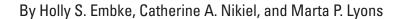


Open-File Report 2021–1104–E



Open-File Report 2021-1104-E

U.S. Geological Survey, Reston, Virginia: 2023

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Conversion Factors

International System of Units to U.S. customary units

| Multiply | Ву | To obtain |
|-------------------------------|-----------|-------------------------------|
| | Length | |
| centimeter (cm) | 0.3937 | inch (in.) |
| meter (m) | 3.281 | foot (ft) |
| meter (m) | 1.094 | yard (yd) |
| kilometer (km) | 0.6214 | mile (mi) |
| | Flow rate | |
| cubic meter per second (m³/s) | 70.07 | acre-foot per day (acre-ft/d) |
| meter per second (m/s) | 3.281 | foot per second (ft/s) |

Temperature in degrees Celsius (°C) may be converted to degrees Fahrenheit (°F) as follows:

$$^{\circ}F = (1.8 \times ^{\circ}C) + 32.$$

Abbreviations

LST lake surface temperature

MACAv2 Multivariate Adaptive Constructed Analogs version 2

NBS net basin supply

RCP4.5 representative concentration pathway 4.5 RCP8.5 representative concentration pathway 8.5

> greater than

< less than

By Holly S. Embke, 1 Catherine A. Nikiel, 2 and Marta P. Lyons 1

Abstract

Acipenser fulvescens (Rafinesque, 1817; lake sturgeon) are the only sturgeon species native to the Great Lakes region and are threatened across most of their range. They are historically vulnerable because of overfishing and habitat fragmentation with the potential for climate change acting as an increasing stressor in the future. Lake sturgeon span multiple habitats during their long lifespans, including high gradient streams, nearshore areas, and deep rivers and lakes. Climate change is projected to strongly affect the suitability of these habitats through increasing precipitation and temperatures and decreasing ice cover and snowmelt. Changes in flow timing and amount can affect movement to spawning and nursery sites, and increased water temperatures are likely to affect species activity patterns, survival, and prey availability. Ultimately, the Great Lakes region is expected to face wide ranging effects from climate change, which may have positive and negative effects on lake sturgeon depending on a variety of factors, including the life stages affected, habitat availability, and interactions with other stressors, but all shifts are likely to affect spawning. Importantly, several areas of further research would be beneficial to understanding the complex effects of climate change on lake sturgeon.

Introduction

Acipenser fulvescens (Rafinesque, 1817; lake sturgeon) have a wide geographical range (fig. 1), and the largest lake sturgeon populations are in the St. Lawrence River east of Montreal, Canada; Lake of the Woods/Rainy River, United States; Lake Winnebago system, United States; and Lake St. Clair, Canada/United States; however, other recovering and remnant populations exist (Kampa and others, 2014; Bruch and others, 2016; Schloesser and Quinlan, 2019; GLIFWC Climate Change Team, 2023). For example, a lake sturgeon population is recovering in the St. Louis River, United States (Stults and others, 2016). Currently, 2 Canadian

Provinces have commercial fisheries, whereas 3 U.S. States and 5 Canadian Provinces have recreational fisheries, and 2 U.S. States (Michigan and Wisconsin) and 5 Canadian Provinces support indigenous fisheries (Bruch and others, 2016). Harvest, where practiced, is highly regulated, resulting in predominantly legal harvest.

Lake sturgeon are strongly tied to multiple habitats used during different life stages. Lake sturgeon make large (greater than [>] 200 kilometers [km]) spawning migrations to fast-moving streams in the spring; however, these fish spend a brief part of their lifetime spawning given they are late maturing and intermittently reproduce (Welsh and others, 2017). Most of their time is spent in nearshore lakes and rivers, which provide critical habitat for larvae and juveniles, or in deeper lakes and rivers where adults predominantly spend their nonspawning time. These different life stages are dependent on suitable habitat conditions, which are likely to be affected by climate (fig. 2).

Lake sturgeon have great ecological, subsistence, and cultural value in the Great Lakes region (Stults and others, 2016). Given the expansive movements lake sturgeon make throughout their long lifetimes, these fish connect habitats and food webs across a wide spatial extent. Additionally, for some indigenous nations including the Menominee Indian Tribe of Wisconsin and the Little River Band of Ottawa Indians, lake sturgeon are a foundational component of indigenous history and considered a close relative responsible for supporting food sovereignty (GLIFWC Climate Change Team, 2023). For recreational anglers, lake sturgeon support a highly popular ice fishery in some areas (for example, the Lake Winnebago system). However, climate change is expected to disproportionately affect many of these relations among lake sturgeon, indigenous communities, and anglers because of geographically limited boundaries of reservations and ceded territories, which may limit communities' ability to follow species' shifts (Stults and others, 2016).

The purpose of this report is to provide an overview, based on peer-reviewed literature, government reports, and analysis of regional climate data where appropriate, of the direct and indirect effects of climate change on lake sturgeon. This report is structured around lake sturgeon life stages, habitat needs, and associated stressors. Because of the wide geographic range and mobility of the species, we limit the analysis to a broad consideration of climate effects in the

¹U.S. Geological Survey.

²Oak Ridge Institute for Science Education Research Participation Program.



Figure 1. Acipenser fulvescens (Rafinesque, 1817; lake sturgeon) current range, historical range expanse, and U.S. Great Lakes Basin (adapted from the Committee on the Status of Endangered Wildlife in Canada [2006] and NatureServe [2022]).

U.S. Great Lakes Basin because this region supports the largest native lake sturgeon populations (fig. 1). We do not provide specific information on Lake St. Clair given limited

data despite its importance to lake sturgeon; however, these trends are somewhat captured in the Lake Huron and Lake Erie information.

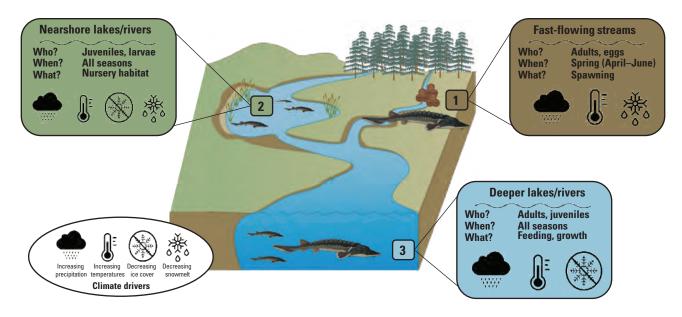


Figure 2. Summary of Acipenser fulvescens (Rafinesque, 1817; lake sturgeon) habitat use and climate drivers.

Data and Methods

To explore the potential range of future climate changes, we used downscaled projections for this region based on the Intergovernmental Panel on Climate Change Coupled Model Intercomparison Project Phase 5 models for two emissions scenarios: a moderate emissions scenario (representative concentration pathway 4.5 [RCP4.5]) and a high emissions scenario (representative concentration pathway 8.5 [RCP8.5]) (Taylor and others, 2012). The output from 20 global climate models with the necessary variables is statistically downscaled using the Multivariate Adaptive Constructed Analogs version 2 (MACAv2) method; this multistep constructed analog approach establishes relations between global climate model output and historical climate observations. Outputs are bias corrected by climate variable to develop higher resolution and localized projections with about a 4-km spatial resolution. The first realization (rlilp1) of each model is used, except for model CCSM4, which uses the sixth realization (r6i1p1). Output can be accessed through https://climate.northwestknow ledge.net/MACA/data portal.php (Abatzoglou and Brown, 2012), and the full list of models used is provided in table 1. MACAv2 was chosen because it is widely vetted and used for impact studies, is available at the appropriate spatial and temporal resolution for the region of interest, and has been documented to better capture signals in rainfall extremes and frequency compared to other downscaled products (Wang and others, 2020; Wootten and others, 2021). Although the distribution of lake sturgeon extends into Canada, data analysis in this report focuses on climate changes and effects in the United States.

Current climate modeling efforts have challenges in the Great Lakes region. Some current global climate models do not represent the Great Lakes (Mallard and others, 2015;

Briley and others, 2021), and higher resolution modeling studies often use simple lake models (Bennington and others, 2014; Mallard and others, 2014; Xiao and others, 2016; Xue and others, 2017; Sharma and others, 2018; Wang and others, 2022). Projections of hydroclimatic variables such as precipitation and evaporation in the basin may be particularly affected by the inclusion of lake models that better capture energy, momentum, and moisture fluxes and feedback, especially during the winter months when low ice cover enhances lake effect precipitation (Notaro and others, 2021, 2022). Analysis in this report uses statistically downscaled projections from the MACAv2 product. The MACAv2 product provides a larger ensemble of projections at a higher resolution than other regional datasets and is used here as a conservative estimate of the full range of potential risk. We do so with an understanding of the potential drawbacks of traditional statistical downscaling methods in regions like the Great Lakes where lake-atmosphere processes driving energy, momentum, and moisture fluxes are critical and likely nonstationary under climate change (Lanzante and others, 2020; Briley and others, 2021). However, large changes in modeled precipitation and temperature because of lake model coupling are often detected overlake and to the north and east of the Great Lakes, which is outside the domain for this study (Bennington and others, 2014; Mallard and others, 2014; Xiao and others, 2016). Where appropriate, projections of change are compared with dynamical downscaling literature to provide a more complete picture of regional change. This comparison includes analysis of two recent dynamical downscaling efforts: an ensemble downscaled using RegCM4 (Notaro and others, 2015a, b) and the multimodel ensemble NA-CORDEX (Mearns and others, 2017).

Table 1. Downscaled global climate models from Multivariate Adaptive Constructed Analogs version 2-METDATA used in this study (Abatzoglou and Brown, 2012).

| Letter ¹ | Model | Citation | |
|---------------------|----------------|--|----------------------------|
| A | BNU-ESM | College of Global Change and Earth System Science, Beijing Normal University (GCESS) | Ji and others (2014) |
| В | CCSM4 | National Center for Atmospheric Research (NCAR) | Gent and others (2011) |
| C | CNRM-CM5 | Centre National de Recherches Meteorologiques/Centre Européen de Recherche et Formation Avances en Calcul Scientifique (CNRM-CERFACS) | Voldoire and others (2013) |
| D | CSIRO-Mk3-6-0 | Commonwealth Scientific and Industrial Research Organisation in collaboration with the Queensland Climate Change Centre of Excellence (CSIRO–QCCCE) | Rotstayn and others (2010) |
| Е | CanESM2 | Canadian Centre for Climate Modeling and Analysis (CCCma) | Arora and others (2011) |
| 7 | GFDL-ESM2G | Geophysical Fluid Dynamics Laboratory (NOAA-GFDL) | Dunne and others (2012) |
| G | GFDL-ESM2M | Geophysical Fluid Dynamics Laboratory (NOAA-GFDL) | Dunne and others (2012) |
| Н | HadGEM2-CC365 | Met Office Hadley Centre (MOHC) | Collins and others (2011) |
| I | HadGEM2–ES365 | Met Office Hadley Centre (additional HadGEM2–ES realizations contributed by Instituto Nacional de Pesquisas Espaciais) (MOHC) | Collins and others (2011) |
| J | IPSL-CM5A-LR | Institut Pierre-Simon Laplace (IPSL) | Marti and others (2010) |
| K | IPSL-CM5A-MR | Institut Pierre-Simon Laplace (IPSL) | Marti and others (2010) |
| L | IPSL-CM5B-LR | Institut Pierre-Simon Laplace (IPSL) | Marti and others (2010) |
| M | MIROC-ESM-CHEM | Japan Agency for Marine-Earth Science and Technology, Atmosphere and Ocean Research Institute (The University of Tokyo), and National Institute for Environmental Studies (MIROC) | Watanabe and others (2010) |
| N | MIROC-ESM | Japan Agency for Marine-Earth Science and Technology, Atmosphere and Ocean Research Institute (The University of Tokyo), and National Institute for Environmental Studies (MIROC) | Watanabe and others (2010) |
| Ο | MIROC5 | Atmosphere and Ocean Research Institute (The University of Tokyo), National Institute for Environmental Studies (MIROC), and Japan Agency for Marine-Earth Science and Technology | Watanabe and others (2010) |
| P | MRI-CGCM3 | Meteorological Research Institute (MRI) | Yukimoto and others (2012) |
| Q | NorESM1-M | Norwegian Climate Centre (NCC) | Bentsen and others (2013) |
| R | bcc-csm1-1-m | Beijing Climate Center, China Meteorological Administration (BCC) | Xiao-Ge and others (2013) |
| S | bcc-csm1-1 | Beijing Climate Center, China Meteorological Administration (BCC) | Xiao-Ge and others (2013) |
| T | inmcm4 | Institute for Numerical Mathematics (INM) | Volodin and others (2010) |

¹Letters index the models and are used to identify panels in figures 3, 4, and 1.1–1.8.

Climate data are presented for four mean periods—a historical period (1971–2000) and three future periods (2010–39, 2040–69, and 2070–99). These 30-year periods provide a long enough climate window to complete a robust analysis of period means and changes. At each future period, projections are provided for two Intergovernmental Panel on Climate Change emissions scenarios: RCP4.5 and RCP8.5. Both of these scenarios are considered equally plausible. Future projections are presented by selecting a spread of three models

that represent the 25th percentile, median, and 75th percentile change in a given variable averaged over the U.S. extent of the Great Lakes Basin (fig. 1). The minimum and maximum change models are shown in appendix 1 (figs. 1.1–1.8). These chosen models indicate a spatially consistent projection from an individual model rather than presenting a combined median and range calculated at the individual grid cells. This presentation highlights how the magnitude and the spatial pattern of changes vary among model projections. Thus, for example, the

driest or warmest model in the early future period (2010–39) may not show the same climate trends as the driest or warmest model in the late period (2070–99) and this presentation reflects the different rates of change among models and over the full future period. For clarity, the model used in each panel is marked by a letter in the bottom right-hand corner of each panel, and the corresponding model is detailed in table 1.

Species Background

Lake sturgeon are long lived (50–150 years) and late maturing (mean age at maturity of 15–25 years) and may spawn intermittently, making them sensitive to disturbances like climate and land-use change (Stults and others, 2016). Additionally, lake sturgeon have multiple thermal and hydrological requirements that may limit their ability to respond to change. Lake sturgeon are potamodromous (freshwater migratory fish) and fluvial dependent, requiring drainage basins that have diverse and unobstructed habitat (>250–300 km) to support their life cycle (Bruch and others, 2016). As opportunistic feeders, lake sturgeon mostly feed on benthic prey (for example, invertebrates and fish eggs) but will also feed on pelagic prey when available (for example, zooplankton and fish; Bruch and others, 2016).

Historically, lake sturgeon experienced mass declines in the late 1800s to early 1900s after European colonization because of habitat fragmentation and extensive overharvest (Bruch and others, 2016; Stults and others, 2016). Despite the closure of commercial and recreational fisheries in many areas, lake sturgeon failed to rebound because of two primary factors: habitat fragmentation and their life history. The construction of dams has fragmented habitat to prevent access to spawning and rearing areas, reducing natural reproduction (Baker, 2006). Additionally, the lake sturgeon life history means that recruitment can be low even with large populations and may not compensate for high adult mortality (Baker, 2006; Stults and others, 2016). The history of decline, persistent effects of habitat fragmentation, and life history make lake sturgeon especially sensitive to additional human disturbance from climate change.

Spawning

Lake sturgeon may spawn intermittently (once every 2–6 years), which presents tradeoffs for population stability and resilience (Baker, 2006). Although unfavorable conditions may be avoided, intermittent spawning may result in fewer opportunities for population recovery from disturbances, leading to overall population decline. The climate connections discussed in this report focus on multidecade trends; however, interannual variability is likely and may affect spawning success in any given year.

Movement to Spawning Sites

Adult lake sturgeon migrate upstream to spawning grounds in response to environmental cues related to the interaction among river flow, water level, and water temperature (Bruch and others, 2016; Ecclestone and others, 2020). Specifically, lake sturgeon move from lakes or deeper, slower river areas to rapid river environments when river flows increase and water temperatures reach 9–18 degrees Celsius (°C; Bruch and others, 2016; Baril and others, 2018; Ecclestone and others, 2020). The rate of water temperature increase also affects spawning timing (Bruch and others, 2016). Historically, the spawning period has been between April and June for lake sturgeon in the Great Lakes region (Baril and others, 2018). Changes in precipitation amount and timing, including extreme events, can alter runoff, lake levels, stream depth, stream velocity, and the timing of peak and low flow. These changes in precipitation can affect the timing and length of the spawning season (Baril and others, 2018).

Warmer air temperature could alter lake sturgeon metapopulation dynamics, for example, resulting in changes in the spawning timing of populations in the same areas (Buchinger and others, 2022). Based on a negative correlation between temperature when spawning happens and site latitude, lake sturgeon populations may be adapted to their local climate, but further research is needed to understand these relations (Baril and others, 2018).

Streamflow

Lake sturgeon are dependent on instream conditions throughout their life cycles; therefore, variability in streamflow may affect population resilience. In the Great Lakes Basin, streamflow is projected to change in velocity and timing. Stream depth and velocity are affected by runoff, which is in turn driven substantially by precipitation amount in this region. Timing and variability of streamflow are affected if the precipitation seasonality and amount change or if snowmelt contribution is reduced or shifted earlier.

Precipitation is expected to increase annually in the basin with possible drying in the summer (June–August; Easterling and others, 2017; fig. 3*A*–*F*; appendix 1, figs. 1.1–1.4). Although many climate models and the mean of the model ensemble indicate increasing precipitation in most seasons, several plausible models project future drying, especially at the mid- and late century in theRCP8.5 scenario. In general, uncertainty in precipitation projections increases with time and emissions scenario. The U.S. Great Lakes Basin mean precipitation changes in MACA2-METDATA are listed in table 2.

The projections documented in table 2 are largely consistent with recent literature projecting increases in annual precipitation, more consistent increases in spring (March–May), and a possibility of decreased precipitation in the summer (Notaro and others, 2015b; Mailhot and others, 2019; Channell and others, 2022a, b, c, d; Kayastha and others, 2022; Shrestha and others, 2022; Xue and others, 2022). Variations

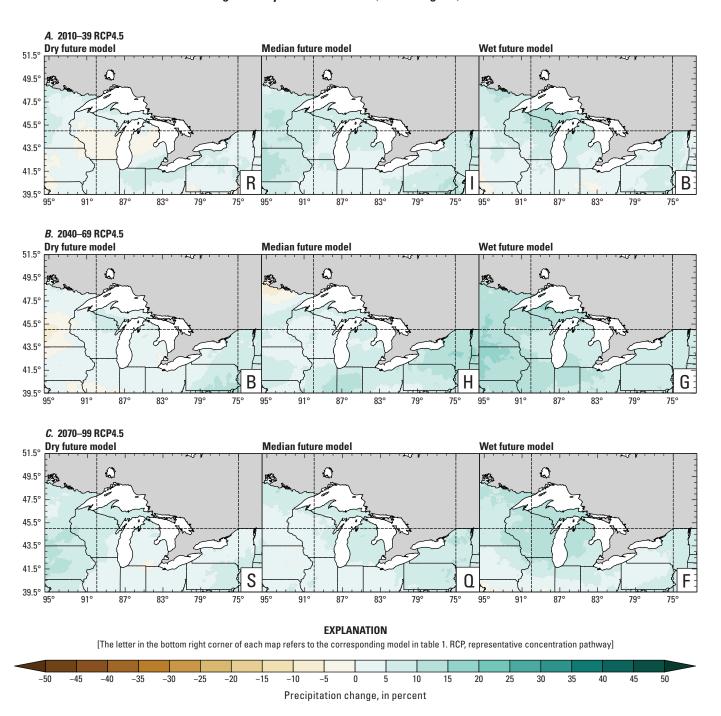


Figure 3. Annual mean total precipitation change (in percent) in the downscaled climate simulations under emissions scenarios for representative concentration pathway 4.5 (RCP4.5) and representative concentration pathway 8.5 (RCP8.5) from the historical period (1971–2000) to the future periods. *A*, early 21st century (2010–39) for RCP4.5; *B*, mid-21st century (2040–69) for RCP4.5; *C*, late 21st century (2070–99) for RCP4.5; *D*, early 21st century (2010–39) for RCP8.5; *E*, mid-21st century (2040–69) for RCP8.5; *F*, late 21st century (2070–99) for RCP8.5.

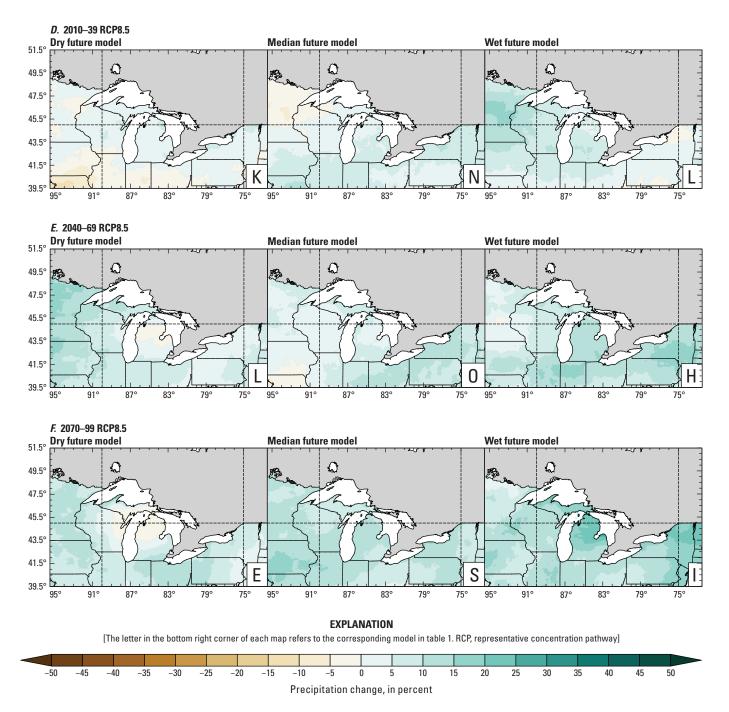


Figure 3. Annual mean total precipitation change (in percent) in the downscaled climate simulations under emissions scenarios for representative concentration pathway 4.5 (RCP4.5) and representative concentration pathway 8.5 (RCP8.5) from the historical period (1971–2000) to the future periods. *A*, early 21st century (2010–39) for RCP4.5; *B*, mid-21st century (2040–69) for RCP4.5; *C*, late 21st century (2070–99) for RCP4.5; *D*, early 21st century (2010–39) for RCP8.5; *E*, mid-21st century (2040–69) for RCP8.5; *F*, late 21st century (2070–99) for RCP8.5.—Continued

Table 2. Change in seasonal and annual precipitation (in percent) averaged over the U.S. Great Lakes Basin from the historical period (1971-2000) for the three future periods under representative concentration pathways 4.5 and 8.5 using Multivariate Adaptive Constructed Analogs version 2 downscaled global climate model data (Abatzoglou and Brown, 2012). Median model values are given.

[Bold values indicate range spans of zero. The range between the wettest and driest model is given in parentheses. RCP, representative concentration pathway]

| Davied | RCP4.5 | | | RCP8.5 | | |
|----------------------------|---------------|----------------|---------------|---------------|----------------|----------------|
| Period | 2010–39 | 2040-69 | 2070–99 | 2010–39 | 2040–69 | 2070–99 |
| Annual (January–December) | +5.2 | +6.6 | +6.9 | +4.1 | +6.4 | +10.4 |
| | (-1.1, +8.8) | (+1.3, +15.0) | (-2.0, +18.2) | (+0.6, +9.9) | (-2.4, +18.5) | (-1.8, +22.9) |
| Winter (December–February) | +7.6 | +9.8 | +10.7 | +9.4 | +14.3 | +21.0 |
| | (-3.4, +18.1) | (-2.4, +31.9) | (-4.3, +33.0) | (-5.4, +28.6) | (-5.3, +33.3) | (-18.0, +55.4) |
| Spring (March–May) | +7.6 | +10.2 | +12.0 | +6.7 | +14.2 | +21.3 |
| | (-3.4, +15.8) | (+0.6, +23.8) | (-4.6, +31.0) | (+3.4, +16.6) | (+2.3, +27.5) | (+11.5, +45.5) |
| Summer (June–August) | +2.5 | -0.1 | -1.1 | +1.1 | -3.8 | -7.2 |
| | (-9.2, +11.1) | (-11.6, +11.6) | (-9.6, +18.9) | (-4.1, +6.3) | (-23.2, +13.5) | (-28.0, +18.4) |

in projections can be attributed to differences in area considered, periods, and model variability. The high likelihood of increased precipitation in the winter and spring is important because it may contribute to earlier and increased streamflows.

In contrast, the depth of snowpack is projected to decrease with earlier snowmelt because of increasing air temperatures (Dyer and Mote, 2006; Notaro and others, 2011; Mailhot and others, 2019; Shrestha and others, 2022). This melt timing may push snowmelt-associated peak streamflow earlier (Dudley and others, 2017; Byun and others, 2019; Dhakal and Palmer, 2020) while reducing the number of snowmelt-driven peak events and the amount of snowmelt runoff (Notaro and others, 2013; Burn and Whitfield, 2023). Additionally, flooding events may decrease because of reduced snowpack but, when they do happen, are likely to happen earlier in the year (Li and others, 2019). This finding is consistent with historical data for some areas west of Lake Michigan (not shown) where annual streamflow has decreased between 1960 and 2015 (Norton and others, 2019), although another study indicates no consistent trends from 1951 to 2009 (McCabe and Wolock, 2014).

Overall, changing precipitation seasonality, increasing precipitation in most seasons (with more precipitation falling as rain versus snow in most areas), reduced snowpack amount, and altered snowmelt timing may concentrate high flow events earlier in the season. Decreases or increases in spring flow are dependent on snowmelt and rainfall, but flow may increase with increases in winter and spring precipitation. In general, projections point to increases in winter and spring runoff in the Great Lakes Basin, coincident with increased precipitation and earlier snowmelt (Notaro and others, 2015a; Mailhot and others, 2019; Channell and others, 2022a, b, c, d; Kayastha and others, 2022). Lake sturgeon spawning may be affected by increased spring flows because this season is a critical reproductive period. Earlier and more intensive high flow events may result in shifted spawning timing for adult lake sturgeon

leading to unusual or protracted spawning events, inaccessible habitat, and unsuitable conditions for eggs and larvae (see more in the "Spawning and Egg Survival" section).

Water Temperature

Increased lake and stream temperatures, especially in the spring, resulting from changes in air temperature, ice cover, shifting stratification, and snowmelt may lead to earlier spawning and prolong the exposure of lake sturgeon to warmer conditions. Climate change may lead to increased air temperatures across all seasons for the full range of lake sturgeon habitats (fig. 4A-F; appendix 1, figs. 1.5-1.8). Northern basins (for example, Ontario) are expected to warm more than southern areas in the winter and spring, whereas southern and central areas (for example, Michigan) are expected to warm more in the summer. Air temperatures in the northern Great Lakes Basin have already increased more than other parts of the basin, with the largest increases in the winter (Vose and others, 2017). Models consistently project these changes, with winter continuing to warm more than other seasons (fig. 4A-F; appendix 1, figs. 1.5-1.8; Xue and others, 2022). The Great Lakes Basin mean temperature change in MACA2-METDATA is listed in table 3.

Earlier ice melt is linked to warming air temperatures. Historically, total ice cover has decreased. Although this change can be quantified linearly as a decrease of 71 percent from 1973 to 2010 in the Great Lakes region (Wang and others, 2012), it is more appropriate to view the change as a regime shift from high mean ice cover (69 percent) to low mean ice cover (36 percent) in 1998 (Van Cleave and others, 2014). Ice-out dates happened 2–5 days earlier over the 1950-2012 period, and the freeze-free season has lengthened on inland lakes across northeastern Minnesota (Stults and others, 2016); however, low ice in one year can lead to high ice conditions the next year by modifying warming and evaporation rates (Wang and others, 2012; Spence and others,

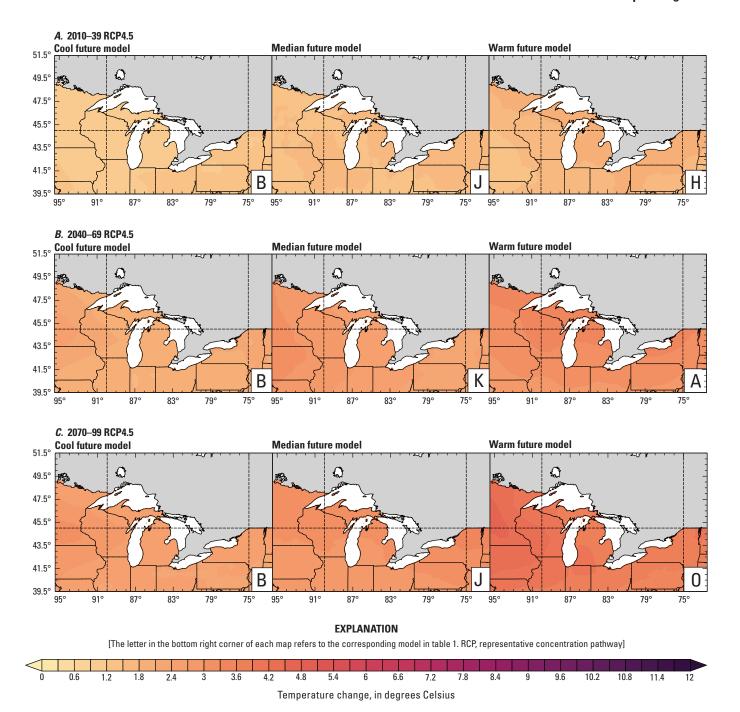


Figure 4. Annual mean 2-meter air temperature change (in degrees Celsius) in the downscaled climate simulations under emissions scenarios for representative concentration pathway 4.5 (RCP4.5) and representative concentration pathway 8.5 (RCP8.5) from the historical period (1971–2000) to the future periods. *A*, early 21st century (2010–39) for RCP4.5; *B*, mid-21st century (2040–69) for RCP4.5; *C*, late 21st century (2070–99) for RCP4.5; *D*, early 21st century (2010–39) for RCP8.5; *E*, mid-21st century (2040–69) for RCP8.5; *F*, late 21st century (2070–99) for RCP8.5.



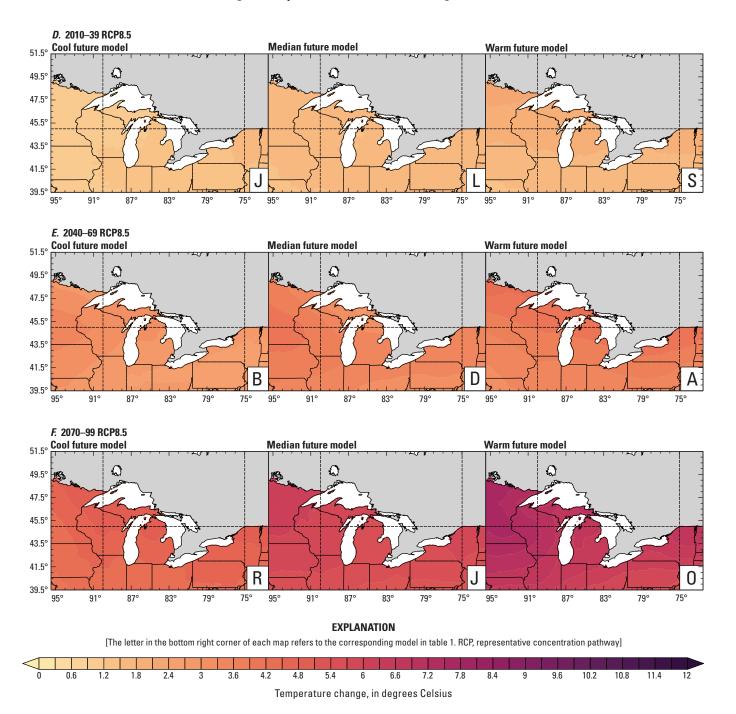


Figure 4. Annual mean 2-meter air temperature change (in degrees Celsius) in the downscaled climate simulations under emissions scenarios for representative concentration pathway 4.5 (RCP4.5) and representative concentration pathway 8.5 (RCP8.5) from the historical period (1971-2000) to the future periods. A, early 21st century (2010-39) for RCP4.5; B, mid-21st century (2040-69) for RCP4.5; C, late 21st century (2070–99) for RCP4.5; D, early 21st century (2010–39) for RCP8.5; E, mid-21st century (2040–69) for RCP8.5; F, late 21st century (2070-99) for RCP8.5.—Continued

Table 3. Change in seasonal and annual temperature (in degrees Celsius) averaged over the Great Lakes Basin from the historical period (1971–2000) for the three future periods under representative concentration pathways 4.5 and 8.5 using Multivariate Adaptive Constructed Analogs version 2 downscaled global climate model data (Abatzoglou and Brown, 2012). Median model values are given.

| [The range between the coolest and warmest model is given in parenth | neses. RCP, representative concentration pathy | vay] |
|--|--|------|
|--|--|------|

| Daviad | RCP4.5 | | | RCP8.5 | | |
|----------------------------|--------------|--------------|--------------|--------------|--------------|--------------|
| Period | 2010–39 | 2040-69 | 2070–99 | 2010–39 | 2040-69 | 2070–99 |
| Annual (January–December) | +1.5 | +2.6 | +3.1 | +1.6 | +3.4 | +5.7 |
| | (+0.4, +2.3) | (+1.5, +4.2) | (+1.7, +5.0) | (+0.7, +2.5) | (+2.2, +5.2) | (+3.7, +8.2) |
| Winter (December–February) | +1.6 | +3.1 | +3.7 | +2.0 | +3.8 | +6.2 |
| | (+0.1, +2.9) | (+1.5, +5.5) | (+1.9, +6.5) | (+0.4, +3.4) | (+2.1, +6.4) | (+3.8, +9.9) |
| Spring (March-May) | +1.3 | +2.3 | +2.8 | +1.1 | +2.8 | +4.6 |
| | (+0.3, +2.9) | (+1.1, +5.1) | (+1.7, +5.7) | (+0.4, +3.2) | (+1.6, +6.2) | (+3.2, +8.5) |
| Summer (June–August) | +1.4 | +2.6 | +3.2 | +1.5 | +3.6 | +5.6 |
| | (+0.4, +2.0) | (+0.9, +3.5) | (+1.5, +5.0) | (+0.5, +2.5) | (+1.7, +4.7) | (+2.8, +9.0) |

2013). Additionally, natural variability and very cold winters can contribute to higher ice cover in some years (Wang and others, 2012; Spence and others, 2013); therefore, projections of decreased mean ice coverage and duration will be accompanied by strong interannual variability (Austin and Colman, 2007; Xue and others, 2022).

Earlier ice-free days and resulting earlier stratification are likely to lead to increased spring lake temperatures (Xue and others, 2022). Although lake temperatures may not be substantially affected by spring runoff, reduced contribution of snowmelt to streamflow may increase the diurnal variability and mean of stream temperature and increase sensitivity to changing air temperatures (Constantz, 1998; Ficklin and others, 2013; Lisi and others, 2015). Warming spring water temperatures are a key spawning signal for adult lake sturgeon to begin migratory movements; therefore, these changes may combine with other signals (for example, high flows) to trigger earlier and potentially prolonged spawning periods.

Spawning and Egg Survival

Lake sturgeon spawning typically happens on unobstructed gravel shoals and stream rapids from April to June (Stults and others, 2016). From a lake sturgeon range-wide assessment, four habitat characteristics affect spawning success: flow velocity, water depth, water temperature, and substrate size (Baril and others, 2018). Specifically, peak spawning suitability was observed at velocities of 0.6 meter per second, depths of 0.55–0.85 meter in small rivers (less than [<] 100 cubic-meter-per-second mean annual streamflow) and 0.75-5.25 meter in large rivers (>100 cubic meters per second), and over cobble, and suitable water temperatures decreased with increasing latitude (Baril and others, 2018). Once released, eggs drift downstream and yolk-sac larvae emerge about 11-19 days after postpeak spawn, sometimes drifting more than 61 km downstream (Baker, 2006; Bruch and others, 2016). Eggs may have limited hatch success when impeded by sediment or algae (Bruch and others, 2016).

Extreme Events and the Interaction Between Flow and Temperature

Flow and runoff in the Great Lakes Basin are projected to increase in the spring and potentially decrease in the summer, with the potential for increased extreme flood and drought events. Historically, precipitation associated with heavy rainfall events has increased, and more of the increase is associated with higher frequency of events than intensity (Mallakpour and Villarini, 2015; Hoerling and others, 2016). In the Midwest, extreme precipitation events have increased over time, especially in the autumn, with a 42-percent increase in extreme precipitation (99th percentile) from 1958 to 2016 (Easterling and others, 2017). Continued warmer temperatures are expected to contribute to increased heavy precipitation events in the future, although interactions between increasing and variable temperatures and precipitation are difficult to quantify, especially considering model overestimation of historical winter precipitation (Diffenbaugh and others, 2013; Feng and others, 2016; Notaro and others, 2021).

Extreme precipitation and temperature events have a substantial effect on instream conditions experienced by lake sturgeon and their prey. Extreme precipitation events can cause an influx of contaminants and sediments, especially for sites near developed areas, which may coincide with lake sturgeon spawning locations. Large precipitation and runoff events may flush fertilizer and other effluent from surrounding agricultural areas into stream basins. Toxic effluents have been cited as a source of lake sturgeon mortality in some sites such as the St. Lawrence River (Dumont and others, 1987). Effluents also indirectly affect juvenile and adult lake sturgeon by decreasing forage availability (see more in the "Larvae and Juvenile Survival" section; Dumont and others, 1987).

Lake sturgeon eggs may be buried with increased sedimentation from extreme events, resulting in reduced hatching success (Bruch and others, 2016; Stults and others, 2016). In addition, lake sturgeon are vulnerable to low oxygen conditions that may result from reduced water quality (Harkness

and Dymond, 1962). Extreme high and low flow events can contribute to flushing or desiccation risk of eggs. In general, decreased streamflow is associated with decreased survival, growth, and abundance of fish populations and shifts in community composition; therefore, the increased variability in streamflows associated with climate change may result in adverse conditions for lake sturgeon (van der Lee and others, 2022).

Although runoff and streamflow may increase in spring because of increasing precipitation and shifting snowmelt, summer flows may be reduced because of a possible reduction in precipitation, increasing temperatures, and earlier snow melt. Increased air temperatures and heat waves can cause rapid increases in stream temperature. Daily maximum temperatures have increased in the Great Lakes Basin and are projected to become more frequent, with a substantial increase in the number of extremely hot days (Lopez and others, 2018; Zhang and others, 2018; Xie and others, 2021). In addition to signaling the start of spawning migrations, flow variability and water temperature also provide cues for spawning cessation (Baril and others, 2018). Lake sturgeon tend to spawn as river flow decreases and will cease spawning activity if streamflow or water temperature change rapidly (Baril and others, 2018); therefore, increased variability in river flows and temperature conditions may result in abbreviated spawning events with uncertain outcomes for egg and larval survival.

For lake sturgeon, outmigration to a lentic environment could be a response to potentially uncertain conditions, ensuring that fish are not stranded in unfavorable conditions such as complete dewatering or extensive freezing (Ecclestone and others, 2020). If climatic changes lead to lower river levels, spawning areas could become inaccessible to fish, along with reduced overall suitable habitat (Stults and others, 2016). For example, flash droughts caused by high temperatures and low precipitation (Mo and Lettenmaier, 2020) could contribute to spawning period stranding, spawning cessation caused by a rapid change in conditions, or a potential transition to other spawning locations (Moore and others, 2022).

General Water Conditions

Maintenance of adequate river conditions during spawning season is crucial for lake sturgeon reproductive success. Results demonstrate that the mean and variability in temperatures experienced during egg incubation have considerable effects on larval body size and timing of development whereas increasing mean stream temperatures have mixed effects on spawning success (Dammerman and others, 2019). Mean June air temperature was positively correlated with increased abundances of young lake sturgeon (Adams and others, 2006). This correlation may be linked to food availability because early lake sturgeon life stages are dependent on zooplankton and benthic invertebrates; however, increased temperatures are linked to increased fungal transmission to eggs and may

promote the development of benthic algal blooms, resulting in reduced hatching success (Bruch and others, 2016). Therefore, projected increases in temperature may result in enhanced growth and survival of larval lake sturgeon under certain conditions but may adversely affect egg hatching and larval development in other contexts when optimal thresholds are exceeded.

Spawning Conclusions

Given the importance of riverine habitats for lake sturgeon throughout various life stages, shifting streamflow may have implications for adult spawning timing and success. Increasing winter and spring precipitation in combination with earlier snowmelt may result in earlier spring high streamflow events. Spring high streamflow events are key spawning signals for adult lake sturgeon to make migrations from deep-water habitats to fast-flowing, shallow spawning areas; therefore, earlier high streamflow events may result in earlier and prolonged flows suitable for spawning for lake sturgeon. Additionally, increased water temperature resulting from increased air temperatures, less ice and earlier melt times, and earlier spring stratification could also trigger earlier spawning. Although earlier spawning events may result in increased growth periods for eggs and larvae, conditions may not be favorable to support successful egg and larval development given adverse effects from increased extreme events (for example, flood waters scouring eggs). Further, an earlier spawning season, in combination with increased variability in spring conditions (for example, increased extreme precipitation events), could lead to more frequent cues to cease spawning, resulting in condensed spawning events.

High or extreme flow conditions can also lead to increased sedimentation and contaminant influxes, which could adversely affect lake sturgeon spawning ability and be detrimental to prey. Reduced summer precipitation and increased temperatures may lead to low flow conditions and hot periods in some years, which can cue spawning cessation, negatively affecting lake sturgeon and prey, and contribute to algal blooms, leading to reduced egg hatch success. Increased variability in extreme events has the potential to limit habitat availability for adult lake sturgeon and present challenges for egg survival; however, in systems where ample habitat is available, spawning site fidelity has been documented to be weak, and fish may spawn in numerous locations and over protracted periods (Bruch and others, 2016). Diversified spawning sites may lead to increased resilience to protect against adverse conditions in a single year or location (Forsythe and others, 2011; Paradis and others, 2022). The effect shifted spawning timing and conditions may have on population growth, individual body condition, and survival is unclear and highlights key knowledge gaps.

Larvae and Juvenile Survival

After hatching and moving downstream, young lake sturgeon primarily inhabit shallow nearshore wetland and marsh habitat. Water quality issues (for example, nutrients from agricultural runoff, contaminants from industry, and algal blooms) are a particular concern for nearshore areas, especially under low lake-level conditions and changes in the amount and timing of runoff (Abdel-Fattah and Krantzberg, 2014). Given potential increased variability in water levels, wetland water quantity and quality may be adversely affected such that suitable lake sturgeon nursery habitat is reduced.

Larval Movement

Larvae hatch from eggs soon (11–19 days) after egg release and are dependent on flows for movement downstream to nursery sites. Therefore, larval lake sturgeon are dependent on similar water temperature and flow dynamics as spawning conditions. As a result of more variable and earlier peak flows, larvae may be subjected to additional challenges finding suitable nursery habitat because of extreme precipitation events.

Habitat Requirements

Rapid change and interannual variability in Great Lakes Basin lake water levels can cause ecosystem disturbance to crucial larval and juvenile nursery habitat. Lake levels are dictated by the components of net basin supply (NBS; the balance among overlake precipitation, evaporation, and runoff; Deacu and others, 2012). Rising lake levels in the spring from increasing runoff from precipitation and snowmelt are sustained during the summer by low evaporation and decrease in the autumn as evaporation reaches its annual peak.

Great Lakes water levels have gone through several periods of decrease and increase in the last 50 years, with substantial reductions until 2014 when water levels began to rise (Assel and others, 2004; Gronewold and others, 2016; Gronewold and Rood, 2019). The latest models, which have the benefit of dynamical downscaling (Notaro and others, 2015b), lake model coupling (Xue and others, 2017, 2022; Sharma and others, 2018; Notaro and others, 2021; Kayastha and others, 2022), and more accurate energy balance considerations for lake evaporation (Lofgren and others, 2011; Lofgren and Rouhana, 2016), project stable or increasing mean lake levels in the future (Mailhot and others, 2019; Channell and others, 2022a, b, c, d; Kayastha and others, 2022; VanDeWeghe and others, 2022). Runoff seasonality shifts may also lead to an earlier annual increase in lake levels (see more in the "Streamflow" section; VanDeWeghe and others, 2022).

Although the mean change in lake levels is less certain, the interannual variability of the NBS, and consequent lake levels, is likely to increase in the future (Gronewold and Rood, 2019; Seglenieks and Temgoua, 2022; VanDeWeghe and others, 2022). Extreme low and high water levels may have

positive effects on coastal habitat by promoting wetland plant germination and preventing desiccation, respectively (Assel and others, 2004). Extended periods of low water levels or rapid changes in water level, however, may cause degradation of nearshore habitat. Ultimately, evidence indicates that the intensity and rate of fluctuations in lake levels can be harmful to nearshore habitats, and planning for potential increased variability may be critical (Gronewold and Rood, 2019; Theuerkauf and Braun, 2021).

Nearshore ice cover is important to insulate and decrease evaporation in nearshore areas. Sublake-level analysis indicates that ice coverage has decreased more quickly in nearshore areas (Mason and others, 2016) and is generally projected to decrease with strong variations from year to year (Xiao and others, 2016; Xue and others, 2022). Ice cover duration has important implications for lake temperature, the rate of warming, and resulting stratification. Increased air temperatures and decreasing lake ice are likely to lead to earlier and more rapid warming in the spring and summer, which could reduce suitable habitat for juvenile lake sturgeon and lead to increased mortality (see more in the "Water Temperature" section).

Indirectly, shifting water temperatures can affect food availability and cause a zooplankton phenological mismatch when larval lake sturgeon hatch before zooplankton emergence and do not have access to this important food source. Increased temperatures (sometimes from reduced ice cover and water levels) in nearshore areas may also increase diel hypoxic events, resulting in adverse habitat conditions for larvae and juveniles (Tellier and others, 2022).

Larvae and Juvenile Survival Conclusions

The direct and indirect effects of changing ice cover, water levels, and lake temperatures are likely to affect larval and juvenile lake sturgeon in the future. Climate change is likely to affect lake sturgeon larvae and juvenile survival through changes in development timing, habitat conditions, and prey availability. Increasing water temperatures and earlier spring precipitation events may cause earlier hatching, with uncertain consequences for lake sturgeon development and growth. Further, shifts in hatch timing may be misaligned with prey emergences (for example, zooplankton), resulting in decreased food availability for early life stages.

Nearshore wetlands provide key nursery habitats for larval and juvenile lake sturgeon, but increased spring extreme events may flush eggs and larvae away from these habitats. Indirectly, changes in precipitation and decreased ice cover may lead to lower and more variable water levels, further reducing the availability of suitable refuge habitat for larvae and juveniles. The effect projected changes in precipitation, water levels, temperature, and winter conditions may have on early life stages of lake sturgeon is unclear; therefore, considering the interactions among these factors when assessing future habitat conditions and potential refuge is beneficial.

Adult Foraging and Survival

After spawning, adults commonly migrate back to deeper slow flow areas in lakes and rivers (Ecclestone and others, 2020) where they spend most of their life. Higher water temperatures may cause acute mortality or increase stress and thereby affect condition, growth, and mortality rates for adult lake sturgeon. Warmer water temperatures may also negatively affect benthic productivity, specifically of cool-water invertebrates, and increase the incidence of parasites and diseases of sturgeon (Haxton, 2008); however, some prey items, including warm-water invertebrates, may benefit from warmer water temperatures, resulting in increased food resources for lake sturgeon.

Physiological Preference and Limits

Lake sturgeon are a cool-water fish, where adults prefer water temperatures <25 °C (Cech and Doroshov, 2004) and have an optimal temperature for growth at <20 °C (Haxton and others, 2016). Lake sturgeon are highly mobile (for example, spawning migrations >100 km) and spend limited time in hypolimnetic waters (deep areas below the thermocline); therefore, lake sturgeon are more likely to be exposed to warming surface temperatures.

Lake temperatures in the Great Lakes Basin are expected to warm in the future, although warming will not be equal across lakes. Increases in lake surface temperature (LST) are

related, but not equal, to increases in air temperature. For example, LST in Lake Superior increased at double the rate of the overlying air temperature from 1979 to 2006 (Austin and Colman, 2007; O'Reilly and others, 2015). Projections agree that the Great Lakes will continue to warm in the future, but the rate of that warming remains uncertain (Zhong and others, 2016) and requires a broader set of investigations using coupled lake modeling. Warming temperatures, a potential reduction in lake ice, and an earlier onset of thermal lake stratification are expected to enhance summer LST warming (Austin and Colman, 2007; Zhong and others, 2016). Model projections indicate that annual mean lake temperature increases will continue to be higher than surrounding air temperature increases (table 4; Xue and others, 2022). Despite recent warming, LST temperatures are not projected to exceed optimal and preferred ranges for lake sturgeon on average in the summer except slightly in Lake Erie; however, extreme heat events in any given year can further boost summer LST greater than the ranges preferred by lake sturgeon (table 4). Linear trends in LST are complicated by short data spans and strong spatial heterogeneity in warming within the lakes (Mason and others, 2016).

Increased water temperature affects ice cover, evaporation, and stratification onset and duration. Historically, more northern and deeper lakes increasingly warmed in part because of shifting stratification dynamics and shallow lakes warmed from increased spring temperatures and solar radiation (Zhong and others, 2016). Lake stratification affects water circulation

 Table 4.
 Historical and projected summer mean lake surface temperatures.

[LST, lake surface temperature; GLERL/GLSEA, Great Lakes Environmental Research Laboratory/Great Lakes Surface Environmental Analysis; RCP, representative concentration pathway]

| Region | 1995–2021 mean summer LST (GLERL/GLSEA)¹, in degrees Celsius | Historical summer lake warming trends, in degrees Celsius per year; period; and reference | End of century (2000–19 to 2080–99) change in LST², in degrees Celsius, and RCP model |
|---------------|--|--|---|
| Lake Superior | 11.4 | 0.09, (1979–2006), Austin and Colman (2007) 0.14, (1994–2013), Mason and others (2016) | 3.2 (RCP4.5) 6.1 (RCP8.5) |
| Lake Ontario | 19.1 | 0.05, (1968–2002), Dobiesz and Lester (2009) 0.1, (1994–2013), Mason and others (2016) | 2.5 (RCP4.5) 5 (RCP8.5) |
| Lake Huron | 16.8 | 0.08, (1968–2002), Dobiesz and Lester (2009) 0.09, (1994–2013), Mason and others (2016) | 2.5 (RCP4.5) 4.5 (RCP8.5) |
| Lake Erie | 21.7 | 0.06, (1994–2013), Mason and others (2016) | 2 (RCP4.5) 4 (RCP8.5) |
| Lake Michigan | 17.9 | 0.09, (1994–2013), Mason and others (2016) | 2.5 (RCP4.5) 4.5 (RCP8.5) |

¹Historical LST available for download at https://coastwatch.glerl.noaa.gov/statistic/.

²End of century change in LST data are from Xue and others (2022).

and hypolimnetic oxygen availability and amplifies warming. Additionally, high water temperatures can combine with low oxygen levels to cause mass fish mortality events, reducing population abundances and resilience (Hupfeld and others, 2015). Although the Great Lakes Basin is projected to warm, these dynamics will likely not be uniform amongst and within a given lake given interactions between ice cover and stratification, which may affect conditions for organisms such as lake sturgeon.

Effect on Activity

Lake sturgeon are considered a cool-water fish but can live in water temperatures warmer than most of their current habitat (Moore and others, 2022). Warmer temperatures may increase the abundance of lake sturgeon in some areas in which they were once limited (for example, northern Canada; Lemmen and Warren, 2004). Therefore, apart from the shallowest inland lakes and rivers, the projected nearterm warming may extend the growing season and increase productivity for lake sturgeon, although indirect effects in timing of reproduction and overall ecosystem health could result in adverse effects (Magnuson and others, 1997; Stults and others, 2016). Most river-lake networks have cooler segments, large tributaries, or lakes that might provide temporary escape from potentially less suitable temperatures (Lyons and Stewart, 2014). Temperature also affects individual movement, with populations at lower latitudes and those that experience warmer temperatures having lower summer movement relative to spring movement, potentially indicating local adaptation or behavioral plasticity (Moore and others, 2021). Understanding suitable thermal habitat given climate change projections can inform how lake sturgeon may respond to further warming.

Food Availability

As opportunistic, benthic feeders, juvenile and adult lake sturgeon mostly feed on invertebrates and fish eggs, although they may feed on pelagic prey when available (for example, zooplankton and fish; Bruch and others, 2016). Increasing temperatures and associated changes to oxygen may affect the phenological timing and abundance of important prey (for example, zooplankton; Magnuson and others, 1997; Adams and others, 2006); therefore, the seasonal dynamics of projected climate changes in relation to lake sturgeon life stages are important to consider.

Changes to water temperature and precipitation may increase algal blooms, which may exacerbate low oxygen availability caused by an extended stratification period in deep ponds and even in shallower waterbodies that do not achieve full seasonal stratification (Chapra and others, 2017; Shinohara and others, 2023). Algal blooms may also be affected by changes in flow regime and associated inputs of nutrients and contaminants (Chapra and others, 2017). Algal blooms and

changing stratification dynamics may lead to food web shifts such that lake sturgeon are indirectly affected by limnological changes in areas they do not directly inhabit.

Other Biotic Interactions

Rising temperatures can also contribute to higher parasite and disease transmission (Haxton, 2008) and increased interaction with nonnative species. Effects of nonnative species on lake sturgeon are somewhat unknown and difficult to predict, but species introductions could affect lake sturgeon populations through shifts in the food web or alterations to habitat. For lake sturgeon, *Dreissena polymorpha* (Pallas, 1771; zebra mussels) and Neogobius melanostomus (Pallas, 1814; round goby) are of particular concern given their increasing prevalence in some parts of the Great Lakes Basin (Escobar and others, 2018). Climate change effects on river flow (Rahel and Olden, 2008) and temperature may make habitats more suitable for nonnative species, increasing their abundances and distributions leading to further interactions with native fishes such as lake sturgeon (Escobar and others, 2018). Nonnative species may increase competition for prey and directly predate on different lake sturgeon life stages. For example, round gobies have preyed on lake sturgeon eggs, reducing hatch success (Kornis and others, 2012).

Newly introduced species may provide an alternate food source. Zebra mussels may affect the feeding efficiency of lake sturgeon (McCabe and others, 2006) but can also contribute an important dietary component for lake sturgeon larger than >70 centimeters (Jackson and others, 2002). Further, round gobies can be a preferred bait species for lake sturgeon indicating round goby may be a preferred food source (Thomas and Haas, 2002); therefore, lake sturgeon increases in round goby may inadvertently benefit lake sturgeon. Lake sturgeon, being benthivores (feeding predominantly on bottom resources), may offset the negative ecological effects of some nonnative introductions if lake sturgeon populations were of substantial size (Haxton, 2008), but these novel interactions are a key knowledge gap.

Adult Foraging and Survival Conclusions

Climate change has the potential to affect adult foraging and survival in positive and negative ways. Changes in water basin levels and thermocline depth, ice cover reductions, LST increases, summer stratification duration, and hypolimnetic hypoxia increases are projected for the future. These changes may directly affect fish, prey, habitat, and other biotic interactions (Mulvaney and others, 2014). Effects of climate change (for example, warming water temperatures) may lead to increased individual growth and nonnative prey availability for adult lake sturgeon; however, this growth may be offset by losses in native prey and limited habitat availability. Specifically, further research is needed on the potential for algal blooms and hypoxia in deeper lakes; much of the current

research is focused on shallower lakes, particularly Lake Erie (Tellier and others, 2022). Given the potential effects of climate change, whether lake sturgeon will have viable habitat in the future and whether that habitat will align with historical occupancy are two questions with varying implications.

Conclusions

Adult lake sturgeon span multiple habitats during their long lifespans, including high gradient streams, nearshore areas, and deep rivers and lakes (Moore and others, 2021). However, climate change is projected to strongly affect the suitability of these habitats through increasing precipitation and temperatures and decreasing ice cover and snowmelt. These changes may lead to myriad effects on lake sturgeon, including changes in spawning initiation and cessation, egg survival, larval growth, juvenile foraging, and adult movement; however, the effects of the interactions of complex climate drivers on lake sturgeon populations across their range is still unclear.

Higher air temperatures are projected to coincide with reduced ice cover to drive increases in LST, stream temperature, and lake stratification, which may contribute to increased hypoxia and algal blooms. These changes are linked to altered mortality risk and prey availability. Changes in precipitation and snow seasonality, volume, and distribution of extreme events affect streamflow, stream temperature, runoff, and lake levels. These changes are linked to movement and hatching success of eggs and larvae and availability of larval and juvenile nursery habitat. All projected changes are likely to affect spawning timing and duration.

Many of these changes are already being observed, especially in communities where lake sturgeon are economically and culturally important. For example, indigenous fishers and managers have noticed shifts to shorter winters and reduced ice cover on Lake Superior over their careers, contributing to warmer water temperatures, earlier ice out on inland lakes, shifting species compositions in lakes, increasing nonnative species, and an alteration in precipitation timing and intensity (Stults and others, 2016). Despite an understanding of historical and future climatic conditions for the Great Lakes Basin, the effect on lake sturgeon that may be caused by the interaction between shifting dynamics is still unclear.

Several gaps in knowledge have been highlighted in this report that would aid in linking climate drivers to potential ecological responses for lake sturgeon. First, a deeper understanding of the potential adaptive capacity of lake sturgeon across spatial and temporal scales is needed. Lower latitude populations spawn at higher temperatures, but it is unclear if higher latitude populations may adapt to spawning at higher temperatures, especially given the rapid pace of climate change.

Other gaps in knowledge include the following:

• How climate effects on one life stage (for example, a loss of juveniles or larvae in a given spawning year) propagate to overall population trajectories.

- The interaction between increased precipitation and variability on sedimentation and runoff in known spawning areas, which may affect adult spawning habitat and egg hatching success.
- The potential for and effects of nonnative species as a food source and predator, which may affect juvenile and adult abundance, growth, and condition.
- The effect of shifting spawning timing and duration on population dynamics and prey availability, which may affect larval, juvenile, and adult abundance, growth, and condition.
- The effect of changing water levels and reduced ice cover on future wetland conditions, which may affect larval, juvenile, and adult abundance, growth, and condition.

Despite the potential for adverse effects of climate change, some projected shifts may benefit lake sturgeon. Many populations currently live in areas where water temperatures are generally less than their preferred or optimal ranges; therefore, increased temperatures may increase the activity, movement, and prey availability in some areas and promote growth. Earlier snowmelt and peak flow events may lengthen the spawning season in some years. Additionally, lake levels are not likely to drop substantially, sparing wetland habitats from this stressor. Identifying locations where adverse climate effects may be limited could provide opportunities for increased population resilience.

Additional local drivers of change can interact with climate effects to exacerbate shifting dynamics; climate shifts may allow agricultural expansion to formerly unsuitable areas for certain crops and this and other land-use changes may interact with climate drivers (temperature and precipitation) to affect runoff of contaminants and sediment into waterbodies (Chu and Jones, 2011; Burn and Whitfield, 2023). Flow modification using dams, a potential reaction to hydrologic variability or increased anthropogenic water demand, can separate lake sturgeon from spawning grounds. In many areas, lake sturgeon have strong spawning site fidelity and return to their natal river, but dams have fragmented habitat to prevent access to spawning and rearing areas, reducing natural reproduction (Baker, 2006; Bruch and others, 2016). Further, water releases from control structures can affect streamflow, oxygen, and water temperature, which may have subsequent effects on spawning behavior and egg and larval survival (Haxton, 2008). Additional local drivers such as industrial development, deforestation, and groundwater withdrawal may affect habitat suitability for lake sturgeon (Chu and Jones, 2011). Lake sturgeon are a culturally, ecologically, and economically important fish to the Great Lakes Basin; therefore, understanding how climate may interact with other drivers is necessary to support population adaptation and resilience.

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Appendix 1. Climate Maps

In this appendix, climate maps (figs. 1.1–1.8) are provided. These maps show the mean precipitation (figs. 1.1–1.4) and temperature (figs. 1.5–1.8) changes under all emissions scenarios and time periods for winter (December-February), spring (March–May) and summer (June–August) in addition

to annual mean changes. The appendix figures expand on the three panels shown in figures 3 and 4 and show the models that represent the maximum and minimum projected change in a given variable averaged over the U.S. extent of the Great Lakes Basin (fig. 1).

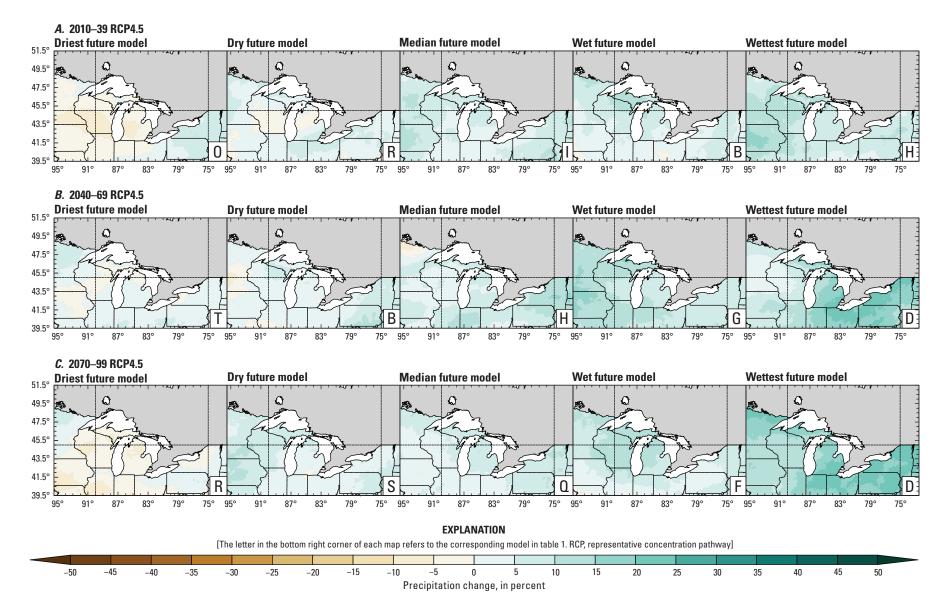


Figure 1.1. Annual mean total precipitation change (in percent) in the downscaled climate simulations under emissions scenarios for representative concentration pathway 4.5 (RCP4.5) and representative concentration pathway 8.5 (RCP8.5) from the historical period (1971–2000) to the future periods. *A*, early 21st century (2010–39) for RCP4.5; *B*, mid-21st century (2040–69) for RCP4.5; *C*, late 21st century (2070–99) for RCP8.5; *E*, mid-21st century (2040–69) for RCP8.5; *F*, late 21st century (2070–99) for RCP8.5.

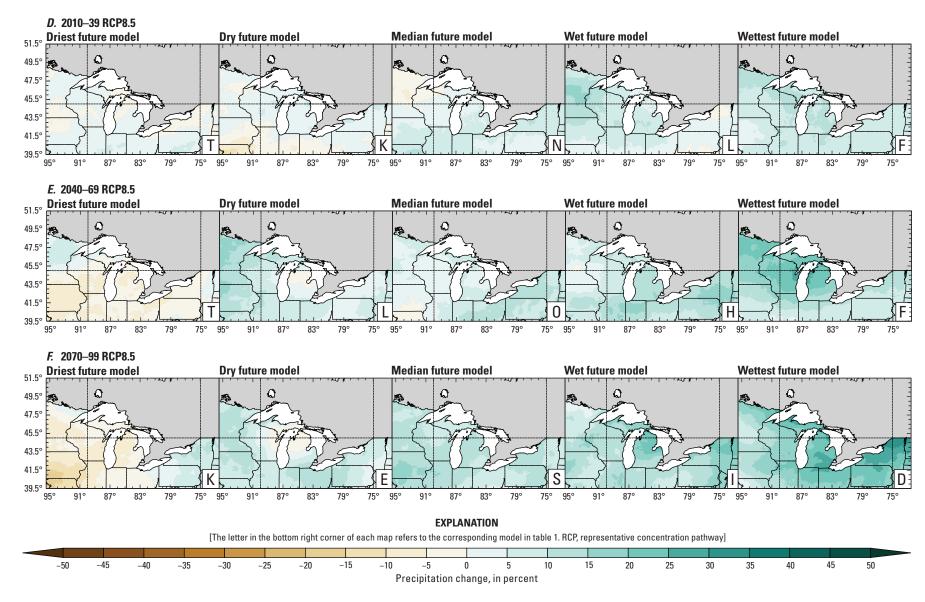


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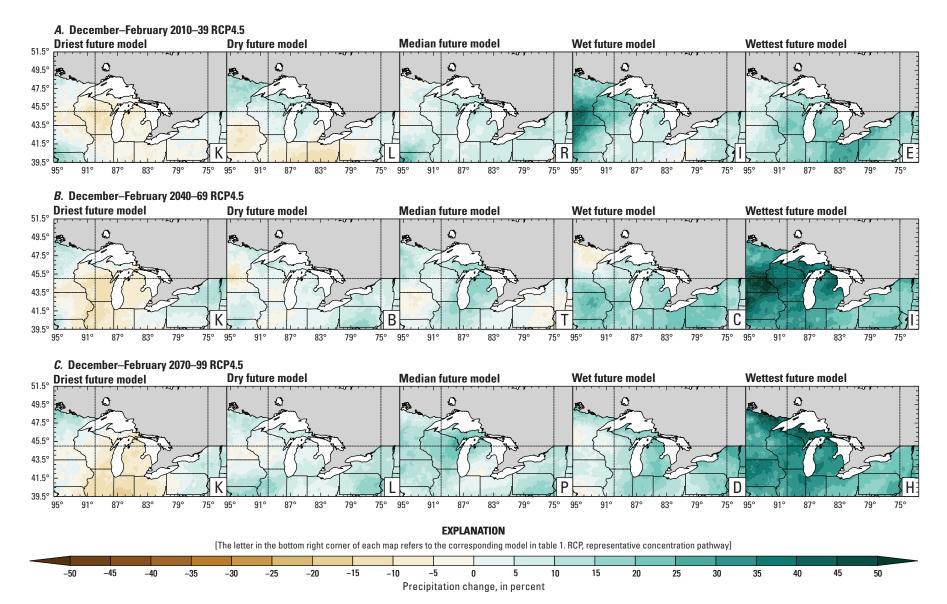


Figure 1.2. December—February mean total precipitation change (in percent) in the downscaled climate simulations under emissions scenarios for representative concentration pathway 4.5 (RCP4.5) and representative concentration pathway 8.5 (RCP8.5) from the historical period (1971–2000) to the future periods. *A*, early 21st century (2010–39) for RCP4.5; *B*, mid-21st century (2040–69) for RCP4.5; *C*, late 21st century (2070–99) for RCP4.5; *D*, early 21st century (2010–39) for RCP8.5; *E*, mid-21st century (2070–99) for RCP8.5.

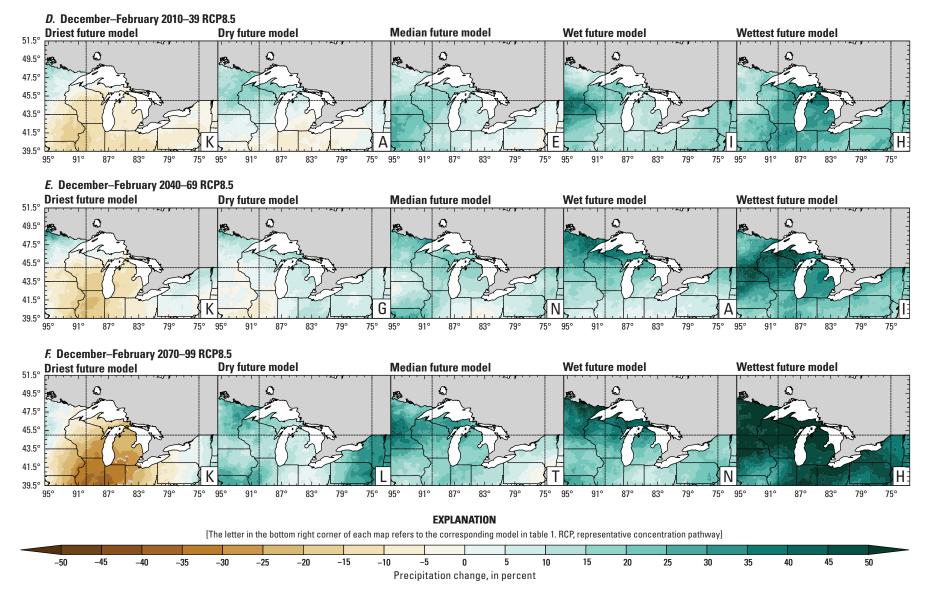


Figure 1.2. December—February mean total precipitation change (in percent) in the downscaled climate simulations under emissions scenarios for representative concentration pathway 4.5 (RCP4.5) and representative concentration pathway 8.5 (RCP8.5) from the historical period (1971–2000) to the future periods. *A*, early 21st century (2010–39) for RCP4.5; *B*, mid-21st century (2040–69) for RCP4.5; *C*, late 21st century (2070–99) for RCP4.5; *D*, early 21st century (2010–39) for RCP8.5; *E*, mid-21st century (2070–99) for RCP8.5.—Continued

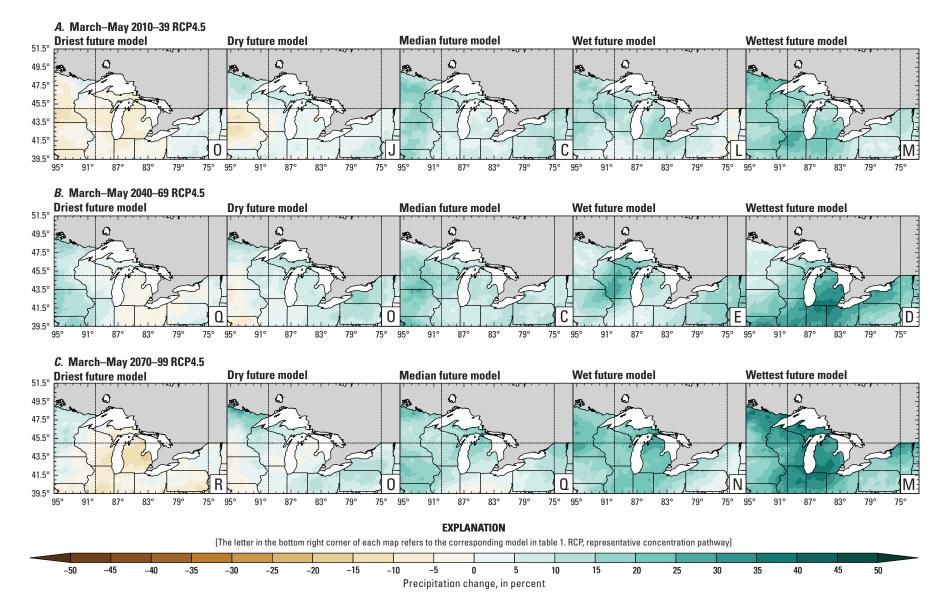


Figure 1.3. March—May mean total precipitation change (in percent) in the downscaled climate simulations under emissions scenarios for representative concentration pathway 4.5 (RCP4.5) and representative concentration pathway 8.5 (RCP8.5) from the historical period (1971–2000) to the future periods. *A*, early 21st century (2010–39) for RCP4.5; *B*, mid-21st century (2040–69) for RCP4.5; *C*, late 21st century (2070–99) for RCP4.5; *D*, early 21st century (2010–39) for RCP8.5; *E*, mid-21st century (2040–69) for RCP8.5.

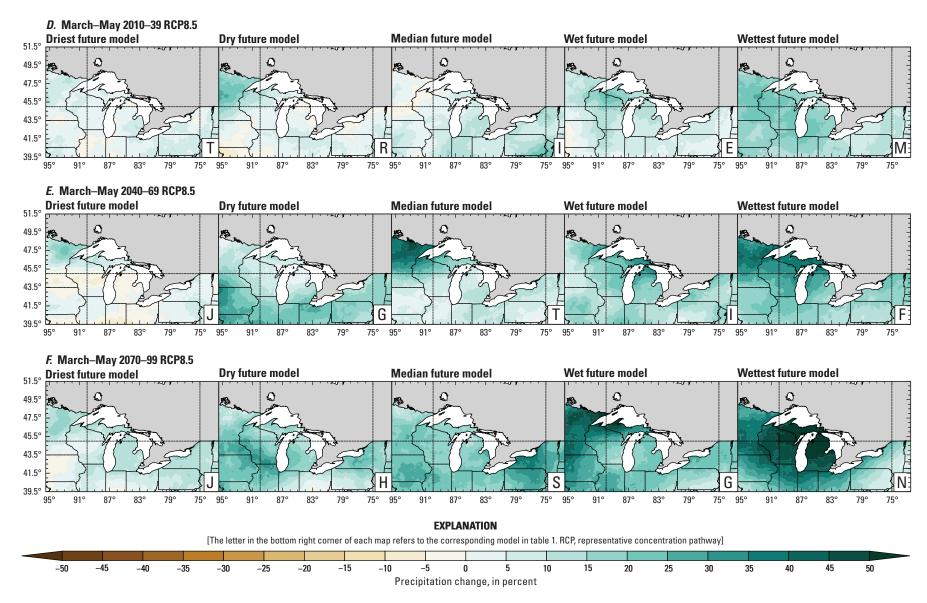


Figure 1.3. March—May mean total precipitation change (in percent) in the downscaled climate simulations under emissions scenarios for representative concentration pathway 4.5 (RCP4.5) and representative concentration pathway 8.5 (RCP8.5) from the historical period (1971–2000) to the future periods. *A*, early 21st century (2010–39) for RCP4.5; *B*, mid-21st century (2040–69) for RCP4.5; *C*, late 21st century (2070–99) for RCP4.5; *D*, early 21st century (2010–39) for RCP8.5; *E*, mid-21st century (2040–69) for RCP8.5.—Continued

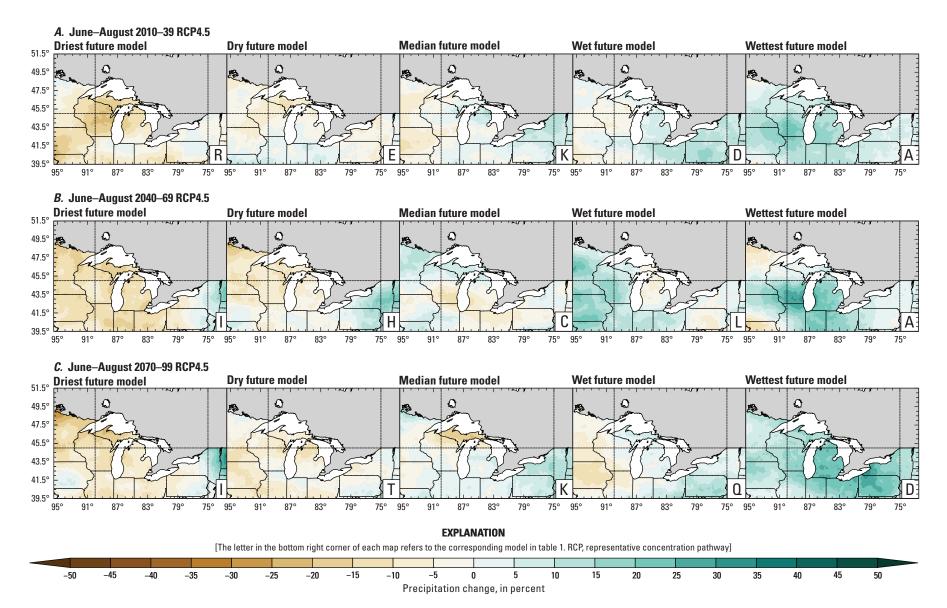


Figure 1.4. June—August mean total precipitation change (in percent) in the downscaled climate simulations under emissions scenarios for representative concentration pathway 4.5 (RCP4.5) and representative concentration pathway8.5 (RCP8.5) from the historical period (1971–2000) to the future periods. *A*, early 21st century (2010–39) for RCP4.5; *B*, mid-21st century (2040–69) for RCP4.5; *C*, late 21st century (2070–99) for RCP4.5; *D*, early 21st century (2010–39) for RCP8.5; *E*, mid-21st century (2070–99) for RCP8.5.

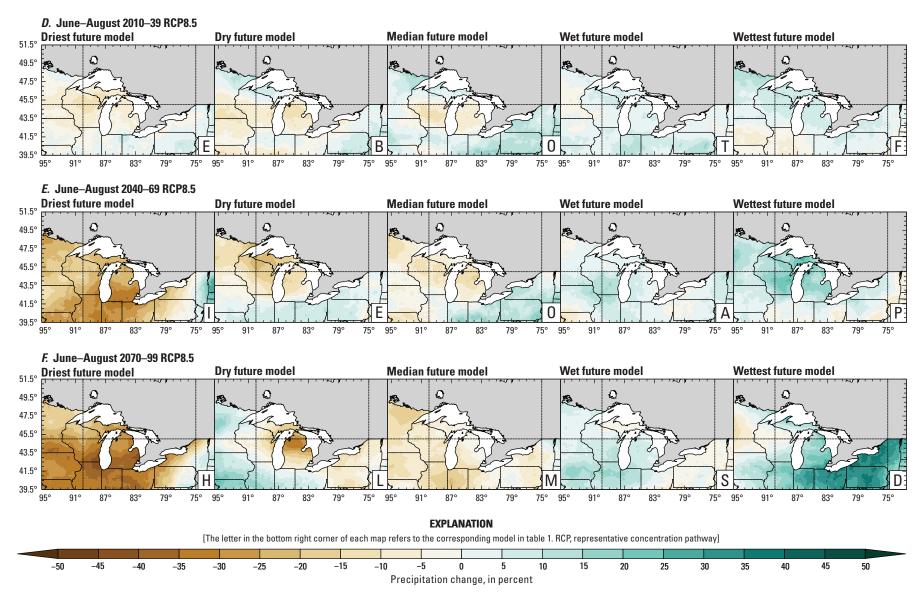


Figure 1.4. June—August mean total precipitation change (in percent) in the downscaled climate simulations under emissions scenarios for representative concentration pathway 4.5 (RCP4.5) and representative concentration pathway8.5 (RCP8.5) from the historical period (1971–2000) to the future periods. *A*, early 21st century (2010–39) for RCP4.5; *B*, mid-21st century (2040–69) for RCP4.5; *C*, late 21st century (2070–99) for RCP8.5; *E*, mid-21st century (2040–69) for RCP8.5; *F*, late 21st century (2070–99) for RCP8.5.—Continued

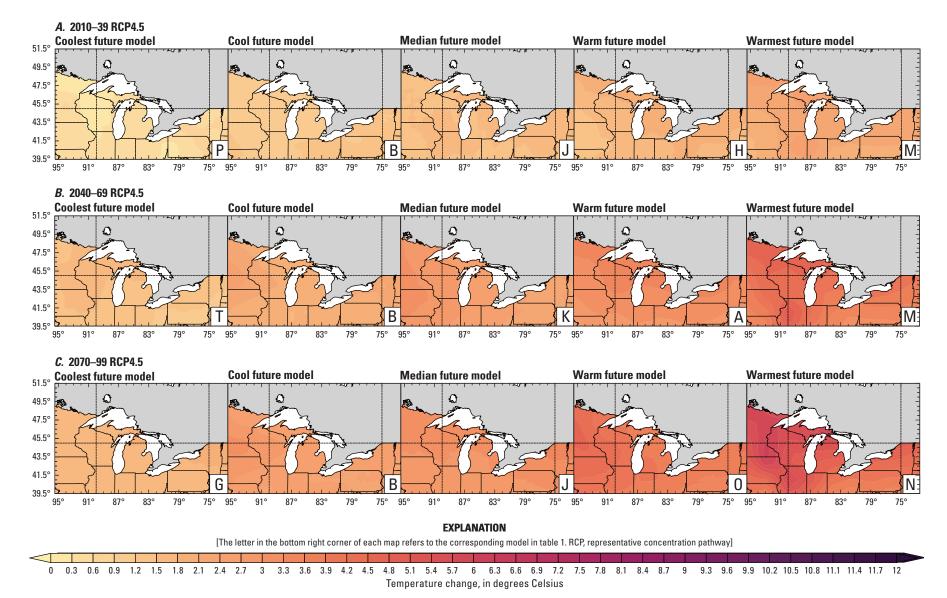


Figure 1.5. Annual mean 2-meter air temperature change (in degrees Celsius) in the downscaled climate simulations under emissions scenarios for representative concentration pathway 4.5 (RCP4.5) and representative concentration pathway 8.5 (RCP8.5) from the historical period (1971–2000) to the future periods. *A*, early 21st century (2010–39) for RCP4.5; *B*, mid-21st century (2040–69) for RCP4.5; *C*, late 21st century (2070–99) for RCP4.5; *D*, early 21st century (2010–39) for RCP8.5; *E*, mid-21st century (2040–69) for RCP8.5; *F*, late 21st century (2070–99) for RCP8.5.

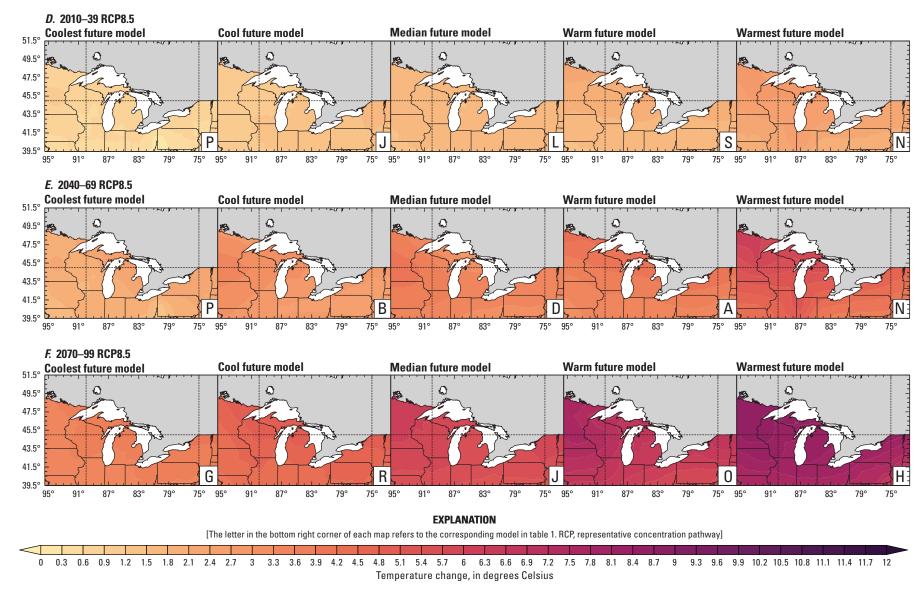


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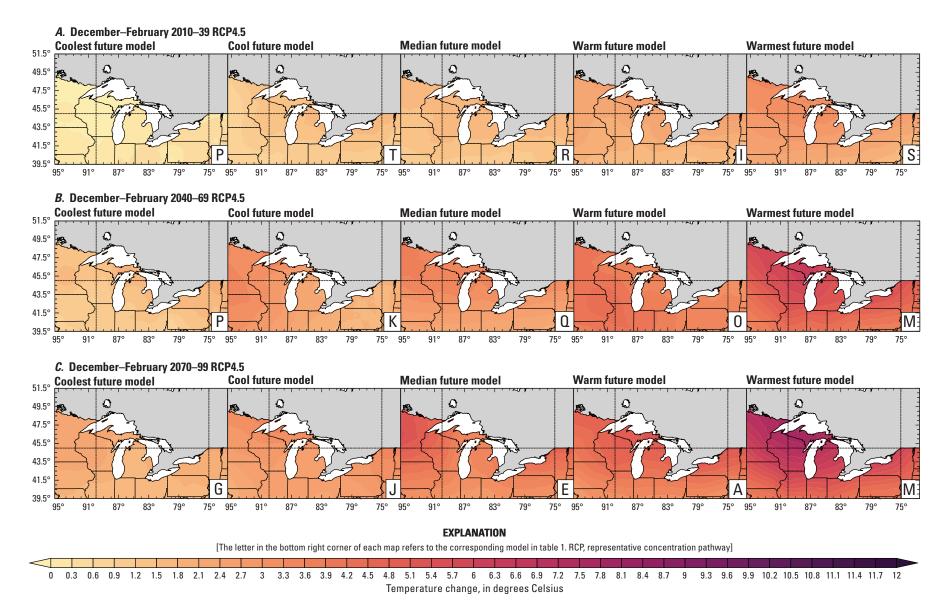


Figure 1.6. December—February mean 2-meter air temperature change (in degrees Celsius) in the downscaled climate simulations under emissions scenarios for representative concentration pathway 4.5 (RCP4.5) and representative concentration pathway 8.5 (RCP8.5) from the historical period (1971–2000) to the future periods. *A*, early 21st century (2010–39) for RCP4.5; *B*, mid-21st century (2040–69) for RCP4.5; *C*, late 21st century (2070–99) for RCP4.5; *C*, late 21st century (2070–99) for RCP8.5; *E*, mid-21st century (2070–99) for RCP8.5.

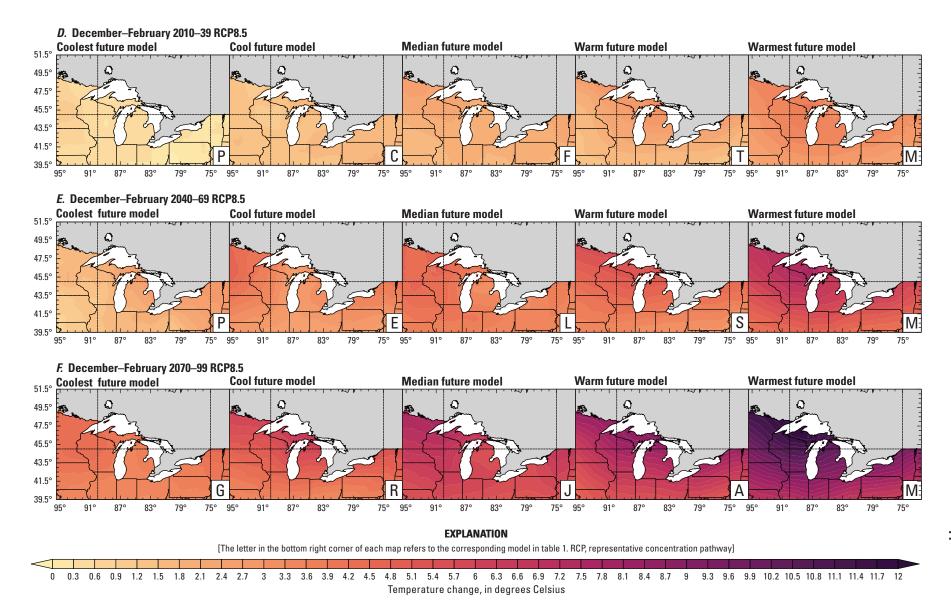


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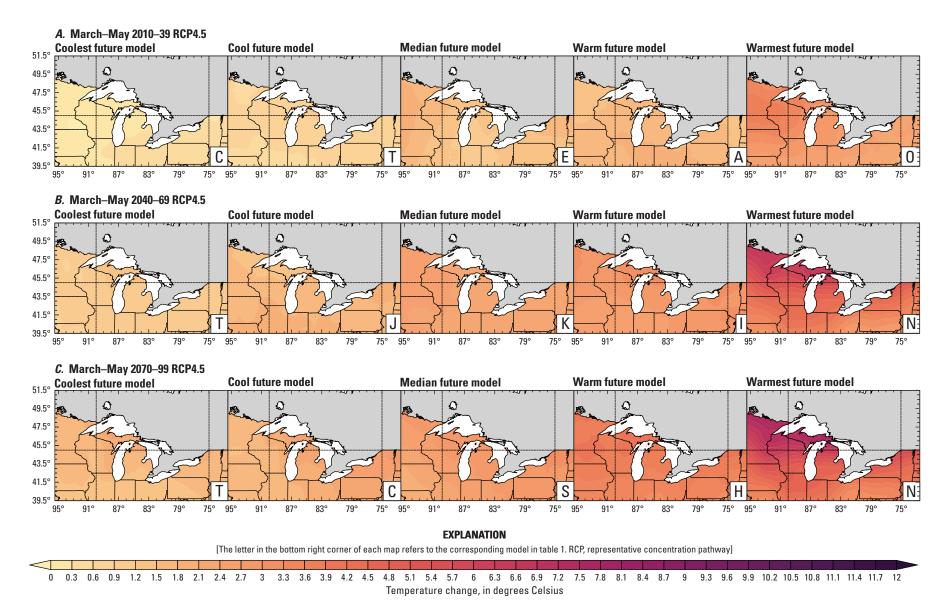


Figure 1.7. March—May mean 2-meter air temperature change (in degrees Celsius) in the downscaled climate simulations under emissions scenarios for representative concentration pathway 4.5 (RCP4.5) and representative concentration pathway 8.5 (RCP8.5) from the historical period (1971–2000) to the future periods. *A*, early 21st century (2010–39) for RCP4.5; *B*, mid-21st century (2040–69) for RCP4.5; *C*, late 21st century (2070–99) for RCP4.5; *D*, early 21st century (2010–39) for RCP8.5; *E*, mid-21st century (2070–99) for RCP8.5.

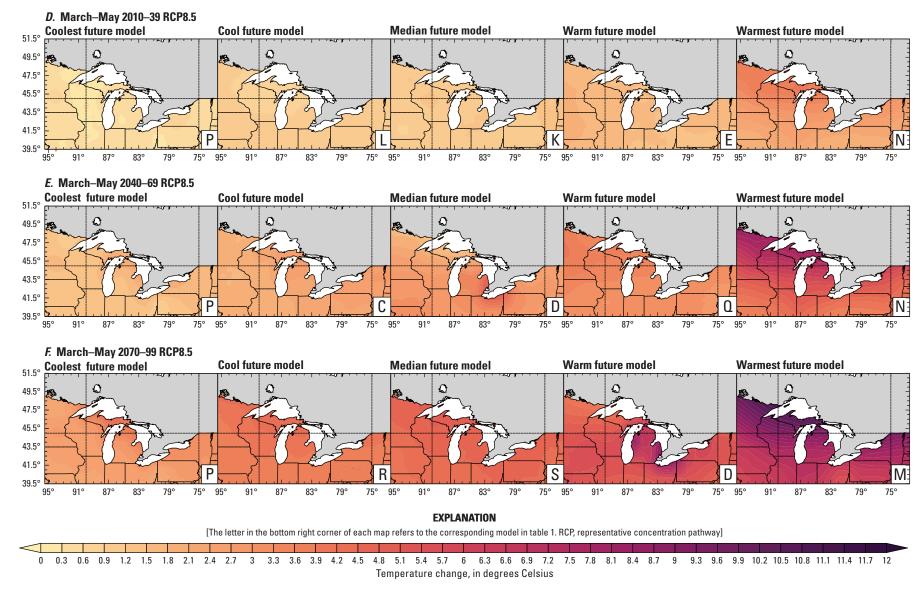


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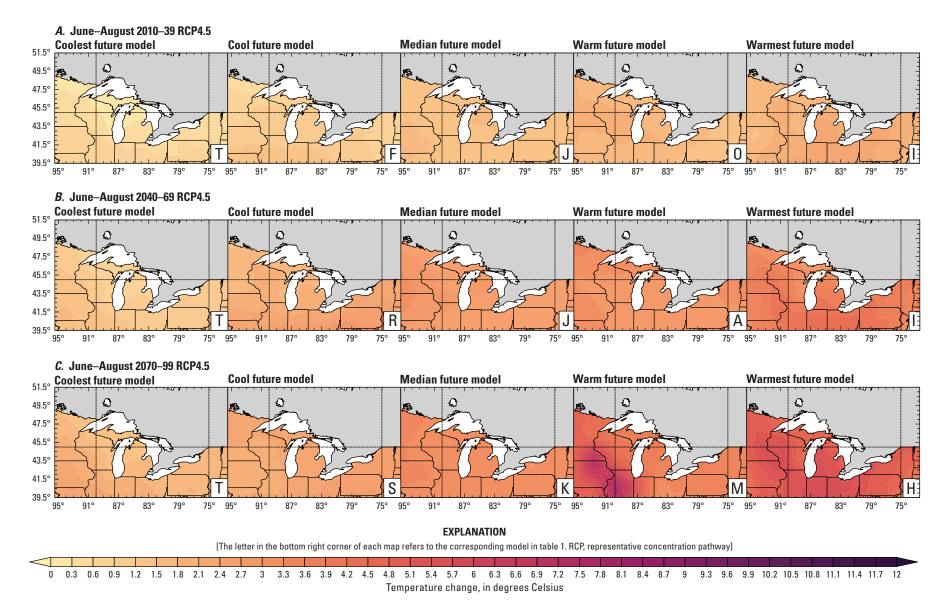


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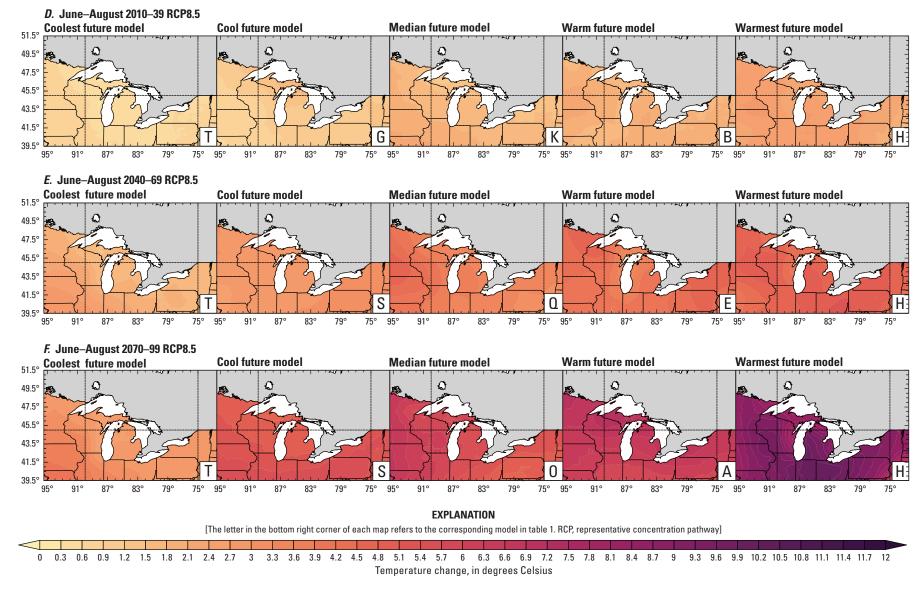


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