfish
Centrarchidae	<i>Lepomis symmetricus</i>	Bantam sunfish
Centrarchidae	<i>Micropterus dolomieu</i>	Smallmouth bass
Centrarchidae	<i>Micropterus punctulatus</i>	Spotted bass
Centrarchidae	<i>Micropterus salmoides</i>	Largemouth bass
Centrarchidae	<i>Pomoxis annularis</i>	White crappie
Centrarchidae	<i>Pomoxis nigromaculatus</i>	Black crappie
Clupeidae	<i>Dorosoma cepedianum</i>	Gizzard shad
Clupeidae	<i>Dorosoma petenense</i>	Threadfin shad
Cottidae	<i>Cottus carolinae</i>	Banded sculpin
Cyprinidae	<i>Campostoma anomalum</i>	Central stoneroller

Family	Scientific name	Common name
Cyprinidae	<i>Campostoma oligolepis</i>	Largescale stoneroller
Cyprinidae	<i>Carassius auratus</i>	Goldfish
Cyprinidae	<i>Cyprinella camura</i>	Bluntnose shiner
Cyprinidae	<i>Cyprinella lutrensis</i>	Red shiner
Cyprinidae	<i>Cyprinella spiloptera</i>	Spotfin shiner
Cyprinidae	<i>Cyprinella venusta</i>	Blacktail shiner
Cyprinidae	<i>Cyprinella whipplei</i>	Steelcolor shiner
Cyprinidae	<i>Cyprinus carpio</i>	Common carp
Cyprinidae	<i>Erimystax x-punctatus</i>	Gravel chub
Cyprinidae	<i>Hybognathus placitus</i>	Plains minnow
Cyprinidae	<i>Hybopsis amblops</i>	Bigeye chub
Cyprinidae	<i>Hybopsis amnis</i>	Pallid shiner
Cyprinidae	<i>Luxilus cardinalis</i>	Cardinal shiner
Cyprinidae	<i>Luxilus chrysocephalus</i>	Striped shiner
Cyprinidae	<i>Luxilus pilsbryi</i>	Dusky stripe shiner
Cyprinidae	<i>Lythrurus fumeus</i>	Ribbon shiner
Cyprinidae	<i>Lythrurus snelsoni</i>	Ouachita Mountain shiner
Cyprinidae	<i>Lythrurus umbratilis</i>	Redfin shiner
Cyprinidae	<i>Macrhybopsis australis</i>	Prairie chub
Cyprinidae	<i>Macrhybopsis storeriana</i>	Silver chub
Cyprinidae	<i>Nocomis asper</i>	Redspot chub
Cyprinidae	<i>Notemigonus crysoleucas</i>	Golden shiner
Cyprinidae	<i>Notropis atherinoides</i>	Emerald shiner
Cyprinidae	<i>Notropis atrocaudalis</i>	Blackspot shiner
Cyprinidae	<i>Notropis bairdi</i>	Red River shiner
Cyprinidae	<i>Notropis blennius</i>	River shiner
Cyprinidae	<i>Notropis boops</i>	Bigeye shiner
Cyprinidae	<i>Notropis buechanani</i>	Ghost shiner
Cyprinidae	<i>Notropis greeniei</i>	Wedgespot shiner
Cyprinidae	<i>Notropis nubilus</i>	Ozark minnow
Cyprinidae	<i>Notropis ortenburgeri</i>	Kiamichi shiner
Cyprinidae	<i>Notropis percobromus</i>	Carmine shiner
Cyprinidae	<i>Notropis perpallidus</i>	Peppered shiner
Cyprinidae	<i>Notropis potteri</i>	Chub shiner
Cyprinidae	<i>Notropis shumardi</i>	Silverband shiner
Cyprinidae	<i>Notropis stramineus</i>	Sand shiner
Cyprinidae	<i>Notropis suttkusi</i>	Rocky shiner
Cyprinidae	<i>Notropis volucellus</i>	Mimic shiner
Cyprinidae	<i>Opsopoeodus emiliae</i>	Pugnose shiner
Cyprinidae	<i>Phenacobius mirabilis</i>	Suckermouth minnow
Cyprinidae	<i>Phoxinus erythrogaster</i>	Southern redbelly dace
Cyprinidae	<i>Pimephales notatus</i>	Bluntnose minnow

Family	Scientific name	Common name
Cyprinidae	<i>Pimephales promelas</i>	Fathead minnow
Cyprinidae	<i>Pimephales tenellus</i>	Slim minnow
Cyprinidae	<i>Pimephales vigilax</i>	Bullhead minnow
Cyprinidae	<i>Semotilus atromaculatus</i>	Creek chub
Cyprinodontidae	<i>Cyprinodon rubrofluviatilis</i>	Red River pupfish
Elassomatidae	<i>Elassoma zonatum</i>	Banded pygmy sunfish
Esocidae	<i>Esox americanus</i>	Grass pickerel
Esocidae	<i>Esox niger</i>	Chain pickerel
Fundulidae	<i>Fundulus blaire</i>	Western starhead topminnow
Fundulidae	<i>Fundulus catenatus</i>	Northern studfish
Fundulidae	<i>Fundulus chrysotus</i>	Golden topminnow
Fundulidae	<i>Fundulus kansae</i>	Northern plains killifish
Fundulidae	<i>Fundulus notatus</i>	Blackstripe topminnow
Fundulidae	<i>Fundulus olivaceus</i>	Blackspotted topminnow
Fundulidae	<i>Fundulus sciadicus</i>	Plains topminnow
Fundulidae	<i>Fundulus zebrinus</i>	Plains killifish
Hiodontidae	<i>Hiodon alosoides</i>	Goldeye
Ictaluridae	<i>Ameiurus melas</i>	Black bullhead
Ictaluridae	<i>Ameiurus natalis</i>	Yellow bullhead
Ictaluridae	<i>Ictalurus furcatus</i>	Blue catfish
Ictaluridae	<i>Ictalurus punctatus</i>	Channel catfish
Ictaluridae	<i>Noturus eleutherus</i>	Mountain madtom
Ictaluridae	<i>Noturus exilis</i>	Slender madtom
Ictaluridae	<i>Noturus flavus</i>	Stonecat
Ictaluridae	<i>Noturus gyrinus</i>	Tadpole madtom
Ictaluridae	<i>Noturus miurus</i>	Brindled madtom
Ictaluridae	<i>Noturus nocturnus</i>	Freckeled madtom
Ictaluridae	<i>Noturus placidus</i>	Neosho madtom
Ictaluridae	<i>Pylodictis olivaris</i>	Flathead catfish
Lepisosteidae	<i>Lepisosteus oculatus</i>	Spotted gar
Lepisosteidae	<i>Lepisosteus osseus</i>	Longnose gar
Lepisosteidae	<i>Lepisosteus platostomus</i>	Shortnose gar
Moronidae	<i>Morone chrysops</i>	White bass
Moronidae	<i>Morone chrysops x Morone saxatilis</i>	Wiper
Moronidae	<i>Morone mississippiensis</i>	Yellow bass
Moronidae	<i>Morone saxatilis</i>	Striped bass
Percidae	<i>Ammocrypta clara</i>	Western sand darter
Percidae	<i>Ammocrypta vivax</i>	Scaly sand darter
Percidae	<i>Crystallaria asprella</i>	Crystal darter
Percidae	<i>Etheostoma asprigene</i>	Mud darter
Percidae	<i>Etheostoma blennioides</i>	Greenside darter
Percidae	<i>Etheostoma chlorosoma</i>	Bluntnose darter

Family	Scientific name	Common name
Percidae	<i>Etheostoma cragini</i>	Arkansas darter
Percidae	<i>Etheostoma flabellare</i>	Fantail darter
Percidae	<i>Etheostoma gracile</i>	Slough darter
Percidae	<i>Etheostoma histrio</i>	Harlequin darter
Percidae	<i>Etheostoma microperca</i>	Least darter
Percidae	<i>Etheostoma nigrum</i>	Johnny darter
Percidae	<i>Etheostoma proeliare</i>	Cypress darter
Percidae	<i>Etheostoma punctulatum</i>	Stippled darter
Percidae	<i>Etheostoma radiosum</i>	Orangebelly darter
Percidae	<i>Etheostoma spectabile</i>	Orangethroat darter
Percidae	<i>Etheostoma stigmaeum</i>	Speckled darter
Percidae	<i>Etheostoma whipplei</i>	Redfin darter
Percidae	<i>Etheostoma zonale</i>	Banded darter
Percidae	<i>Percina caprodes</i>	Logperch
Percidae	<i>Percina copelandi</i>	Channel darter
Percidae	<i>Percina fulvitaenia</i>	Ozark logperch
Percidae	<i>Percina macrolepida</i>	Bigscale logperch
Percidae	<i>Percina maculata</i>	Blackside darter
Percidae	<i>Percina nasuta</i>	Longnose darter
Percidae	<i>Percina pantherina</i>	Leopard darter
Percidae	<i>Percina phoxocephala</i>	Slenderhead darter
Percidae	<i>Percina sciera</i>	Dusky darter
Percidae	<i>Percina shumardi</i>	River darter
Percidae	<i>Sander canadensis</i>	Sauger
Petromyzontidae	<i>Ichthyomyzon castaneus</i>	Chestnut lamprey
Petromyzontidae	<i>Ichthyomyzon gagei</i>	Southern brook lamprey
Poeciliidae	<i>Gambusia affinis</i>	Mosquitofish
Sciaenidae	<i>Aplodinotus grunniens</i>	Freshwater drum