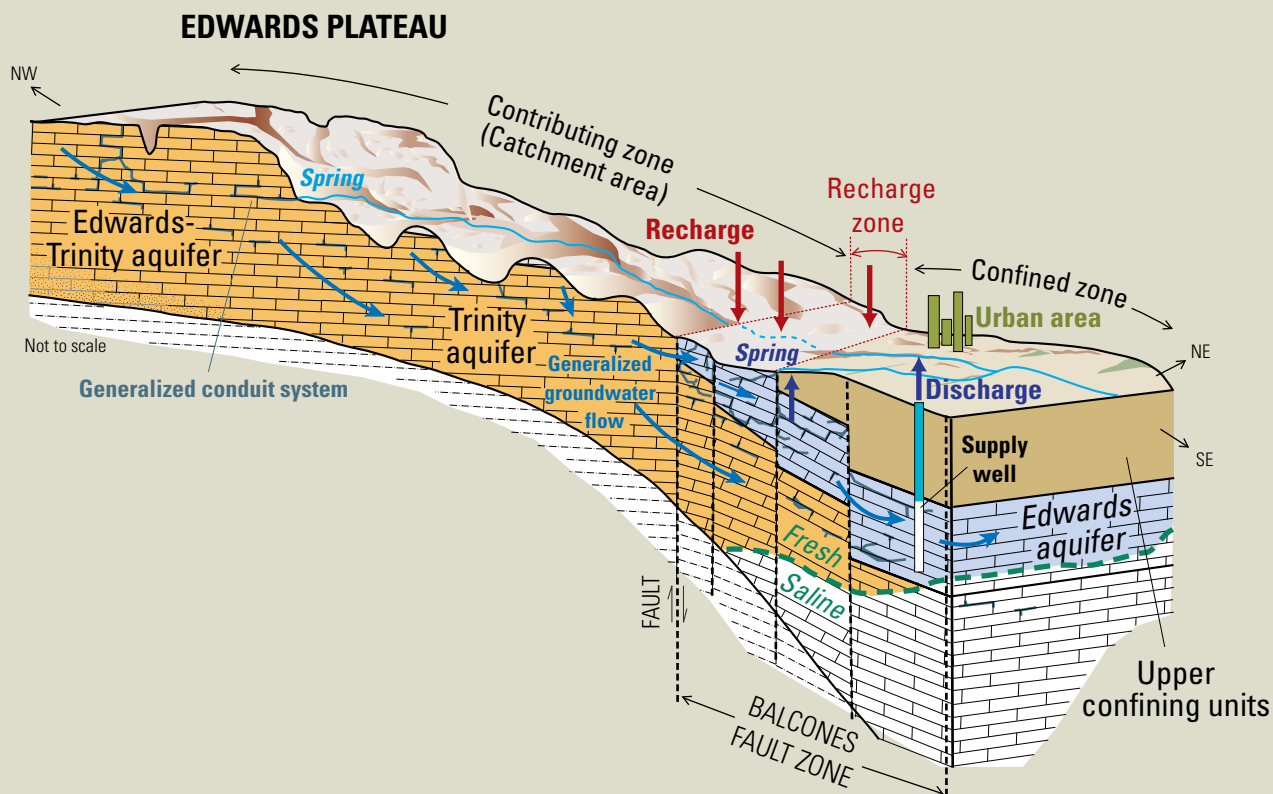


U.S. Geological Survey National Water-Quality Assessment Program

Hydrogeology, Chemical Characteristics, and Water Sources and Pathways in the Zone of Contribution of a Public-Supply Well in San Antonio, Texas



Scientific Investigations Report 2011–5146

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By MaryLynn Musgrove, Lynne Fahlquist, Gregory P. Stanton, Natalie A. Houston,
and Richard J. Lindgren

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Foreword

The U.S. Geological Survey (USGS) is committed to providing the Nation with reliable scientific information that helps to enhance and protect the overall quality of life and that facilitates effective management of water, biological, energy, and mineral resources (<http://www.usgs.gov/>). Information on the Nation's water resources is critical to ensuring long-term availability of water that is safe for drinking and recreation and is suitable for industry, irrigation, and fish and wildlife. Population growth and increasing demands for water make the availability of that water, measured in terms of quantity and quality, even more essential to the long-term sustainability of our communities and ecosystems.

The USGS implemented the National Water-Quality Assessment (NAWQA) Program in 1991 to support national, regional, State, and local information needs and decisions related to water-quality management and policy (<http://water.usgs.gov/nawqa>). The NAWQA Program is designed to answer: What is the quality of our Nation's streams and groundwater? How are conditions changing over time? How do natural features and human activities affect the quality of streams and groundwater, and where are those effects most pronounced? By combining information on water chemistry, physical characteristics, stream habitat, and aquatic life, the NAWQA Program aims to provide science-based insights for current and emerging water issues and priorities. From 1991 to 2001, the NAWQA Program completed interdisciplinary assessments and established a baseline understanding of water-quality conditions in 51 of the Nation's river basins and aquifers, referred to as Study Units (http://water.usgs.gov/nawqa/studies/study_units.html).

National and regional assessments are ongoing in the second decade (2001–12) of the NAWQA Program as 42 of the 51 Study Units are selectively reassessed. These assessments extend the findings in the Study Units by determining water-quality status and trends at sites that have been consistently monitored for more than a decade, and filling critical gaps in characterizing the quality of surface water and groundwater. For example, increased emphasis has been placed on assessing the quality of source water and finished water associated with many of the Nation's largest community water systems. During the second decade, NAWQA is addressing five national priority topics that build an understanding of how natural features and human activities affect water quality, and establish links between sources of contaminants, the transport of those contaminants through the hydrologic system, and the potential effects of contaminants on humans and aquatic ecosystems. Included are studies on the fate of agricultural chemicals, effects of urbanization on stream ecosystems, bioaccumulation of mercury in stream ecosystems, effects of nutrient enrichment on aquatic ecosystems, and transport of contaminants to public-supply wells. In addition, national syntheses of information on pesticides, volatile organic compounds (VOCs), nutrients, trace elements, and aquatic ecology are continuing.

The USGS aims to disseminate credible, timely, and relevant science information to address practical and effective water-resource management and strategies that protect and restore water quality. We hope this NAWQA publication will provide you with insights and information to meet your needs and will foster increased citizen awareness and involvement in the protection and restoration of our Nation's waters.

The USGS recognizes that a national assessment by a single program cannot address all water-resource issues of interest. External coordination at all levels is critical for cost-effective management, regulation, and conservation of our Nation's water resources. The NAWQA Program, therefore, depends on advice and information from other agencies—Federal, State, regional, interstate, Tribal, and local—as well as nongovernmental organizations, industry, academia, and other stakeholder groups. Your assistance and suggestions are greatly appreciated.

William H. Werkheiser
USGS Associate Director for Water

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Conversion Factors and Abbreviations

Multiply	By	To obtain
Length		
inch (in.)	25.4	millimeter (mm)
foot (ft)	0.3048	meter (m)
mile (mi)	1.609	kilometer (km)
millimeter (mm)	0.03937	inch (in.)
micrometer (μm)	1,000	millimeter (mm)
Area		
square mile (mi^2)	2.590	square kilometer (km^2)
Volume		
gallon (gal)	3.785	liter (L)
million gallons (Mgal)	3,785	cubic meter (m^3)
cubic centimeter (cm^3)	0.06102	cubic inch (in^3)
milliliter (mL)	0.03381	fluid ounce (oz)
liter (L)	0.2642	gallon (gal)
cubic meter (m^3)	264.2	gallon (gal)
Flow rate		
gallon per minute (gal/min)	0.06309	liter per second (L/s)
million gallons per day (Mgal/d)	0.04381	cubic meter per second (m^3/s)
Mass/weight		
gram (g)	0.3527	ounce avoirdupois (oz avdp)
kilogram (kg)	2.205	pound (lb)
Transmissivity*		
foot squared per day (ft^2/d)	0.09290	meter squared per day (m^2/d)

*Transmissivity: The standard unit for transmissivity is cubic foot per day per square foot times foot of aquifer thickness [$(\text{ft}^3/\text{d})/\text{ft}^2$ ft. In this report, the mathematically reduced form, foot squared per day (ft^2/d), is used for convenience.

Vertical coordinate information is referenced to the North American Vertical Datum of 1988 (NAVD 88) and to the National Geodetic Vertical Datum of 1929 (NGVD 29) for Bexar County index well (J-17).

Horizontal coordinate information is referenced to the North American Datum of 1983 (NAD 83).

Altitude, as used in this report, refers to distance above the vertical datum.

Temperature in degrees Celsius (°C) may be converted to degrees Fahrenheit (°F) as follows:

$$^{\circ}\text{F}=(1.8\times^{\circ}\text{C})+32$$

Specific conductance is given in microsiemens per centimeter at 25 degrees Celsius ($\mu\text{S}/\text{cm}$ at 25°C).

Concentrations of chemical constituents in water are given either in milligrams per liter (mg/L) or in micrograms per liter ($\mu\text{g}/\text{L}$).

Explanation of Isotope Units

The δ values for stable-isotope ratios discussed in this report are referenced to the following standard materials:

Element	R	Standard identity and reference
Hydrogen	Hydrogen-2/hydrogen-1	Vienna Standard Mean Ocean Water (VSMOW) (Casciotti and others, 2002)
Oxygen	Oxygen-18/oxygen-16	Vienna Standard Mean Ocean Water (VSMOW) (Casciotti and others, 2002)
Nitrogen	Nitrogen-15/nitrogen-14	Nitrogen in air

Hydrogeology, Chemical Characteristics, and Water Sources and Pathways in the Zone of Contribution of a Public-Supply Well in San Antonio, Texas

By MaryLynn Musgrove, Lynne Fahlquist, Gregory P. Stanton, Natalie A. Houston, and Richard J. Lindgren

Abstract

In 2001, the National Water-Quality Assessment (NAWQA) Program of the U.S. Geological Survey initiated a series of studies on the transport of anthropogenic and natural contaminants (TANC) to public-supply wells (PSWs). The main goal of the TANC project was to better understand the source, transport, and receptor factors that control contaminant movement to PSWs in representative aquifers of the United States. Regional- and local-scale study areas were selected from within existing NAWQA study units, including the south-central Texas Edwards aquifer. The local-scale TANC study area, nested within the regional-scale NAWQA study area, is representative of the regional Edwards aquifer. The PSW selected for study is within a well field of six production wells. Although a single PSW was initially selected, because of constraints of well-field operation, samples were collected from different wells within the well field for different components of the study. Data collected from all of the well-field wells were considered comparable because of similar well construction, hydrogeology, and geochemistry. An additional 38 PSWs (mostly completed in the confined part of the aquifer) were sampled throughout the regional aquifer to characterize water quality. Two monitoring well clusters, with wells completed at different depths, were installed to the east and west of the well field (the Zarzamora and Timberhill monitoring well clusters, respectively). One of the monitoring wells was completed in the overburden to evaluate potential hydrologic connectivity with the Edwards aquifer. Geophysical and flowmeter logs were collected from one of the well-field PSWs to determine zones of contribution to the wellbore. These contributing zones, associated with different hydrogeologic units, were used to select monitoring well completion depths and groundwater sample collection depths for depth-dependent sampling. Depth-dependent samples were collected from the PSW from three different depths and under three different pumping conditions. Additionally, selected monitoring wells and one of the well-field PSWs were sampled several times in response to a rainfall and recharge event to assess short-term (event-scale)

temporal variations in water quality. For comparison purposes, groundwater samples were categorized as being from regional aquifer PSWs, from the well field (wellhead samples), from the monitoring wells (excluding the overburden well), from the overburden well, from the PSW depth-dependent sampling, and from temporal sampling. Groundwater samples were analyzed for inorganic, organic, isotopic, and age-dating tracers to characterize geochemical conditions in the aquifer and provide understanding of the mechanisms of mobilization and movement of selected constituents from source areas to a PSW. Sources, tracers, and conditions used to assess water quality and processes affecting the PSW and the aquifer system included (1) carbonate host rock composition; (2) physicochemical constituents; (3) major and trace element concentrations; (4) saturation indices with respect to minerals in aquifer rocks; (5) elemental ratios, such as magnesium to calcium ratios, that are indicative of water-rock interaction processes; (6) oxidation-reduction conditions; (7) nutrient concentrations, in particular nitrate concentrations; (8) the isotopic composition of nitrate, which can point to specific nitrate sources; (9) strontium isotopes; (10) stable isotopes of hydrogen and oxygen; (11) organic contaminant concentrations, including pesticides and volatile organic compounds; (12) age tracers, apparent-age distribution, and dissolved gas data used in age interpretations; (13) depth-dependent water chemistry collected from the PSW under different pumping conditions to assess zones of contribution; and (14) temporal variability in groundwater composition from the PSW and selected monitoring wells in response to an aquifer recharge event.

Geochemical results indicate that the well-field and monitoring well samples were largely representative of groundwater in the regional confined aquifer. Constituents of concern in the Edwards aquifer for the long-term sustainability of the groundwater resource include the nutrient nitrate and anthropogenic organic contaminants. Nitrate concentrations (as nitrogen) for regional aquifer PSWs had a median value of 1.9 milligrams per liter, which is similar to previously reported values for the regional aquifer. Nitrate-isotope compositions for groundwater samples collected from the well-field PSWs and monitoring wells had a narrow range, with values

indicative of natural soil organic values. A comparison with historical nitrate-isotope values, however, suggests that a component of nitrate in groundwater from biogenic sources might have increased over the last 30 years. Several organic contaminants (the pesticide atrazine, its degradate deethylatrazine, trichloromethane (chloroform; a drinking-water disinfection byproduct), and the solvent tetrachloroethene (PCE)) were widely distributed throughout the regional aquifer and in the local-scale TANC study area at low concentrations (less than 1 microgram per liter). Higher concentrations of PCE were detected in samples from the well-field PSWs and Zarzamora monitoring wells relative to the regional aquifer PSWs. The urban environment is a likely source of contaminants to the aquifer, and these results indicate that one or more local urban sources might be supplying PCE to the Zarzamora monitoring wells and the well-field wells. Samples from the well field also had high concentrations of chloroform relative to the monitoring wells and regional aquifer PSWs. For samples from the regional aquifer PSWs, the most frequently detected organic contaminants generally decreased in concentration with increasing well depth. Deeper wells might intercept longer regional flow paths with higher fractions of older water or water recharged in rural recharge areas in the western part of the aquifer that have been less affected by anthropogenic contaminants. A scenario of hypothetical contaminant loading was evaluated by using results from groundwater-flow-model particle tracking to assess the response of the aquifer to potential contamination. Results indicate that the aquifer responds quickly (less than 1 year to several years) to contaminant loading; however, it takes a relatively long time (decades) for concentrations to reach peak values. The aquifer also responds quickly (less than 1 year to several years) to the removal of contaminant loading; however, it also takes a relatively long time (decades) to reach near background concentrations.

Interpretation of geochemical age tracers in this well-mixed karst system was complicated by contamination of a majority of measured tracers and complexities of extensive mixing. Age-tracer results generally indicated that groundwater samples were composed of young, recently recharged water with piston-flow model ages ranging from less than 1 to 41 years, with a median of 17 years. Although a piston-flow model is typically not valid for karst aquifers, the model ages provide a basis for comparing relative ages of different samples and a reference point for more complex hydrogeologic models for apparent-age interpretations. Young groundwater ages are consistent with particle-tracking results from hydrogeologic modeling for the local-scale TANC study area. Age-tracer results compared poorly with other geochemical indicators of groundwater residence time and anthropogenic effects on water quality, indicating that hydrogeologic conceptual models used in groundwater age interpretations might not adequately account for mixing in this karst system. Groundwater samples collected from the well field under a variety of pumping conditions were relatively homogeneous and well mixed for numerous geochemical constituents (with the notable exception of age tracers). Groundwater contributions

to the PSW were dominated by well-mixed, relatively homogeneous groundwater, typical of the regional confined aquifer. Zones of preferential flow were determined for the PSW, but groundwater samples from different stratigraphic units were not geochemically distinct. Variations in chemical constituents in response to a rainfall and aquifer recharge event occurred but were relatively minor in the PSW and monitoring wells. This observation is consistent with the hypothesis that the response to individual recharge events in the confined aquifer, unless intersecting conduit flow paths, might be attenuated by mixing processes along regional flow paths. Results of this study are consistent with the existing conceptual understanding of aquifer processes in this karst system and are useful for water-resource development and management practices.

Introduction

About a third of the population of the United States receives drinking water from public groundwater systems (Toccalino and Hopple, 2010). The vulnerability of drinking-water public-supply wells (PSWs) to contamination has raised numerous health concerns. Anthropogenic contaminants—such as solvents, drinking-water disinfection byproducts, other volatile organic compounds, and pesticides—have been detected in low concentrations in water from PSWs across the United States (Gilliom and others, 2006; Ivahnenko and Zogorski, 2006; Schaap and Zogorski, 2006; Zogorski and others, 2006; Carter and others, 2007; Hopple and others, 2009). Groundwater contamination from nutrients such as nitrate (NO_3) is widespread (Nolan and others, 1998). Naturally occurring contaminants such as radon, uranium (U), and arsenic (As) also are present in groundwater from many drinking-water wells (Focazio and others, 1999; Sowerby and others, 2000; Welch and others, 2000; Jurgens and others, 2008; Landon and others, 2008). In response to these concerns, the U.S. Geological Survey (USGS) National Water-Quality Assessment (NAWQA) Program began a series of studies in 2001 to assess the vulnerability of PSWs to contamination from compounds commonly found in the environment, specifically with regard to the transport of anthropogenic and naturally occurring contaminants (TANC) to PSWs at regional and local scales at several sites across the United States (Eberts and others, 2005). The TANC project was built on previous NAWQA studies that found low levels of contaminants and mixtures of contaminants in groundwater beneath urban areas (Hamilton and others, 2004). TANC study sites have included a range of hydrogeologic settings and management practices that were selected to be representative of drinking-water resources used by a majority of the U.S. population.

The goal of the TANC project is to understand the source, transport, and receptor factors that control the movement of contaminants to PSWs in representative aquifers of the United States (Eberts and others, 2005). Specific objectives are to (1) identify the dominant contaminants and their sources in PSWs; (2) assess the effects of natural processes and human

activities on the occurrence of contaminants in PSWs; (3) identify the factors that are most important for PSW vulnerability assessments in different settings and at different scales; (4) develop simple methods and models for evaluating PSWs for vulnerability to contamination in unstudied areas and from emerging contaminants; and (5) increase understanding of the potential effects of water-resources development and management decisions on the quality of water from PSWs (Eberts and others, 2005). Because subsurface processes and management practices differ among aquifers and public-water systems, PSWs in different parts of the United States are not equally susceptible or vulnerable to contamination (Focazio and others, 2002), even where similar contaminant sources might exist. Concentrations of anthropogenic or natural contaminants can change along groundwater flow paths from recharge or source areas to PSWs as a result of many factors. These factors can be broadly grouped into those concerning sources, including contaminant loading or mobilization; transport, including migration and reaction; and receptors (wells), including PSW operations (Franke and others, 1998; Focazio and others, 2002).

TANC studies were conducted at regional scales of up to thousands of square miles (mi^2) at 11 locations and at local scales of less than $1,100 \text{ mi}^2$ at six additional locations; these were delineated from within six NAWQA study units of differing hydrogeology (Connecticut, glacial deposits; Nebraska, unconsolidated sediment; California, unconsolidated sediment; Florida, conduit-dominated karstic carbonate; New Mexico, unconsolidated sediment; and Texas, mixed carbonate). Local-scale study areas were selected that met the following criteria: location in an aquifer previously studied under the NAWQA Program; availability of existing source (raw water) water-quality data from PSWs; previous detection of natural or anthropogenic contaminants in source waters; and previously developed regional groundwater model(s). One of the selected local-scale study areas was the San Antonio segment of the Edwards aquifer, within the South-Central Texas (SCTX) regional study unit of the NAWQA Program (Bush and others, 2000). This report describes the hydrogeology and geochemistry of the $1,098\text{-mi}^2$ SCTX local-scale study area, which contains the zone of contribution to a selected PSW, and discusses factors controlling the transport of selected anthropogenic and natural contaminants to PSWs.

Purpose and Scope

The purposes of this report are to document the hydrogeologic setting, the occurrence and distribution of selected chemical characteristics, and factors controlling transport of selected anthropogenic and natural contaminants to PSWs in the local-scale TANC study area at San Antonio, Texas. A description of the design of the local-scale monitoring well network and methods of data collection and analysis are included. The hydrogeologic setting of the local-scale study area is described, as are general water geochemistry characteristics such as field parameters, oxidation-reduction

(redox) conditions, the occurrence and distribution of selected inorganic and organic constituents, and the distribution of interpreted groundwater ages. The occurrence and distribution of contaminants and contaminant point sources are described to help assess factors influencing transport of these anthropogenic and natural contaminants to PSWs in the local-scale TANC study area. This report is intended to serve as a foundation for analyses comparing results from this local-scale TANC study area with other TANC study areas.

Description of the Study Area

The regional-scale NAWQA and local-scale TANC study areas considered in this report (hereinafter, regional-scale study area and local-scale study area, respectively) are in south-central Texas within the Edwards aquifer (fig. 1). The regional-scale study area includes the extent of the San Antonio segment of the Edwards aquifer, which is within a narrow band along the Balcones fault zone from the western groundwater divide in Kinney County to the groundwater divide north of San Marcos Springs in Hays County (fig. 1). The Balcones escarpment is the surface manifestation of the Balcones fault zone (fig. 1). Late Cenozoic faulting of the predominantly flat-lying region along the Balcones fault zone formed a series of high-angle normal en echelon down-toward-the-coast faults. The faulting resulted in a series of blocks of Edwards aquifer rocks that are partially to completely offset and divide the confined and unconfined parts of the aquifer (Maclay and Small, 1983) (fig. 2). The local-scale study area is nested within the larger regional-scale study area (Lindgren and others, 2011a). Although this report focuses on the results of the local-scale study, an understanding of the characteristics of the regional-scale aquifer provides insight into movement of water and solutes along groundwater flow paths to PSWs at the local scale. Characteristics of the regional-scale study area around San Antonio are described by Lindgren and others (2011a). Characteristics of the Edwards aquifer that influence water quality have been previously described by numerous studies, including Bush and others (2000), Fahlquist and Ardis (2004), and Musgrove and others (2010). A hydrogeologic model constructed and calibrated to simulate groundwater flow and age distribution in the local-scale study area is described in a companion report by Lindgren and others (2011b).

The local-scale study area is predominantly in Bexar and Medina Counties and includes the northwestern part of San Antonio (fig. 1). The local-scale study area and model boundary were determined by using the regional-scale hydrogeologic model (Lindgren and others, 2011b) to identify the area contributing recharge (ACR) to the PSW and by bounding the likely ACR with the local-scale model area. The local-scale study area includes the active area of the local-scale model (recharge and confined zones) and a small part of the catchment area (contributing zone) north of the Edwards aquifer recharge zone (fig. 3). The local-scale study area is bounded on the west and east by the western and eastern

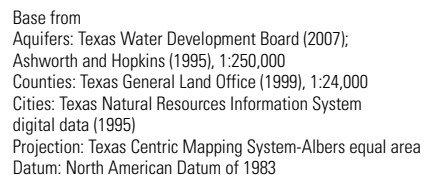


Figure 1. Hydrogeologic setting and location of regional aquifer public-supply wells (ranked by pumping quartile) sampled for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells in the San Antonio segment of the Edwards aquifer and the local-scale study area, near San Antonio, south-central Texas, 2004–9.

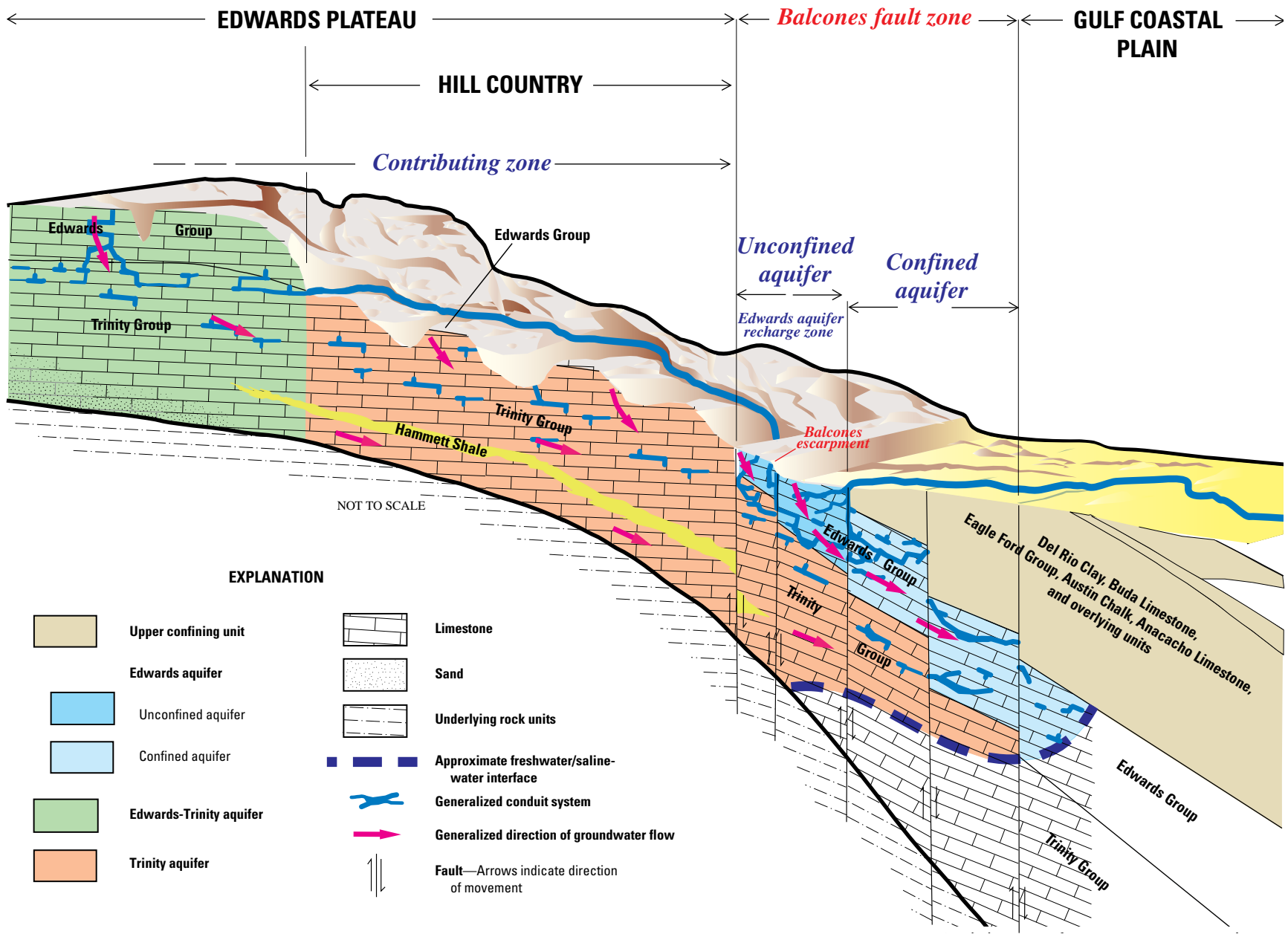


Figure 2. Diagrammatic north-northwest-to-south-southeast section showing hydrogeologic framework and generalized groundwater flow directions, Edwards Plateau to Gulf Coastal Plain, San Antonio region, Texas (modified from Barker and Ardis, 1996, plates 1 and 3).

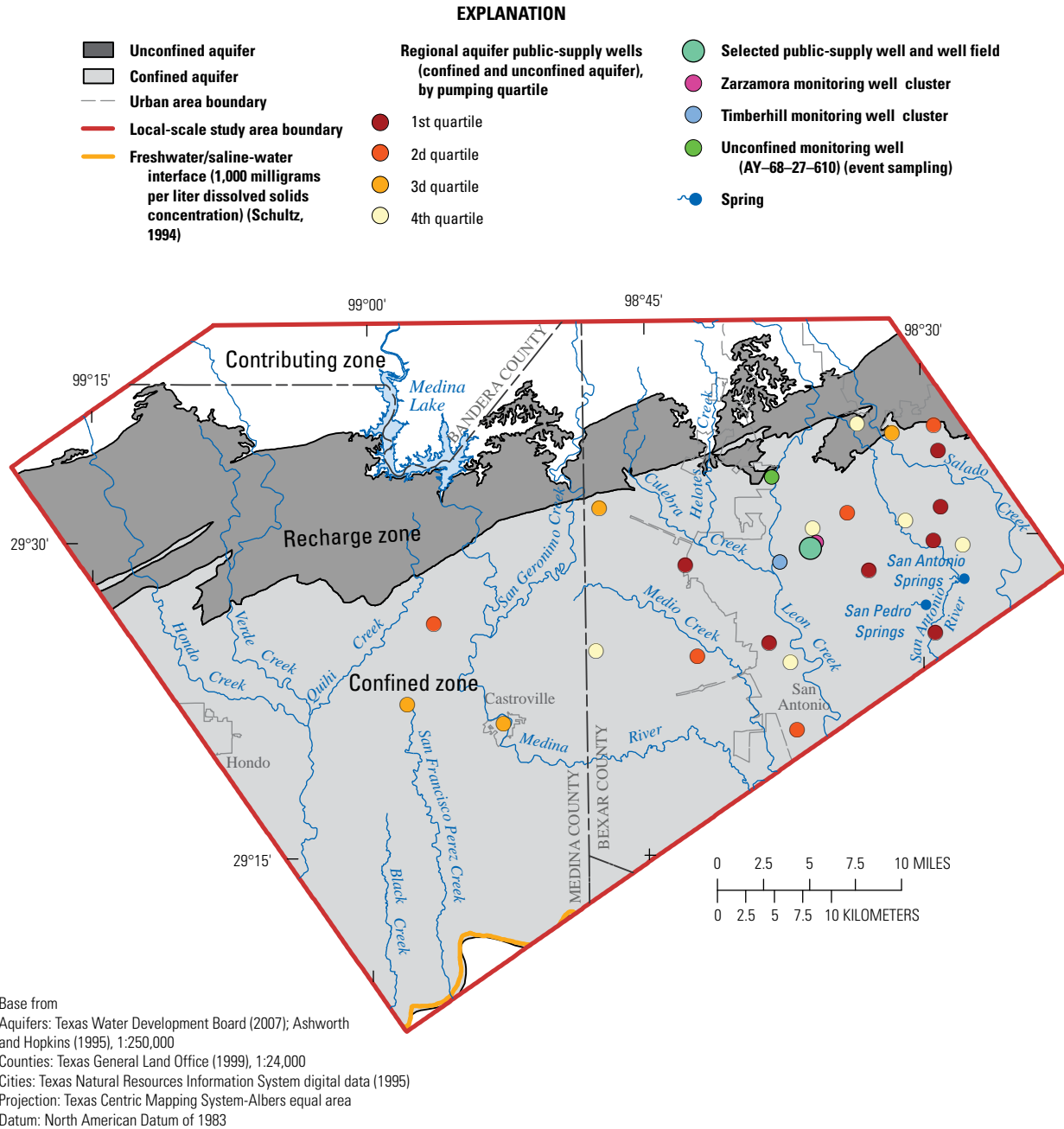


Figure 3. Local-scale study area for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells showing the selected public-supply well and well field, monitoring wells, and regional aquifer public-supply wells (ranked and shown by pumping quartile) near San Antonio, south-central Texas, 2004–9.

extents of the ACR to the selected PSW; these boundaries are near Hondo Creek in Medina County and near Salado Creek in Bexar County, respectively (fig. 1). It is bounded on the south approximately by the 1,000-milligrams-per-liter (mg/L) dissolved-solids concentration line (updip boundary of the freshwater/saline-water transition zone). The northern boundary is north of the updip limit of contiguous, outcropping rocks of the Edwards Group, Georgetown Formation, and

their westward stratigraphic equivalents (Edwards rocks). The local-scale model area is oriented with the regional direction of groundwater flow and the major faults in the Balcones fault zone (Lindgren and others, 2011b).

The construction and operational practices of the selected PSW are representative of many PSWs that use the Edwards aquifer for supplying the population of the San Antonio metropolitan area. The selected PSW is a primary

community-water-system PSW, which a public-water purveyor made available for the study. NAWQA-study analyses of groundwater samples from the PSW have detected multiple contaminants at low concentrations (substantially below drinking-water standards). The combination of low concentrations of detected contaminants, typical well construction and operational practices, and cooperation of the water utility prompted the selection of this PSW for the local-scale study. The selected PSW is within a well field of six wells in northwest San Antonio.

Probabilistic assessments of zones of contribution to the selected PSW, based on forward and backward particle tracks that were generated by using two regional groundwater flow models (Lindgren and others, 2004; Lindgren, 2006), were considered in delineating the local-scale study area (fig. 3). Although such determinations are likely unrealistic at a local scale in a fractured karst system, the study area was sufficiently large (1,098 mi²) to reasonably approximate probable zones of flow to the selected PSW. The probable zones of contribution selected from models is similar to those represented by previous conceptualizations (Maclay and Land, 1988; Clark and Journey, 2006).

In the local-scale study area, the layered fractured karstic carbonate rocks are approximately 450–500 feet (ft) thick (Rose, 1972) and consist of eight hydrogeologic units (HGU) previously defined for the Edwards aquifer on the basis of hydrogeologic properties (Maclay, 1995; Stein and Ozuna, 1995). The regional paleogeographic and structural features associated with the Edwards aquifer, and the previously defined stratigraphic units associated with these features, are detailed by Lindgren and others (2011b). A conceptual cross-sectional schematic of the hydrogeology of the local-scale study area and well-field and monitoring well sites in the context of HGUs is shown in figure 4. The local-scale model includes four additional HGUs, including one confining HGU in the overburden overlying the aquifer and one confining HGU in the Trinity aquifer underlying the Edwards aquifer (Lindgren and others, 2011b) (fig. 4). Among the eight previously identified units (Stein and Ozuna, 1995), the Kirschberg evaporite member and the leached and collapsed members (composed of the leached and collapsed members, undivided, hereinafter, leached and collapsed) generally are the most porous and permeable units (fig. 4). Two highly conductive HGUs (5 and 9) were added to the base of these two members to represent highly transmissive zones within these two members in the local-scale hydrologic model (Lindgren and others, 2011b) (fig. 4). The cyclic and marine HGU (composed of the cyclic and marine members, undivided; hereinafter cyclic and marine (HGU 3)) within the Person Formation is also considered moderately permeable (Stein and Ozuna, 1995) but was not divided into multiple HGUs.

Land surface elevation in the local-scale study area ranges from 1,745 ft in the northwestern part of the study area in Medina County to 551 ft along the Medina River in southern Bexar County (National Hydrography Dataset Plus,

2010). Mean annual precipitation decreases across the region from east to west, ranging from 32.5 inches (in.) in the eastern part to 27.5 in. in the western part of the local-scale study area (Daly and Taylor, 2000). The region is prone to climatic and hydrologic extremes (Griffiths and Strauss, 1985; Jones, 1991); mean annual rainfall at San Antonio International Airport is 29 in., but annual rainfall has ranged from 10 to 52 in. (1893–2009) (National Oceanic and Atmospheric Administration, 2010). Some of the most extreme 1-day-duration storms in the world have occurred along the Balcones escarpment (Slade, 1986). Storms often produce rapid runoff that provides recharge to the aquifer. Droughts lasting from many months to years have been documented in the region since the earliest settlers began keeping records (Texas State Historical Association, 2009). The mean annual maximum temperature (1970–2000), measured in July, is 36.1 degrees Celsius (°C), and the average annual minimum temperature, measured in January, is 1.7°C (U.S. Department of Agriculture, 2010). Land use that overlies the (recharge and confined zones of the) Edwards aquifer in the local-scale study area is 35.6 percent rangeland, 27.8 percent forest, 23.1 percent urban/built-up, 11.8 percent agricultural, and 1.7 percent water, wetland, or barren land (Homer and others, 2001). Most of the rangeland, forest, and agricultural land is to the west of San Antonio in Medina County. Most of the outcrop (unconfined zone) is rangeland and forest, and most of the agricultural area overlies the confined part of the aquifer. Nearly all urban/built-up land overlies unconfined and confined parts of the aquifer in the San Antonio area in Bexar County.

The city of San Antonio, the principal urban area in the local-scale study area, includes much of Bexar County in the east-central part of the aquifer region. San Antonio is the seventh largest city in the United States, with a 2007 population of about 1.3 million (U.S. Census Bureau, 2009). Although pumping in the San Antonio area has increased about fivefold since the 1930s, aquifer water-level altitudes have not shown long-term declines because the aquifer readily recharges during periods of rainfall (Bush and others, 2000).

Potential sources of contamination and the fractured karstic hydrogeologic conditions of the Edwards aquifer in the San Antonio area are characteristic of many communities that withdraw water from carbonates and fractured rocks. Potential contamination sources include chlorinated organic solvents from historical chemical use at small commercial facilities typically found in urban areas, NO₃ and pesticides from irrigated cultivated and urban land, and As and U from carbonate or marine sediments. Movement of these contaminants through the aquifer system might be enhanced by increased groundwater flow velocities caused by temporal response to rainfall, increased pumping during summer or drought, and increased recharge from precipitation. Moreover, the common practice of constructing supply wells across large (greater than 100 ft) open intervals of the aquifer to improve well yields can contribute to rapid movement of contaminants (Zinn and Konikow, 2007).

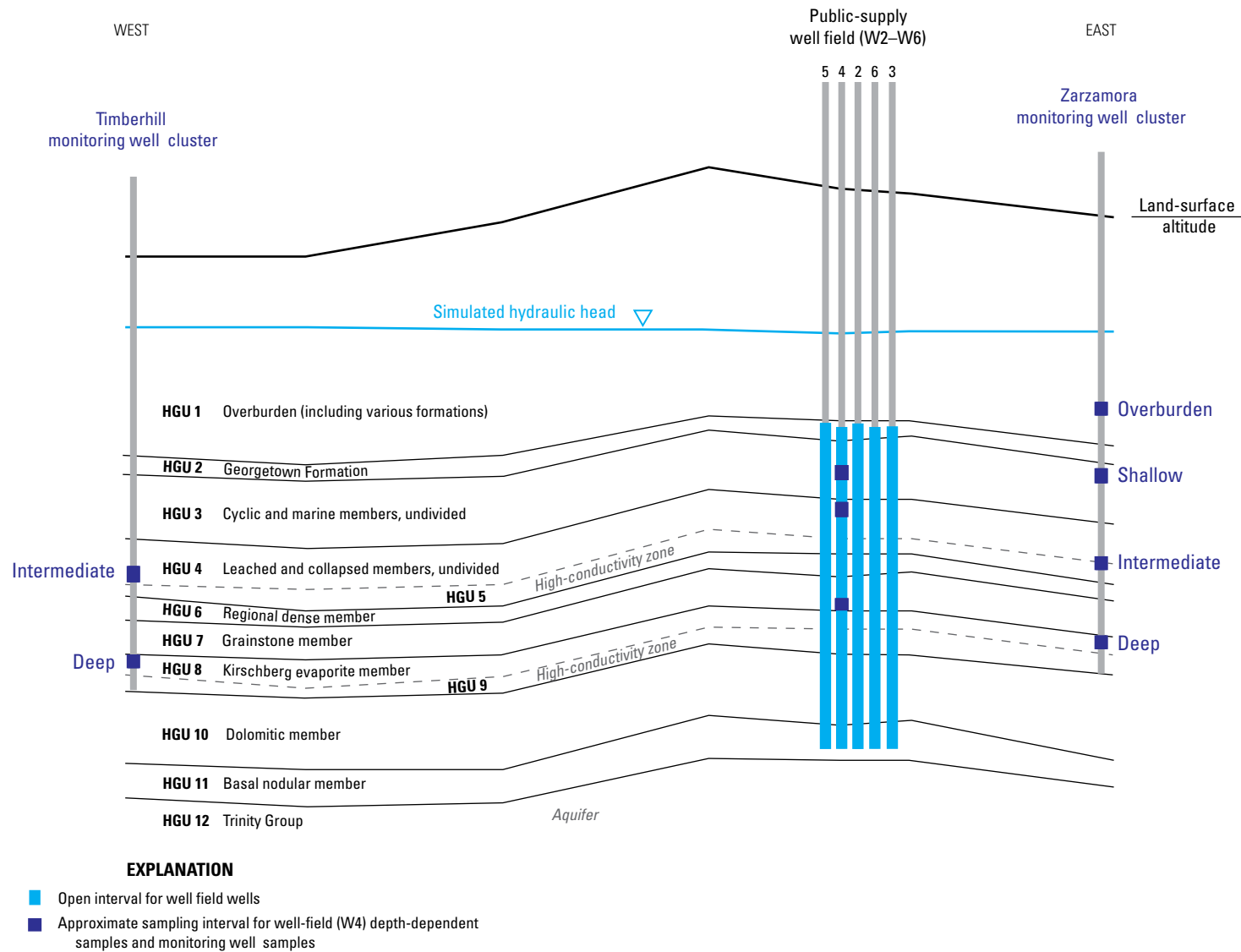


Figure 4. Schematic east-west (strike) cross section showing location of study area hydrogeologic units (HGUs) as modeled by Lindgren and others (2011b) and relation of HGUs at well-field, monitoring well, and depth-dependent (W4) sample sites, near San Antonio, south-central Texas. HGUs 5 and 9 are highly conductive zones at the base of HGUs 4 and 8, respectively. Relatively small (less than 80 percent) displacement modeled faults are not shown.

Overview of Karst Aquifers

Karst aquifers, such as the Edwards aquifer, are composed of soluble host rocks, usually carbonates, that have dissolved preferentially to form large voids and conduits (White, 1988). Karst aquifers tend to be extremely heterogeneous, with orders of magnitude ranges of porosity and permeability. Groundwater transport can be rapid through high-porosity voids and conduits, with little opportunity for subsurface filtration, sorption, or degradation of dissolved or particulate constituents (White, 1988). As a result, karst aquifer systems are particularly susceptible to contamination.

Karst aquifer permeability ranges from matrix permeability, characterized by high storage and low transmissivity, to conduit permeability, characterized by low storage and high transmissivity (Schuster and White, 1971; Atkinson, 1977). Karst aquifers commonly are conceptualized as dual-flow systems that comprise a continuum-flow system with Darcian flow comparable to porous media and a discrete conduit network frequently with turbulent flow conditions (Birk and others, 2003). Most groundwater storage occurs within the aquifer matrix, but most transport occurs within conduits, which often dominate groundwater flow where present (White, 2002). Fluid flow in karst aquifers can be highly variable and difficult to measure or predict.

Recharge to karst aquifers occurs as direct infiltration into karst features such as sinkholes and fractures (focused recharge) and as distributed infiltration into the rock matrix (diffuse recharge); the proportion of focused and diffuse recharge can vary greatly in time and space depending on a number of variables (Worthington, 2003; Mahler and others, 2006). In wells and springs that are dominantly affected by conduit flow, physicochemical properties, concentrations of geochemical constituents, and discharge (from springs) or water-level altitudes (in wells) can change rapidly (minutes to hours to days) following rainfall and aquifer recharge. In contrast, water-level altitudes in wells and springs that are dominantly affected by matrix flow might show little response to recharge events and less variability in physicochemical properties and geochemistry. Natural discharge in karst aquifers often occurs through large springs, which represent a composite (integrator) of all water moving through the aquifer; as a result, variations in spring discharge, physicochemical constituents, and geochemistry provide insight into karst aquifer hydrogeology (White, 2002).

Summary of Previous Studies

PSW contamination is a potential health threat and has had negative economic effects on communities that need to treat contaminated water or find alternative water supplies. Investigations at multiple spatial scales across the United States have indicated that shallow groundwater quality is affected by land-use activities in many locations but that water

quality is less frequently affected at the greater depths where PSWs typically are screened (Nolan and Stoner, 2000; Gilliom and others, 2006; Zogorski and others, 2006). In 1991, the USGS implemented the NAWQA Program to characterize, in a nationally consistent manner, water quality of major surface-water and groundwater resources of the Nation, to determine natural and human factors that affect water quality, and to define trends in water quality (Gilliom and others, 1995). Between 1996 and 2009, the NAWQA Program collected and analyzed groundwater samples from numerous sites throughout the San Antonio segment of the Edwards aquifer in the SCTX study unit; specific NAWQA studies include an analysis of samples collected between 1996 and 2006 (Musgrove and others, 2010) and samples collected as part of the local-scale study, which is the focus of this report. NAWQA analyses of water samples for anthropogenic contaminants in the San Antonio segment of the Edwards aquifer have detected numerous anthropogenic contaminants, mostly at low concentrations, less than 1 microgram per liter ($\mu\text{g/L}$) (Bush and others, 2000; Fahlquist and Ardis, 2004). Two regional-scale groundwater flow models were previously developed for the regional aquifer by using MODFLOW2000 (Harbaugh and others, 2000): (1) a model that includes simulated, continuously connected conduits and barrier faults that influence flow (Lindgren and others, 2004) and (2) a model without simulated conduits in which flow occurs through small fractures and openings (Lindgren, 2006). Both models were calibrated for steady-state and transient conditions. A hydrogeologic model has been constructed for the local-scale study area, which is described in a companion report by Lindgren and others (2011b).

Previous studies of the Edwards aquifer provide insight into its hydrology and geochemistry. Geochemical processes that affect groundwater geochemistry and water quality include interaction with overlying soils, mineral-solution reactions, and mixing with saline waters from downdip and underlying hydrostratigraphic units (Clement and Sharp, 1988; Oetting and others, 1996; Groschen and Buszka, 1997; Musgrove and Banner, 2004; Musgrove and others, 2010). Geochemical tracers of mineral-solution reactions (such as the molar ratios of magnesium to calcium (Mg/Ca) concentrations and strontium to calcium (Sr/Ca) concentrations and dolomite mineral saturation indices) can be indicative of groundwater residence time, which is a function of both flow path and water flux (Musgrove and Banner, 2004; Wong, 2008; Musgrove and others, 2010). Previous studies in the San Antonio and Barton Springs segments of the Edwards aquifer have documented extensive surface-water-groundwater interaction, relations between variable hydrologic conditions and water quality, and movement of anthropogenic contaminants and dissolved constituents through the aquifer (Hauwert, Johns, and others, 2004; Hauwert, Samson, and others 2004; Mahler and others, 2006; Garner and Mahler, 2007; Mahler and Massei, 2007; Mahler and Garner, 2009; Musgrove and others, 2010).

Acknowledgments

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Methods

Methods used for this study are described in the sections below, including the design of the well network; drilling, monitoring well installation, coring, and core analysis; collection and analysis of water samples; collection of depth-dependent flow and chemistry data; quality assurance of water-quality data; and a description of statistical and geochemical modeling methods.

Design of Well Network

The well network was designed so that an existing PSW would be the focus of the study. Monitoring wells were installed to allow monitoring of groundwater at different depths along likely flow paths from potential source areas contributing to the PSW. An additional 38 PSWs that tap the regional aquifer were sampled to characterize water quality throughout the aquifer and over a wide range of pumping conditions (fig. 1; table 1 at end of report).

In the selected well field, at least one of the six wells is pumping in the well field at any time. Two wells are routinely pumping during summer months when demand increases, and more than two wells might be pumping during peak demand periods. Annual pumpage from the well field was about 7,840 million gallons (Mgal) in 2008 (TANC study water purveyor, written commun., 2008), and the well field produced about 11 percent of total pumpage provided by the purveyor (one of several purveyors in the greater San Antonio area). The first well in the well field was installed in 1963 (one of four installed in the 1960s), and the sixth well was installed in 1984 (one of two installed in the 1980s). The PSWs are cased with steel in the overlying confining units and are open hole through most or all of the Edwards aquifer. The open interval of the wells were acidified at the time of construction to increase production. Depths to the top of the open interval in the wellbores (at or near the top of the aquifer) at the well field range from 321 to 360 ft. Open-interval lengths range from 454 to 495 ft. Open-hole diameters range from approximately 30 to 36 in. Open-hole diameters are variable because of vuggy porosity and collapsed limestone. Rated pump

capacities for the well-field wells range from 11.2 to 20.7 million gallons per day (Mgal/d) (TANC study water purveyor, written commun., 2008).

One or more wells in the well field were not operational at any given time during the study period. Well 3 (W3) was initially selected for study. However, well 4 (W4) was sampled for depth-dependent sampling because there was no pump in the well, which permitted access for depth-dependent sampling equipment, and well 6 (W6) was sampled for temporal characterization because W3 was not operational at the time of the hydrologic event during which temporal changes were evaluated. As a result, although a single PSW was initially selected, different components of the study collected samples from different PSWs within the well field. Well construction, hydrogeology, and geochemistry among all the PSWs in the well field are similar, and data collected from each of the wells are presumed to be comparable for the purposes of this study.

It was not possible to characterize karstic flow paths to the selected PSW because of the large capture zone for the well and because of the karstic nature of the system. Therefore, two vertically nested monitoring well clusters were installed in proximity to the PSW to evaluate processes close to the PSW. The locations of the two clusters were selected on the basis of the two potential flow zones to the PSW from groundwater modeling results (Lindgren and others, 2011b). The monitoring well clusters were drilled as close to the PSW as possible. The Zarzamora well cluster was about 0.5 mile (mi) to the east-northeast of the PSW, and the Timberhill well cluster was about 1.5 mi to the west (fig. 4).

The Zarzamora monitoring well cluster, consisting of four wells, was installed in August–September 2007 (fig. 4). The first monitoring well (Z–DED) was drilled with the goal of being completed near the top of the karstic Kirschberg evaporite member (HGU 8) in the lower part (Kainer Formation) of the Edwards aquifer. This unit is known to be a prolific water-bearing zone (Maclay, 1995). Information learned from drilling and logging this first deep Edwards monitoring well (Z–DED) was used to determine the target depths of the other two Edwards monitoring wells completed at this location. An intermediate-depth well (Z–IED) was drilled to target the bottom of the leached and collapsed members of the Edwards aquifer (HGUs 4 and 5), also known to be a prolific water-bearing zone (Maclay, 1995). The shallowest well (Z–SED) was completed at the top of the Edwards Group in the cyclic and marine members of the Edwards aquifer (HGU 3). A fourth monitoring well (Z–OVB) was completed in an overburden unit (HGU 1), the overlying water-bearing Buda Limestone (fig. 4). This well (Z–OVB) was constructed to evaluate the possibility of a hydrologic connection between the Edwards aquifer and overlying formations at this location.

The Timberhill monitoring well cluster, consisting of two wells, was installed in April 2008 (fig. 4). Similar to the deep- and intermediate-depth wells at the Zarzamora site, the Timberhill wells were completed near the top of the Kirschberg evaporite member (T–DED) (HGU 8) and near the top of the leached and collapsed members (T–IED) (HGU 4).

PSWs throughout the regional aquifer were identified, and a subset was selected for sampling to characterize water quality in PSWs throughout the aquifer. A total of 299 PSWs were identified throughout the regional extent of the San Antonio segment of the Edwards aquifer and were separated into four quartiles based on pumpage. Thirty-nine wells (including the selected PSW) were randomly selected for sampling by quartile, spaced at least 0.62 mi apart within a quartile (fig. 1). Fifteen wells from the largest quartile of pumping were selected, and eight were selected from each of the lower three quartiles of pumping. From the largest pumping quartile, 13 of the wells were in Bexar County, 1 in Uvalde County, and 1 in Comal County. Wells from the lower three pumping quartiles were in Uvalde, Medina, Bexar, and Comal Counties. Water samples were collected from the top quartile wells (in 2004) and the lower three quartile wells (in 2005) and were analyzed for major ions, nutrients, trace elements, dissolved organic carbon (DOC), ultraviolet (UV) absorbance, pesticide compounds (including breakdown products), and volatile organic compounds (VOCs). A subset (8 of 15) of the top quartile wells were resampled for a subset of organic compounds in 2005. Results from the top quartile wells are described in Carter and others (2007) and Hopple and others (2009), and results from all of the 39 PSWs are described in Musgrove and others (2010).

Drilling, Well Installation, Coring, and Core Analysis

Monitoring well design was reviewed and approved by Edwards Aquifer Authority and San Antonio Water System permitting authorities and met requirements of the Texas Department of Licensing and Regulation. Monitoring wells were installed by USGS by using an air rotary drill rig and generally followed procedures modified from Lapham and others (1995) for drilling in karstic limestone. Equipment was decontaminated by steamcleaning between drill sites.

The design was similar for all Edwards aquifer monitoring wells. A 14-in.-diameter hole was drilled to a depth of approximately 20 ft, where a 12.75-in.-diameter surface protector casing was installed and cemented. This protector casing was used to prevent excavated materials from entering the borehole during drilling. An 11-in.-diameter borehole was then drilled through overlying confining units (Austin Chalk, Eagle Ford Group, Buda Limestone, and Del Rio Clay) to a 10-ft depth into the Georgetown Formation at the top of the Edwards aquifer. Thickness of confining units was approximately 320 ft at the Zarzamora location and 385 ft at the Timberhill location. A 6.63-in.-diameter steel casing was installed to this depth to isolate the overlying units from the Edwards aquifer. Cement was subsequently installed in stages by tremie technique to prevent loss of cement to voids. After waiting 24 hours for cement surrounding the steel casing to cure, a 6-in. borehole was drilled to the desired depth in the Edwards aquifer. A schedule 80, 2.5-in.-diameter polyvinyl chloride

(PVC) casing was installed in the borehole to a depth of about 10 ft above the targeted zone. The PVC casing was reduced to 2 in. diameter at this depth. About 3 ft of 2-in.-diameter PVC casing was added below the reducer. Two bell-shaped rubber formation packers (shale traps) were clamped around this casing to make a seal between the PVC pipe and formation above the screen. A screen of desired length was installed below the formation packers. Screen slot size was 0.04 in. (0.040 slot). Screen lengths ranged from 10 to 30 ft depending on the presence and sizes of voids identified from geophysical logs and whether the well could be completed in the desired zone. Steel casing was not required for the overburden well (Z-OVB). Bentonite pellets were poured from land surface to seal the zone between the depth of shale traps to a depth inside steel casing. Bentonite and Portland cement were used to seal the well from about 20 ft below to about 3 ft above land surface. Wells were completed with a surface cement wellpad, locking metal shroud, and locking cap for PVC pipe. Wells were developed by jetting after drilling was completed. Completed wells were subsequently pumped until stabilization of field properties (water temperature, specific conductance, pH, dissolved oxygen, and turbidity).

Cores were collected by using a 4-in.-diameter, 10-ft-long core barrel at eight selected intervals: three in the overburden units and five in the Edwards aquifer. The intervals in the overburden units included Austin Chalk/Eagle Ford Group contact, Eagle Ford Group/Buda Limestone contact, and the Buda Limestone/Del Rio Clay contact (fig. 2). Intervals in the Edwards aquifer included Del Rio Clay/Georgetown Formation contact, leached and collapsed members, regional dense member/grainstone member contact, grainstone member, and Kirschberg evaporite member HGUs (fig. 4).

Core samples were collected aerobically, and then subsets of material were collected from each core and placed in airtight jars under a nitrogen (N) atmosphere to minimize oxidation reactions. Samples were stored chilled to about 4°C until treatment. Solid materials to be analyzed for inorganic (selected major ion and trace element) chemical analysis were removed from the airtight containers and were crushed to less than 2 millimeters (mm) diameter at the University of Texas at San Antonio Geoscience Laboratory. Crushed samples were returned to a N atmosphere and sent to the USGS National Research Program Inorganic Chemistry Laboratory in Reston, Va., for chemical analysis of selected ions. Crushed samples under a N atmosphere also were sent to the USGS National Water Quality Laboratory in Denver, Colo., for analysis of percent total carbon and percent inorganic carbon. Bulk samples were sent to the USGS National Research Program Clay Mineralogy X-ray Diffraction Laboratory in Reston, Va., for X-ray diffraction (XRD) analysis. Selected major, minor, and trace element concentrations of core samples were determined from two acid extractions (leaches): 6 normal (N) hydrochloric acid (HCl) (Cornwell and Morse, 1987; Brugam and others, 1988), and 10 percent nitric acid (HNO₃) (Rabinowitz and Bellinger, 1988). Identical extraction procedures for each acid were used; 5.0 ± 0.5 grams (g) of crushed core sample was

weighed into a 50-milliliter (mL) polycarbonate centrifuge tube. One 15-mL aliquot of acid was added. The tube was left open to the atmosphere for 30 minutes to permit the evolution of carbon dioxide (CO_2) gas from sediments with high carbonate content and then was tightly capped. Extractions were performed on an “as received” basis; that is, wet. Analytical concentrations were corrected to dry weight after calculating for moisture content. Sample tubes were placed on a shaker table at room temperature for 6 hours to ensure complete mixing. Samples were centrifuged, and the supernatant (leachate) was withdrawn with 20-mL polycarbonate disposable syringes and filtered through 0.45-micrometer (μm) disposable polycarbonate filters into 10-mL centrifuge tubes. The leachate was analyzed for selected major ions and trace elements by using inductively coupled plasma-emission spectrophotometry (ICP-ES). Leachates derived from the 6 N HCl extractions were diluted to 1.2 N HCl, and leachates from the 10 percent HNO_3 extractions were diluted to 2 percent HNO_3 prior to analysis with the ICP-ES. Further dilution (5–20 times) was required for analysis of As, lead, selenium, and U with inductively coupled plasma-mass spectroscopy (ICP-MS). Concentrations for As are not reported for the HCl extractions because of mass interference of As with chloride (Cl). Concentrations derived from the leachate analyses are semiquantitative because a small amount of residual solids remained after extractions. The methods for XRD scans, carbonate digestions, and calculations of semiquantitative weight percents are described in detail in Webster and Jones (1994). XRD was performed on bulk samples and on samples that were digested to remove carbonate so that residual particulates of less than 1 mm size could be analyzed. In some cases, a second digestion was performed. The qualitative carbonate digestion technique was designed to release clay minerals bound within grains of carbonate. Excess carbonate was abundant after chemical treatments, but enough clay was separated to make reasonable phase identifications. Any mineral with quantifiable principal or secondary peaks resolved significantly above background was identified. Some trace identifications could not be quantitatively discriminated from background and were identified as trace. Although qualitative weight percent abundance is provided in the data (table 6 at end of report), identifications are listed in the order of the relative intensity of principal peaks where not quantified. Peak intensity is proportional to weight percent in an indirect and sometimes unreliable way. XRD weight percent calculations have large errors because peak intensities routinely vary by 10 percent or more from one scan to the next, even when the same slide is analyzed.

Geophysical and Depth-Dependent Flow Measurements

Borehole geophysical logs such as caliper, normal resistivity, induction conductivity, and natural gamma, as well as fluid resistivity and temperature, and vertical flow rates were collected in PSW W4 and in the deepest monitoring wells

in the two monitoring well nests (Z–DED and T–DED). W4 was logged through 30-in.-diameter steel casing from land surface datum (LSD) to 320 ft below LSD and through open hole from 321 ft to the depth of 753 ft (total depth of hole is 816 ft) below LSD with the production pump removed. For flow measurements in PSW W4, a temporary turbine pump was lowered below the existing water-level depth such that the intake was a few feet above the bottom of the steel casing, allowing for drawdown and a pumping rate of 650–850 gallons per minute (gal/min). This pump was used to stress the well to simulate production conditions during the pumped flowmeter logging and depth-dependent water sample collection. This pumping rate, however, is considerably lower (by one to two orders of magnitude) than typical well production rates and rated pump capacities (TANC study water purveyor, written commun., 2008). The monitoring wells were logged in open hole prior to completion under ambient conditions.

Fluid resistivity logs provide a record of the capacity of the borehole fluid to conduct electrical current (Keys, 1997). Changes of fluid resistivity are measured by ring electrodes inside a housing that allows borehole fluid to flow through it. The best fluid resistivity logging results are achieved when logging downward into boreholes containing ambient water that has had sufficient time to stabilize. Ideally, fluid resistivity logs are collected as the first logging run to record ambient conditions before other probes have passed through the borehole and vertically mixed the borehole fluid. Curve deflections on the fluid resistivity log might indicate horizontal or vertical flow movement, stratification of borehole fluid, or screen openings in cased wells. The fluid resistivity values can also be used in conjunction with other geophysical measurements and logs. Fluid conductivity, the reciprocal of fluid resistivity, was compared with measured specific conductance values. The fluid conductivity values for this study were recorded at the ambient borehole temperature and were not corrected to a standard temperature. A Century model 8044 multiparameter E-log probe (Century Geophysical Corporation, Tulsa, Okla.) was used to log fluid resistivity and temperature. Calibration of the fluid resistivity logging probes was completed with solutions of known conductivity/resistivity in a two-point calibration. Fluid resistivity and temperature logs were completed in W4 during ambient conditions (well field off) and with the well field pumping (well field on).

Natural gamma logs provide a record of gamma radiation detected at depth in a borehole and are useful in determining lithologies and contact depths of the strata penetrated by the borehole. Fine-grained sediments that contain abundant clay tend to be more radioactive than quartz-rich sandstones or carbonates (Keys, 1997). The natural gamma log was run simultaneously with the fluid resistivity log and was recorded in natural gamma counts per second by using a natural gamma sensor with a sodium iodide detector built into a Century 8144c multiparameter probe. The natural gamma sensor, factory calibrated, requires no field calibration. Natural gamma count rates, which typically increase in the proximity of clay and shale, might also increase slightly when adjacent to any

bentonite seals in the wells. The caliper log was measured by using a three-arm caliper, which measures the diameter of the borehole by using a three-arm configuration that can be opened downhole.

The electromagnetic (EM) flowmeter probe measures the vertical flow rate and direction of vertical flow in a borehole by using the principle of Faraday's Law of Induction. The EM flowmeter probe consists of an electromagnet and two electrodes oriented 180 degrees apart and 90 degrees to the magnetic field inside a hollow cylinder or tube. The voltage induced by a conductor moving at right angles through the magnetic field is directly proportional to the velocity of the conductor (water) through the field (Century Geophysical Corporation, 2010). The direction of vertical water flow is determined by the polarity of the response, with upward flow being positive and downward flow being negative. Downward and upward flow were calibrated at a rate of 1 gal/min through the tool in a controlled environment with a bypass factor calculated for high flow within the wellbore. Flowmeter and fluid properties data for W4 were collected for four different pumping conditions: (1) no pumping in the PSW or well field (an operationally rare condition) ("ambient 1"); (2) no pumping in the PSW, but with pumping nearby well(s) in the well field (greater than 7,000 gal/min) ("ambient 2"); (3) PSW pumping (850 gal/min), no other pumping in the well field ("moderate"); (4) PSW pumping (850 gal/min) and with pumping of nearby well(s) in the well field (greater than 7,000 gal/min) ("normal").

Collection of Water Samples

Groundwater samples were collected from PSWs in the regional aquifer, monitoring wells (Timberhill and Zarzamora wells and an overburden well), and well-field PSWs during 2004–9 (table 1). Well-field samples included wellhead samples (W2, W3, W4, W5, and W6) and depth-dependent (W4) and temporal (time series in response to a rainfall/recharge event) samples (W6). Temporal samples also were collected from the Zarzamora Edwards aquifer monitoring wells and an upgradient unconfined monitoring well (AY-68-27-610). For the purposes of comparing water chemistry, groundwater samples were categorized as being from regional aquifer PSWs, from the well field (wellhead samples), from the monitoring wells (Zarzamora and Timberhill monitoring well clusters, excluding the overburden well), from the overburden monitoring well, from the PSW depth-dependent sampling, and from temporal sampling (from the PSW and selected monitoring wells (including an upgradient unconfined monitoring well)). Constituents analyzed for are summarized in table 2 at end of report. It was not possible to collect a surface discharge (wellhead) sample during the depth-dependent sample collection from W4; alternatively, this sample was approximated by placing the sample pump intake just below the bottom of the turbine pump intake (near the bottom of the casing at 320 ft of depth). The water sample collected from this intake point was

assumed to represent discharge at the surface and is considered hereinafter as the wellhead sample from W4. Depth-dependent and temporal samples are described in more detail in the two following sections.

Sampling protocols used for this study are described in detail in Koterba and others (1995), Koterba (1998), Lapham and others (1995), and U.S. Geological Survey (variously dated). Groundwater samples were collected and processed in a mobile water-quality laboratory. Existing pumps, permanently installed in each PSW (except W4), were used to deliver water to the land surface. In most cases, a permanent spigot near the wellhead was used as the sampling point. In a few instances, plumbing connections near the wellhead were modified temporarily for sampling purposes. Water was transferred from the wellhead spigot to the mobile laboratory by using Teflon® (DuPont™) tubing with stainless steel fittings. To minimize risk of contamination, sample collection and preservation took place in chambers consisting of clear polyethylene bags supported by plastic frames, much like laboratory glove boxes. Equipment that came in contact with the water sample was cleaned between each well sampling by using a progression of (1) dilute phosphate-free detergent to remove both organic and inorganic residues, (2) deionized water rinse to remove detergent, (3) methanol wash to remove organic residue, and (4) deionized water rinse to remove any remaining methanol residue. Monitoring wells were sampled by using a portable positive displacement groundwater sample pump.

Wells were purged before sampling. During the initial pumping period, measurements of field parameters (specific conductance, pH, temperature, dissolved oxygen) were routinely monitored in a sealed flowthrough cell. Turbidity was monitored by collecting periodic water samples from a discharge tube and analyzing with a portable field turbidimeter. Water samples were collected after stable measurements of these properties were achieved (Koterba and others, 1995; U.S. Geological Survey, variously dated). Water was redirected by way of a manifold to the sample chamber inside the mobile laboratory for sample collection.

Additional parameters measured at the time of sample collection were carbonate alkalinity, ferrous and total iron, and sulfide. Alkalinity was measured by using an inflection point titration method (Rounds, 2003). Iron species and sulfide were analyzed by field spectrophotometric methods, and results are considered qualitative. Additional samples were collected for analysis of major ions (mineral analysis), nutrients (N and phosphorus compounds), DOC, UV absorbance (organic carbon), trace elements (mostly metals), pesticide compounds (pesticides and their degradates), VOCs, and wastewater indicator compounds. Samples for analysis of selected isotopes, radon gas, radium-226 and radium-228, and age-dating compounds and isotopes (chlorofluorocarbons (CFCs, specifically CFC-11, -12, and -113), sulfur hexafluoride (SF₆), tritium (³H), helium-3 (³He), and associated dissolved gases) were collected at a subset of wells. Most analyses were performed at the USGS National Water Quality Laboratory. Dissolved gases,

CFCs, and SF₆ were analyzed at the USGS Reston Chloro-fluorocarbon Laboratory in Virginia. Helium-3 was analyzed at the Gas Extraction Laboratory of Lamont-Doherty Earth Observatory in Palisades, N.Y. Tritium was analyzed at the Gas Extraction Laboratory of Lamont-Doherty Earth Observatory or at the USGS Isotope Tracers of Biogeochemical and Hydrologic Processes Laboratory in Menlo Park, Calif. (Thatcher and others, 1977). Age-dating analyses and methods are further described below in the section “Groundwater Age-Dating Methods.”

Depth-Dependent Samples

Geophysical and flowmeter logs were collected from W4 to determine zones of contribution to the wellbore prior to sampling; the W4 borehole was selected because it did not contain a production pump. Prior to logging and sampling, a “test” turbine pump, capable of pumping up to 850 gal/min, was installed in the borehole of W4 to approximate PSW pumping conditions. Although the test pump did not pump at the capacity of production pumps, it was deemed sufficient to stress the well for the purpose of this study. As described earlier in the section “Geophysical and Depth-Dependent Flow Measurements,” four pumping conditions were logged with the flowmeter: ambient 1, ambient 2, moderate, and normal. The normal pumping condition (for example, with one or two wells in the well field pumping at total discharge rates from approximately 7,778 to 28,270 gal/min) was likely to be most representative of pumping conditions associated with the production of supply water. Each of the pumping conditions produced different results for stratigraphic zones that contributed water to the well. This information was used to select three depth intervals for depth-dependent sampling under three of the four pumping conditions: ambient 1 (hereinafter, ambient), moderate, and normal.

Selected depths for sampling based on zones of flow within the well were (1) shallow (380 ft), which was just above the high-conductivity zone, (2) intermediate (430 ft), and (3) deep (540 ft), which are associated with three different HGU's, specifically, (1) the cyclic and marine members (HGU 3), (2) the leached and collapsed members (HGU's 4 and 5), and (3) above the Kirschberg evaporite member (HGU's 8 and 9) in the grainstone member of the Edwards Group (HGU 7), respectively (fig. 4). Depth-dependent samples were collected by suspending 2.5-in.-diameter PVC casing (tubing) to the three selected depths. A positive displacement sample pump (1.8-in. outside diameter) was placed inside the PVC casing and used to collect water samples from the three selected depths. This approach is similar to that used by Izicki (2004) to collect samples from discrete depths in production wells under pumping conditions. The PVC tubing was purged of three casing volumes prior to sampling. Water samples were collected at the selected depth intervals for each of the three pumping conditions as defined above (ambient, moderate, and normal). An additional sample was collected

under the moderate condition from below the existing casing of the wellbore to approximate the wellhead sample for W4. Depth-dependent sampling details for the PSW, as well as sampling of the monitoring wells completed at different depths, are shown in table 3 at end of report.

Temporal Water Samples

Selected monitoring wells and one of the well-field PSW's (W6) were sampled several times in response to a rainfall and recharge event that occurred in October 2009 to assess short-term (event-scale) temporal variations in water quality for the TANC study area. Sampled monitoring wells were three Zarzamora wells (Z-SED, Z-IED, and Z-DED) and an upgradient unconfined monitoring well (AY-68-27-610) (table 1). Well AY-68-27-610 is one of the shallow/urban unconfined monitoring wells described by Musgrove and others (2010). It is approximately 4.5 mi to the north of the well field (fig. 3).

A baseline sample was collected from these five wells on April 8 through 14, 2009. At that time, it was believed that heavy spring rains were likely. However, an extended drought continued in the region throughout the summer of 2009, as evidenced by declining water levels in the Bexar County index well (J-17) (fig. 1), indicative of very dry regional conditions (fig. 5). From June through August 2009, a total of approximately 1.4 in. of rain was recorded at the San Antonio airport. The sampled storm event was preceded by 6.35 in. of rain in September, which resulted in an increase in water level close to that observed when the baseline sample was collected in April (fig. 5). Between October 3 and 4, 2009, 6.06 in. of rain was recorded at the San Antonio airport, and no additional rain was recorded until October 9, 2009. The first event samples were collected from October 4 through October 6, 2009. A total of six samples, including the baseline sample, were collected from each designated well through November 19, 2009 (table 4 at end of report). An additional 6.73 in. of rain was recorded at the San Antonio airport between October 9 and November 19, 2009. As a result, the response in the study area to the designated rain and recharge event cannot be distinguished from subsequent rain and recharge during the event sampling interval.

Laboratory Analytical Methods

Depending on sampling event, samples were analyzed for a variety of constituents, including major ions, nutrients, UV absorbance, DOC, pesticides and pesticide degradates, VOCs and wastewater indicator compounds, fuel oxygenates, stable and radiogenic isotopes (N, oxygen, hydrogen, strontium), trace elements, radon-222 gas, or radium-226 and -228 (table 2). The above constituents were not collected in every sample; rather, they were collected for particular study objectives. Selected groundwater samples were analyzed for dissolved gases and geochemical age-dating compounds and isotopes: CFCs, SF₆, ³H, and ³He.

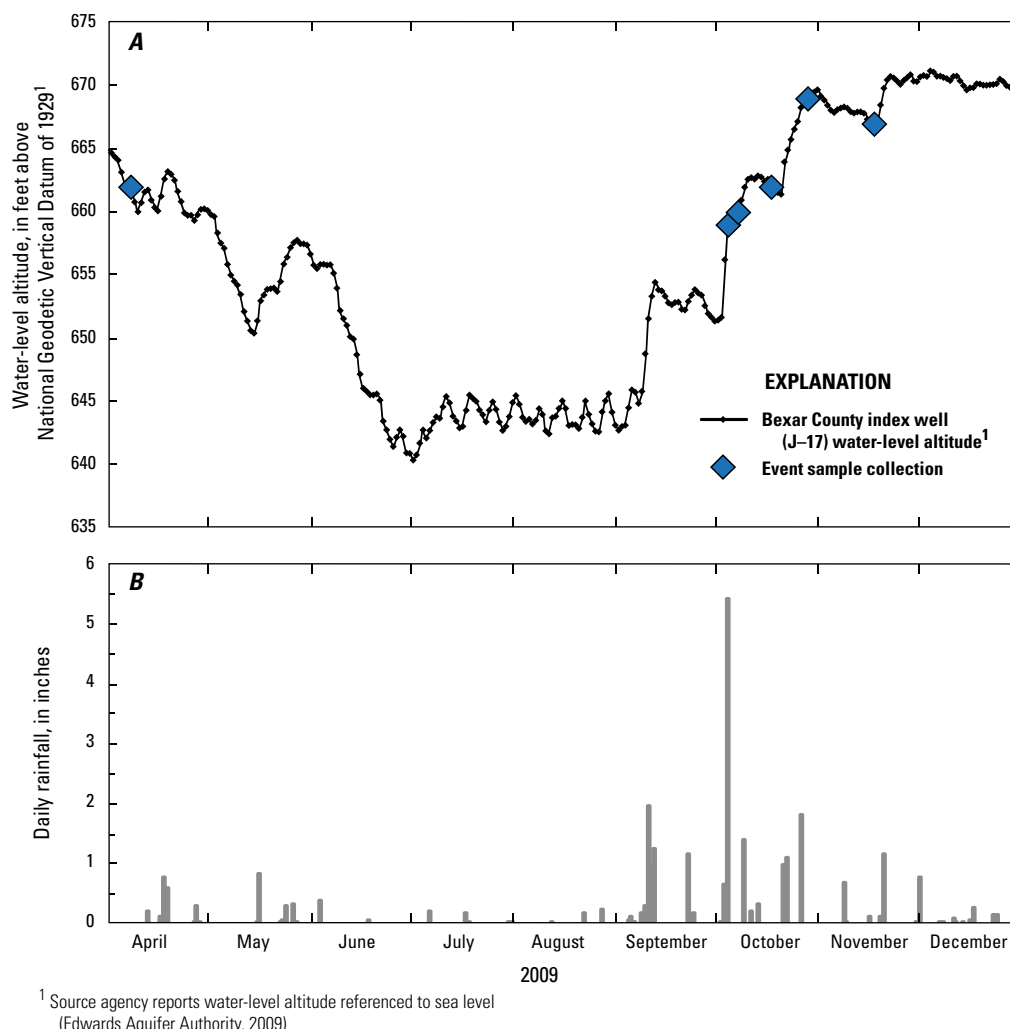


Figure 5. Time series (April 1 to December 31, 2009) climatic and hydrologic data for the south-central Texas study area. *A*, Bexar County index well (J-17) water-level altitude (Edwards Aquifer Authority, 2009), and dates of sample collection for event sampling (well field W6). *B*, San Antonio rainfall (recorded at the San Antonio International Airport) (National Oceanic and Atmospheric Administration, 2010).

Samples for major ions, trace elements, and nutrients were filtered through a 0.45- μ m capsule filter and analyzed by using ICP-ES, ICP-MS, ion-exchange chromatography, and colorimetry procedures described in Fishman and Friedman (1989), Faires (1993), Fishman (1993), McClain (1993), Garbarino (1999), Garbarino and others (2006), and Patton and Kryskalla (2003). Hydrogen sulfide was measured in the field by using colorimetric methods (Hach Company, 2002). Water samples for radon-222 were collected in a syringe prior to contact with the atmosphere and injected into a mineral-oil-based scintillation solution in a 30-mL glass scintillation vial. The sample was shipped overnight to the laboratory and analyzed by liquid-scintillation counting methods (Prichard and Gesell, 1977; American Society for Testing and Materials, 1996). Samples for stable isotopes of hydrogen (δ D), oxygen

(δ^{18} O), and NO_3 ($\delta^{15}\text{N}$ and $\delta^{18}\text{O}$ of NO_3) were analyzed by the USGS Stable Isotope Laboratory in Reston, Va. Isotope data are presented in delta (δ) notation as the ratio of the heavy to the light isotope, normalized to a standard (eq. 1).

$$\delta_{\text{sample}} = 1,000 [(R_{\text{sample}}/R_{\text{standard}}) - 1] \quad (1)$$

where

R_{sample} is the ratio of the heavy to the light isotope in the sample and

R_{standard} is the ratio of the heavy to the light isotope in the standard.

δ D and δ^{18} O were analyzed by using an isotope ratio-mass spectrometer by hydrogen gas-water equilibration and CO_2 -water equilibration techniques, respectively, and are

reported relative to Vienna Standard Mean Ocean Water (Epstein and Mayeda, 1953; Coplen and others, 1991; Coplen, 1994; Révész and Coplen, 2008a, 2008b). Nitrate isotopes ($\delta^{15}\text{N}$ and $\delta^{18}\text{O}$ of NO_3) were analyzed by bacterial conversion of NO_3 to nitrous oxide and reported relative to N_2 gas in air and Vienna Standard Mean Ocean Water, respectively (Sigman and others, 2001; Casciotti and others, 2002; Révész and Casciotti, 2007).

VOCs, pesticides, pesticide degradates, DOC, major ions, trace elements, nutrients, and radon-222 were analyzed by the USGS National Water Quality Laboratory. Samples for VOCs and fuel oxygenates were unfiltered and determined by gas chromatography and mass spectrometry by using methods described in Connor and others (1998) and Rose and Sandstrom (2003). Pesticides samples were filtered by using a 0.7- μm glass-fiber filter and analyzed by carbon-18 solid-phase extraction and capillary column gas chromatography and mass spectrometry (Zaugg and others, 1995; Lindley and others, 1996; Furlong and others, 2001; Sandstrom and others, 2001; Lee and Strahan, 2003; Madsen and others, 2003). Wastewater indicator compounds were collected similarly to pesticide samples and were analyzed as described in Zaugg and others (2002).

DOC samples were collected by forcing raw water samples through a 0.45- μm glass-fiber filter by using high-purity N_2 gas. DOC samples were analyzed by methods described by Brenton and Arnett (1993). Nitrogen species concentrations are reported as N (for example, NO_3 as N; hereinafter, NO_3), and phosphorus species concentrations are reported as phosphorus (P).

Groundwater Age-Dating Methods

To facilitate comparison of results between groundwater age tracers, multiple tracers were measured in selected groundwater samples, including concentrations of CFCs, SF_6 , ^3H , and the ratio of the parent/daughter isotopes $^3\text{H}/^3\text{He}$. Because the solubility of atmospheric tracers varies as a function of temperature, dissolved gas concentrations (argon, CO_2 , methane, N_2 , and oxygen) were also measured to estimate the temperature of groundwater at the time of recharge, as well as to calculate the amount of excess air and excess N. Excess air is that which is trapped in pores (for example, after a rapid water table rise) and subsequently dissolved in groundwater to concentrations above ambient conditions. Excess N reflects reactions with N species, such as denitrification. Helium-4 (^4He) was also measured to determine whether the $^3\text{H}/^3\text{He}$ age-tracer method was suitable.

Anthropogenic activities have released SF_6 , CFCs, and ^3H into the atmosphere in low concentrations since the 1940s and 1950s (Schlosser and others, 1988, 1989; Busenberg and Plummer, 1992, 2000; Solomon and others, 1992). Precipitation incorporates these constituents from the atmosphere, and as a result, groundwater recharge carries a chemical or isotopic signature that can be related to atmospheric conditions at the time of recharge. The $^3\text{H}/^3\text{He}$ method is based on

the radioactive decay of ^3H to ^3He . Each age-tracer technique has associated uncertainties and might be affected differently by aquifer processes. Because all physical and chemical processes that might affect age-tracer results cannot be accounted for, results using these techniques are reported as apparent ages. CFC and SF_6 dating techniques are described in Busenberg and Plummer (1992, 2000, respectively). Techniques for $^3\text{H}/^3\text{He}$ dating are described by Schlosser and others (1988, 1989) and Solomon and Cook (1999).

Dissolved helium (He) in groundwater is derived from both atmospheric and terrigenic (from earth) sources. The atmospheric component includes air-soluble He and excess-air He (He derived from excess air) that results from dissolution of air bubbles trapped just below the water table. Another atmospheric source of He (as ^3He) is from decay of ^3H in recently (less than 60 years old) recharged water; ^3H concentrations in rainfall, some of which became recharge, during the mid-20th century were increased substantially by atmospheric testing of nuclear weapons (Michel, 1989). The He derived from ^3H decay is the He component of interest for the $^3\text{H}/^3\text{He}$ age-tracer method. Helium is also produced by the radioactive decay of U and thorium in crustal rocks and aquifer solids (crustal He) and from upward diffusion or advection of He from the mantle (mantle He). Together, crustal and mantle He are termed terrigenic He (Solomon, 2000). Potential sources of terrigenic He in the Edwards aquifer include igneous rock intrusions, oil and gas hydrocarbons that are widely present in Edwards Group rocks downdip from the freshwater zone of the Edwards aquifer, rocks composing the underlying and adjacent Trinity aquifer, and rocks underlying the Trinity aquifer. Corrections for atmospherically derived excess He are included in estimating ages. If excess He concentrations are higher than those which result from atmospheric sources, then it is generally not possible to estimate an age with the $^3\text{H}/^3\text{He}$ age-tracer method.

Samples for age-dating and dissolved gas analyses were collected according to protocols described by the U.S. Geological Survey Reston Chlorofluorocarbon Laboratory (2010). Dissolved gas samples, including samples to be screened for ^4He , were collected in preweighed, previously evacuated 150-mL glass septum bottles. Upon removal of the rubber stopper, a sample discharge tube was placed at the bottom of the sample bottle. After the sample bottle was initially filled, it was placed in a 1-liter (L) beaker filled with sample water such that the top of the sample bottle was under water. The sample discharge tube flowed continuously during this process until three or more sample-bottle volumes had flowed from the sample discharge tube into the septum bottle. The sample discharge tube was removed, and a rubber stopper containing a needle was inserted into the top of the septum bottle while still submerged. The needle was removed after the stopper was securely placed. Septum bottles were stored inverted and chilled until analysis. CFC samples were collected in 125-mL glass bottles that were sealed with screw caps lined with aluminum foil. Sample bottles were filled by using a similar method as the dissolved gas samples. Bottle caps were kept

under water until after they were applied to the top of the sample bottle. Bottle caps were taped to prevent loosening, and samples were stored inverted at room temperature until analysis. SF_6 samples were collected in 1-L glass bottles and sealed with polyseal screw caps. The sample discharge tube was placed at the bottom of the sample bottles, and sample water flowed into the bottles for three or more sample-bottle volumes. The sample tube was removed, and samples were capped such that no headspace was allowed to form prior to sealing the sample bottle. Sample caps were taped, and samples were stored at room temperature until analysis. ^3H samples were collected in 1-L polyethylene bottles. The sample discharge tube was placed at the bottom of the sample bottle until the bottle was filled such that a small headspace remained. Sample bottles were sealed with polyseal screw caps, and the caps were taped. Samples were stored at room temperature until analysis. ^3He samples were collected in copper tubes mounted on trays that contained clamps at each end. The sample discharge tube was attached to the inlet side of the copper tube, and a plastic tube (Tygon®, Saint-Gobain Performance Plastics) containing a valve was attached at the outlet side of the copper tube. Sample water flowed through the copper tube with some back pressure applied by way of the attached valve. The copper tube was gently tapped to remove any bubbles that might have formed during the sampling procedure. After the copper tube was purged with three sample volumes, the clamps on the copper trays were tightly closed to seal the copper tubes, and the external tubing was removed. Samples were stored at room temperature until analysis.

Groundwater samples were analyzed for dissolved gases, SF_6 , and CFCs at the USGS Reston Chlorofluorocarbon Laboratory in Virginia. Dissolved gases were extracted from headspaces of glass samplers and analyzed by gas chromatography (Busenberg and others, 1998). Analytical methods for SF_6 followed those described by Busenberg and Plummer (2000). CFCs were analyzed by using purge and trap gas chromatography with electron-capture detector (GC-ECD) (Bullister, 1984; Bullister and Weiss, 1988; Busenberg and Plummer, 1992). Groundwater samples for ^3H and ^3He were analyzed by the Gas Extraction Laboratory of Lahmont-Doherty Earth Observatory in Palisades, New York, by using quantitative gas extraction followed by mass spectrometric techniques; methods are described by Clark and others (1976), Ludin and others (1998), and Schlosser and others (1989). The $^3\text{H}/^3\text{He}$ methods included analysis of concentrations of ^4He and neon (Ne), which are used in $^3\text{H}/^3\text{He}$ age-tracer calculations.

In spite of the analytical precision available for measurement of many geochemical age tracers, there are numerous uncertainties associated with groundwater age interpretations (Phillips and Castro, 2003; Bethke and Johnson, 2008). Age interpretation for a sample that consists of water of a single age might be relatively simple but becomes increasingly complex for a sample that consists of a mixture of water; that is, a mixture of water “parcels,” each of which has a discrete age (Böhlke, 2006). Age-interpretation complexities associated with mixing might be especially important in karst aquifers,

where each component of a complex network of conduits, fractures, and matrix porosity might have a different groundwater age and a well or spring might discharge water from any combination of porosity types (Plummer and Busenberg, 2005; Long and Putnam, 2006). Models for the hypothetical concentrations of age tracers in water samples with different age distributions that consider the effects of hydrologic processes such as dispersion, exponential mixing, and binary mixing of shallow and deep groundwater have been developed to facilitate interpretation of geochemical age tracers (Böhlke, 2006; Long and Putnam, 2006). Age-tracer concentrations were compared with results expected from different lumped-parameter models by using atmospheric input data for tracers and using age-date tracer modeling software developed by Böhlke (2006). Lumped-parameter models assume a steady-state flow system and also assume that selected tracers behave like a water molecule.

Quality Control

Quality-control (QC) samples are collected and analyzed to evaluate bias and variability of environmental samples. Bias is systematic error inherent in a method or caused by some artifact of the measurement system. Variability is the degree of variation in independent measurements that results from repeated application of the process under specified conditions (T.L. Schertz, U.S. Geological Survey, written commun., 1997). QC data are particularly useful for analyzing and interpreting water quality data analyzed for small concentrations, including those measured at $\mu\text{g/L}$ levels for many constituents in this study. QC samples collected and analyzed for this study consisted of three types of samples collected onsite: blanks, replicate environmental samples, and environmental matrix-spike samples. In addition to QC samples submitted from the field, the USGS National Water Quality Laboratory analyzes laboratory blanks, replicates, and spikes. QC data stored in USGS databases are available upon request.

Blank samples test for bias from introduction of contaminants to environmental samples. Three types of blank samples were analyzed in this study: source-solution, equipment, and field. Source-solution blanks consist of high-purity organics-free or inorganics-free water, prepared and tested by the USGS National Water Quality Laboratory for the NAWQA Program. Source-solution blanks were collected with associated equipment or field blanks and were occasionally analyzed for VOCs, DOC, trace elements, or nutrients. Source-solution blanks document the purity of the high-purity waters used to collect field and equipment blanks. A total of 22 blanks (combination of inorganic, organic, field, and equipment) were collected during this study in comparison with 97 environmental samples.

Equipment and field blanks were analyzed for concentrations of a combination of major ions, nutrients, DOC, trace elements, pesticide compounds, or VOCs. Two equipment blanks were collected during the study period prior to collection of the first environmental samples to indicate whether

the equipment or sample collection process would bias the data. Both equipment blanks contained several inorganic constituents, primarily trace metals. Trace metals detected at low concentrations (less than 1 µg/L) included aluminum, cadmium, chromium, manganese, nickel, and vanadium. Zinc was detected in both equipment blanks at concentrations of about 2 and 3 µg/L, and lead was detected in one equipment blank at a concentration of about 2 µg/L. The source of these metals, though at low concentrations, is unknown. Stainless steel fittings were replaced as a result of these detections, and these metals were not evaluated for this study. Calcium (Ca), sodium (Na), silica (Si), sulfate (SO₄), and iodide (I) also were detected at low concentrations (less than 1 mg/L) in one equipment blank, well below concentrations observed in environmental samples. DOC and phenol were detected at low concentrations (less than 1 mg/L and less than 1 µg/L, respectively) in one equipment blank. Occurrences of DOC and phenol are discussed below.

Field blanks were collected at periodic intervals onsite to indicate whether analytical bias occurred from sample collection, cleaning protocols, or contamination of the mobile laboratory. Eleven (organic and inorganic) field blank suites were collected following environmental sample collection and subsequent equipment cleaning. Field blanks were processed the same as environmental samples—by pumping high-purity water through cleaned sample collection equipment, including filter apparatuses, where applicable. DOC was detected in two field blanks. DOC was detected at very high concentrations (greater than 100 mg/L) in one of the field blanks collected in 2005 and was attributed to insufficient removal of methanol during cleaning. Several environmental samples had DOC concentrations that were higher (greater than 10 mg/L) than expected, and it is presumed that the cleaning procedure during 2004–5 affected the concentrations of these samples. Environmental DOC concentrations measured in 2004–5 are considered suspect and are marked with a “V” in data tables, indicating possible contamination. The cleaning process was corrected during the sampling period with additional deionized water rinsing to remove methanol from sampling equipment. Among samples collected in 2004–5 and analyzed for the wastewater schedule, phenol, benzophenone, and 4-nonylphenol were sometimes detected, mostly below method reporting levels, in field blanks and environmental samples. Concentrations in three field blank samples were higher than those in environmental samples. Results for these organic compounds and for caffeine and camphor, on USGS National Water Quality Laboratory Schedule 1433, were not considered to reflect environmental concentrations because of analytical limitations during the study period (Zaugg and Leiker, 2006). The VOCs toluene and styrene were detected at low concentrations (less than 0.09 µg/L) in two field blanks and one source-solution blank. The high-purity blank water might have been the source for toluene and styrene, which were not evaluated in this study. Benzene was detected at a low concentration (0.018 µg/L) in one field blank; that blank also contained

toluene, indicating a possible contamination with fuel used to power field generators. One field blank contained chloroform (trichloromethane) at a concentration of 0.017 µg/L, which is below the long-term method reporting level (0.03 µg/L) and in the range of environmental data evaluated in this study. Chloroform is rarely found in blanks when examining all groundwater QC blanks collected during 1996–2009 by the NAWQA Program in this study area; therefore, the presence of chloroform at such a low concentration, in one field blank, although unexplained, is not considered problematic for interpretation in this study. No pesticide compounds were detected in blanks.

Some trace elements were detected at low concentrations (less than 1 µg/L) in field blanks: chromium, cobalt, manganese, nickel, silver, strontium, and zinc. Aluminum was detected in two field blanks; one sample contained aluminum (3.8 µg/L) above the method reporting level. Aluminum and vanadium were each detected in a source-solution blank. The high-purity blank water might have been a source for low concentrations of these constituents in some blank samples. The more likely source for some trace elements in blank samples is corrosion of stainless steel fittings. One field blank sample contained a low concentration (less than 1 mg/L) of Ca, and two field blanks contained low concentrations of silica. Concentrations of these constituents are orders of magnitude below those of environmental samples. Three field blank samples contained total N, and two of the sample concentrations were above the method reporting level. The detections in field blanks were well below those measured in environmental samples.

Replicate samples are used to evaluate variability in measurements associated with analytical processes, sample collection protocols, or natural variation. Eleven replicate samples were collected sequentially after environmental samples and analyzed for a combination of major ions, nutrients, DOC, trace elements, radon, pesticide compounds, VOCs, or wastewater indicator compounds. Precision can be evaluated by comparing concentrations between environmental and replicate samples. Precision is expressed as relative percent difference (RPD) and is computed as

$$RPD = \{|C_1 - C_2| \div [(C_1 + C_2) \div 2]\} \times 100, \quad (2)$$

where

- C_1 is the concentration of an analyte measured in the environmental sample, in concentration units; and
- C_2 is the concentration of the analyte measured in the replicate sample, in the same concentration units.

Consideration of the magnitude of concentrations being evaluated should be given when evaluating RPDs. For example, the RPD computed from environmental and replicate concentrations of 100 and 101 mg/L is 1.0, the RPD computed from environmental and replicate concentrations of 10 and 11 mg/L is 9.5, and the RPD computed from environmental and replicate concentrations of 1 and 2 mg/L is 67. Although the

difference between environmental and replicate concentrations is the same in each of the examples, the difference evaluated relative to the mean of the measured concentrations increases as the concentrations being evaluated decrease. Constituents measured at small concentrations might yield a large RPD. RPDs were computed only from pairs of analytes that were detected. If one or both analyses in a constituent pair were reported as less than the applicable detection limit, then that constituent pair was not evaluated.

Most mean RPDs computed for the analytes were less than 5, indicating excellent reproducibility. In general, mean RPDs computed for this study indicate that the amount of variability in measurements between environmental and replicate samples is acceptable. Relatively large RPDs (greater than 20 percent) resulted for a number of constituents that occurred at low concentrations: aluminum (two replicates), boron, chromium, copper, fluoride (F), iron, lead, manganese (two replicates), selenium, thallium, zinc (three replicates), radium-226, atrazine, deethylatrazine, and metsulfuron-methyl. Although the RPDs were relatively high, concentrations for these replicate pairs were low. Concentrations of copper, lead, nickel, zinc, and DOC in one replicate-sample pair resulted in large RPDs (greater than 30 percent) for those constituents. Even though concentrations were relatively low, the cause of the large sample variability in these constituents might be indicative of the natural variability of aquifer chemistry at this location or might be a result from sample collection procedures at this site or analytical variability at low concentrations close to the Minimum Reporting Level (MRL). These constituents are not described in this report.

Field spikes are used to assess bias and variability from groundwater matrix interference or degradation of organic constituent concentrations during sample processing, storage, and analysis. Field spikes are environmental samples to which a known volume and concentration of analytes is added. Analytical recoveries of the spiked constituents are expressed as percentages of expected (theoretical) concentrations. Computed field-spike recoveries are compared to theoretical and lab recoveries to evaluate matrix interferences or degradation of organic compounds (pesticides and VOCs).

Recovery (percentage) is computed as

$$\text{Recovery} = [(C_{\text{spiked}} - C_{\text{unspiked}}) \div C_{\text{expected}}] \times 100, \quad (3)$$

where

C_{spiked} is the measured concentration of analyte in the spiked environmental sample, in micrograms per liter; and

C_{unspiked} is the measured concentration of analyte in the unspiked environmental sample, in micrograms per liter.

C_{expected} is the theoretical concentration of analyte in the spiked environmental sample, in micrograms per liter, and is computed as

$$C_{\text{expected}} = C_{\text{solution}} \times V_{\text{spike}} \div V_{\text{sample}}, \quad (4)$$

where

C_{solution} is the concentration of analyte in spiked environmental solution, in micrograms per milliliter;

V_{spike} is the volume of spike added to environmental sample, in milliliters; and

V_{sample} is the volume of environmental sample, in liters.

Organic compound spikes and spike replicates were added to environmental samples and their replicates, resulting in 11 spiked samples for a combination of VOCs or pesticide compounds. Mean percent recoveries were computed for the most frequently detected compounds: atrazine, 98.7 percent; deethylatrazine, 44.0 percent; tetrachloroethene (PCE), 97.9 percent; chloroform, 113 percent. The mean percent recoveries for spikes of atrazine and PCE were similar to laboratory surrogate recoveries; therefore, concentrations measured in environmental samples probably represent actual concentrations. A high (greater than 100 percent) mean percent recovery for chloroform indicates a high bias for this compound. Therefore, actual concentrations of chloroform probably were less than reported concentrations. Results for deethylatrazine were qualified as estimated ("E") by the USGS National Water Quality Laboratory under laboratory reporting conventions used during the study period. A remark code of "E" was used when compounds were identified, but concentrations could not be accurately quantified. Analytical recoveries typically were lower than expected for deethylatrazine; therefore, concentrations were qualified (Sandstrom and others, 2001). Actual concentrations of deethylatrazine probably were higher than reported concentrations.

In addition to field spikes, surrogate compounds are added to environmental samples in the laboratory at the time of analysis. Surrogates are compounds that have a range of expected (observed in the laboratory in 2004–9) recoveries, usually near 100 percent, and are used to verify that analyses are acceptable. Surrogate compounds, which generally do not occur in nature, are similar in physical and chemical properties to analytes of interest. At the laboratory, they are added to samples to provide information on matrix effects and potential gross sample-processing problems. Surrogate compound concentrations were evaluated for analytical schedules containing the most frequently detected organic compounds (atrazine, deethylatrazine, PCE, and chloroform). Organic compound analyses for the most part were within laboratory QC limits, indicating that analytical instrument performance was satisfactory. Three surrogates (1,2-dichloroethane- d_4 ; 1,4-bromofluorobenzene; and toluene- d_8) were added to the VOC analytical schedule, and two surrogates (alpha-HCH- d_6 and diazinon- d_{10}) were added to the selected pesticide schedule. The VOC surrogates were within the acceptable laboratory QC limits for all VOC samples. Recoveries for environmental surrogates of alpha-HCH- d_6 ranged from 53.3 to 129 percent. The highest and lowest environmental surrogate alpha-HCH- d_6 values

were outside the acceptable limit range of 64–121 percent. Eleven environmental surrogate samples for diazinon- d_{10} were below acceptable QC limits, indicating that recoveries for diazinon (and similar compounds) might be suspect. Diazinon was not frequently detected; therefore, it was not evaluated in this report.

For samples with complete ion analyses, the percent difference in the charge balance between cations (positively charged) and anions (negatively charged) is used as a QC check on the quality of the major ion analyses; the percent difference was less than 2 percent in all but three samples. In two of these samples the percent difference was about 4.5 percent, still within acceptable limits of about 5 percent difference. The percent difference for one sample was considered unacceptable at 12 percent. The alkalinity measurement for that sample was low when compared to the dataset and was considered as estimated rather than quantified.

Statistical Methods and Geochemical Modeling

Nonparametric statistical tests were used for all data interpretation. A nonnormal distribution is a common characteristic of water-resource data (Helsel and Hirsch, 2002), and as a result, nonparametric methods are generally better suited than are traditional parametric methods to evaluate water-resource data. The Kendall's tau coefficient is a nonparametric, rank-based test used to measure the strength of the monotonic relation between x and y (linear and nonlinear) and is resistant to the effects of outliers (Helsel and Hirsch, 2002). The tau coefficient ranges from -1 to 1; a value of 0 indicates no correlation, and values approaching -1 or 1 indicate an increasing strength of correlation. The tau coefficient values generally will be lower than values of traditional correlation coefficients for linear associations of the same strength; for example, a strong linear correlation of greater than or equal to (\geq) 0.9 corresponds to a tau value of about ≥ 0.7 (Helsel and Hirsch, 2002). Independent groups of data were compared for differences by using the Mann-Whitney U test. For this report, statistical results with p -values less than 0.05 were considered statistically significant and are reported. Estimated values (less than the laboratory method reporting level) were considered at the estimated concentration.

The geochemical modeling program PHREEQC (Parkhurst, 1995) was used to calculate ion speciation and mineral saturation states. The precipitation or dissolution of mineral phases can be an important control on fluid composition. In a carbonate aquifer, such as the Edwards aquifer, the dissolution or precipitation of minerals such as calcite and dolomite, which compose the aquifer rocks, will add or remove carbonate constituents. Water-rock interaction with gypsum, a trace mineral present in the Edwards Group (Maclay and Small, 1983), might affect groundwater SO_4/Cl ratios and SO_4 concentrations; downdip saline-zone groundwater, for example, shows evidence of incongruent gypsum dissolution (Oetting and others, 1996). A saturation index (SI) of 0 indicates equilibrium conditions, a value greater than 0

indicates saturation with respect to a mineral phase and the potential for precipitation, and a value less than 0 indicates undersaturation with respect to a mineral phase and the potential for dissolution.

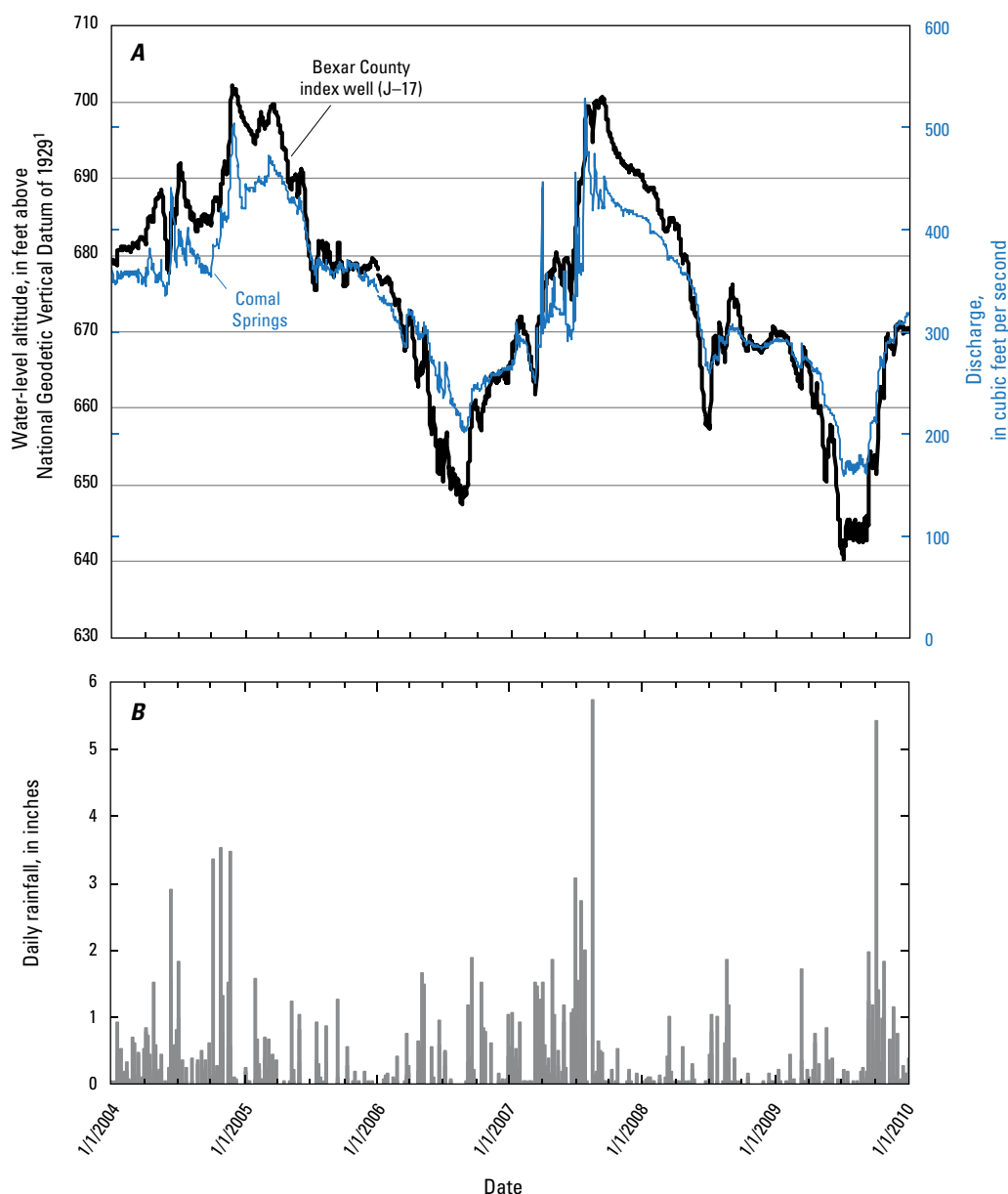
Hydrogeologic Setting

The south-central Texas Edwards aquifer developed in Early Cretaceous-age limestone and dolomite rocks that are extensively faulted, fractured, and karstified. Like many karst aquifers, the Edwards aquifer is characterized by hydrologic variability. Water-level altitudes in the aquifer can rise rapidly in response to rainfall and corresponding recharge and can be accompanied by increases in springflow; conversely, water-level altitudes and springflow decrease during periods of low rainfall and recharge. Records of San Antonio rainfall (recorded at the San Antonio International Airport) (National Oceanic and Atmospheric Administration, 2010), discharge at Comal Springs, and water-level altitudes at well J-17 (Edwards Aquifer Authority, 2009) illustrate the hydrologic variability of the Edwards aquifer (fig. 6).

Aquifer Framework and Lithology

The Cretaceous strata of south-central Texas regionally include three aquifers: the Edwards aquifer in the Balcones fault zone and the Trinity and Edwards-Trinity aquifers in the Edwards Plateau (fig. 2). The aquifer is composed of carbonate rocks of the Georgetown Formation (not shown in fig. 2) and Edwards Group (or their stratigraphic equivalents), which range in thickness from approximately 450 to 550 ft (Burchett and others, 1986) (fig. 2). The Georgetown Formation overlies the Edwards Group and is part of the Edwards aquifer. The less permeable upper member of the Glen Rose Limestone, which is the uppermost part of the Trinity aquifer, underlies the Edwards aquifer and acts as a lower confining unit across much of the region. Where confined, the aquifer is overlain by the Del Rio Clay and other units with relatively low permeability. Depositional stratigraphy of the carbonate facies are discussed in more detail in Lindgren and others (2011b).

The San Antonio segment of the Edwards aquifer is about 180 mi long from west to east, ranges from 5 to 40 mi wide from north to south (Maclay, 1995), and is characterized by relatively high transmissivities (summarized by Lindgren, 2006; Lindgren and others, 2011a, 2011b). The northern limit of the recharge zone (outcrop) defines the northern aquifer boundary. The aquifer is bounded to the south by the downdip limit of potable water in the aquifer, below which groundwater salinity rises rapidly downgradient (Abbott, 1975; Sharp and Banner, 1997). The freshwater/saline-water interface (Schultz, 1994) represents the transition (1,000 mg/L dissolved-solids concentration) from a zone of more active flow with fresh groundwater to a downdip zone of less active flow with saline groundwater. The downdip saline zone is characterized by



¹ Source agency reports water-level altitude referenced to sea level (Edwards Aquifer Authority, 2009)

Figure 6. Time series (2004–9) climatic and hydrologic data for the south-central Texas study area. A, Bexar County index well (J-17) water-level altitude and Comal Springs discharge (Edwards Aquifer Authority, 2009). B, San Antonio daily rainfall (recorded at the San Antonio International Airport) (National Oceanic and Atmospheric Administration, 2010).

numerous complex saline hydrochemical facies (Clement and Sharp, 1988; Oetting and others, 1996) which might locally influence groundwater geochemistry in the freshwater zone (Musgrove and others, 2010). A small number of volcanic and intrusive igneous rocks locally cross-cut the aquifer, which might also affect groundwater geochemistry on a local scale.

Three core samples were collected from the overburden (Austin Chalk/Eagle Ford Group contact, Buda Limestone/Del Rio Clay contact), one sample from the contact of the overburden units with the top of the Edwards aquifer (Del Rio Clay/Georgetown Formation contact); and four samples were

collected from selected HGUs of the Edwards Group (leached and collapsed members, regional dense member/grainstone member contact, grainstone member, Kirschberg evaporite member) (fig. 4; tables 5 and 6 at end of report). For the most part, bulk samples were largely dominated by the carbonate mineral calcite (42.2–96.5 percent, based on relative peak intensity), with lesser amounts of quartz (trace–44.3 percent where present), clays (3.2–26.3 percent), and dolomite (trace to 31.5 percent) (table 6). These minerals were present in varying proportions in each of the core samples. Original deposits in the region of aragonite, calcite, dolomite, and gypsum

have been extensively replaced by calcite within the aquifer and form a highly porous and heterogeneous limestone rock. Dolomite and dolomitic limestone abundance varies spatially as well as stratigraphically (Maclay and Small, 1983, 1984). Mixing-zone dissolution near the downdip freshwater/saline-water interface, driven by gypsum dissolution, has created high porosity and high permeability where dolomite has been dissolved (Maclay and Small, 1984; Hovorka and others, 1998).

One or two digestions (dissolution) were performed on bulk core samples to remove most of the calcite so that mineral phases of less than 1 mm could be identified. Residual clays were present in all samples and, where identified, included kaolinite, illite, smectite, and chlorite. Kaolinite was generally the most abundant clay. Other minerals present in small amounts in some digested samples included goethite (Del Rio Clay/Georgetown Formation contact, possibly Kirschberg evaporite member), pyrite (Buda Limestone/Del Rio Clay contact, regional dense member/grainstone member contact), possibly hornblende (Del Rio Clay/Georgetown Formation contact, regional dense member/grainstone member contact), feldspar (Austin Chalk/Eagle Ford Group contact), and possibly anhydrite or oriented feldspar (Buda Limestone/Del Rio Clay contact).

Total carbon and inorganic carbon were analyzed in core samples (table 5). Solid phase particles (less than 2 mm) contained concentrations of total carbon that ranged from 59 to 129 grams per kilogram (g/kg) (median 113) and inorganic carbon that ranged from 54 to 118 g/kg (median 110). Almost no organic carbon was present in several of the samples.

Estimation of Aquifer Properties

Extensive data have been collected on the hydraulic properties of the Edwards aquifer in the San Antonio region, in particular, permeability, hydraulic conductivity, transmissivity, anisotropy, and storativity. An interpretation of the distribution of relative permeability by HGU of the Edwards aquifer and its confining units is based on field observations, stratigraphic studies by Rose (1972), and data from test-drilling and geophysical-logging programs (Maclay and Small, 1983). These distributions apply to the confined zone and might not be strictly applicable to the unconfined zone. Permeability in the Edwards aquifer—which includes matrix, fracture, and conduit permeability—varies over eight orders of magnitude (Hovorka and others, 1998). The carbonate matrix of the aquifer is very permeable; however, in many intervals, the very high permeabilities resulting from conduits and fractures dwarf the matrix contribution. The highest matrix permeability occurs in downdip parts of the aquifer, where the mixing of freshwater and saline water has maximized dolomite dissolution (Hovorka and others, 1998).

Transmissivity and hydraulic conductivity of the Edwards aquifer each vary over several orders of magnitude (detailed in Lindgren and others, 2011b). Local-scale variability in specific capacity, transmissivity, and hydraulic conductivity is likely

quite large; even closely spaced measurements might differ by a factor of 1,000 (Hovorka and others, 1998). The difference likely reflects the variable contributions of matrix, fracture, and conduit permeability to the measured average values obtained in aquifer tests. Vertical variations in permeability in the Edwards aquifer indicate that the entire aquifer is highly permeable and highly variable. Anisotropy of an aquifer is indicated when the permeability shows variations with the direction of measurement at any given point in a geologic formation. An anisotropic aquifer will have a dominant permeability in one or more directions depending upon geologic and hydrologic conditions. The quantitative magnitude of anisotropy of the Edwards aquifer is largely unknown but might be influenced by the presence of barrier faults with large vertical displacements and the development of conduits. The amount and distribution of water in the Edwards aquifer are related to the development of porosity and the storage characteristics of the aquifer. Hovorka and others (1996) estimated that Edwards aquifer porosity varies vertically from lows of 4–12 percent to highs of 20–42 percent, with an average for the entire aquifer of 18 percent. Kuniansky and others (2001) found that an effective porosity of 1–3 percent was needed to match estimated traveltimes derived from geochemical mixing models.

Flowmeter and geophysical data provide insight into hydraulic properties of the Edwards aquifer, specifically with respect to head differences and vertical gradients within the aquifer, as well as with respect to well-field dynamics. An understanding of the effect of pumping conditions on flow in a PSW might contribute to knowledge of how contaminants migrate to the PSW and how contaminant concentrations might vary in time and space. Vertical flow was measured at several discrete depths throughout the portion of the W4 borehole that is open to the Edwards aquifer. Paillet (2000) showed that two quasi-steady flow profiles measured in a borehole can be combined with a measurement of the water-level difference between pumping and nonpumping conditions (drawdown) to calculate estimates of zone transmissivity and far-field hydraulic head conditions. The flowmeter measurements in W4 were calculated by using the drawdown during pumping, the measured flow through the flowmeter tool, and a flow bypass factor by using the diameter of the tool and the borehole diameter at the depth of the measurement. The corrected measurements were entered into an analytical model which utilizes the Thiem equation (Thiem, 1906) and plotted with lines of model fit representing the four dynamic conditions of flow (ambient 1, ambient 2, moderate, normal). With these lines of model-fitted flow and observed and calculated drawdown, transmissivity was calculated to approximate aquifer conditions at four selected zones in the well. The estimated transmissivity values with increasing depths were

Zone 1: 340–415 ft below LSD, estimated to be 33,901 square feet per day (ft^2/d),

Zone 2: 415–600 ft below LSD, estimated to be 2,396 ft^2/d ,

Zone 3: 600–675 ft below LSD, estimated to be 17,693 ft^2/d , and

Zone 4: 675–755 ft below LSD, estimated to be 8,339 ft^2/d .

The total transmissivity for the well, estimated as the sum of these four zones, was 62,329 ft²/d. This estimate is consistent with transmissivities estimated by other methods (Lindgren and others, 2004; Lindgren, 2006) for this general area.

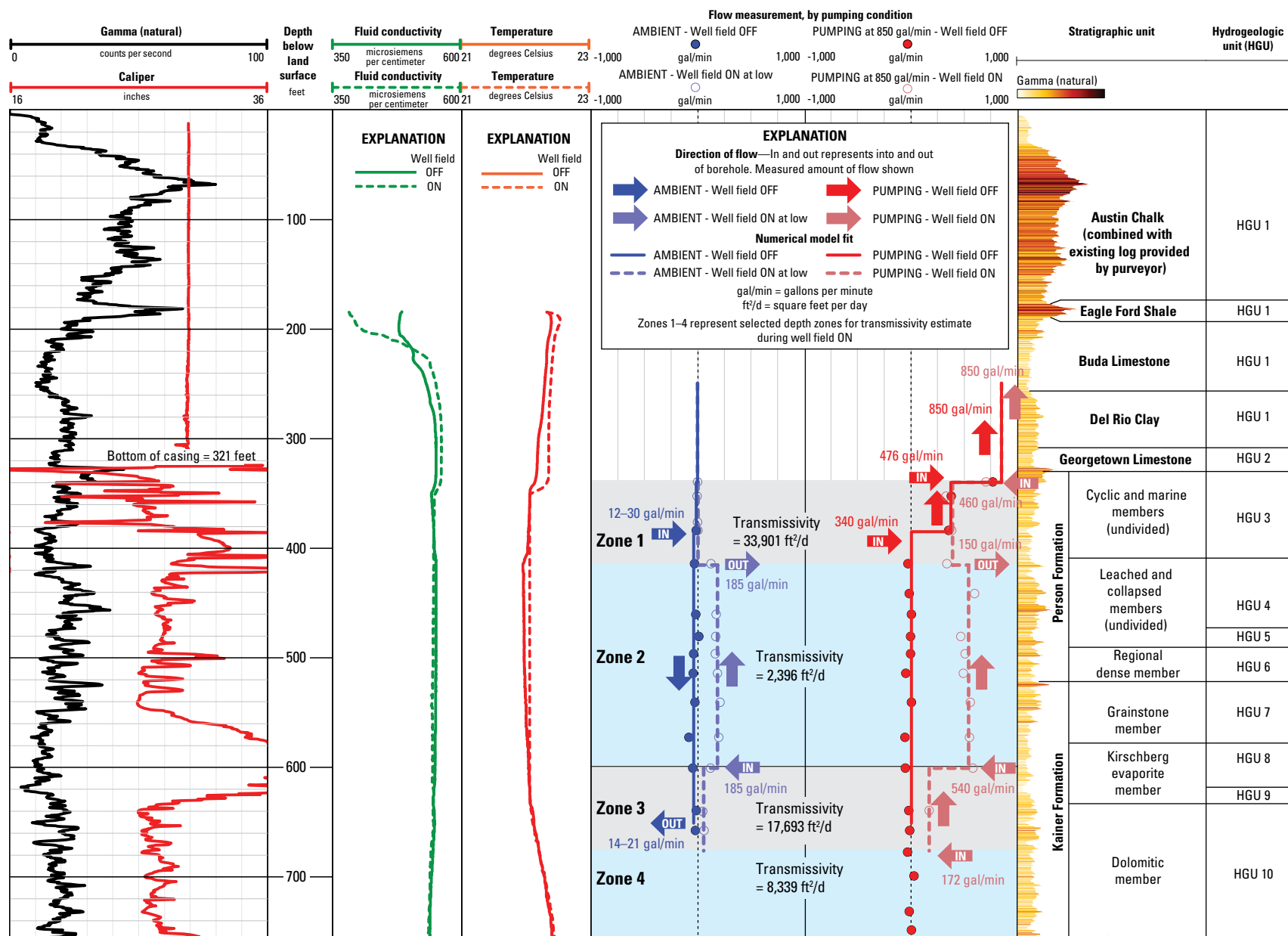
The movement of water in the PSW borehole was variable under different pumping conditions. Flowmeter data and resulting estimated hydraulic properties for W4 are shown in figure 7. The large casing diameter (30 in.) and variable borehole diameter (15–36 in.) are shown on the caliper log and represent the enlarged borehole after acidification. The large diameter has an effect on other logs such as the natural gamma and flowmeter logs. The natural gamma counts per second were somewhat lower than expected in the strata penetrated because of the large spherical zone of detection of the natural gamma tool. Flowmeter values also were affected by the variable borehole diameter; therefore, a correction factor was generated for flowmeter values on the basis of the borehole diameter. The correction factor is a calculated flow bypass, which calculates the surface area of the borehole by using caliper data and the surface area of the skirt/tool; the ratio of these is the percentage of flow that the tool measures. Because the borehole diameter varied throughout the borehole, an individual correction factor was calculated for each measurement/depth. Fluid conductivity and temperature probes are generally used only to measure the fluid flowing through the tool and, thus, do not represent any horizontal variation of the fluid qualities across a large diameter borehole.

Four lines of model fit of the flowmeter data, representing the four dynamic conditions of flow (ambient 1, ambient 2, moderate, normal) are also shown (fig. 7). Arrows indicate the direction of flow under each of the pumping conditions toward the borehole (toward the model line) or away from the borehole (away from the model line) and up or down the borehole. During condition 1 (“ambient 1”), there were small amounts of mostly downward vertical flow in the wellbore (fig. 7). Water entered the borehole from the cyclic and marine members (HGU 3) (between zones 1 and 2), flowed downward, and exited the borehole in the dolomitic member (HGU 10). During condition 2 (“ambient 2”) flow in the wellbore was mostly upward. Water was pulled into the borehole from the Kirschberg evaporite member (HGU 8) (zone 3). Flow was upward and exited the borehole in a highly transmissive zone (zone 2) at the top of the leached and collapsed members (HGU 4). During condition 3 (“moderate”), nearly all water entered the borehole from the cyclic and marine members (HGU 3) in the upper part of the well. That flow was subdivided into two zones, coming from zone 1 and above zone 2. During condition 4 (“normal”), flow in the wellbore was complex, with water both entering and exiting. The borehole gained water from the dolomitic member (HGU 10) (zone 4) and Kirschberg evaporite member (HGU 8) (zone 3) in the lower part of the borehole. Some water exited the borehole at the top of the leached and collapsed members (HGU 4) (zone 2). The borehole gained a large amount of water again at the top of the cyclic and marine members (HGU 3) (zone 1).

Fluid conductivity and temperature were logged in the vertical extent of the W4 borehole in ambient conditions, both with the well field off (“ambient 1”) and with the well field on (“ambient 2”); differences in fluid conductivity and temperature were minor (less than 1 percent) and close to analytical uncertainties. These results are consistent with other geochemical data that indicate that water from W4 and other wells in the well field is relatively well mixed and homogeneous under a range of pumping conditions. The two deepest monitoring wells in each of the monitoring well clusters (Z–DED and T–DED) were logged in open hole under ambient conditions prior to completion; results are summarized in figures 8 and 9, respectively. Although logs were collected during ambient conditions, some downward flow (8.6 gal/min) was detected in the Z–DED borehole. This flow might have been caused by nearby pumping stress from the PSW well field (approximately 0.5 mi to the west). As discussed above, most of the flow in the PSW entered the wellbore from the leached and collapsed members. Ambient flow in Z–DED also entered the borehole in the leached and collapsed members and exited into the Kirschberg evaporite member. If the two wells were hydraulically connected, then a linkage of outflow to the Kirschberg evaporite member at the Z–DED borehole and inflow to W4 is plausible. Throughout the zone of highest flow in Z–DED, fluid conductance consistently measured about 300 microsiemens per centimeter ($\mu\text{S}/\text{cm}$), and temperature was about 21.0°C. These results are considerably different from measured values for the monitoring wells (table 7 at end of report) and might reflect differences in measurement tools and calibrations. In the deep Timberhill monitoring well (T–DED), approximately 1.5 mi west of the PSW well field, there was little or no measureable flow throughout the entire section of the well open to the Edwards aquifer. The fluid conductance consistently measured about 500 $\mu\text{S}/\text{cm}$, and temperature was about 21.4°C. This well appeared to have little or no connection to outside pumping stresses, including those from the PSW well field.

Groundwater Occurrence and Movement

The groundwater flow system of the Edwards aquifer in the San Antonio region includes the Edwards Plateau catchment or contributing zone, the recharge (outcrop) zone, and the confined zone (fig. 1). Streams flowing south and east toward the Gulf of Mexico drain the Edwards Plateau contributing zone and recharge the aquifer by streamflow losses across the Balcones fault zone, where the porous and permeable limestone of the Edwards Group outcrops. The recharge zone consists of approximately 1,250 mi² of Edwards Group limestone exposed at the surface (Hamilton and others, 2008) (fig. 1). Most stream base flow and much of the stormflow recharges the aquifer through open solution channels (Maclay and Land, 1988). The majority of recharge to the regional aquifer occurs as losing streams flow across the recharge zone, with estimates of recharge from streams ranging from 60 to 80 percent (Klemm and others, 1979; Maclay and Land, 1988;



All depths are reported in feet below land surface

Figure 7. Borehole geophysical data for the public-supply well (W4) near San Antonio, south-central Texas, including, from left to right, natural gamma and caliper logs, fluid conductivity, temperature, flowmeter (under ambient and pumping conditions), stratigraphic unit, and hydrogeologic unit (HGU).

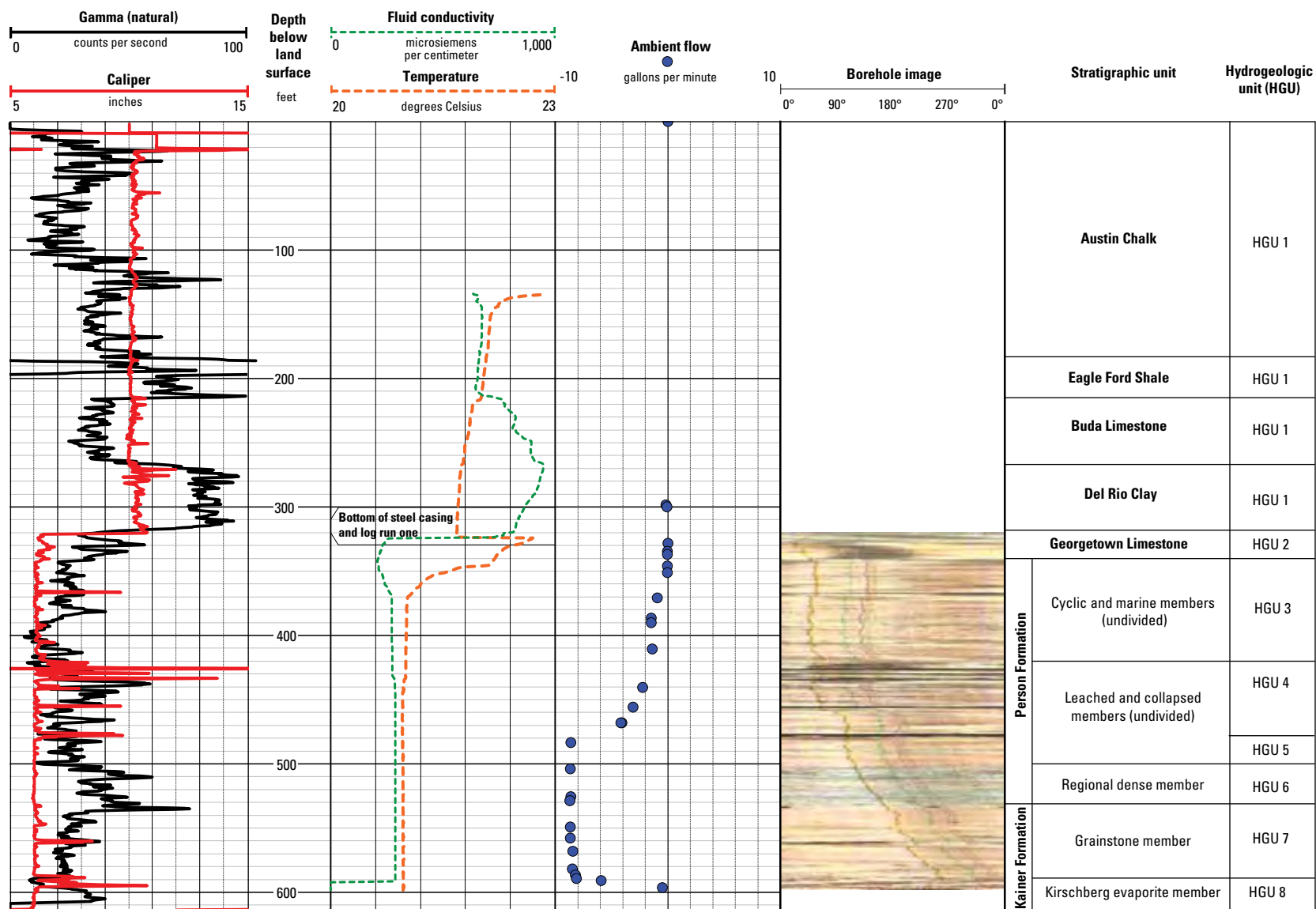


Figure 8. Borehole geophysical data for the deep Zarzamora monitoring well (Z-DED) near San Antonio, south-central Texas, including, from left to right, natural gamma and caliper logs, fluid conductivity and temperature, flowmeter, borehole image, stratigraphic unit, and hydrogeologic unit (HGU).

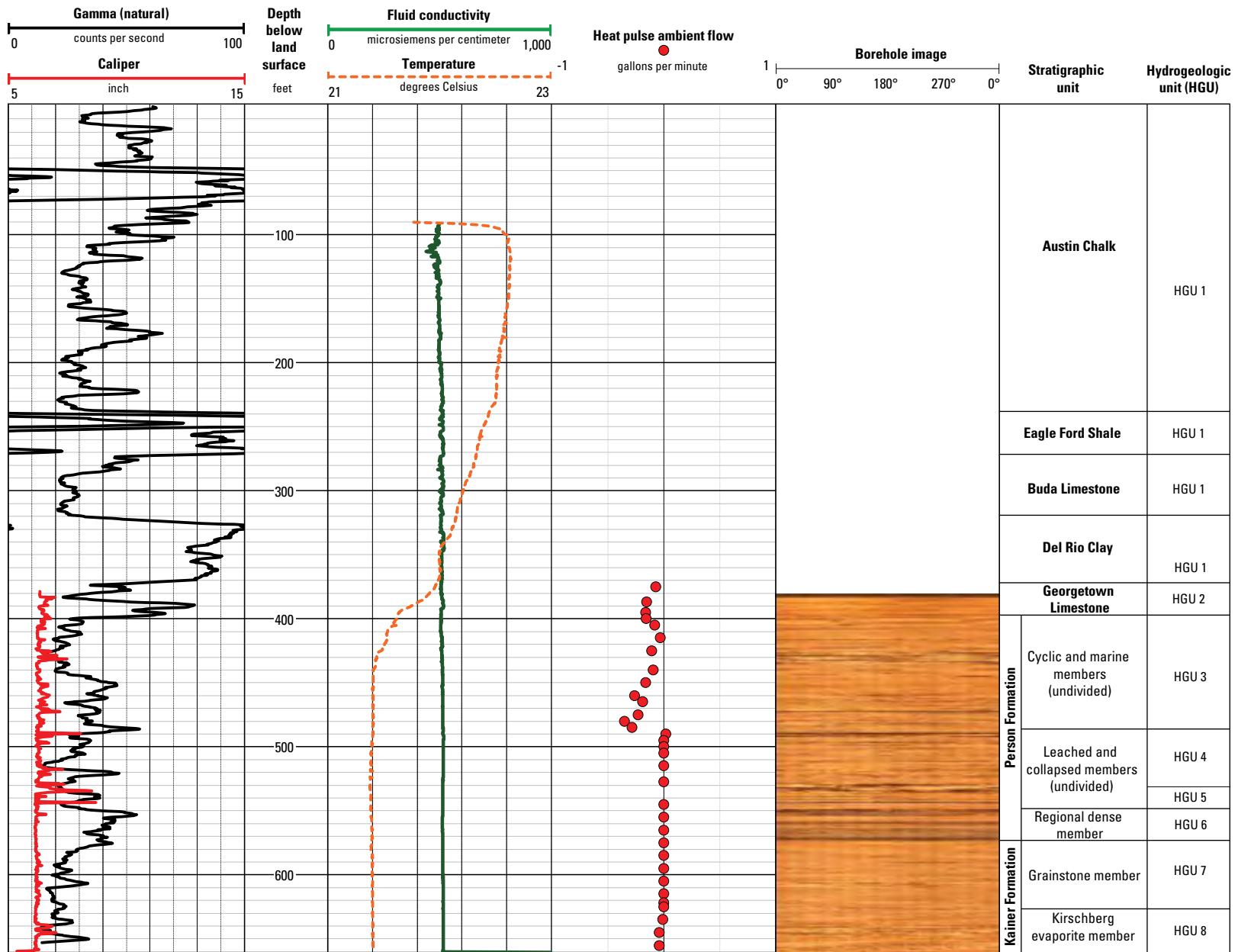


Figure 9. Borehole geophysical data for the deep Timberhill monitoring well (T-DED) near San Antonio, south-central Texas, including, from left to right, natural gamma and caliper logs, fluid conductivity and temperature, flowmeter (no appreciable measured flow), borehole image, stratigraphic unit, and hydrogeologic unit (HGU).

Thorkildsen and McElhaney, 1992; Ockerman, 2005). Most of the remaining recharge occurs as direct infiltration on the recharge zone, with some recharge occurring by leakage from the underlying Trinity aquifer (Sharp and Banner, 1997; Mace and Hovorka, 2000; Lindgren and others, 2004). Estimates of the combined recharge to the San Antonio segment of the Edwards aquifer from stream seepage and infiltration of rainfall range from a low of 53.9 million cubic meters (m^3) during 1956 to a high of 3,066 million m^3 during 1992 (Hamilton and others, 2003). The long-term (1934–2002) mean annual recharge to the Edwards aquifer is 862.2 million m^3 (median 688.1 million m^3) and for 1993–2002 is 979.6 million m^3 (median 710.9 million m^3) (Hamilton and others, 2003). In the local-scale study area, major streams crossing the recharge zone are the Medina River and Hondo and Salado Creeks (fig. 3). Much of the annual flow of the Medina River is impounded in Medina Lake (U.S. Army Corps of Engineers, 1965). Of the volume impounded, about one-half seeps into the Edwards aquifer from the lake and its irrigation facilities (U.S. Army Corps of Engineers, 1965).

Karst features and faults provide pathways for water to move from stream channels to the water table of the unconfined aquifer in the recharge zone. Water that recharges in the recharge zone moves from the unconfined to confined parts of the aquifer through generally southeasterly flow paths. Regional groundwater flow in the confined zone is to the east and northeast with natural discharge occurring at large springs, predominantly Comal and San Marcos Springs (fig. 1). Conduits are likely major contributors to flow (Hovorka and others, 1998, 2004; Worthington, 2004). Faults can either increase or decrease total transmissivity (Hovorka and others, 1998) and might serve as either conduits or barriers to flow (George, 1952; Holt, 1959; Maclay and Small, 1983, 1984; Maclay and Land, 1988; Maclay, 1995; Groschen, 1996; Clark and Journey, 2006). The potentiometric surface varies considerably depending on hydrologic conditions (fig. 6). Domestic, livestock, municipal, agricultural, and industrial withdrawals, by way of wells, also contribute to groundwater discharge throughout the region (Hamilton and others, 2008). Recharge and discharge are not evenly distributed across the aquifer; most recharge occurs in the westernmost catchments, whereas most discharge occurs in the eastern area. The proportion of discharge by wells relative to springs varies from year to year; springflow typically exceeds well discharge in wet years, and well discharge typically exceeds springflow in dry years (Hamilton and others, 2008). Comal Springs (fig. 1), one of the major aquifer discharge points, is hydrologically isolated from large sources of local recharge (Ogden and others, 1986). As such, Comal Springs represents the integration of the regional flow system of the San Antonio segment of the aquifer, and its flow rate and geochemistry provide an indicator of regional hydrologic conditions. Several wells in the aquifer, such as the Bexar County index well (J-17), intersect major flow paths and respond quickly to changes in recharge and pumping; water-level altitude variations in well J-17 are assumed to reflect Edwards aquifer water-level changes on a

regional scale and are also indicative of regional hydrologic conditions. Previous studies indicate that groundwater flow is focused in highly permeable units and is affected by faulting throughout the aquifer (Abbott, 1975; Woodruff and Abbott, 1979; Maclay and Small, 1986; Maclay and Land, 1988; Sharp, 1990).

Potential Sources of Contamination

Natural contaminants such as As, U, or radon generally are not problematic with respect to public drinking-water quality for the Edwards aquifer. Although these contaminants are commonly detected, they rarely occur at concentrations of concern for drinking-water quality. Nitrate can result from natural and anthropogenic sources, and concentrations might be increasing with the increase in human activities in the region (Musgrove and others, 2010); however, concentrations are not high with respect to drinking-water quality. A small number of organic contaminants are routinely detected in the Edwards aquifer at low concentrations (Musgrove and others, 2010).

Land use often is considered when evaluating potential sources of contaminants to groundwater (Brawley and others, 2000; Hiscock and others, 2007; McMahon and others, 2008). Land use in the recharge zone for the PSW (Lindgren and others, 2011b) is 91.6 percent rangeland and forest, 7.6 percent urban/built-up, and 0.8 percent water, wetland, barren, and agricultural (fig. 10). Agricultural land use in the recharge zone for the PSW is insignificant at 0.04 percent. In the local-scale study area, urban/built-up land use is the most likely nonpoint source of potential contaminants.

A database containing locations of points of potential sources of contaminants (PSOCs) was previously compiled by the Texas Commission on Environmental Quality (TCEQ) (written commun., 2006) and examined for the local-scale study area; more than 10,000 PSOCs were identified (fig. 11). Potential sources of contaminants include, but are not limited to, locations that manufacture or store chemicals, cemeteries, retail centers that store and sell chemicals, oilfield-related activities, water and wastewater treatment facilities, landfills, and pipelines. Most of the PSOCs were in the San Antonio metropolitan area, with an additional large number of oilfield-related PSOCs in the southern part of the study area overlying the down dip limit of the freshwater zone of the aquifer.

The TCEQ PSOC database (written commun., 2006) was examined specifically for the most frequently detected anthropogenic compounds: atrazine, chloroform, and PCE (figs. 12, 13, and 14, respectively), results of which indicate that the urban environment dominates potential sources for these compounds. There were 688 potential point sources of atrazine identified, and most were in the urban San Antonio area. Most of the potential locations were sites of manufacture, sales, or application of pesticides; municipal solid waste facilities; and industrial wastewater outfalls, among others. There were 831 potential point sources of chloroform identified (fig. 13), and most sources were in the San Antonio urban area. Sources of chloroform included municipal solid waste facilities; metal

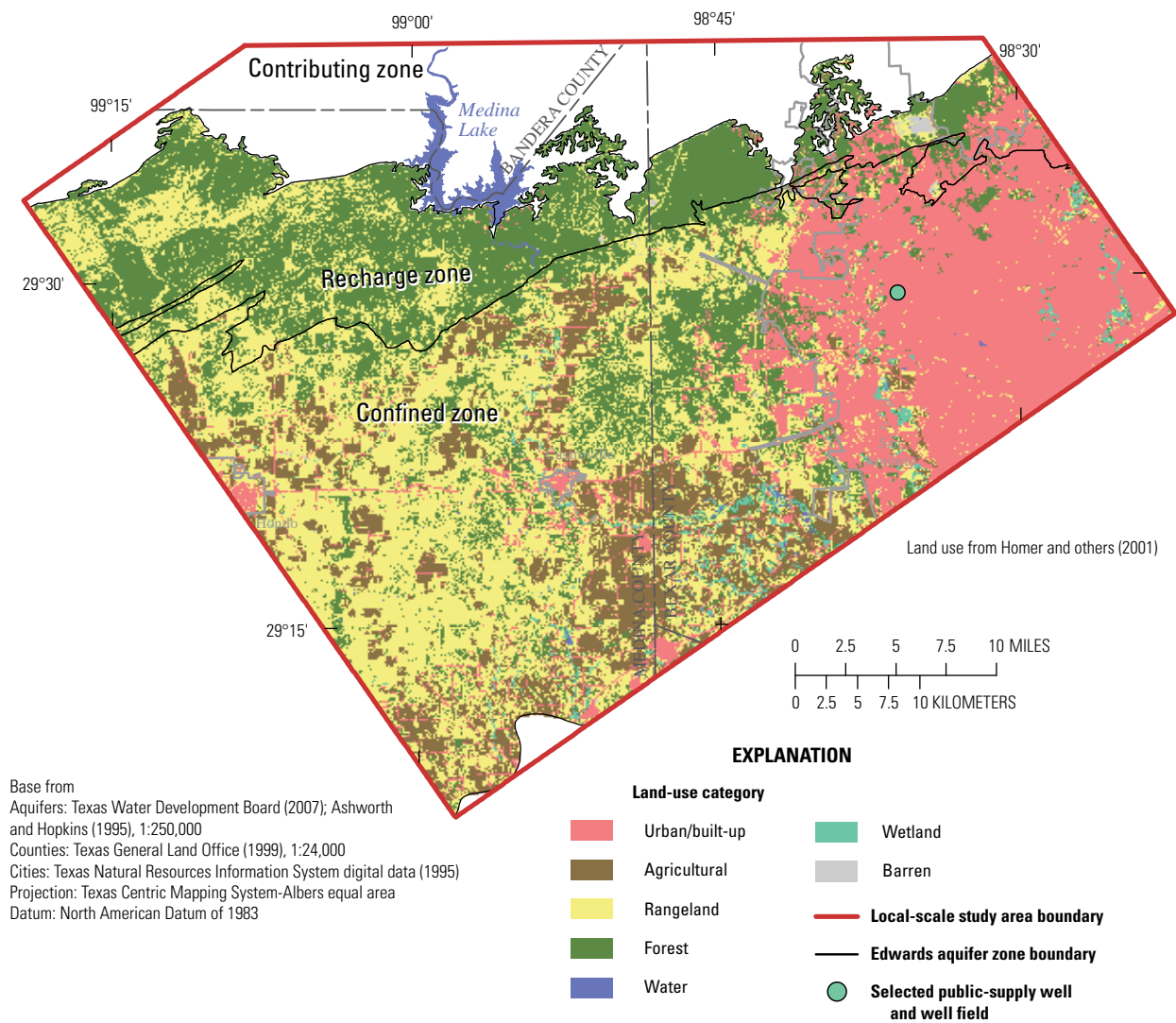


Figure 10. Land use in the local-scale study area for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, and location of selected public-supply well and well field, near San Antonio, south-central Texas.

plating businesses; industrial wastewater outfalls; pesticide manufacturing, sales, or application locations; pipelines; plastic manufacturing; and wastewater treatment, among others. There were 1,991 potential sources of PCE identified (fig. 14). Like the other compounds, most potential sources of PCE were in the urban area, and a few were in the San Antonio area recharge zone. Potential sources of PCE included air transportation facilities, auto parts and repair businesses, dry cleaners, industrial wastewater outfalls, municipal solid waste facilities, new and used oil sites, paint shops, sites of manufacture or sales of pesticides, and sites of plastic manufacturing, among others. The TCEQ PSOC database provides insight into potential sources of contaminants but is not necessarily comprehensive of all potential sources; additional contaminant sources, for example, include application of pesticides on urban lawns, return flows from treated wastewater applied on the landscape, septic systems, and swimming pools.

An additional potential source of PCE, a U.S. Environmental Protection Agency (USEPA) Superfund site about 2.6 mi northwest of the PWS well field with PCE- and TCE-contaminated groundwater in overburden units, was identified during the time period of the TANC project (U.S. Environmental Protection Agency, 2011; hereinafter, Bandera Road Groundwater Plume). Some groundwater wells, three of which were completed at the top of the Edwards aquifer, contained PCE at concentrations above the USEPA public drinking-water standard (U.S. Environmental Protection Agency, 2011) (5 µg/L). Model particle tracks indicate that groundwater flow paths in this area are from the contamination site to the south-southeast toward the well field; however, water-level and dye-tracer data in the vicinity of the Bandera Road Groundwater Plume indicate that water might be moving more westward in the area immediately upgradient of the well field and that barrier faults in the vicinity may inhibit flow (C. Villarreal,

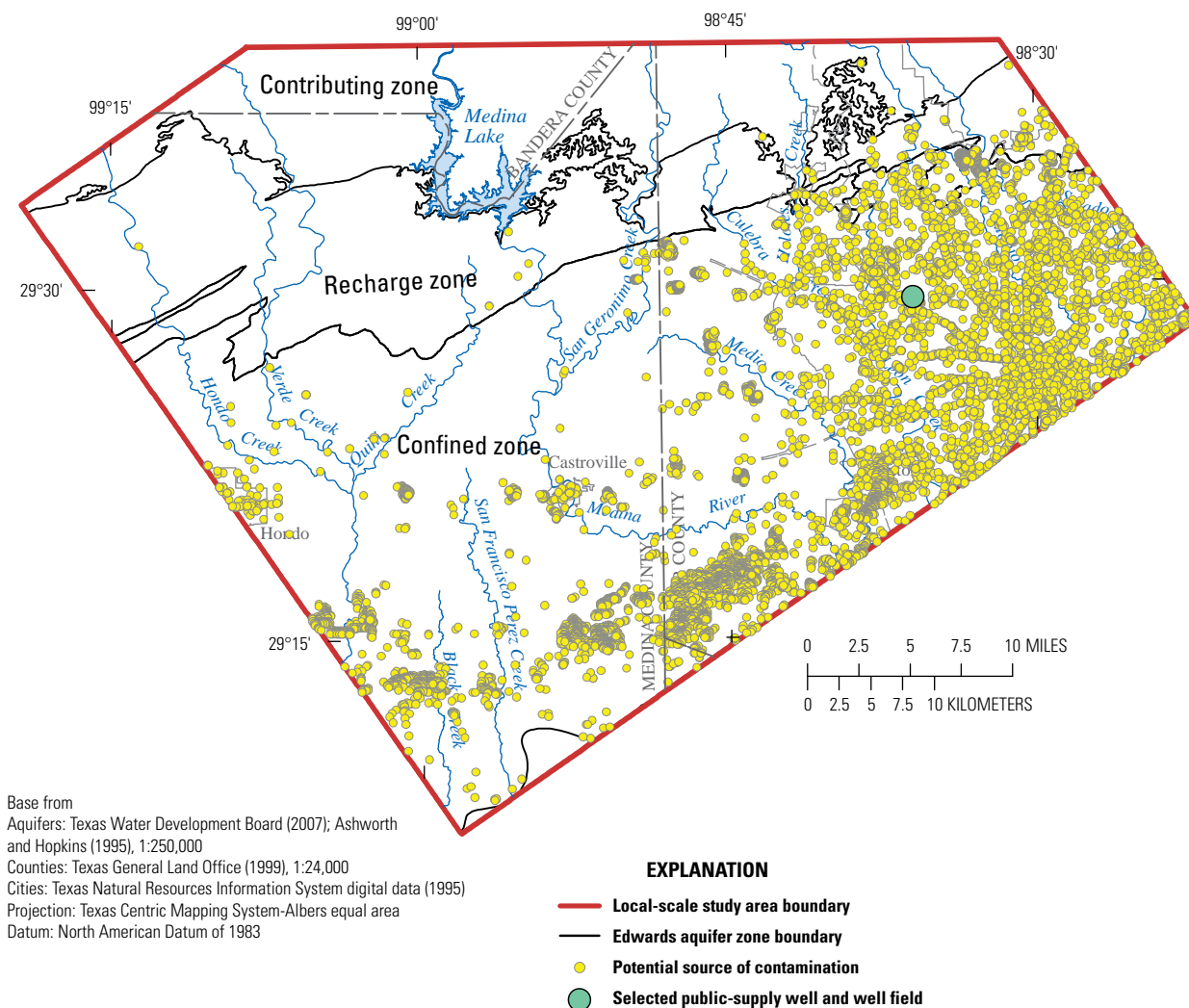


Figure 11. Location of potential sources of contamination (PSOCs) in the local-scale study area for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, and location of selected public-supply well and well field, near San Antonio, south-central Texas; based on Texas Commission of Environmental Quality potential sources of contamination (PSOC) database (written commun., 2006).

U.S. Environmental Protection Agency, written commun.). The Bandera Road Groundwater Plume is currently (2011) the only identified USEPA superfund site with PCE contamination that is upgradient of the well field. No hydrologic connection between the Bandera Road Groundwater Plume and the well field has been documented, however, and other sources are possible.

Groundwater Chemistry

This section describes the chemical characteristics of groundwater samples with respect to water types, field parameters, SIs for selected minerals, reduction-oxidation (redox) conditions, dissolved gases, major and trace element

and nutrient concentrations, isotopic compositions, organic constituents, and age tracers. Special emphasis is placed on understanding groundwater chemistry in the context of water-rock interaction processes, groundwater flow patterns and flow paths, residence time, and mixing, and temporal variations in groundwater chemistry.

Field Parameters

Field parameter (physicochemical) data indicated that well-field and monitoring well samples are representative of groundwater in the regional confined aquifer. The well-field and monitoring well samples had relatively narrow ranges in field parameters (table 7) with median values for specific conductance, pH, temperature, dissolved oxygen concentration,

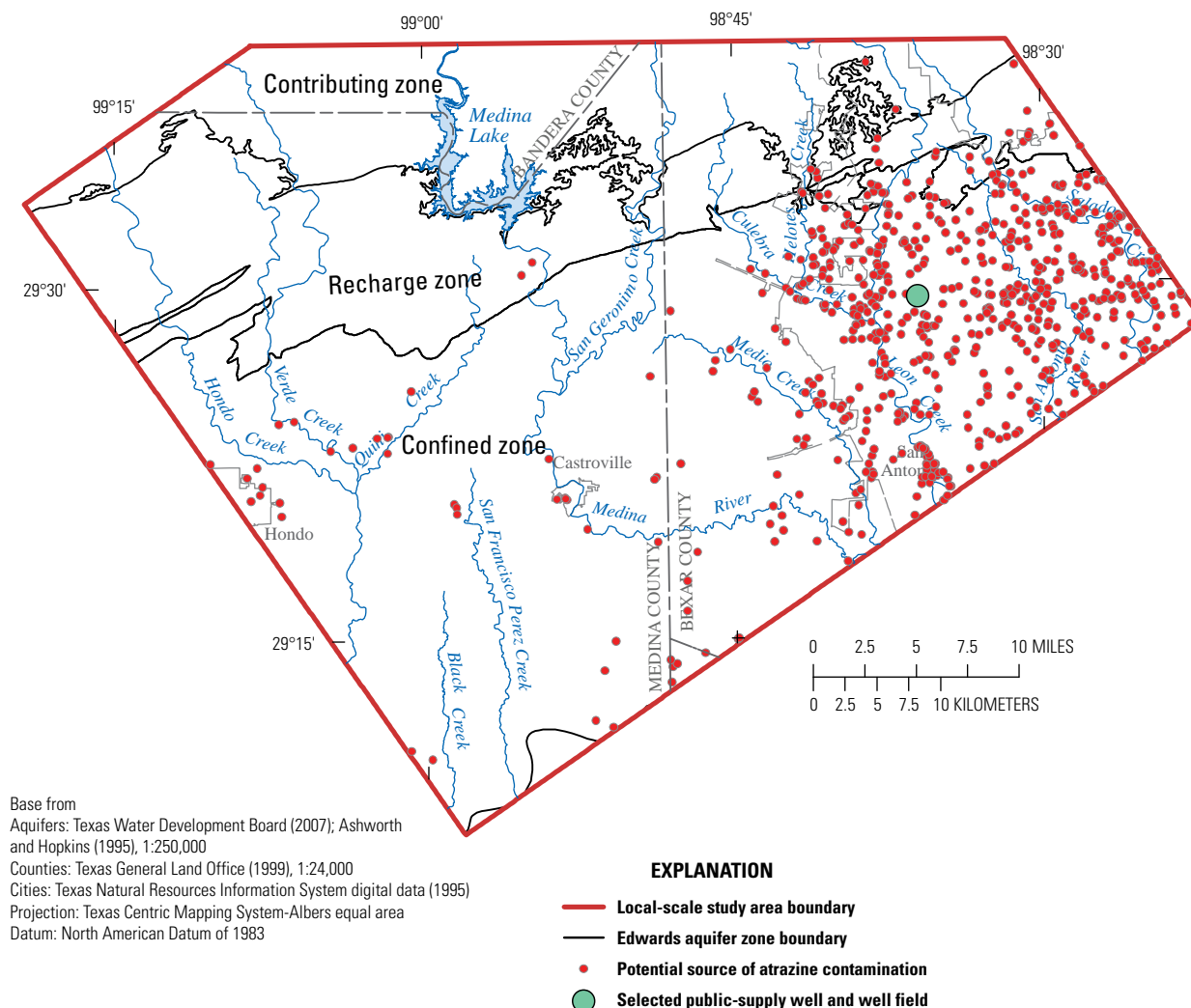


Figure 12. Location of potential sources of atrazine contamination in the local-scale study area for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, and location of selected public-supply well and well field, near San Antonio, south-central Texas; based on Texas Commission of Environmental Quality potential sources of contamination (PSOC) database (written commun., 2006).

and alkalinity (CaCO_3) of 586 $\mu\text{S}/\text{cm}$, 7.1, 24.0°C, 4.3 mg/L, and 258 mg/L, respectively. The sample collected from the overburden well (Z-OVB) was chemically distinct from the well-field and monitoring well samples, with higher specific conductance (971 $\mu\text{S}/\text{cm}$) and temperature (28.3°C) values. These results indicate that the overburden unit is likely not a dominant source of water to the PSW or the aquifer in the study area. Measured values for field parameters for the well-field and monitoring well samples were otherwise similar to those measured for regional aquifer PSWs (table 7), as well as those described by Musgrove and others (2010) for the regional aquifer.

Results for several field parameters for samples from regional (confined) aquifer PSWs were correlated with well depth; both pH and temperature increase with well depth

(Kendall's tau = 0.26 and 0.39, respectively), whereas specific conductance and alkalinity decrease with well depth (Kendall's tau = -0.35 and -0.33, respectively) (fig. 15). The relation between dissolved oxygen concentration and well depth was not statistically significant. These relations between field parameters and well depth were not evident for the monitoring wells, although the monitoring wells cover a more limited depth range than do the regional aquifer PSWs. The observed relation of decreasing specific conductance with depth for samples from the regional (confined) aquifer PSWs is somewhat counterintuitive. In the absence of other sources of salinity, specific conductance values would be expected to generally increase along aquifer flow paths as a result of progressive mineral-solution reactions. Additionally, deeper groundwater in the Edwards aquifer can be affected by mixing with more

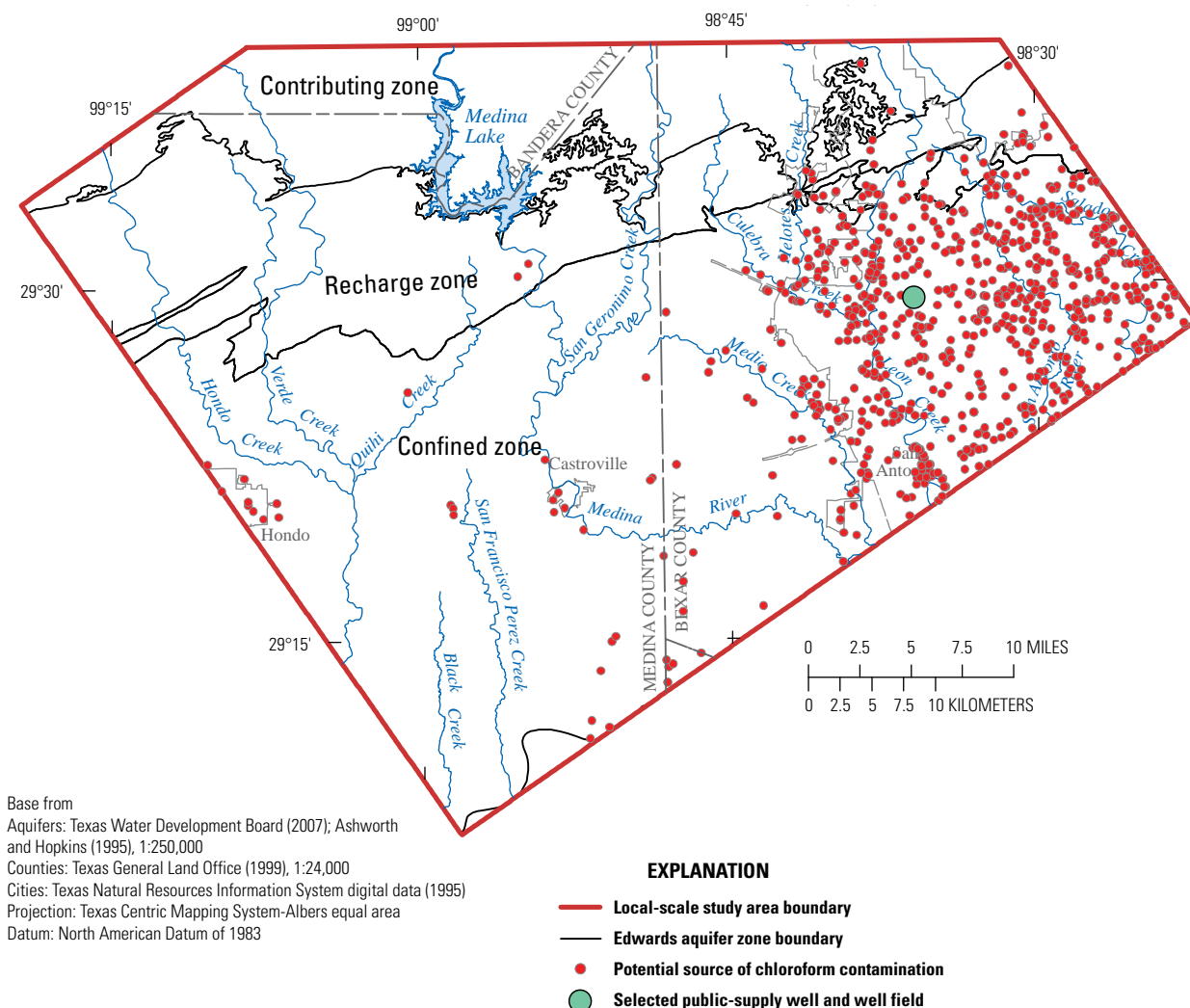


Figure 13. Location of potential sources of chloroform contamination in the local-scale study area for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, and location of selected public-supply well and well field, near San Antonio, south-central Texas; based on Texas Commission of Environmental Quality potential sources of contamination (PSOC) database (written commun., 2006).

saline downdip water. However, at the regional aquifer scale, specific conductance values in shallow/urban unconfined Edwards aquifer groundwater samples are higher relative to deeper unconfined and confined groundwater samples, which indicates that urban land use affects water quality and might be a potential source of dissolved constituents to groundwater (Musgrove and others, 2010).

Chemical Composition and Mineral Saturation Indexes

Groundwater samples from the well-field and monitoring wells and regional aquifer PSWs exhibited a relatively narrow range in major ion compositions. A trilinear (“Piper”) diagram

(Piper, 1953) is a useful tool for evaluating the relative abundance of major cations and anions and classifying major water types. Groundwater samples are mostly a dilute calcium bicarbonate facies, with similar compositions in all sample types and locations (well-field wells, monitoring wells, and regional aquifer PSWs) except the overburden monitoring well (Z-OVB) (fig. 16). Regional aquifer PSW samples show the most variability, with cation concentrations ranging between about 20–40 percent Mg and between about 60–80 percent Ca (fig. 16). With respect to anions, groundwater composition is less variable. Cl concentration exceeds 20 percent in only the overburden monitoring well.

Saturation indices (SIs) with respect to calcite, dolomite, and gypsum reflect the carbonate mineralogy of the aquifer host rocks (fig. 17; table 8 at end of report). All groundwater

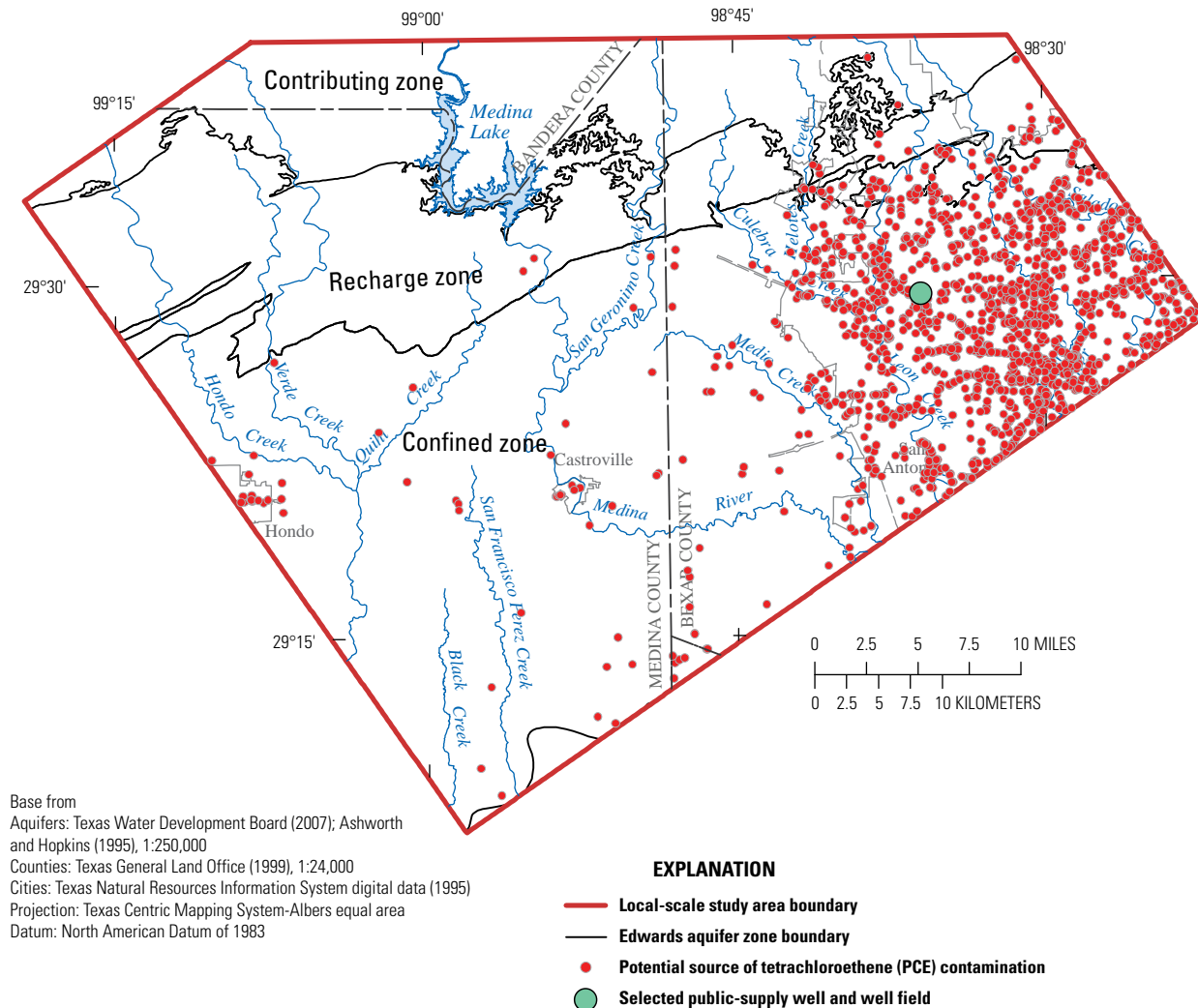


Figure 14. Location of potential sources of tetrachloroethene (PCE) contamination in the local-scale study area for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, and location of selected public-supply well and well field, near San Antonio, south-central Texas; based on Texas Commission of Environmental Quality potential sources of contamination (PSOC) database (written commun., 2006).

samples were near saturation with respect to calcite (-0.5 to 0.5) and very undersaturated with respect to gypsum (less than -1.7). Dolomite SIs range from undersaturated to near saturation. Dolomite SIs for samples from the well-field and monitoring wells were mostly near saturation to undersaturated (-0.48 to -0.11, median = -0.28), whereas results for the regional aquifer PSW samples have a larger range (-1.36 to 0.26, median = -0.34) and are correlated with well depth (fig. 17; Kendall's tau = 0.25). Musgrove and others (2010) reported that water from the confined part of the aquifer is generally closer to equilibrium than is water from the unconfined part of the aquifer with respect to calcite and dolomite, reflecting greater extents of water-rock interaction. Increasing dolomite SIs in the regional Edwards aquifer have been previously proposed to be an indicator of increasing extents

of progressive water-rock interaction processes and relative groundwater residence time (Musgrove and others, 2010). This hypothesis is further supported by the correlation between dolomite SIs and Mg/Ca molar ratios for regional groundwater as discussed in Musgrove and others (2010) and the correlation shown for regional aquifer PSWs (a subset of the regional groundwater wells detailed in Musgrove and others, 2010) in fig. 18 (Kendall's tau = 0.35). This relation is not evident in the well-field and monitoring wells (fig. 18), which cover a more limited depth range than do the regional aquifer PSWs.

Major Elements and Trace Metals

Major element concentrations and groundwater facies (fig. 16) associated with samples collected from regional

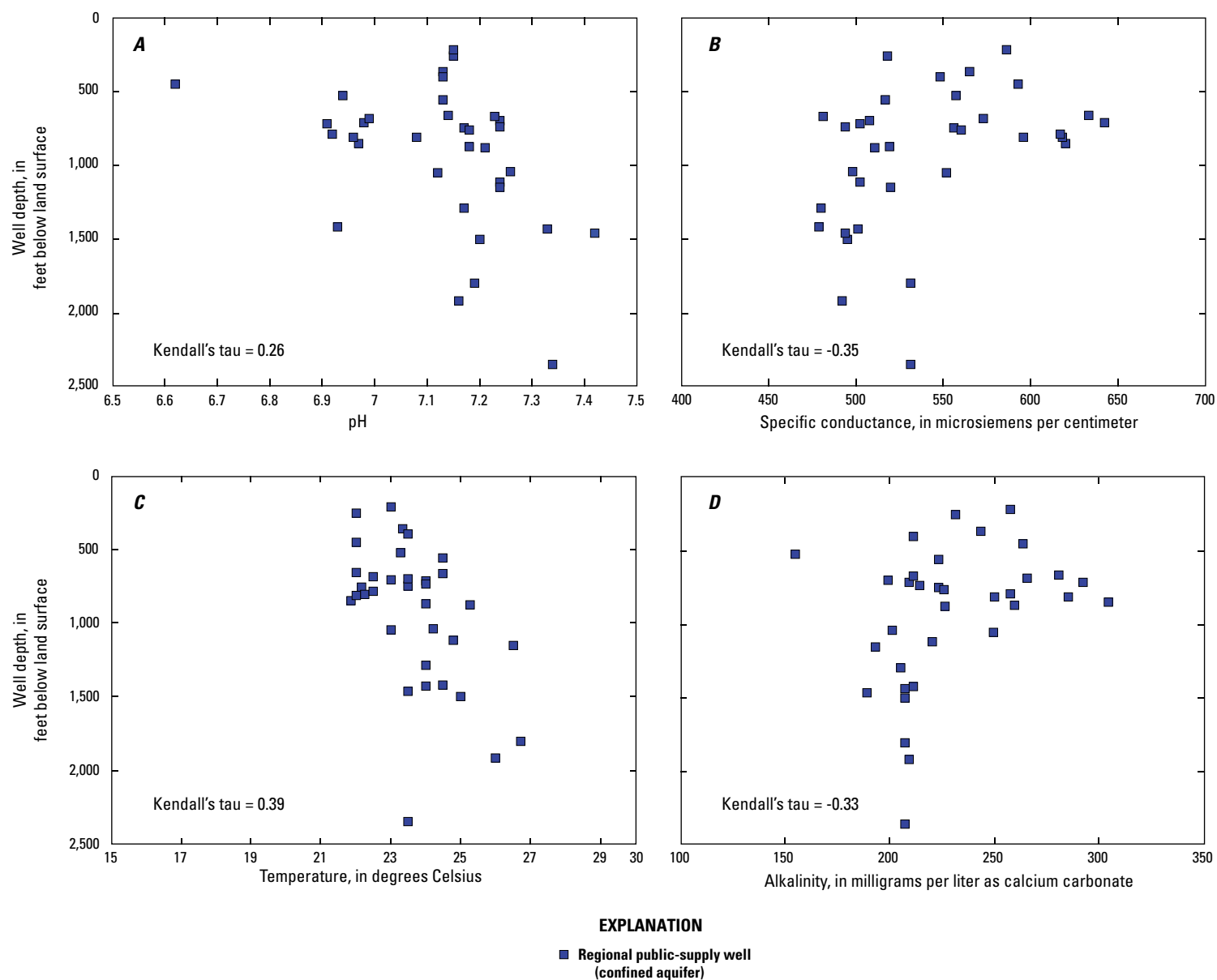


Figure 15. Relation between well depth and selected physicochemical and geochemical constituents for groundwater samples collected from regional aquifer public-supply wells in the confined part of the San Antonio segment of the Edwards aquifer, south-central Texas, 2004–5. *A*, pH. *B*, Specific conductance. *C*, Temperature. *D*, Alkalinity.

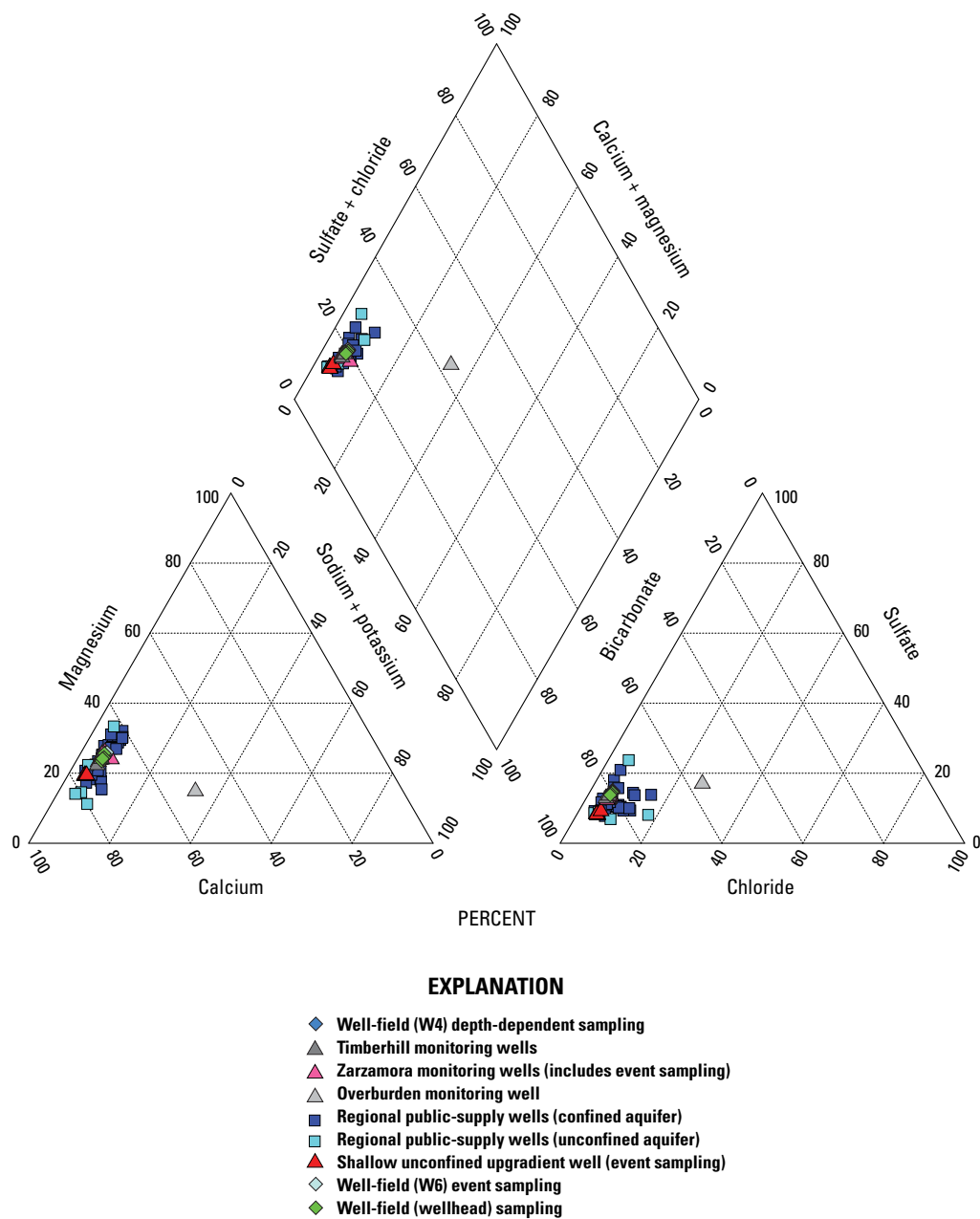


Figure 16. Piper or trilinear diagram showing relations between major cations and anions for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9.

aquifer PSWs, well-field wells, and monitoring wells reflect the carbonate composition of the aquifer host rock and carbonate geochemical reactions. As noted earlier, major element concentrations (table 8) and groundwater facies for TANC study samples, with the exception of the overburden monitoring well, showed little variation (fig. 16). Geochemical processes of water-rock interaction and progressive groundwater evolution in carbonate groundwater (for example, by processes of calcite recrystallization, incongruent dolomite dissolution, and prior precipitation of calcite along flow paths)

can result in increases in Mg/Ca and Sr/Ca molar ratios with increasing groundwater residence time over multiple time scales (Plummer, 1977; Trudgill and others, 1980; Lohmann, 1988; Fairchild and others, 2000; Musgrove and Banner, 2004). Given this evolution pathway, Mg/Ca and Sr/Ca ratios typically increase along flow paths with increasing extents of water-rock interaction. Fluid evolution and water-rock interaction processes noted from the results of this study are consistent with results for the regional Edwards aquifer as discussed in Musgrove and others (2010). Calcium

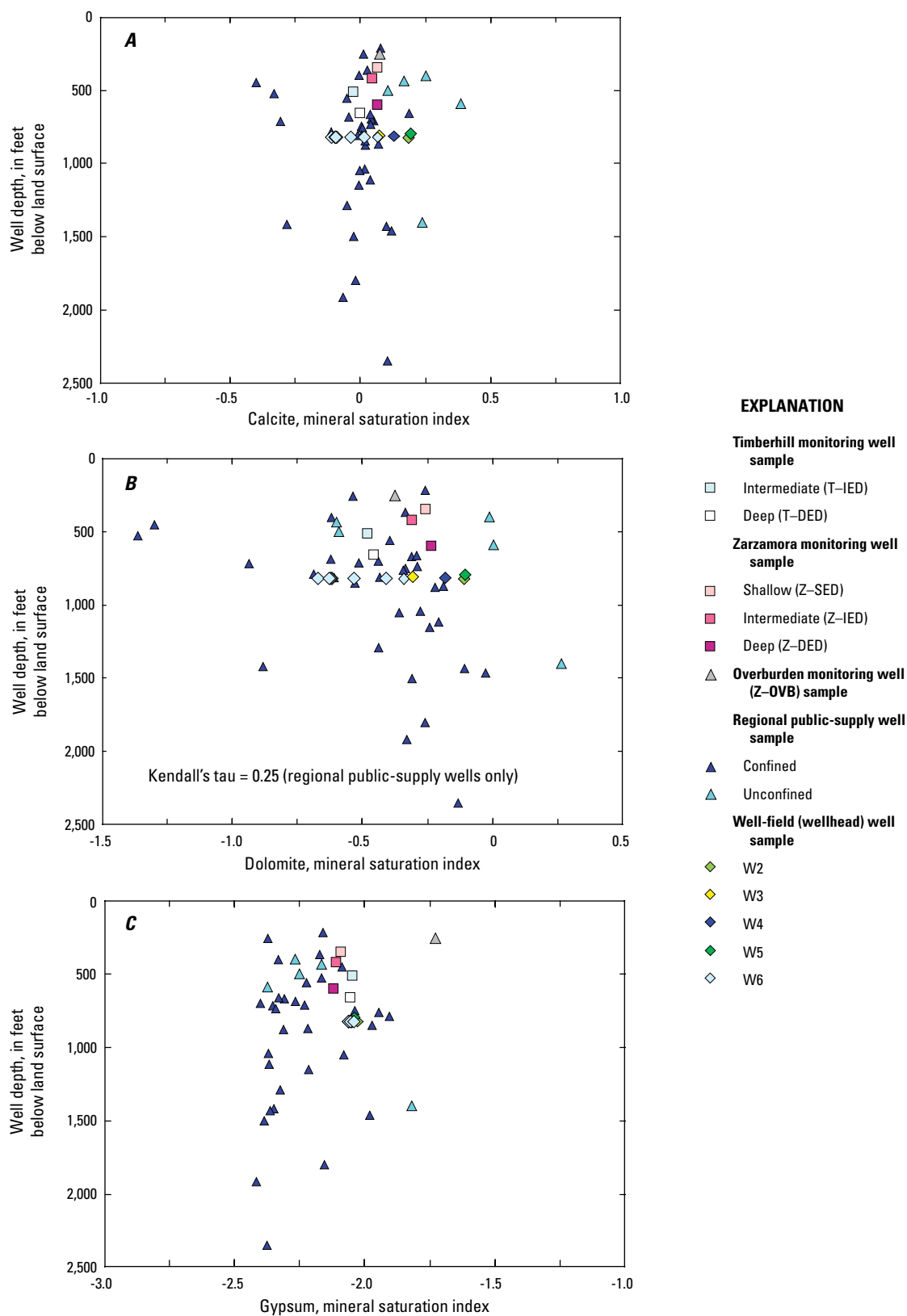


Figure 17. Relation between well depth and selected mineral saturation indexes for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9. *A*, Calcite. *B*, Dolomite. *C*, Gypsum.

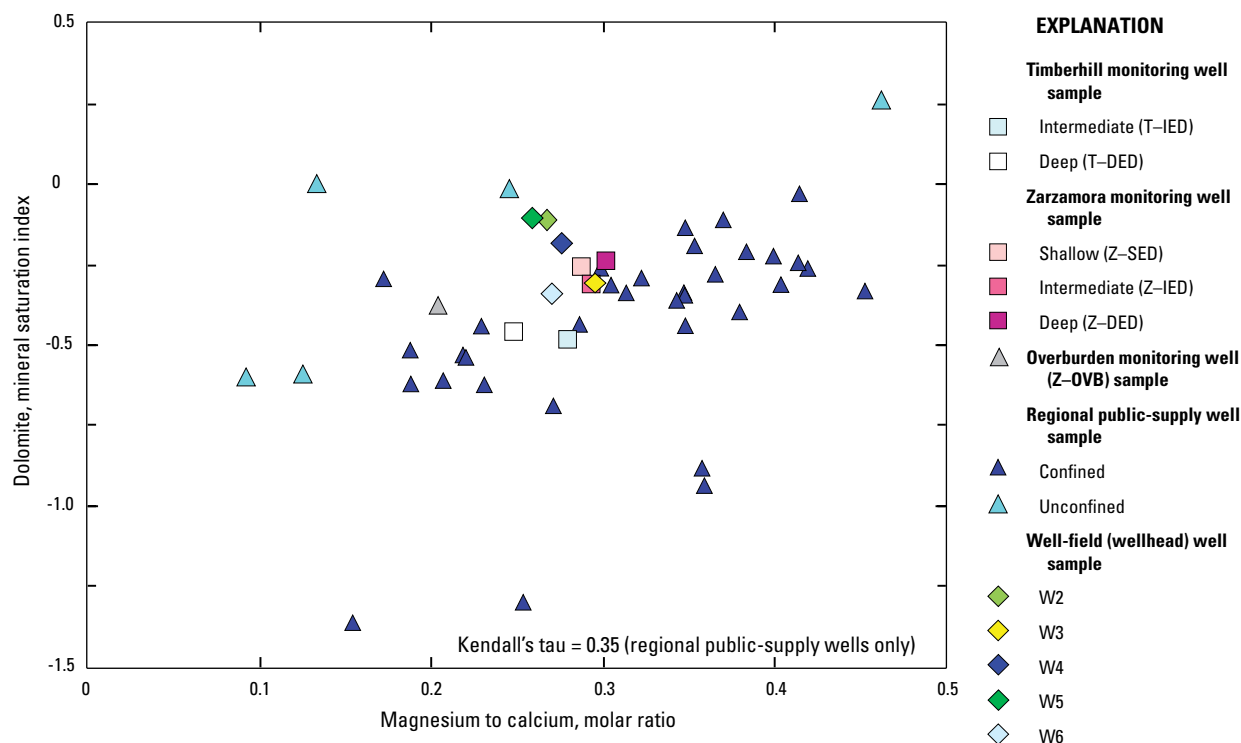


Figure 18. Relation between magnesium to calcium molar ratio and dolomite mineral saturation index for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9.

concentrations in regional aquifer PSW samples decreased with depth and were negatively correlated with Mg/Ca and Sr/Ca ratios (fig. 19). Mineral-solution reactions, including the precipitation of calcite along flow paths, can result in increasing Mg/Ca and Sr/Ca ratios with respect to Ca concentrations (fig. 19). Mg/Ca and Sr/Ca ratios for regional groundwater samples increased with depth (fig. 20), which is indicative of greater extents of water-rock interaction associated with longer flow paths. Mg/Ca and Sr/Ca ratios for TANC groundwater samples were well correlated (Kendall's tau = 0.78). Other than for carbonate-associated elements (Ca, Mg, Sr), major element and trace metal concentrations and elemental ratios (such as Cl to bromide (Cl/Br)) are not strongly correlated with depth. Sr concentrations were well correlated with lithium and F concentrations (Kendall's tau = 0.55 and 0.50, respectively) in regional aquifer PSW samples. These relations might be an indication of small amounts of mixing with downdip saline-zone groundwater, which has relatively high concentrations of these constituents.

Natural contaminants such as U and As are a water-quality concern in some aquifer systems, but in the Edwards aquifer these constituents generally have low and relatively uniform concentrations (table 9 at end of report). The USEPA maximum contaminant level (MCL) is 30 µg/L for U and is 10 µg/L for As. Uranium and As concentrations in well-field

well, monitoring well, and regional aquifer PSW samples are well below these MCLs. Uranium concentrations range from 0.56 to 1.17 µg/L, with a median value of 0.84 µg/L. Arsenic concentrations range from 0.31 to 0.75 µg/L, with a median value of 0.44 µg/L.

Both U and As concentrations in aquifer host rock core samples were generally low (less than 3 milligrams per kilogram (mg/kg)), with one higher As value of 9.9 mg/kg measured in the leached and collapsed members) (table 5). The oxidation of pyrite in limestone is a potential source of As (Appelo and Postma, 2005). Pyrite is a recognized trace mineral in the Edwards Group downgradient of the freshwater/saline-water transition zone but is generally not present in the Edwards Group in the freshwater part of the aquifer (Maclay and Small, 1983). U and As concentrations do not vary significantly with depth in regional aquifer PSWs or in the monitoring wells.

Oxidation-Reduction Conditions

Oxidation-reduction (redox) conditions are an important control on geochemical processes in groundwater that affect a variety of contaminants, including the speciation of elements such as As, iron (Fe), and sulfur and the transformation/biodegradation of NO₃ and some organic compounds. Redox

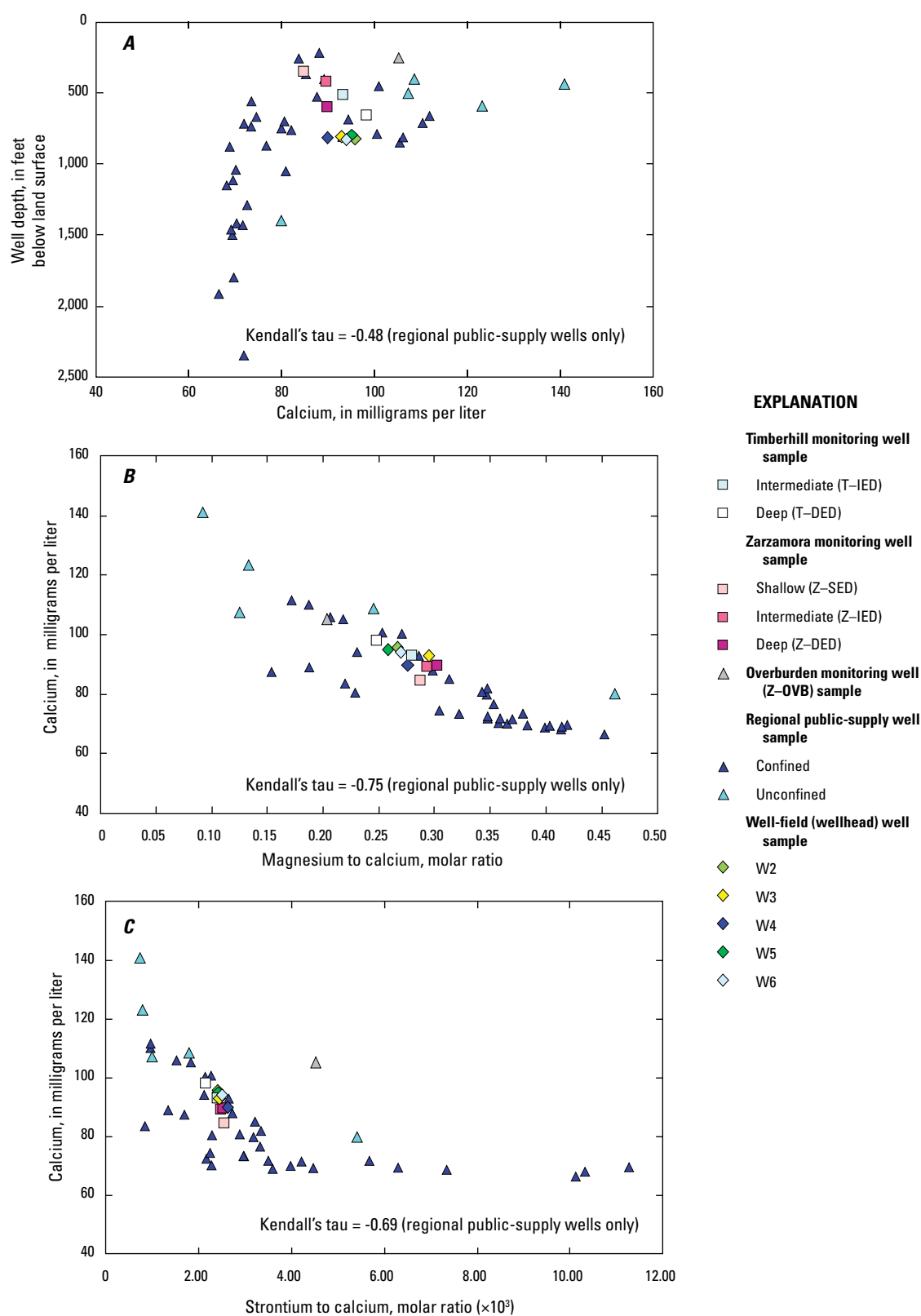


Figure 19. Relation between well depth and selected geochemical constituents or between calcium and selected geochemical constituents for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9. *A*, Relation between well depth and calcium concentration. *B*, Relation between magnesium to calcium molar ratio and calcium concentration. *C*, Relation between strontium to calcium molar ratio ($\times 10^3$) and calcium concentration.

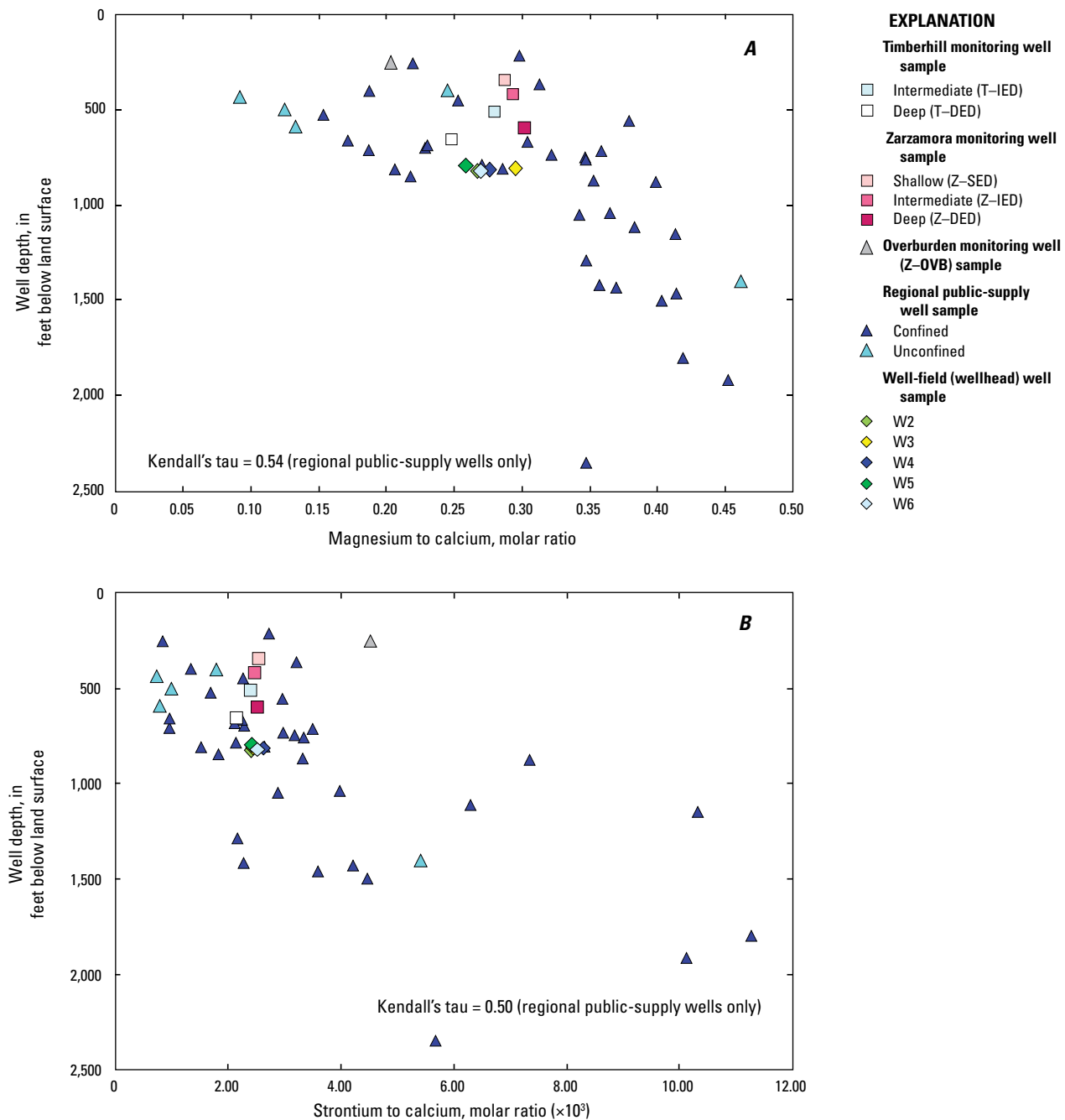


Figure 20. Relation between well depth and selected geochemical constituents for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9. *A*, Magnesium to calcium molar ratio. *B*, Strontium to calcium molar ratio ($\times 10^3$).

conditions in groundwater were evaluated with the classification of McMahon and Chapelle (2008) by using the software of Jurgens and others (2009), which is based on concentrations of dissolved oxygen, NO_3^- , manganese (Mn), Fe, SO_4^{2-} , and sulfide (H_2S). On the basis of this redox classification, it was found that the freshwater part of the San Antonio segment of the Edwards aquifer is generally oxic (oxygen reducing), which is defined by concentrations of dissolved oxygen

greater than 0.5 mg/L, Mn less than 0.05 mg/L, and Fe less than 0.1 mg/L (NO_3^- and SO_4^{2-} concentrations are not considered when dissolved oxygen concentration is greater than 0.5 mg/L). Dissolved oxygen concentrations in the regional aquifer PSW samples ranged from 1.53 to 9.39 mg/L, with a median value of 5.1 mg/L (table 7). Conditions are oxic in the local-scale study area, with a median dissolved oxygen value similar to that for the regional aquifer PSW samples,

but with a narrower range; dissolved oxygen concentrations for well-field and monitoring well samples ranged from 3.65 to 5.26 mg/L, with a median value of 4.6 mg/L (table 7). The dissolved oxygen concentration in the overburden monitoring well (Z-OVB) was similar, with a value of 4.98 mg/L (table 7).

In the vicinity of the freshwater/saline-water interface (fig. 1), dissolved oxygen concentrations decrease, and NO_3^- -reducing/Mn-reducing conditions are widely present (described by concentrations of less than 0.5 mg/L dissolved oxygen, less than 0.5 mg/L NO_3^- , greater than 0.1 mg/L Mn, greater than 0.5 mg/L SO_4^{2-} , and an Fe to H_2S ratio between 0.3 and 10). It is possible that mixed redox or anoxic conditions could occur locally in the freshwater part of the aquifer as a result of point source occurrences of contamination from spills, leaking septic areas, or wastewater lines, or from mixing with less oxic water from the Trinity Group or overlying units (for example, the Buda Limestone or Austin Chalk).

Dissolved Gases

Dissolved gas concentrations (argon, methane, CO_2 , N_2 , oxygen, and He) were measured to estimate properties used in age-tracer calculations, including the temperature of groundwater at the time of recharge and the amounts of excess air. Measured and calculated results, excluding He, are shown in table 10 at end of report; He results are described in the section "Groundwater Age Tracers and Age Distributions" later in the report. Concentrations of N_2 and argon are consistent with atmospheric equilibration during groundwater recharge with excess air added during recharge. Calculated excess air concentrations range from 1.9 to 11.0 cubic centimeters per kilogram at standard temperature and pressure ($\text{cm}^3\text{STP/kg}$), with a median value of 5.8 $\text{cm}^3\text{STP/kg}$. Most groundwater contains excess air, which is generally attributed to air entrainment during recharge (Heaton and Vogel, 1981). Excess air concentrations are typically from near 0 to 3 $\text{cm}^3\text{STP/kg}$, although higher values have been measured (Busenberg and Plummer, 2000) and might be associated with some flow conditions in fractured rocks or in aquifers with rapid variations in water table elevation such as karst (Busenberg and Plummer, 1992; L.N. Plummer, U.S. Geological Survey, oral commun., 2009). No excess N_2 , such as would occur from denitrification of NO_3^- , was detected in any dissolved gas samples (table 10). These results are consistent with an absence of denitrification in the sampled groundwater based on oxic conditions.

Nutrients and Stable Isotopes of Nitrogen

Nutrients are defined as those elements important to plant growth and survival and include the primary mineral nutrients N, phosphorus, and potassium. Anthropogenic activity has transformed the global N cycle, and elevated nutrient concentrations in water resources are an issue of worldwide concern (Schlesinger and others, 2006; Galloway and others, 2008).

Nitrogen is the most soluble of these nutrients, which generally occurs as nitrate (NO_3^-), nitrite (NO_2^-), and ammonium (NH_4^+) in groundwater (Hem, 1992). Ammonium strongly adsorbs on mineral surfaces, whereas NO_3^- and NO_2^- are readily transported in water (Hem, 1992). Reduction of NO_3^- to NO_2^- (denitrification) occurs in anoxic conditions. Anthropogenic N sources include agricultural fertilizer, animal wastes, septic systems, municipal sewage-treatment systems, leaking sewerlines, and N oxide emissions. Atmospheric deposition and precipitation, decaying organic matter, and soil mineralization processes are also potential N sources. Natural N sources are not common in limestone and dolomitic aquifer rocks; thus, N concentrations greater than the prevailing background concentration likely represent anthropogenic sources. National background concentrations of NO_3^- have been estimated to be 1.0 mg/L for groundwater (Dubrovsky and others, 2010).

Concentrations of NO_2^- were negligible where detected and below laboratory method reporting levels for most samples (table 11 at end of report), which is consistent with generally oxic conditions in the aquifer. As a result, measured concentrations of $\text{NO}_3^- + \text{NO}_2^-$ primarily represent concentrations of NO_3^- ; where NO_2^- was detected, NO_3^- concentrations were calculated as the difference between $\text{NO}_3^- + \text{NO}_2^-$ and NO_2^- .

All measured NO_3^- concentrations were below the MCL of 10 mg/L for drinking water (U.S. Environmental Protection Agency, 1993). The overburden monitoring well (Z-OVB) sample had the highest NO_3^- concentration, with a value of 3.82 mg/L. Groundwater samples from regional aquifer PSWs had NO_3^- concentrations ranging from 0.73 to 3.12, with a median of 1.86 mg/L. The median for samples from regional (confined) aquifer PSWs was 1.87 mg/L, which is similar to the previously reported median values for regional groundwater in the aquifer (1.67 mg/L) (Musgrove and others, 2010). Well-field and monitoring well samples had a slightly narrower range of NO_3^- concentrations (1.65–2.74 mg/L) (table 11) with a similar median value of 2.04 mg/L.

Concentrations of NO_3^- for regional aquifer PSW samples showed a decrease in concentration with depth (Kendall's $\tau = -0.25$) (fig. 21). This relation is not evident for well-field and monitoring well samples, which cover a more limited depth range than do the regional aquifer PSWs (fig. 21). Land use might affect NO_3^- concentrations in Edwards aquifer groundwater; higher NO_3^- concentrations have been measured in groundwater from shallow unconfined monitoring wells in the urban San Antonio area. These patterns are consistent with either greater amounts of anthropogenically derived NO_3^- in the urban environment and (or) dilution of ambient NO_3^- by surface-water recharge (Musgrove and others, 2010).

Nitrate isotopes ($\delta^{15}\text{N}$ and $\delta^{18}\text{O}$ of NO_3^-) have been widely applied to trace the cycling and sources of NO_3^- in hydrologic systems (Kendall, 1998). Different sources or reservoirs of NO_3^- have characteristic isotopic "signatures" that reflect the source of the NO_3^- (Kendall, 1998; Mengis and others, 2001). Values shown on figure 22 reflect typical ranges of major NO_3^- reservoirs, although the boundaries are not absolute (Kendall, 1998), and local endmember values of these sources have not

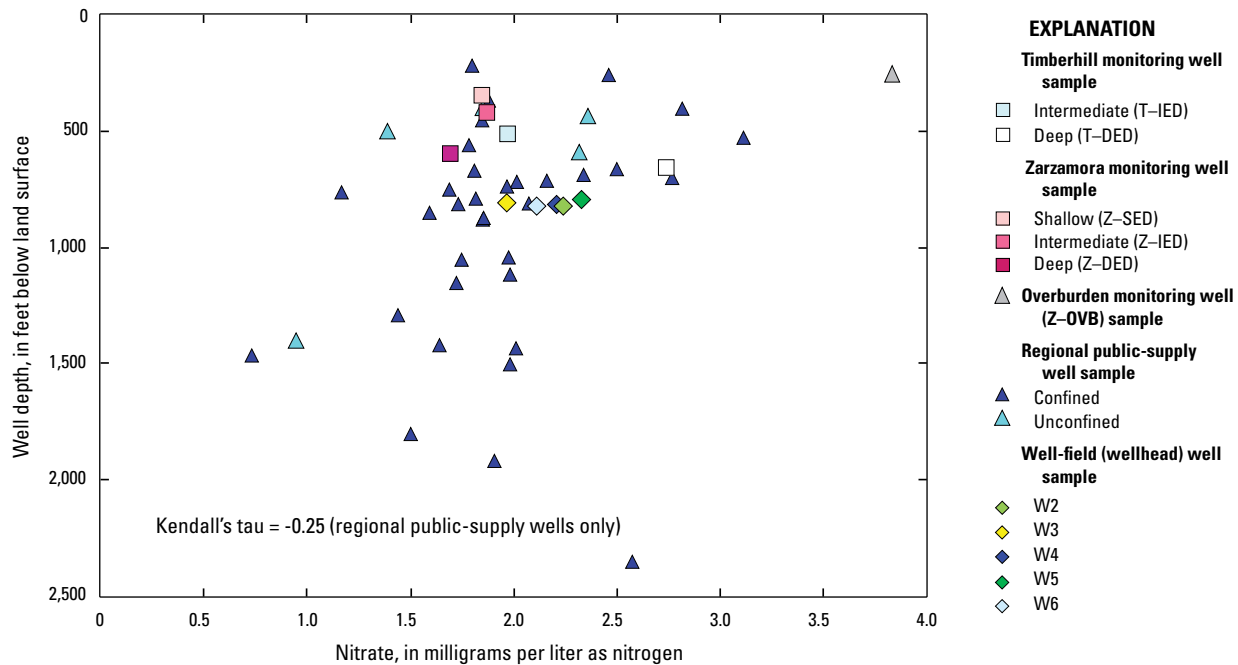


Figure 21. Relation between well depth and nitrate concentration for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9.

been determined. Additionally, the isotopic composition of the NO_3 in a water sample might reflect a single source or a mix of sources with different isotopic compositions. The isotopic signature can also be modified by biogeochemical processes, such as denitrification (fig. 22). Nitrate-isotope samples were collected from well-field and monitoring wells, including depth-dependent sampling in W4, and from the overburden well (Z-OVB) (table 12). Results for $\delta^{15}\text{N}$ range from 6.84 per mil to 8.16 per mil with a median of 7.91 per mil; results for $\delta^{18}\text{O}$ range from 4.59 per mil to 5.52 per mil, with a median of 5.23 per mil. Nitrate-isotope results for all TANC study samples, excluding the overburden well, are within a very narrow range that is within analytical uncertainty (± 0.5 per mil) (fig. 22). The overburden well, in the Buda Limestone, has a slightly lower $\delta^{15}\text{N}$ value (6.84 per mil) (fig. 22). Based on the ranges of different NO_3 sources shown in figure 22 (Kendall, 1998; Mengis and others, 2001), $\delta^{15}\text{N}$ values for TANC groundwater samples are mostly within the range of natural soil organic values, near the higher values that transition to NO_3 sourced from human and (or) animal waste (fig. 22). Samples collected from Barton Springs (the main discharge point for the Barton Springs segment of the aquifer) during 2008–10 under a range of flow conditions have similar median NO_3 -isotope values (7.57 per mil for $\delta^{15}\text{N}$ and 5.58 per mil for $\delta^{18}\text{O}$) (Mahler and others, 2011).

Nitrate-isotope results confirm that denitrification is not an important process affecting NO_3 in the Edwards aquifer. This finding is consistent with oxic conditions throughout the aquifer and the absence of excess N_2 gas, an indicator of

denitrification, in all samples collected for this study. Denitrification has been identified as an important process in other shallow (Panno and others, 2001) and deep karst aquifers (Katz and others, 2007).

The median $\delta^{15}\text{N}$ composition of Edwards aquifer groundwater collected for this study (7.91 per mil) is higher than historical values. A previous study of regional unconfined ($n=33$) and confined ($n=34$) groundwater in the Edwards aquifer measured median $\delta^{15}\text{N}$ values of 6.60 and 6.00 per mil, respectively (Kreitler and Browning, 1983). Higher $\delta^{15}\text{N}$ values are likely indicative of NO_3 from human and (or) animal waste, including septic systems and wastewater sources (fig. 22). Historical regional Edwards aquifer $\delta^{15}\text{N}$ values from Kreitler and Browning (1983) ($n=67$) are statistically different (based on Mann-Whitney U testing) from recent results for samples collected during the TANC study ($n=20$). This comparison suggests that NO_3 sources to groundwater in the Edwards aquifer, although primarily in the soil organic N range, might include a potential component of manure and (or) septic/sewage sources. Kreitler and Browning (1983) did not measure $\delta^{18}\text{O}$ compositions; thus, only historical $\delta^{15}\text{N}$ results were compared. This comparison with historical measurements of $\delta^{15}\text{N}$ should be interpreted with caution, however, because interlaboratory and intermethod comparisons could not be made. Nonetheless, other data are consistent with the possibility that $\delta^{15}\text{N}$ values for TANC study samples are higher than historical values as a result of changes in potential NO_3 sources and loadings to the aquifer. The median NO_3 concentration for regional confined groundwater samples collected by

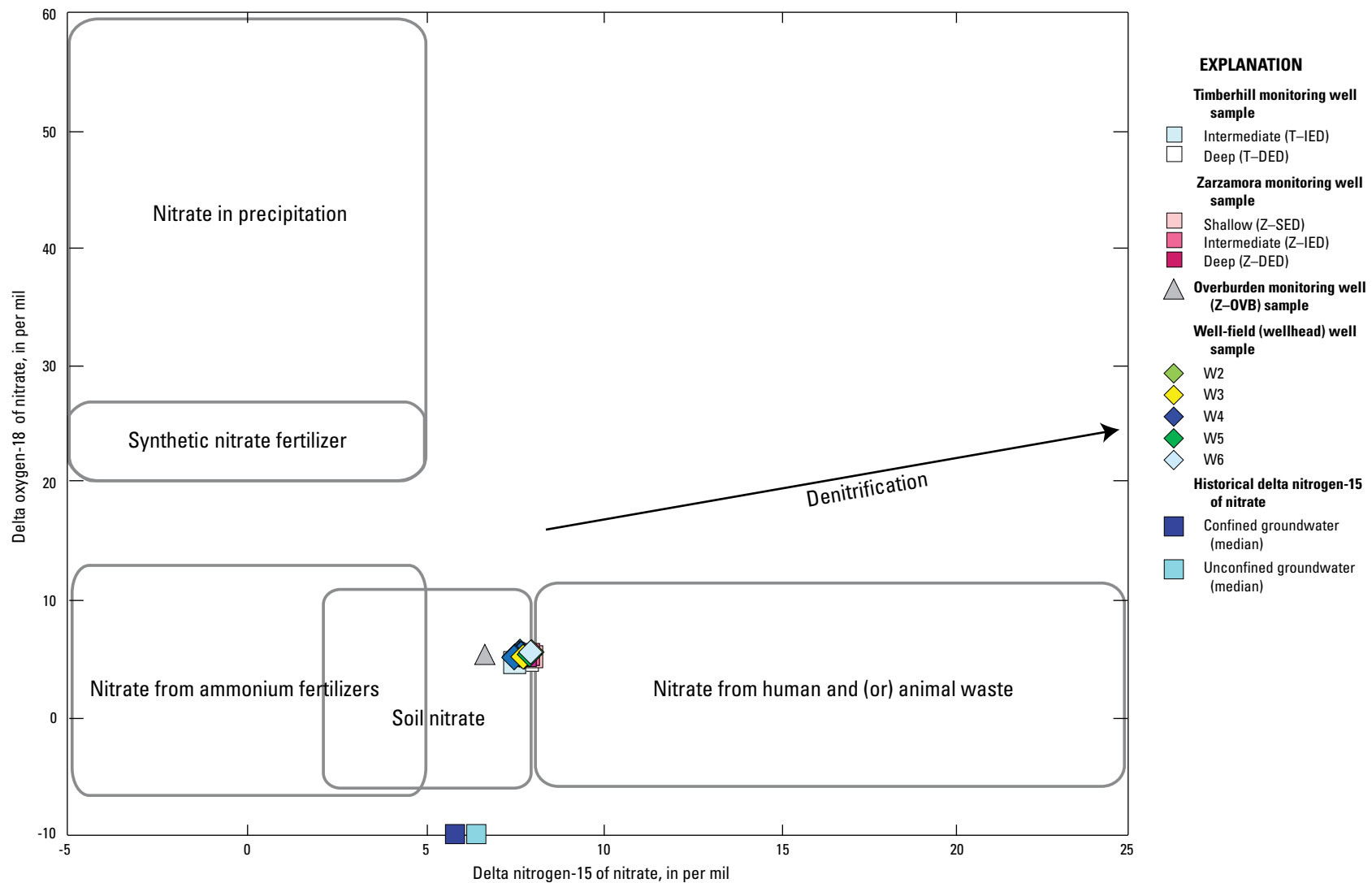


Figure 22. Isotopic composition of nitrate for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9. Historical data for delta nitrogen-15 ($\delta^{15}\text{N}$) of nitrate shown for comparison are based on results from Kreitler and Browning (1983) (delta oxygen-18 of nitrate ($\delta^{18}\text{O}$) was not measured for historical samples). Common fields of $\delta^{15}\text{N}$ and $\delta^{18}\text{O}$ are modified from Kendall (1998), Mengis and others (2001), and Silva and others (2002).

Kreitler and Browning (1983) was 1.40 mg/L, which is significantly lower than that for regional aquifer PSW samples. An upward trend for NO_3 concentrations (1938–2006) in the San Antonio segment of the aquifer was previously proposed by Musgrove and others (2010). Mahler and others (2011) found that nitrate concentrations for recent (2008–10) samples from the Barton Springs segment of the aquifer are elevated relative to historical (1990–2008) samples. Additionally, Mahler and others (2011) found that recent (2008–10) $\delta^{15}\text{N}$ values for samples from Barton Springs were significantly higher than historical values (1990–94).

Strontium Isotopes

Strontium isotopes ($^{87}\text{Sr}/^{86}\text{Sr}$) are a useful tracer in the hydrologic cycle, in particular for sources of dissolved constituents to a fluid (Banner, 2004). Strontium (Sr) isotopes in hydrology have been used to trace flow paths, mineral-solution reactions, and variations in recharge and flow routes in a number of different groundwater settings (McNutt and others, 1990; Banner and others, 1994). In the Edwards aquifer, $^{87}\text{Sr}/^{86}\text{Sr}$ values have been applied previously in conjunction with element ratios (Mg/Ca, Sr/Ca) as a tracer of water-rock interaction, groundwater residence time, recharge, unsaturated zone flow-routes, and the influence of soil composition on groundwater geochemistry (Oetting and others, 1996; Musgrove and Banner, 2004; Garner, 2005; Wong and others, 2011). Sr isotope ratios in the Edwards aquifer generally decrease with increasing water-rock interaction, approaching values similar to those of the host limestone aquifer rocks (Oetting and others, 1996; Sharp and Banner, 1997; Musgrove and Banner, 2004; Musgrove and others, 2010). Cretaceous-age carbonate host rocks that compose the Edwards aquifer have $^{87}\text{Sr}/^{86}\text{Sr}$ values ranging from 0.7074 to 0.7077 (Koepnick and others, 1985; Oetting, 1995). Higher $^{87}\text{Sr}/^{86}\text{Sr}$ values for groundwater relative to the aquifer host rocks are indicative of a source of more radiogenic Sr (enriched in ^{87}Sr) to the groundwater, which has been proposed to be a result of chemical interaction of groundwater with overlying soils and proportionally less interaction with aquifer host rocks (Musgrove and Banner, 2004). The Sr isotope composition of regional Edwards aquifer groundwater demonstrates a progression toward lower $^{87}\text{Sr}/^{86}\text{Sr}$ values with increased mineral-solution reactions as allowed by increased residence time and longer flow paths; Sr isotope results are consistent with a characterization of confined groundwater samples from the well-field and monitoring wells as more geochemically evolved relative to shallow and unconfined groundwater (Musgrove and others, 2010).

Sr isotopes were not measured for regional aquifer PSWs; thus, only values for the well-field and monitoring wells are evaluated. The strontium isotope composition of well-field and monitoring well samples had a relatively narrow range (0.70767–0.70784). The median value of 0.70772 is very close to the range of the carbonate host rocks, indicative

of relatively geochemically evolved groundwater (that is, groundwater that has undergone extensive mineral-solution reaction with the carbonate host rocks). Values for the well-field samples and samples from the Zarzamora monitoring wells are mostly within analytical uncertainty of each other; values for the samples from the Timberhill monitoring wells are slightly elevated in comparison (median of 0.70783). The $^{87}\text{Sr}/^{86}\text{Sr}$ value for the sample from the overburden monitoring well (Z–OVB) (0.70785) was also slightly elevated in comparison, similar to values for the Timberhill monitoring wells. Slightly higher $^{87}\text{Sr}/^{86}\text{Sr}$ values for the Timberhill monitoring wells might reflect differences in flow paths supplying these wells or might indicate that groundwater from these wells are less geochemically evolved than is groundwater from the well-field and Zarzamora monitoring wells. $^{87}\text{Sr}/^{86}\text{Sr}$ values for the Buda Limestone have not been characterized. Both the Zarzamora and Timberhill monitoring well values decrease slightly with depth, although since results for most of the monitoring wells are within analytical uncertainty of each other, any $^{87}\text{Sr}/^{86}\text{Sr}$ trend with depth in the monitoring wells is poorly constrained (fig. 23). A trend of decreasing $^{87}\text{Sr}/^{86}\text{Sr}$ with depth, however, would be consistent with more geochemically evolved groundwater associated with deeper flow paths in the aquifer.

Stable Isotopes of Hydrogen and Oxygen

Stable isotope ratios of hydrogen and oxygen (δD and $\delta^{18}\text{O}$) are commonly used to study hydrologic processes and have been applied to assess groundwater origin, mixing, and evaporation, as well as hydrograph separation and the response of karst systems to storms, aquifer recharge, subsurface flow routing (diffuse relative to conduit), and antecedent conditions (Craig, 1961; Gat, 1981; Lakey and Krothe, 1996; Jones and others, 2000; Desmarais and Rojstaczer, 2002; Maloszewski and others, 2002). Well-field and monitoring well samples were analyzed for δD and $\delta^{18}\text{O}$; results showed little variability, with median values of -22.7 per mil and -4.1 per mil, respectively (table 12 at end of report). A comparison of these data with both the global meteoric water line (GMWL), which represents a regression line for the isotopic composition of rainfall around the globe (Craig, 1961), and the local meteoric water line (LMWL) (Pape and others, 2010) indicates that groundwater samples were meteoric in origin (fig. 24). On the basis of stable isotope and other geochemical results, a small number of regional groundwater samples have been previously interpreted to be affected by mixing with (a) down-dip saline water, (b) recent recharge associated with tropical cyclonic storms, or (c) recharge water that has undergone evaporation (Musgrove and others, 2010). Regional aquifer PSW samples were not analyzed for δD and $\delta^{18}\text{O}$, but previous results for Edwards aquifer groundwater indicate that confined groundwater is less variable relative to shallow/urban unconfined and unconfined groundwater (Musgrove and others, 2010). The absence of variability in δD and $\delta^{18}\text{O}$ results

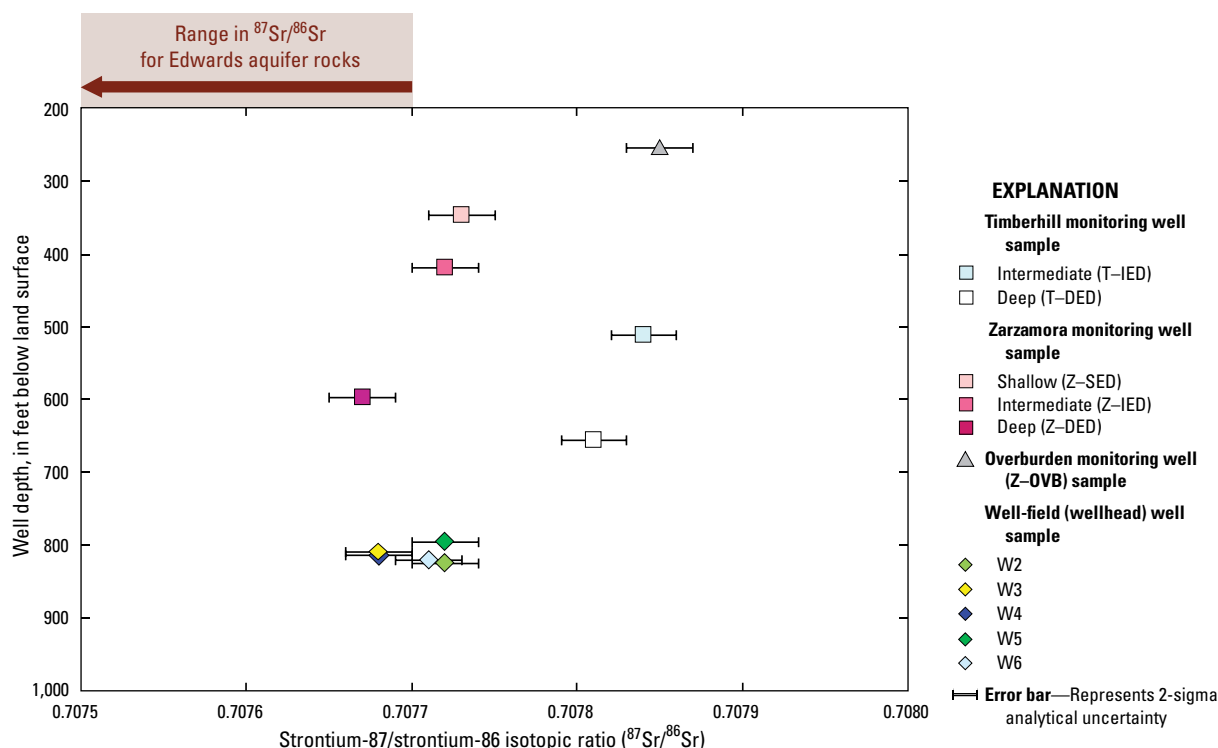


Figure 23. Relation between well depth and strontium isotope composition for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9. Range for Cretaceous-age carbonate rocks that compose the Edwards aquifer (0.7074 to 0.7077) from Koepnick and others (1985) and Oetting (1995).

for samples from TANC well-field and monitoring wells is consistent with the study characterization of the PSW samples as well-mixed, homogeneous, confined groundwater that is typical of the regional aquifer.

Organic Constituents

The presence of organic constituents that do not occur naturally in hydrologic systems, such as pesticides and VOCs, is a direct result of anthropogenic activities. Anthropogenic organic contaminants are of concern in groundwater and drinking-water supplies because of their potential detrimental effects on the environment and to human health. Many of these compounds are relatively soluble and thus can be present in groundwater if anthropogenic sources are present. Pesticides and pesticide degradation products generally include toxic compounds used to kill or control plants (herbicides), insects (insecticides), fungi (fungicides), and rodents (rodenticides). VOCs are carbon-based compounds that readily volatilize and are commonly found in paints, solvents, gasoline-related compounds, refrigerants, fumigants, and other industrial compounds. VOCs include trihalomethanes (THMs), which are a byproduct created by the disinfection of drinking water by chlorination. The frequency of detection

and spatial distribution of anthropogenic contaminants provide insight into their potential sources, as well as into effects of anthropogenic activities on water quality.

Samples for organic constituents on the analytical schedule for wastewater indicator compounds were collected from regional aquifer PSWs (table 13 at end of report). Results from these analyses were almost entirely nondetections, so these constituents are not discussed further.

Pesticides and VOCs were analyzed in samples collected from the well-field and monitoring wells and the regional aquifer PSWs. A subset of the regional aquifer PSWs ($n=39$) that were sampled in 2004 were resampled in 2005 for only pesticides ($n=8$) and VOCs ($n=7$). TANC results for organic contaminants were consistent with previous studies that have documented the occurrence of anthropogenic organic contaminants in the Edwards aquifer at low concentrations (Bush and others, 2000; Fahlgust and Ardis, 2004; Musgrove and others, 2010). Pesticides and VOCs were detected in all analyzed well categories, including well-field wells, monitoring wells, and regional aquifer PSWs. Detected concentrations of organic constituents were mostly less than 1 $\mu\text{g/L}$ and were typically orders of magnitude below USEPA MCLs for those constituents for which MCLs currently exist. The detection of at least one pesticide compound and (or) VOC in samples from most of the regional aquifer PSWs collected in 2004 through 2005

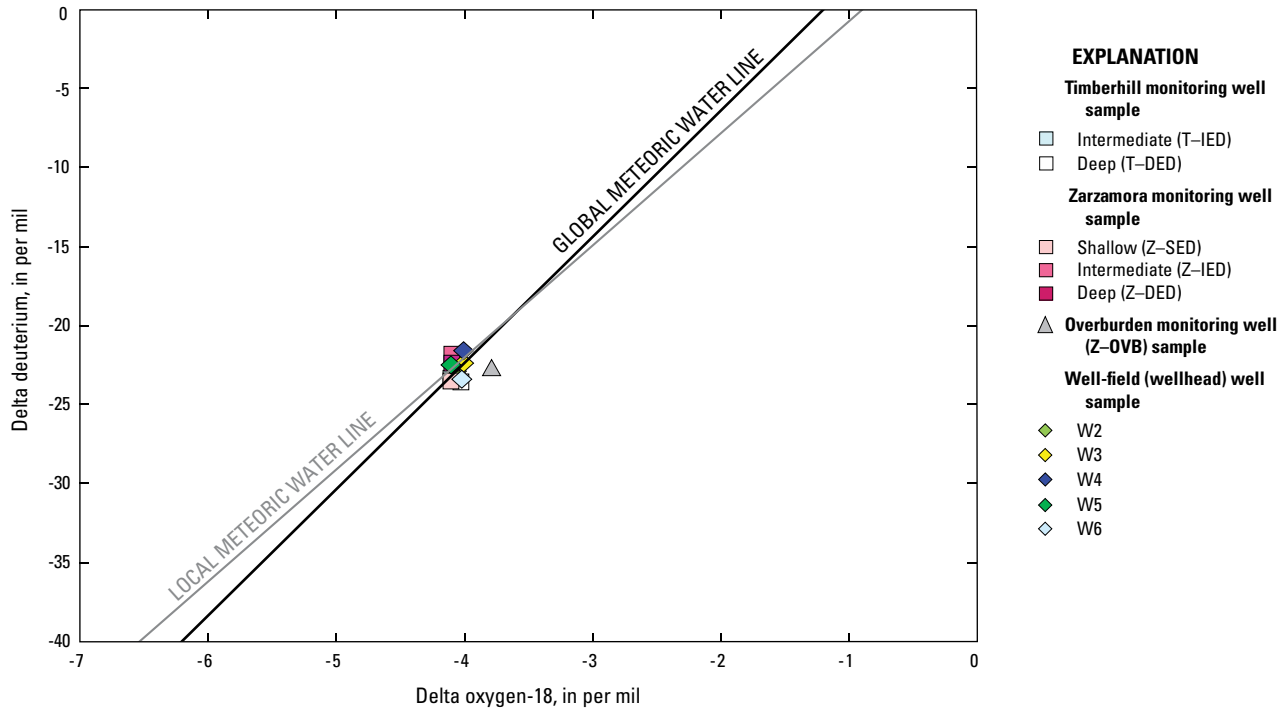


Figure 24. Relation between deuterium and oxygen isotopes for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9. Local and global meteoric water lines are shown for comparison, based on Pape and others (2010) and Craig (1961), respectively.

(tables 14–17 at end of report) indicates that some fraction of groundwater in the regional Edwards aquifer is affected by anthropogenic activities.

One or more pesticide compounds were detected in 79 percent of the regional aquifer PSW samples, and one or more VOCs were detected in 87 percent of these samples. The number of organic compounds detected in any single groundwater sample ranged from 0 to 10, with a median of 5. A total of 17 pesticide compounds and 11 VOCs were detected; most were detected infrequently. A small number of compounds were consistently and frequently detected (≥ 50 percent of samples) in regional aquifer PSW samples, indicating that there are likely widespread, low-level inputs of these constituents from multiple sources. The most frequently detected organic constituents were (1) the pesticide atrazine, (2) the atrazine degradate 2-chloro-4-isopropylamino-6-amino-s-triazine (CIAT, also known as deethylatrazine or, hereinafter, DEA), (3) the THM chloroform, and (4) the solvent PCE, which were detected in 60, 74, 72, and 50 percent of samples from regional aquifer PSWs, respectively (fig. 25). In samples from the regional aquifer PSWs, median concentrations of atrazine, DEA, chloroform, and PCE were 0.005 $\mu\text{g/L}$, 0.006 $\mu\text{g/L}$, 0.020 $\mu\text{g/L}$ and 0.008 $\mu\text{g/L}$, respectively (fig. 26). These concentrations are orders of magnitude below the USEPA MCLs for these compounds, which are 3 $\mu\text{g/L}$ for atrazine, 80 $\mu\text{g/L}$ for total THMs, including chloroform, and 5 $\mu\text{g/L}$ for PCE. Other pesticides that were frequently detected were simazine

and prometon, which were detected in 38 and 30 percent of the regional aquifer PSW samples, respectively (fig. 25). The atrazine degradates deisopropylatrazine (or 2-chloro-6-ethyl-amino-4-amino-s-triazine (hereinafter, CEAT)) and hydroxy-atrazine (or 2-hydroxy-4-isopropylamino-6-ethylamino-s-triazine (hereinafter, OIET)) were analyzed for in only a subset of the regional aquifer PSW samples ($n=15$) and were detected in 25 and 15 percent of those samples, respectively. In the regional aquifer PSW samples, median concentrations of CEAT, OIET, simazine, and prometon were below laboratory method reporting levels of 0.08 $\mu\text{g/L}$, 0.032 $\mu\text{g/L}$, 0.005 $\mu\text{g/L}$, and 0.010 $\mu\text{g/L}$, respectively. The THMs bromoform, dibromochloromethane, and bromodichloromethane were detected in 26, 17, and 17 percent of the regional aquifer PSW samples, respectively. In regional aquifer PSW samples, total THM concentrations ranged from below laboratory method reporting levels to 3.178 $\mu\text{g/L}$, with a median value of 0.023 $\mu\text{g/L}$.

The frequent detection of atrazine, DEA, chloroform, and PCE, as well as the common detection of simazine and prometon at low concentrations, is consistent with constituent occurrence patterns in the Edwards aquifer previously noted by Mahler and others (2006) and Musgrove and others (2010). Lindsey and others (2009) report similar findings for other carbonate aquifers in similar land use settings. Where detected, organic constituents tend to co-occur; that is, concentrations of the most frequently detected organic compounds are correlated. For example, for the regional aquifer PSW samples,

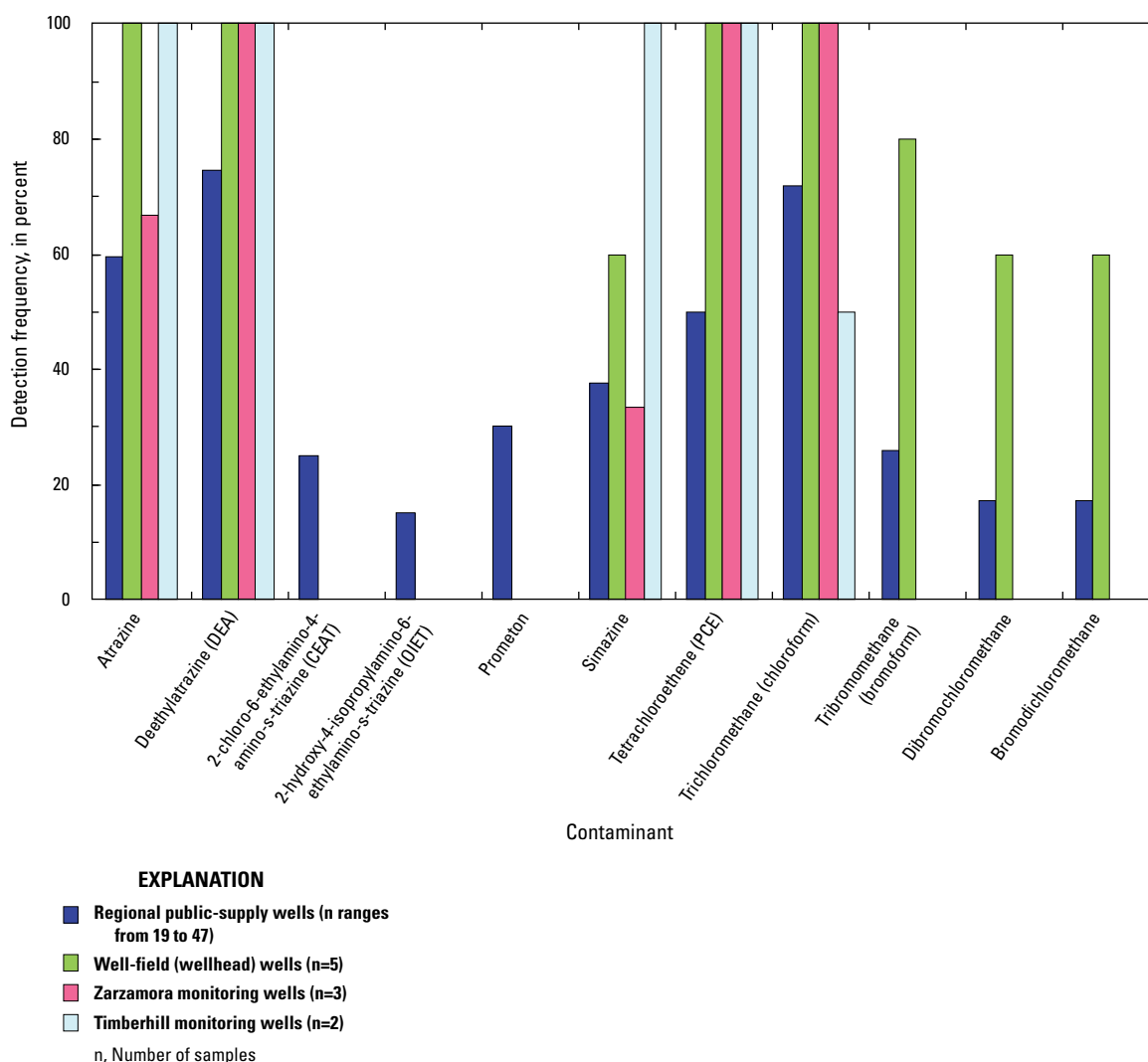


Figure 25. Detection frequency for the most frequently detected anthropogenic contaminants for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9.

concentrations of atrazine were correlated with concentrations of DEA, simazine, PCE, and chloroform (Kendall's tau = 0.69, 0.55, 0.56, 0.46, respectively).

Atrazine, DEA, simazine, chloroform, and PCE were frequently detected in the well-field (n=5) and monitoring well (n=5) samples, though at different frequencies (fig. 25) and median concentrations (fig. 26) than in the regional aquifer PSW samples. DEA and PCE were detected in 100 percent of the well-field and monitoring well samples. Atrazine and chloroform were detected in 100 percent of the well-field samples; simazine was detected in 60 percent of these samples, and prometon was not detected. Bromoform and dibromochloromethane were detected in 80 and 60 percent of the well-field samples, respectively. Detection frequencies for the two sets of monitoring wells were different for some organic constituents (fig. 25); atrazine and simazine were detected in 100 percent of the Timberhill monitoring well samples (n=2) but had

lower detection frequencies in the Zarzamora monitoring well samples (n=3). For the Timberhill monitoring wells, chloroform was detected in only the deeper well sample.

A comparison of median concentrations for the most frequently detected organic constituents in the different well categories provides insight into potential sources and transport of anthropogenic constituents to the well field. Median concentrations of atrazine, DEA, and PCE in the regional aquifer PSW samples (0.005, 0.006, and 0.008 µg/L, respectively) are representative of regional confined groundwater on the basis of their similarity with this groundwater category as detailed by Musgrove and others (2010). Median concentrations of these contaminants, however, are significantly higher (on the basis of Mann-Whitney U testing) in the well-field and monitoring well samples compared to the regional aquifer PSW samples (fig. 26); this result may indicate that the well field is affected by some additional, and likely local, source(s)

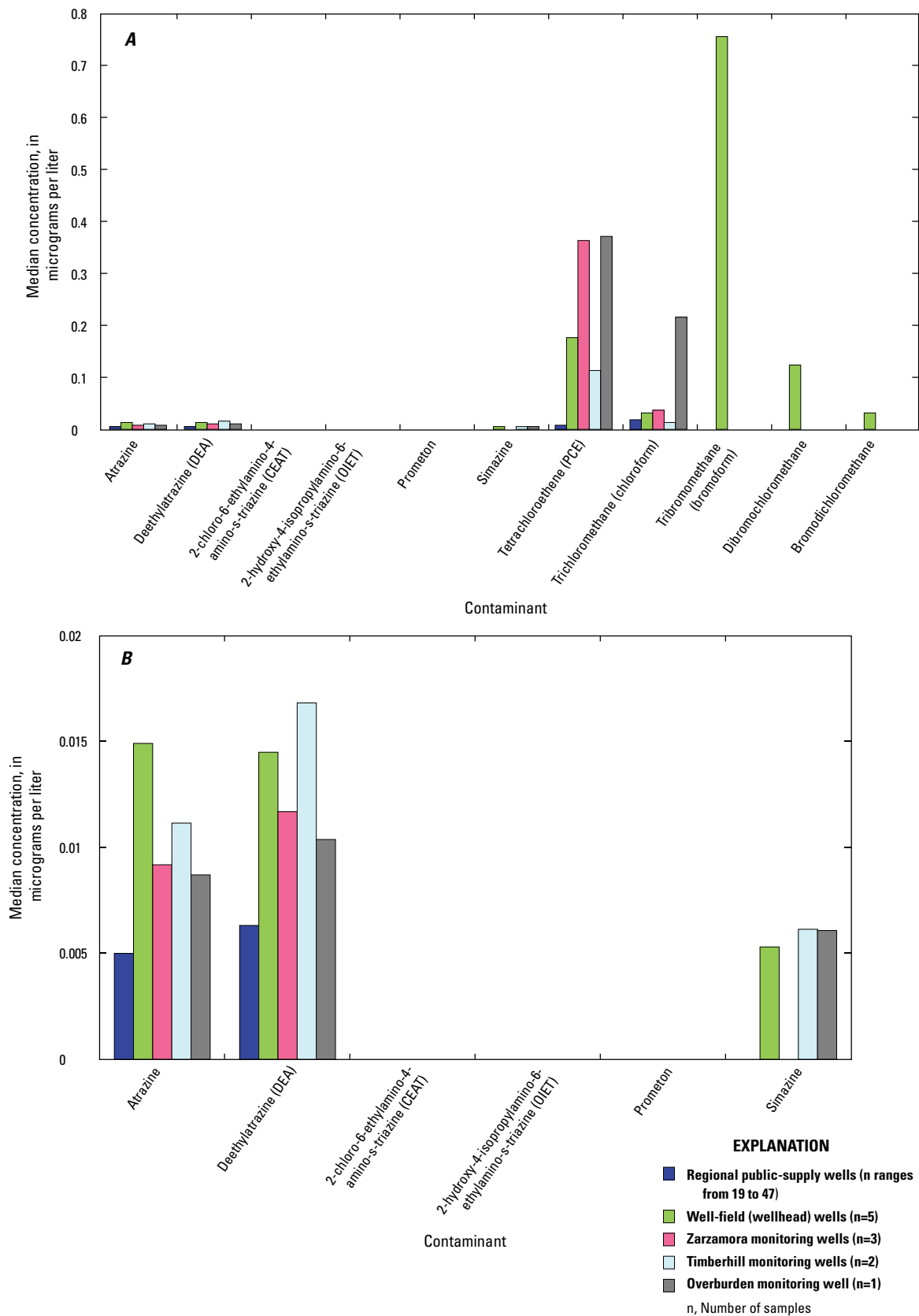


Figure 26. Median concentration of the most frequently detected anthropogenic contaminants for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9. Two scales are shown for selected contaminants for comparison. *A*, Full scale. *B*, Reduced scale.

of these contaminants. A previous evaluation of organic constituents in the Edwards aquifer found that the urban environment is a potential source of anthropogenic contaminants to the aquifer; shallow/urban unconfined groundwater had higher median concentrations of atrazine, DEA, and PCE (0.008, 0.009, and 0.027 $\mu\text{g/L}$, respectively) and a higher frequency of detection than did regional confined groundwater (Musgrove and others, 2010). Median concentrations of PCE are significantly higher (on the basis of Mann-Whitney U testing), in particular, in the well-field samples (0.176 $\mu\text{g/L}$) and the Zarzamora monitoring well samples (0.365 $\mu\text{g/L}$) in comparison with both regional aquifer PSW samples (fig. 26) and shallow/urban unconfined groundwater described by Musgrove and others (2010).

The most frequently detected organic contaminants generally showed a trend of decreasing concentration with depth for the regional aquifer PSW samples; atrazine, simazine, PCE, and chloroform concentrations were correlated with well depth (Kendall's tau = -0.36, -0.37, -0.28, -0.25, respectively). More specifically, the relation with well depth for most of the frequently detected organic contaminants tends to be bimodal (fig. 27). Median concentrations and the frequency of detection for atrazine, DEA, simazine, PCE, and chloroform in the regional aquifer PSW samples are consistently higher for wells that are less than 1,000 ft in depth. For wells that are greater than 1,000 ft deep, median concentrations for these organic contaminants are all less than the laboratory method reporting levels. The concentration difference between the two groups (that is, wells less than 1,000 and greater than 1,000 ft deep) for all of these organic contaminants is statistically significant on the basis of Mann-Whitney U testing. About half of the deep (greater than 1,000 ft deep) regional (confined) aquifer PSWs are near or along the southern boundary of the confined segment of the aquifer (fig. 1). These deeper confined PSWs might intersect downgradient regional flow paths that have been less affected by anthropogenic contaminants. Samples from these deep wells (greater than 1,000 ft deep) show other evidence for a large component of more geochemically evolved and potentially older waters, such as higher Mg/Ca ratios and dolomite SIs, both of which have been shown to be qualitative indicators of residence time in this aquifer (Musgrove and others, 2010). Both Mg/Ca ratios and dolomite SIs were significantly different between samples collected from shallower wells (less than 1,000 ft deep) and those collected from deeper wells (greater than 1,000 ft deep) on the basis of Mann-Whitney U testing. It is possible that groundwater from these deep wells might be affected by mixing with a component of downdip saline-zone water, which might dilute concentrations of anthropogenic contaminants. However, there is no clear indication that such mixing has likely occurred; for example, concentrations of ions associated with the saline zone (for example, Cl, SO_4 , Sr, and F) were not statistically different for samples from the regional (confined) aquifer PSWs with well depths greater than 1,000 ft, with the exception of Sr. These results are consistent with a previous study of low-level VOC concentrations in the San Antonio area that

found fewer VOC detections in deeper wells and proposed that these deep wells might intercept regional flow paths that contain higher fractions of water recharged in rural areas in the western portion of the aquifer (Plummer and others, 2008).

The frequent detection of chloroform at low concentrations in samples from the well-field wells, monitoring wells, and regional aquifer PSWs is consistent with widespread detection of chloroform in other studies in the Edwards aquifer (Mahler and others, 2006; Musgrove and others, 2010), in other carbonate aquifers (Lindsey and others, 2009), and in many aquifers across the United States (Zogorski and others, 2006). This frequent detection in a variety of well types and land-use settings has been partly attributed to the recycling of chlorinated waters (Moran and others, 2002; Zogorski and others, 2006). Anthropogenic sources of chloroform include drinking-water treatment processes, leaking water and wastewater lines, septic systems, and recharge from irrigation with treated water. Public-water-supply sources also include the possible use of treated water as a downhole well-pump lubricant. Natural sources of chloroform include volcanic gases, biomass burning, marine algae, and soil microorganisms (Ivahnenco and Barbash, 2004).

In addition to chloroform, other THMs that are also a byproduct of drinking-water treatment processes—bromoform, dibromochloromethane, and bromodichloromethane—were frequently detected (80, 60, and 60 percent, respectively) in the well-field samples (fig. 25). The median bromoform concentration for the well-field samples (0.756 $\mu\text{g/L}$) was higher than the median concentration of any other organic constituent in any other sample category (fig. 26). Higher concentrations of brominated THMs relative to chloroform is unusual, as chloroform is the THM produced in the highest concentrations during chlorination (Ivahnenco and Barbash, 2004). Ivahnenco and Barbash (2004) indicate that the formation of brominated THMs might occur in the presence of unusually high bromide concentrations, which is not the case for the Edwards aquifer. The source of bromoform to the well-field samples is likely local to the well field but is currently unknown.

Groundwater Age Tracers and Age Distributions

Groundwater age estimates have been used to determine recharge rates, estimate rates of geochemical and microbiological processes, calibrate groundwater flow models, evaluate aquifer susceptibility to contamination, and improve water-resource management (Plummer, 2005). Groundwater age estimates are based on the measurement of a dissolved solute, gas, or isotope in groundwater that has been transported through the aquifer to the point of sampling. As such, groundwater age estimates are model dependent, based on an interpretation of recharge and flow mechanisms, and might be affected by a variety of physical and geochemical processes (Cook and Böhlke, 2000; Plummer, 2005).

Age tracers—concentrations of CFCs (CFC-11, CFC-12, and CFC-113), SF_6 , ^3H and the ratio of the parent/daughter

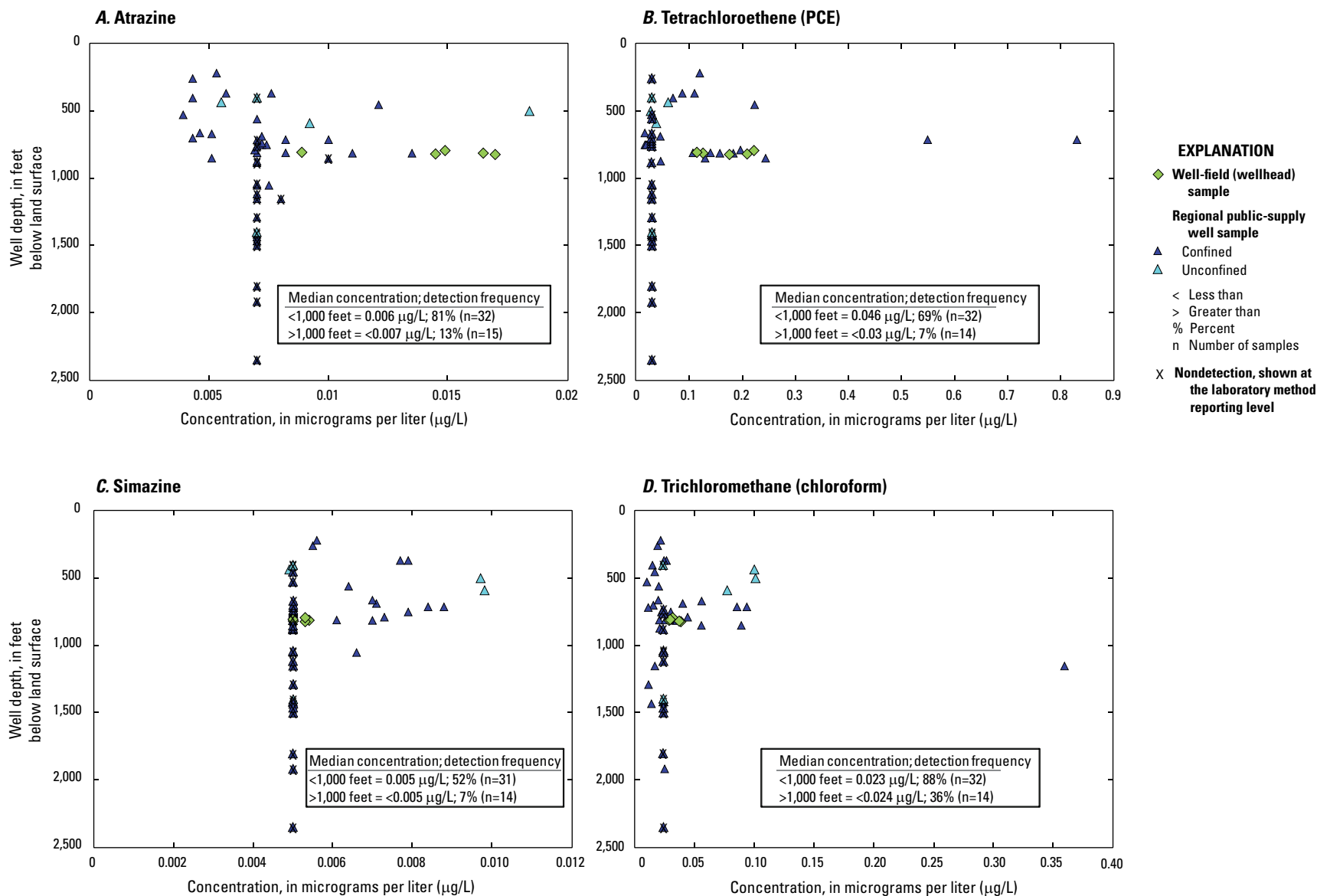


Figure 27. Relation between well depth and concentration for selected organic contaminants for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9. *A*, Atrazine. *B*, Tetrachloroethene (PCE). *C*, Simazine. *D*, Trichloromethane (chloroform).

isotopes $^3\text{H}/^3\text{He}$, and dissolved gases (methane, CO_2 , He, N_2 , oxygen, argon, and Ne)—were measured for most of the well-field and monitoring well samples, as well as for the W4 depth-dependent samples. Groundwater age tracers were not measured on samples from the regional aquifer PSWs, but previous NAWQA studies have measured a variety of age tracers in shallow/urban unconfined groundwater and unconfined groundwater samples, as well as ^3H in confined groundwater from the regional aquifer (Musgrove and others, 2010). Those results, with a median apparent groundwater age of 20 years (based on a piston-flow model interpretation), indicate that the Edwards aquifer is dominated by “young” or “modern” water, with extensive mixing of groundwater, likely of different ages. “Young” groundwater is defined as water recharged within approximately the past 50 years, the time scale of which is largely defined by the dating range of applicable tracers (Plummer and Friedman, 1999; Plummer and Busenberg, 2005). “Modern” groundwater has been similarly defined; that is, as water that has recharged within the past few decades in a system with an active hydrologic cycle (Clark and Fritz, 1997). Similar to Musgrove and others (2010), this report considers modern groundwater to represent samples with apparent ages that belong to the present day; that is, within 10 years of the time of sampling. This report considers “old” groundwater as that recharged prior to 50 years ago. Previous studies of age tracers and dye-tracing tests in the Edwards aquifer have documented a range of possible groundwater travel times and groundwater mixing, largely dominated by young and modern water (Pearson and others, 1975; Campana and Mahin, 1985; Ogden and others, 1986; Johnson and Schindel, 2008; Johnson and others, 2010). Apparent groundwater age in the Edwards aquifer has been demonstrated to vary temporally as a result of changes in hydrologic conditions; that is, variability in hydrologic conditions results in different mixtures of groundwater of different apparent ages (Musgrove and others, 2010). Temporal variability in apparent age results for karst groundwater in response to changes in hydrologic conditions has also been demonstrated in the Floridan aquifer (Katz, 2004; Happell and others, 2006). These results indicate that a one-time sampling of groundwater age tracers in karst might not be representative over a range of hydrologic conditions.

Because of contamination effects and age interpretation complexities (which are described in more detail below), age interpretations were based largely on $^3\text{H}/^3\text{He}$ results, even though other age tracers were measured. Thus, a comparison of independent apparent ages determined by different dating tracers was not possible. Interpretation of age-tracer results were complicated by some natural background or low-level anthropogenic contamination of groundwater samples by non-atmospheric sources of some tracers; that is, the measured tracer concentration in the groundwater sample was greater than that theoretically possible, on the basis of equilibrium of the groundwater sample with the maximum atmospheric concentration. As a result, most of the age-dating tracer results did not yield resolvable apparent groundwater ages. These results are consistent with previous efforts to interpret results

of groundwater age tracers in the Edwards aquifer, which indicated that low-level contamination issues in the aquifer are common, in particular with respect to CFC and SF_6 tracers (Thompson and Hayes, 1979; Musgrove and others, 2010). Contamination from CFCs might be particularly associated with urban areas (Musgrove and others, 2010). Possible urban CFC sources include septic tanks, leaking sewerlines and storage tanks, sewage treatment plant effluent, and industrial wastes (Plummer and Busenberg, 2005). An early study of CFCs identified a plume of CFC-11 in the Edwards aquifer from a proposed point source in the urban San Antonio area to the northeast along a regional flow gradient (Thompson and Hayes, 1979). Age-tracer results for all samples analyzed for SF_6 and all but one location analyzed for CFCs indicated that samples were contaminated with excess tracer and that age interpretations were not resolvable (table 18 at end of report) from these compounds. CFC results for one of the well-field samples (W5) yielded independent age interpretations based on CFC-11, CFC-12, and CFC-113 tracers (table 18). However, these resolvable ages might still be affected by excess CFC from widespread low-level aquifer contamination and, as a result, might still be questionable.

Interpreting groundwater ages with ^3H is qualitative, but ^3H concentrations can provide insight into distinguishing between prebomb and postbomb water based on the history of ^3H input to the atmosphere from nuclear weapons testing. Based on its relatively short half-life (12.3 years), the presence of ^3H in groundwater is indicative of recent groundwater recharge (Clark and Fritz, 1997; Plummer, 2005). The input of ^3H to the atmosphere as a result of nuclear testing dominated the concentrations of ^3H in rainfall for about 50 years, with concentrations increasing above natural levels beginning in the early 1950s and peaking in the 1960s. Concentrations of ^3H in recent rainfall (since 2006) have globally decreased to approximately prebomb background levels (Clark and Fritz, 1997; Phillips and Castro, 2003; R.L. Michel, U.S. Geological Survey, written commun., 2009). Concentrations of ^3H in the well-field and monitoring well samples (excluding the overburden well) (2007–8; $n=19$) ranged from less than 1 to 2.23 ^3H units (TU), with a median of 1.97 TU (table 19 at end of report). An additional sample collected from the shallow Zarzamora monitoring well (Z–SED) in 2009 had a ^3H concentration of 1.74 TU. The overburden monitoring well sample (Z–OVB) had a ^3H concentration of 1.30 TU. Tritium concentrations showed no relation with well depth or sampling depth. Groundwater ^3H concentrations were compared to a local ^3H input function derived from a combination of ^3H concentrations measured in rainfall at Waco, Tex. (1961–87) and ^3H concentrations measured in rainfall at Vienna, Austria (1961–2006) (International Atomic Energy Agency/World Meteorological Organization, 2008). Results for ^3H concentration from this study and from Musgrove and others (2010) are indicative of dominantly young and modern groundwater throughout the Edwards aquifer.

The decay of ^3H produces the stable isotope ^3He ; the ratio of parent to daughter isotopes is used to estimate the time

since recharge (Plummer and others, 1993). Tritium decay, however, is not the only possible source of ^3He , and determining the amount of ^3He from the decay of ^3H (tritogenic ^3He , or $^3\text{H}_{\text{trit}}$) requires an assessment of other possible ^3He sources. Excess air, introduced to groundwater when entrapped air bubbles dissolve, is common in groundwater and is usually the source of most additional ^3He (Busenberg and Plummer, 2000). Ne concentrations are used in excess-air correction calculations. Helium can also be introduced to groundwater from terrigenous sources (He_{terr}) because it is produced by the decay of uranium-series (U-series) radionuclides. Age calculations are sensitive to the amount of terrigenous helium-4 ($^4\text{He}_{\text{terr}}$). $^4\text{He}_{\text{terr}}$ (relative to the total ^4He) for samples which yielded resolvable apparent ages ranged from near 0 to 20.3 percent, with an average value of around 7 percent (table 19). Two samples had negative values for $^4\text{He}_{\text{terr}}$, which might indicate a loss of ^3He and yield an older estimated age. $^4\text{He}_{\text{terr}}$ values for both of these samples, however, were close to 0, and the negative values might also reflect analytical uncertainties for very young samples with low $^4\text{He}_{\text{terr}}$ values. For these two samples, age calculations were not corrected for $^4\text{He}_{\text{terr}}$, and calculated apparent ages were relatively young (0.8 and 3.4 years). A large fraction of $^4\text{He}_{\text{terr}}$ (greater than 10 percent) is an additional source of uncertainty in age calculations, which are dependent on the $^3\text{He}/^4\text{He}$ ratio of the He_{terr} (Plummer, 2005). The highest $\Delta^4\text{He}_{\text{terr}}$ values ($^4\text{He}_{\text{terr}}$ as a percentage of measured He) for samples with resolvable ages were associated with the Timberhill monitoring wells (17.5 and 20.3 percent). There

was no significant relation between $\Delta^4\text{He}_{\text{terr}}$ and U concentration in the water samples, and U concentrations in well-field well and monitoring well samples had a narrow range (0.80–1.17 $\mu\text{g/L}$). Terrigenous ^4He likely accumulates from U-series decay in aquifer host carbonates where U is present in low concentrations (up to 1.7 mg/kg) (table 5) or from migration from other sources such as deep crustal rocks (Hunt and others, 2010). Values of $\Delta^4\text{He}_{\text{terr}}$ (percent) increased with interpreted age (Kendall’s tau = 0.89) (fig. 28), which suggests that $\Delta^4\text{He}_{\text{terr}}$ values might provide an indicator of relative residence time in this aquifer (Hunt and others, 2010).

Results for $^3\text{H}/^3\text{He}$ measurements were used to estimate groundwater ages in 15 samples, including those from the monitoring wells (n=5), well-field (n=4) and depth-dependent sampling from W4 (n=6). Apparent ages (summarized in table 19) were not resolvable for several samples, including the well field sample (W5) that had CFC results. Measured ^3H concentrations in the overburden well (1.30 TU) and in one of the W4 depth-dependent samples (less than 1 TU) were too low to obtain a corresponding $^3\text{H}/^3\text{He}$ interpreted age. Results for several additional samples did not yield resolvable age interpretations because of effects of probable degassing during sample collection (negative $\Delta^4\text{He}_{\text{terr}}$ or ΔNe values) (table 19).

An evaluation of geochemical age-dating tracers in a complex karst system such as the Edwards aquifer requires consideration of flow-system dynamics. Lumped-parameter models were used to represent different conceptual flow-system characteristics, specifically, a piston-flow model,

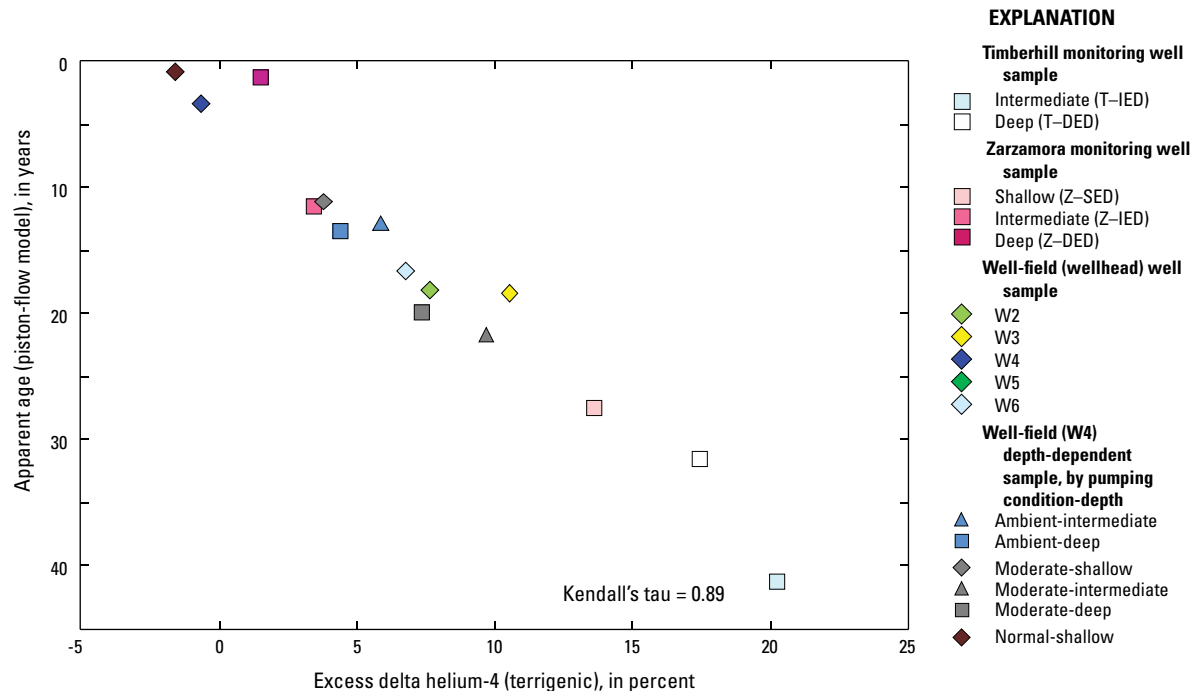


Figure 28. Relation between apparent age and excess delta helium-4 (terrigenous) for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9.

an exponential mixing model, and several combinations of binary mixing models of older water with younger water. A piston-flow model, which is the simplest and most commonly assumed transport model, assumes that the tracer concentrations were not altered by transport processes (Plummer and Busenberg, 2005). This simplified approach assumes that no mixing or dispersion affects the tracer concentration and that all groundwater flow lines have similar velocity. A piston-flow model likely does not adequately represent a mixed groundwater sample with a range of ages, which is especially valid for karst aquifers (Long and Putnam, 2006). In spite of these uncertainties in age interpretation, piston-flow model ages provide an initial reference point for comparison of different age tracers and time scales of flow processes in the aquifer. An exponential mixing model describes a hydrologic transport model in which the mean residence time of groundwater is exponentially distributed; that is, groundwater is a mixture of recharge from past years. Groundwater contributions to well or spring discharge would be dominated by the most recent recharge and decrease exponentially in proportion to older recharge. An exponential mixing model might represent a reasonable approximation of a homogeneous unconsolidated aquifer but, like the piston-flow model, is likely not an adequate representation of a karst aquifer, where traveltimes can vary widely between matrix-flow-dominated and conduit-flow-dominated endmembers. The mean age of the exponential mixture was varied in evaluating exponential models. Binary mixing models assume that groundwater is composed of a mixture of two endmembers of different ages, for example relatively young (modern) water (recharged within the past 10 years) mixed with relatively older water (recharged several or more decades ago with little to no detectable tracer concentrations). Lumped-parameter models of binary mixtures were evaluated for a variety of young and old age combinations; for example, piston-flow models for both endmembers, a piston-flow model for the old endmember and an exponential flow model for the young endmember, and an exponential flow model for the old endmember and a piston-flow model for the young endmember. For modeling purposes, both endmembers of a binary mixture can be any age, but for purposes of this effort it was assumed that one endmember was modern (that is, recharged within the past zero to 5 years) and the other was young to old (that is, recharged 30–70 years ago, which includes that time period for which there would be little to no detectable tracer concentrations). For binary mixtures of modern and old water, the $^3\text{H}/^3\text{He}$ method is particularly useful for determining the age of the younger fraction; the isotopic ratio is nearly independent of dilution with old ^3H -free water (Plummer and others, 1998). Plots of age-frequency distributions based on tracer-tracer concentrations for collected samples relative to theoretical curves for different lumped-parameter models can provide insight into the distribution of apparent ages in an aquifer, age mixtures at wells, and the representativeness of different hydrogeologic concepts of the flow system.

Lumped-parameter-model results for different hydrogeologic conceptual models and different mixtures of age combinations did not clearly indicate that any one conceptual hydrogeologic model can account for the range of measured age-tracer concentrations. For the tracer-tracer plot showing the relation between ^3H and $^3\text{H}/^3\text{H}_0$, most of the samples plot above the piston-flow-model curve and slightly above the exponential model curve (fig. 29). The sample with the lowest $^3\text{H}/^3\text{H}_0$ ratio and oldest interpreted age (T–IED) plots well below the piston-flow-model curve but close to the exponential model curve. The $^3\text{H}_0$ value is the sum of the measured concentrations of ^3H and $^3\text{He}_{\text{trit}}$ (tritogenic ^3H , derived from the decay of ^3H) and represents the initial ^3H concentration at the time of recharge. The ratio of tritium to initial tritium ($^3\text{H}/^3\text{H}_0$) is a proxy for the relative age of water; values near zero represent older water, whereas a value approaching 1 represents younger water (Böhlke, 2002). $^3\text{H}/^3\text{H}_0$ values are correlated with piston-flow groundwater ages (Kendall's tau = -1.0) (fig. 30). The W4 depth-dependent ambient-deep sample deviated from the trend of this relation relative to the rest of the samples; the cause of this deviation is unclear. Other lumped-parameter models shown on tracer-tracer plots (figs. 29, 31, and 32) are for an exponential model with a mean groundwater model age of 3 years, a combined piston-flow/exponential-flow model, and binary mixture models with modern and young endmember ages of 0.5 and 30 years (binary mix 1 with piston-flow models for both endmembers; binary mix 2 with an exponential-flow model for the modern endmember and a piston-flow model for the young endmember; and binary mix 3 with a piston-flow model for the modern endmember and an exponential-flow model for the young endmember). These potential mixture ages are consistent with particle-track results from a calibrated groundwater-flow model for the local-scale study area, described by Lindgren and others (2011b). These modeling results indicate that approximately 48 percent of particle tracks to the study PSW are within a 0- to less than 2-year range of traveltimes, an additional 40 percent are within the 2- to less than 5-year range of traveltimes, an additional 8 percent are within the 5- to less than 10-year range of traveltimes, and 4 percent have a greater than 10 year traveltime (fig. 33). While different lumped-parameter models might account for different samples on the basis of tracer-tracer plots (fig. 29), no prevailing model accounts for the range of age-tracer results and consistently seems to describe the flow system.

An exponential-flow mixing model can be used to estimate the mean age of groundwater in a sample; the W4 well-field sample and the depth-dependent samples for normal pumping conditions fall close to the exponential-flow model (fig. 29). A comparison of model estimates with $^3\text{H}_0$ values suggests that samples from the monitoring wells and well field compose a large range of mean ages, from close to 0 to approaching 50 years (fig. 34A). A similar comparison of ^3H concentrations with model-estimated mean ages does not allow for distinguishing different mean ages between samples as a result of the narrow range of ^3H concentrations and the convergence of model curves for the narrow range of sample

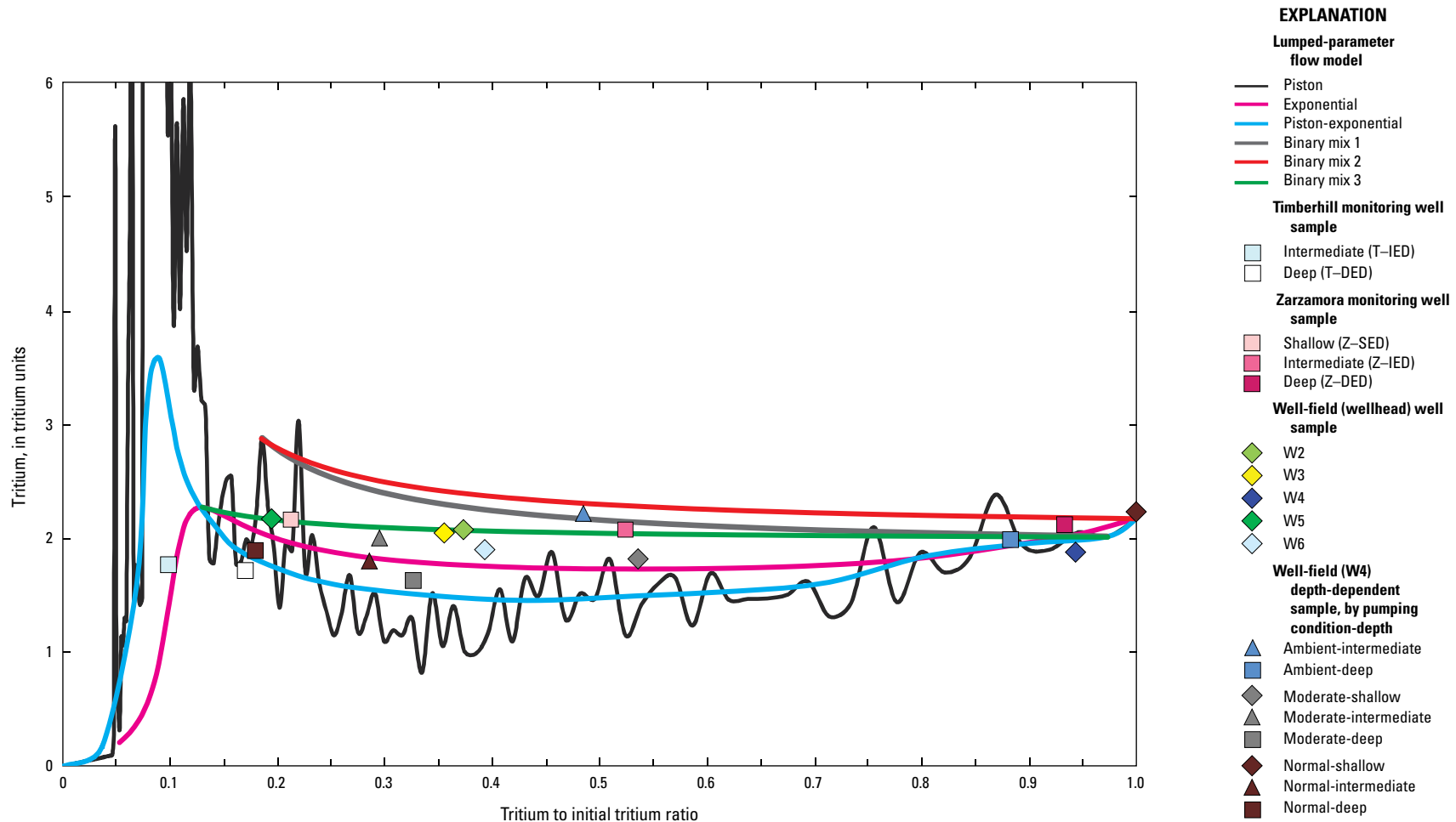


Figure 29. Relation between tritium (^3H) concentration and ^3H to initial tritium ($^3\text{H}_0$) ratio ($^3\text{H}/^3\text{H}_0$) for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9, and lumped-parameter modeled curves for various groundwater age distributions.

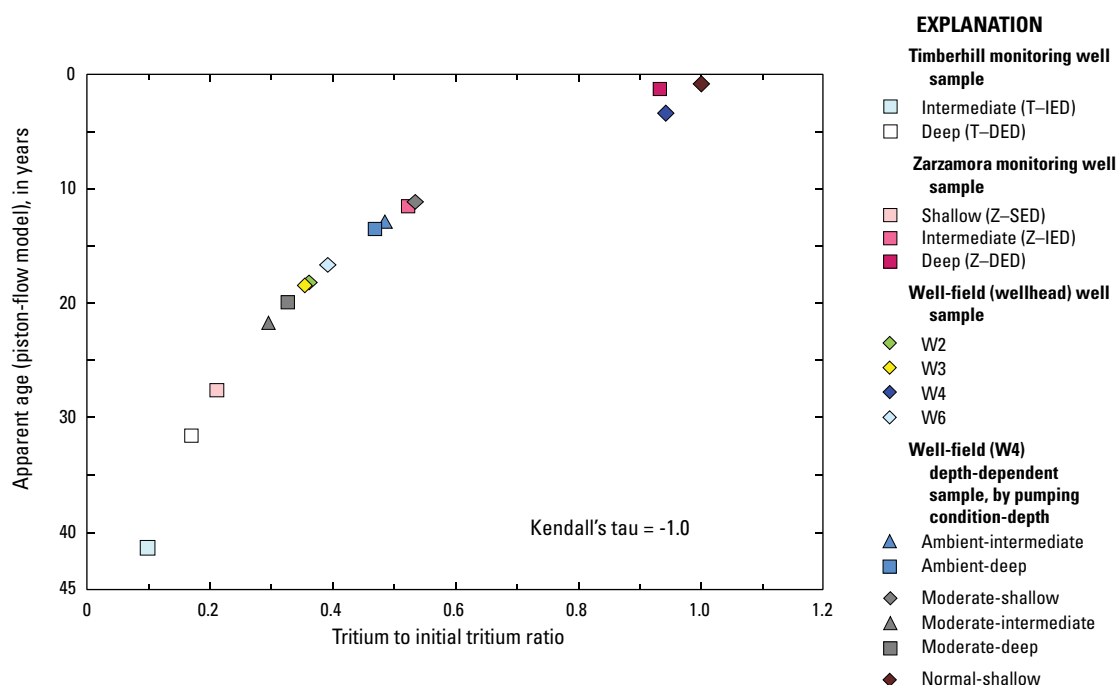


Figure 30. Relation between apparent age and the tritium (^3H) to initial tritium ($^3\text{H}_0$) ratio for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9.

collection dates (fig. 34B). Based on wellbore flow characteristics, flow to W4 under normal pumping conditions (with multiple wells in the well field pumping) is sourced from multiple stratigraphic units in the Edwards section (fig. 7). The mean exponential model age for the W4 well-field sample, which is assumed to represent conditions of water quality for water produced from the PSW, was younger than the other wellhead samples and most of the depth-dependent samples, with a model (apparent) age of approximately 3 years.

Tracer-tracer plots for CFC concentrations show mean results for only the single well-field PSW (W5) for which CFC concentrations were not contaminated for age-tracer purposes. CFC concentrations for CFC-11 relative to CFC-12 and for CFC-11 relative to CFC-113 indicate that tracer concentrations might have been affected by tracer contamination or degradation because sample concentrations plot far outside the area of different flow models (fig. 32). CFC degradation, which can occur in anoxic environments, is unlikely because of the generally oxic condition of the aquifer (Plummer, 2005). The widespread CFC contamination throughout the aquifer (this study; Musgrove and others, 2010) indicates that potential CFC contamination of the W5 well is also plausible. On the tracer-tracer plot for CFC-12 relative to CFC-113, the W5 sample plots in between an exponential-flow model curve and a combined piston-exponential model curve (fig. 32). Results

for the W5 sample on tracer-tracer plots for CFC-12 and CFC-113 concentrations relative to the measured ^3H concentration (fig. 31) are generally consistent with either the piston-flow or exponential-flow lumped-parameter models. It is difficult, however, to evaluate any particular flow model based on a single sample, and results for sample W5 are not consistent with any single lumped-parameter model. These results indicate that apparent ages on the basis of CFC concentrations for the sample from W5 should be interpreted with caution. Piston-flow-model interpreted ages for W5 (mean) were 38.7, 30.9, and 26.4 years based on CFC-11, CFC-12, and CFC-113 tracers, respectively. An exponential model yields a mean age of approximately 17 years for this sample based on the $^3\text{H}_0$ concentration (fig. 34), although this approach is not ideal because a $^3\text{H}/^3\text{He}$ age was not resolvable for this sample.

A comparison of model ages provides insight into flow-system dynamics supplying water to the well field. Piston-flow-model groundwater ages based on $^3\text{H}/^3\text{He}$ results for all samples, including depth-dependent samples (which are discussed in more detail in the subsequent section), range from 0.8 to 41.3 years, with a median of 16.6 years. Piston-flow-model groundwater ages (fig. 35) are inversely stratified for both sets of monitoring wells, with younger ages for the deeper monitoring wells. The oldest apparent ages are for the Timberhill monitoring wells (41.3 and 31.6 years,

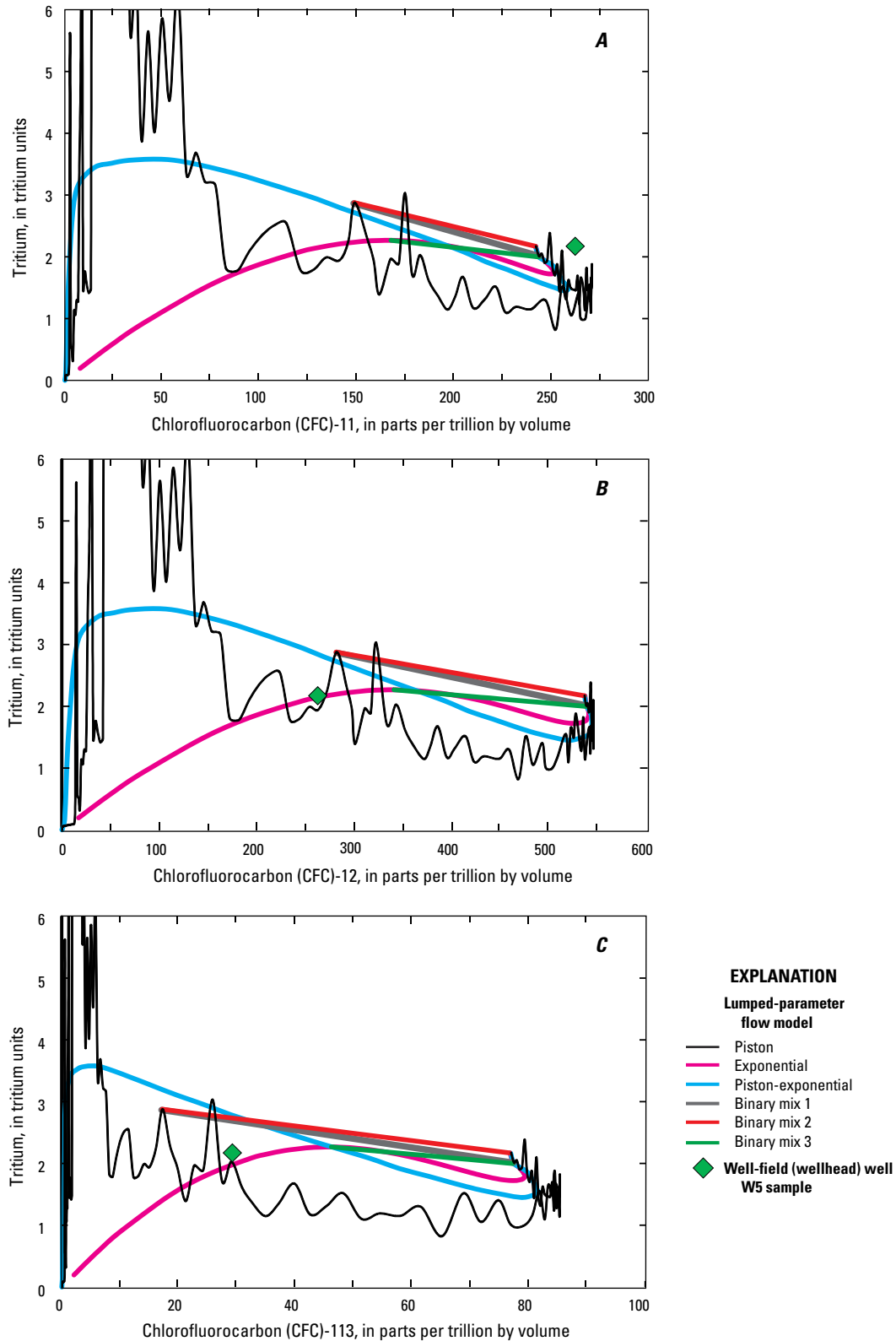


Figure 31. Relation between tritium concentration (^3H) and chlorofluorocarbon (CFC) age tracers for groundwater sample collected from the well-field well, W5, for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9, and lumped-parameter modeled curves for various groundwater age distributions. A, Relation between tritium and CFC-11. B, Relation between tritium and CFC-12. C, Relation between tritium and CFC-113.

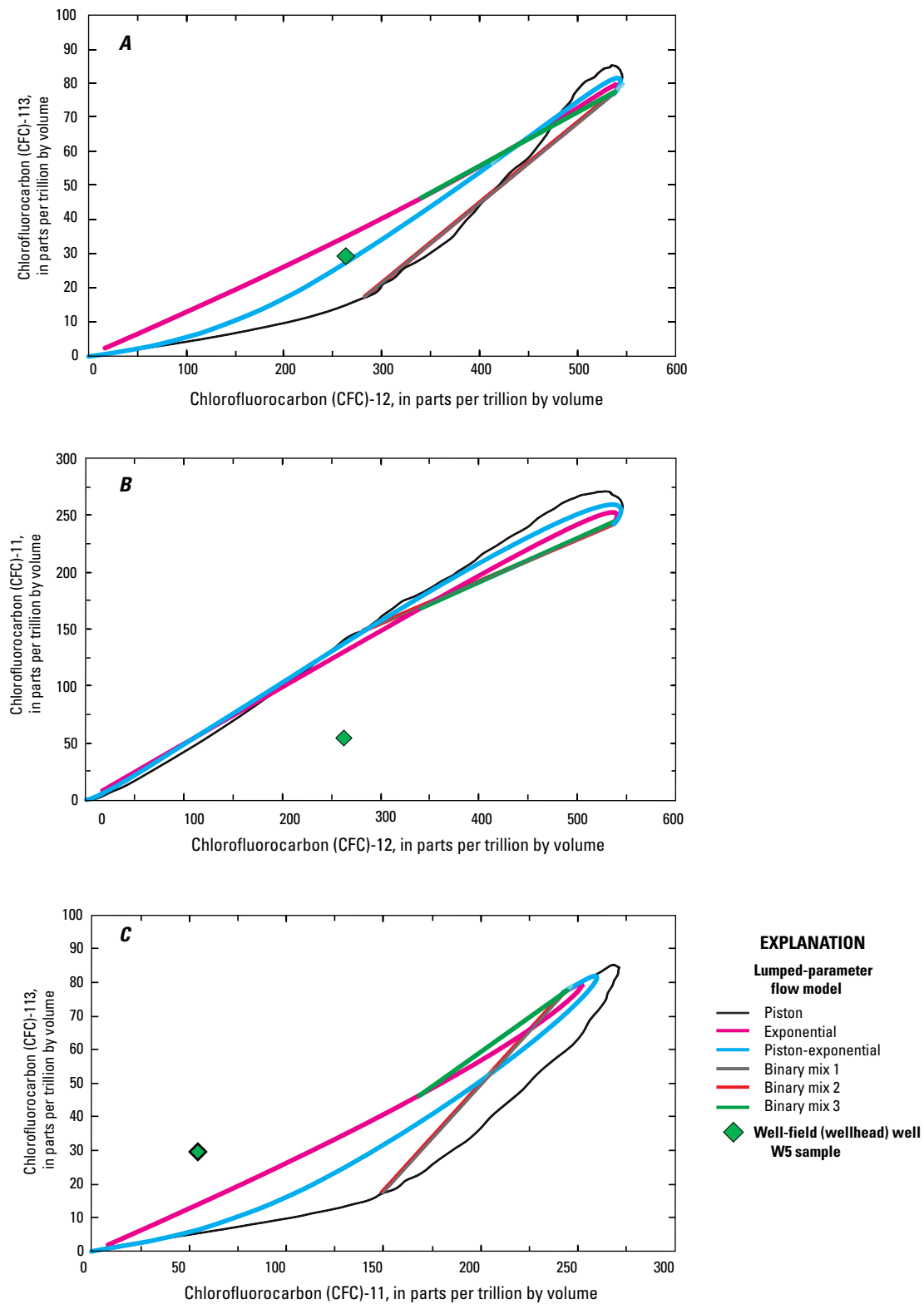


Figure 32. Relation between chlorofluorocarbon (CFC) age tracers (CFC-11, CFC-12, and CFC-113) for groundwater sample collected from the well-field well, W5, for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9, and lumped-parameter modeled curves for various groundwater age distributions. *A*, Relation between CFC-12 and CFC-113. *B*, Relation between CFC-12 and CFC-11. *C*, Relation between CFC-11 and CFC-113.

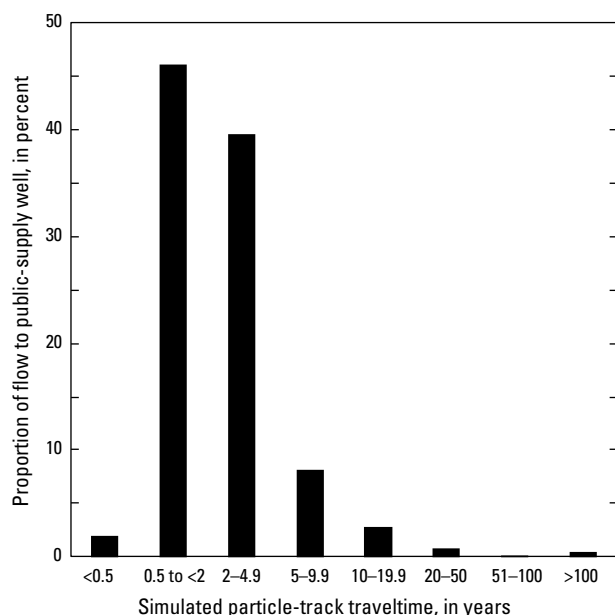


Figure 33. Particle-track age distribution to selected public-supply well for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, based on local-scale study model results described by Lindgren and others (2011b) in the San Antonio segment of the Edwards aquifer, south-central Texas.

respectively, for the intermediate and deep monitoring wells). Apparent ages for the Zarzamora monitoring wells (excluding the overburden well) are quite young in the deepest well (1.3 years); the apparent age for the shallow Zarzamora monitoring well (27.5) approaches the deeper Timberhill monitoring well age. Apparent ages based on $^3\text{H}/^3\text{He}$ results for the well-field samples are similar (16.6–18.4 years), with the exception of W4, which yielded a much younger age (3.4 years). Depth-dependent samples had apparent ages ranging from 0.8 to 21.7 years, but the ages were not consistently related to depth or to pumping condition (fig. 35; table 19). These results are consistent with extensive mixing in this complex karst hydrogeologic system. The range of piston-flow-model ages for TANC study samples is similar to the range reported for Edwards aquifer unconfined groundwater samples by Musgrove and others (2010) of 1 to 52 years and with a median of 20 years.

Selected geochemical constituents were compared with the ratio of $^3\text{H}/^3\text{H}_0$, which, as noted earlier, is a proxy for the relative age of water (with relative ages becoming younger as $^3\text{H}/^3\text{H}_0$ values approach 1) (Böhlke, 2002). Inorganic geochemical parameters compared to $^3\text{H}/^3\text{H}_0$ included specific conductance, Mg/Ca molar ratios, $^{87}\text{Sr}/^{86}\text{Sr}$ values, and dolomite SI. Specific conductance generally increases with increased groundwater residence time. Dolomite SIs, Mg/Ca ratios, and Sr isotopes have been demonstrated to be qualitative indicators of residence time in the Edwards aquifer (Musgrove and others, 2010). With longer residence time and greater extents of water-rock interaction, dolomite SIs approach equilibrium,

Mg/Ca ratios increase, and $^{87}\text{Sr}/^{86}\text{Sr}$ values decrease toward values similar to the host aquifer rocks. A comparison of these constituents with $^3\text{H}/^3\text{H}_0$ ratios, however, does not indicate a consistent relation with interpreted groundwater age (fig. 36). The relation of $^3\text{H}/^3\text{H}_0$ with specific conductance and Mg/Ca ratios was not statistically significant. The relation of $^3\text{H}/^3\text{H}_0$ with dolomite SIs and $^{87}\text{Sr}/^{86}\text{Sr}$ values was significant (Kendall's tau = 0.36 and -0.34, respectively). The relation was inverse, however, to that expected for these constituents with groundwater age: dolomite SIs approach equilibrium in younger apparent-age samples, and $^{87}\text{Sr}/^{86}\text{Sr}$ values approach values similar to the aquifer rocks in younger apparent-age samples. The relation of $^3\text{H}/^3\text{H}_0$ with dolomite SIs and $^{87}\text{Sr}/^{86}\text{Sr}$ values was strongly influenced by values for the Timberhill monitoring wells; if these data were not considered, the relations were no longer statistically significant.

Similar to inorganic constituents, anthropogenic organic constituents were not clearly related to apparent age. Concentrations and detections of anthropogenic contaminants are generally expected to be higher in younger groundwater. However, the relation of $^3\text{H}/^3\text{H}_0$ with the number of detections of organic contaminants and with atrazine, DEA, and PCE concentrations was not significant (fig. 37). The relation with other frequently detected organic contaminants, such as simazine and chloroform, also was not significant.

The absence of significant relations of inorganic geochemical indicators and anthropogenic constituents to apparent age indicates that mixing processes in this aquifer are complex and that hydrogeologic conceptual models used in groundwater age interpretations might not adequately account for mixing in this karst system. Age tracers in this aquifer provide insight into the young nature of the groundwater but might not readily allow for distinguishing the mixed water history or be well suited to distinguish relatively small differences in groundwater age as a result of mixing processes. This hypothesis is consistent with the relatively homogeneous geochemistry for many constituents in groundwater samples collected from the regional aquifer PSWs and from the well-field and monitoring wells.

Depth-Dependent Water Chemistry

Variations in depth-dependent geochemical results can provide insight into water quality of different parts of the aquifer that might contribute water to the PSW, as well as differences in water quality in the Edwards aquifer relative to overburden units. Variations in geochemical results from different pumping conditions can provide insight into the distribution of groundwater flow paths contributing water to the PSW. Previous studies indicate that groundwater flow in the Edwards aquifer is focused in highly permeable units and is affected by faulting (Abbott, 1975; Woodruff and Abbott, 1979; Maclay and Small, 1986; Maclay and Land, 1988; Sharp, 1990; Lindgren and others, 2011b).

As described earlier in the sections “Geophysical and Depth-Dependent Flow Measurements” and

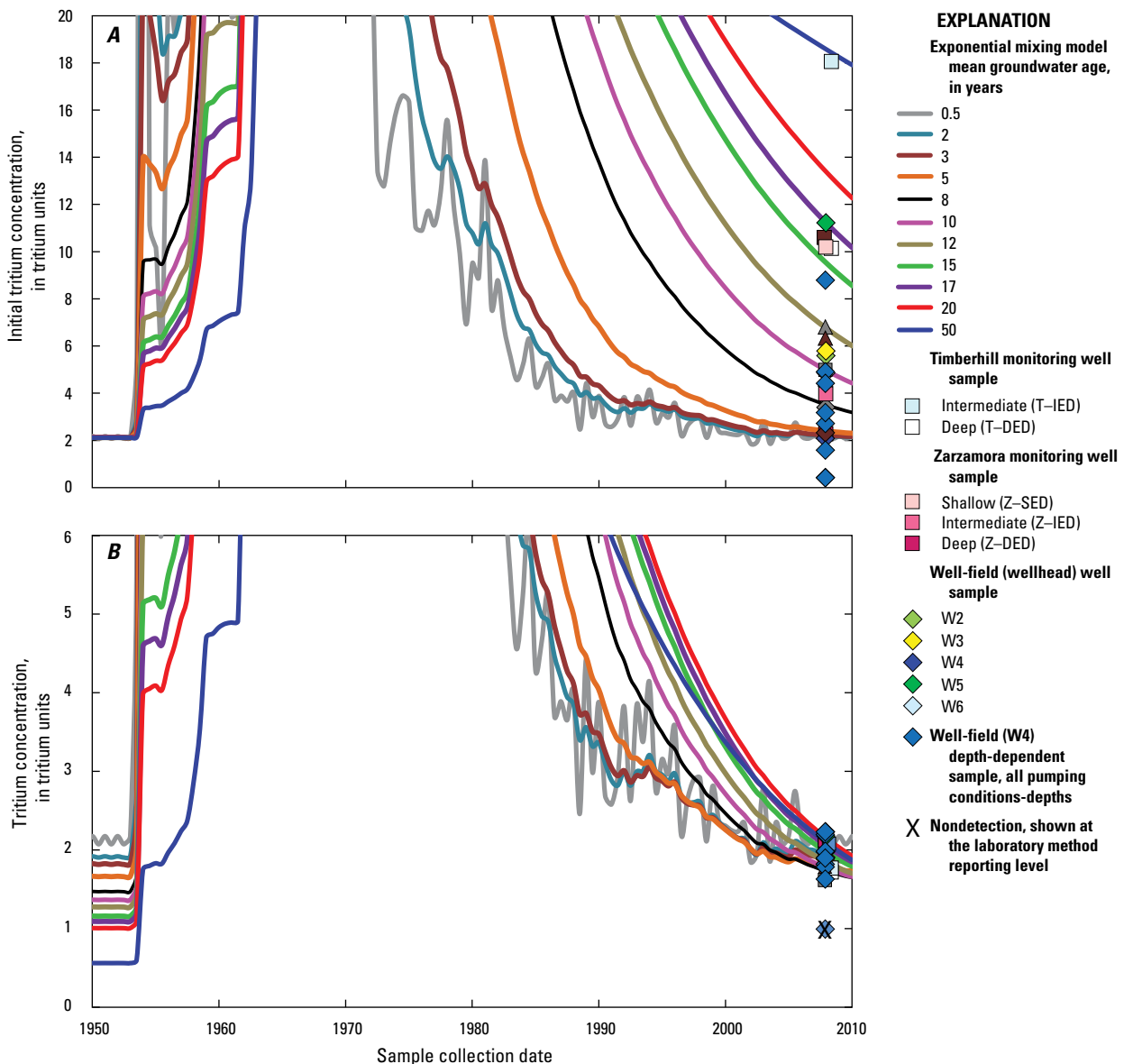


Figure 34. Relation between sample collection date and tritium for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9. *A*, Relation between sample collection date and initial tritium concentration ($^3\text{H}_0$). *B*, Relation between sample collection date and tritium concentration (^3H). Both *A* and *B* are shown with exponential mixing model based trendlines for expected $^3\text{H}_0$ and ^3H concentrations for selected mean groundwater ages.

“Depth-Dependent Samples,” flow measurements and groundwater samples were collected from W4 from three different depths (fig. 4) and under three different pumping conditions (ambient, moderate, and normal) to characterize the water quality of different parts of the aquifer that might contribute water to the PSW. The normal pumping condition is likely to be most representative of pumping conditions associated with the production of supply water and should have chemistry similar to surface discharge samples (wellhead samples) collected from the well field under normal production operations.

As noted earlier, it was not possible to collect a surface discharge (wellhead) sample from W4; a sample was instead collected with the sample pump directly near the bottom of the casing (at a depth of approximately 320 ft).

Flow measurement results indicated that the distribution of the flow of water entering W4 during normal pumping conditions was complex but likely largely originated from the Kirschberg evaporite member (HGU 8 and 9) and the combined leached and collapsed and cyclic and marine members (HGU 3-5), with some contribution from the lower dolomitic

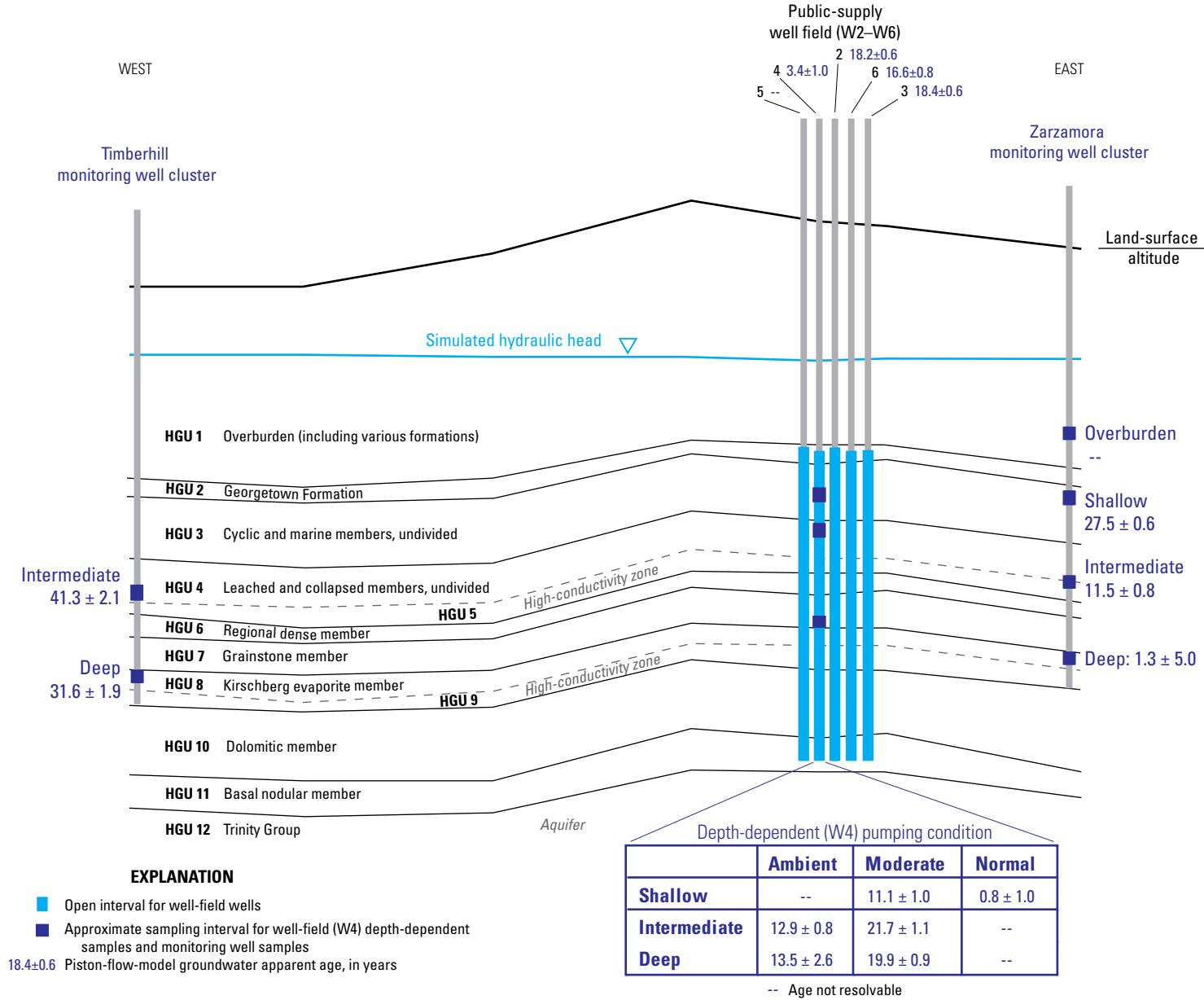
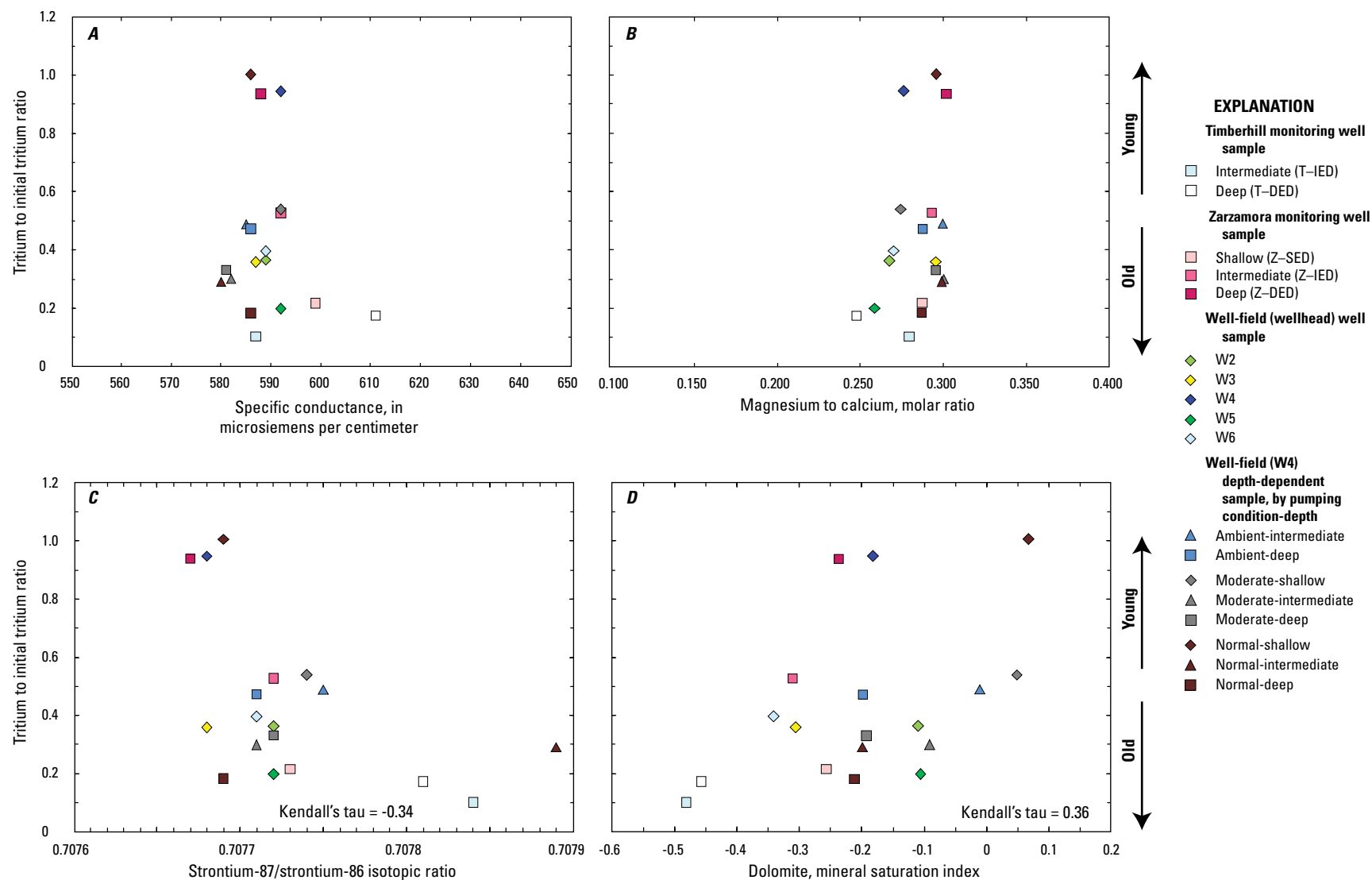


Figure 35. Schematic east-west (strike) cross section showing location of study area hydrogeologic units (HGUs) as modeled by Lindgren and others (2011b), relation of HGUs at selected sample sites, and summary of groundwater apparent-age results based on tritium/helium-3 tracer measurements (piston-flow model ages) near San Antonio, south-central Texas.



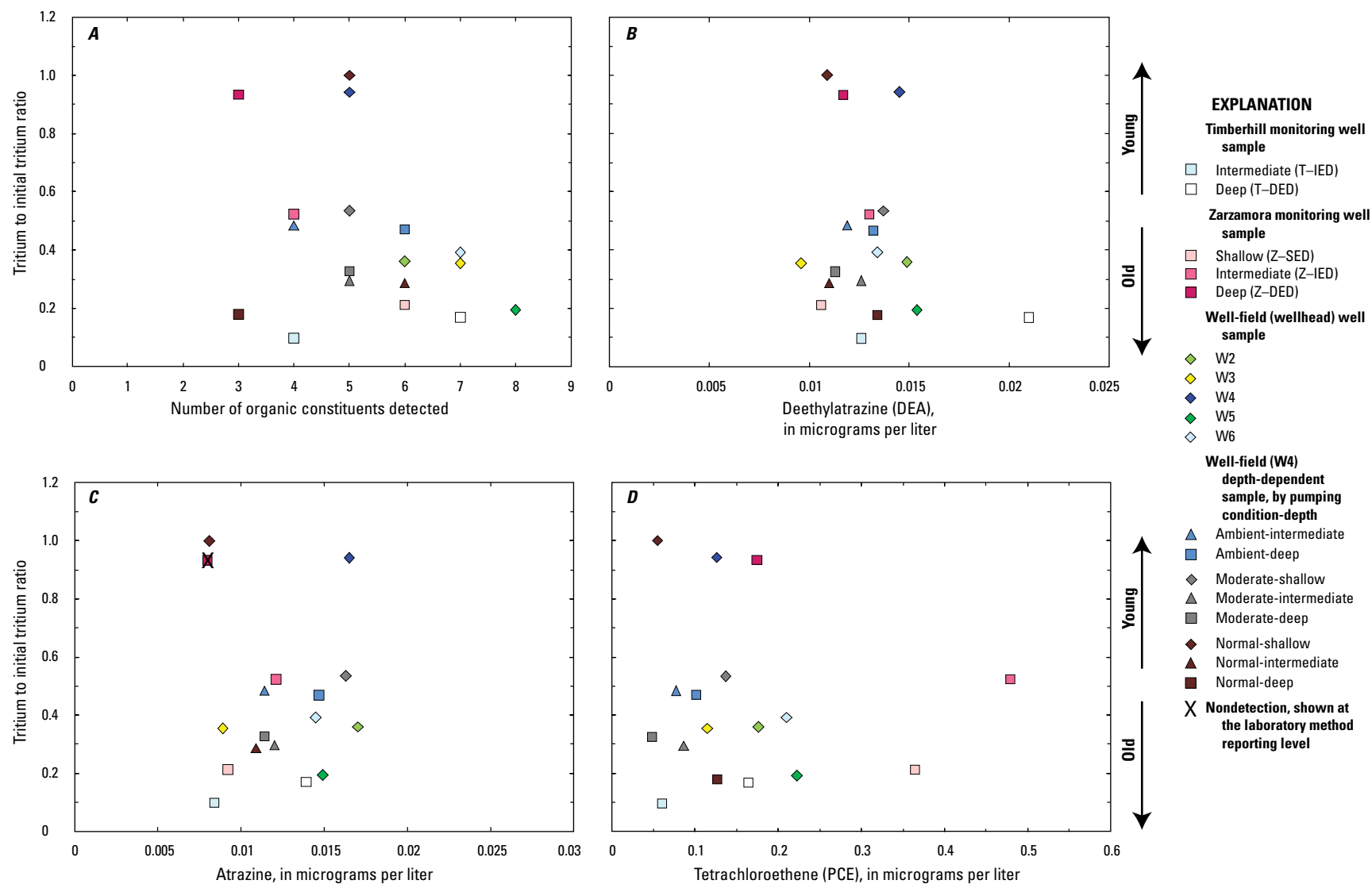


Figure 37. Relation between tritium to initial tritium ratio ($^3\text{H}/^3\text{H}_0$) and selected geochemical constituents for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9. *A*, Number of organic constituents detected. *B*, Deethylatrazine (DEA) concentration. *C*, Atrazine concentration. *D*, Tetrachloroethene (PCE) concentration.

member (HGU 10) (fig. 7). Under one pumping condition (described below), some loss of flow in the leached and collapsed members (HGUs 4 and 5) was observed. Intermediate depth-dependent samples were collected within the leached and collapsed members (HGUs 4 and 5), and the shallow depth-dependent samples were collected within the cyclic and marine members (HGU 3). The deep depth-dependent samples were collected from within the grainstone member (HGU 7) but are considered representative of water flowing upward in the wellbore from the Kirschberg evaporite member (HGUs 8 and 9) and dolomitic member (HGU 10).

Depth-dependent samples collected from W4 during normal pumping conditions indicated that, although much of the groundwater was coming from high-conductivity zones, the water quality throughout different HGUs was relatively homogeneous, indicative of substantial mixing in the aquifer. Geochemical results were compared for different sampling depths and flow conditions, as well as with samples from the monitoring wells. For comparison purposes between the depth-dependent samples and monitoring well samples, results were compared by HGU (fig. 38). A comparison by sample depth for the depth-dependent samples and well depth for the monitoring well samples is complicated by inferred faulting between the well-field and monitoring well locations. Faulting is inferred because of widespread occurrence in the study area and because of observed differences in depths to HGUs between the well-field PSWs and monitoring wells in the Zarzamora monitoring well cluster.

There were generally minor differences in physicochemical properties and chemical and isotopic constituents in depth-dependent samples (fig. 38); this observation is consistent with the hypothesis that water from the PSW is relatively well mixed and homogeneous regardless of pumping conditions. However, several constituents show some differences in depth profiles under different pumping conditions. The sample collected from the intermediate sampling interval (HGUs 4 and 5) under normal pumping conditions had a higher $^{87}\text{Sr}/^{86}\text{Sr}$ value (0.70789) than under ambient or moderate pumping conditions (fig. 38K). This value is higher than those for well-field samples, which had a narrow $^{87}\text{Sr}/^{86}\text{Sr}$ range of 0.70768 to 0.70772. The sample collected from the intermediate and deep sampling intervals also had a higher boron (B) concentration under normal pumping conditions than under ambient and moderate pumping conditions (fig. 38N). Sr isotope compositions in the regional aquifer generally decrease to lower values with increased residence time, whereas B concentrations tend to be higher in the confined part of the aquifer (Musgrove and others, 2010). Although the $^{87}\text{Sr}/^{86}\text{Sr}$ and B concentration results for these depth-dependent samples were different from the remainder of the W4 and monitoring well samples, they were within the range of variability for the regional aquifer (Musgrove and others, 2010). Samples from the Timberhill monitoring wells had similar elevated $^{87}\text{Sr}/^{86}\text{Sr}$ values (fig. 38K). Samples collected from the Timberhill monitoring wells also had lower dolomite SIs, and the sample collected from the deep Timberhill monitoring well (T–DED) had a

slightly lower Mg/Ca ratio. These results are consistent with the Timberhill monitoring well samples representing less geochemically evolved groundwater, although this hypothesis contrasts with apparent-age results for these samples (which had the oldest interpreted ages of 41.3 and 31.6 years for the intermediate and deep Timberhill monitoring well samples, respectively). When compared with the depth-dependent samples, results for the monitoring well samples show some minor differences in major ion concentrations, with slightly higher alkalinity and Ca concentrations and slightly lower SO_4 concentrations and dolomite SIs, although the values are within the range of variability for the regional aquifer (Musgrove and others, 2010).

PCE concentrations in samples from the shallow and intermediate Zarzamora monitoring wells (Z–SED and Z–IED) were elevated with respect to the W4 depth-dependent samples, as well as with respect to regional aquifer PSW samples (fig. 38W; fig. 26; table 14). As discussed earlier (in the section “Organic Constituents”), PCE concentrations in samples from W4 were higher than the median for those from regional (confined) aquifer PSWs. These results suggest that the shallower Zarzamora monitoring wells and W4 have some hydraulic connection with urban contaminant sources. The most notable geochemical variability for the depth-dependent and monitoring well samples is for results of age-tracer data and corresponding model (apparent) ages (fig. 38). As discussed earlier (in the section “Groundwater Age Tracers and Age Distribution”), the variability in apparent-age results and their relation with other geochemical constituents indicates that age interpretations might not adequately account for extensive mixing processes in this complex karst system.

The water sample collected from the overburden monitoring well (Z–OVB) was markedly different in geochemistry from W4 depth-dependent samples, well-field samples, and monitoring well samples, with higher specific conductance, higher alkalinity, and higher concentrations of major ions and trace elements, NO_3 , chloroform, and PCE and a different composition with respect to stable isotopes (fig. 38). These results indicate that there is likely not a substantial hydrologic connection between the overburden units and the Edwards aquifer at this location although, as noted earlier, higher concentrations of PCE in samples from the shallow and intermediate Zarzamora monitoring wells (Z–SED and Z–IED) are similar to the value measured in the sample from the overburden well.

Temporal Variability in Water Chemistry

Karst aquifers are characterized by extreme variability and rapid response to changes in environmental, climatic, and hydrologic conditions, which exert a fundamental control on groundwater geochemistry and water quality (for example, Hess and White, 1993; Desmarais and Rojstaczer, 2002; Vesper and White, 2004; Winston and Criss, 2004). Temporal variability in hydrologic conditions for the Edwards aquifer is apparent in data for rainfall, aquifer recharge, water-level

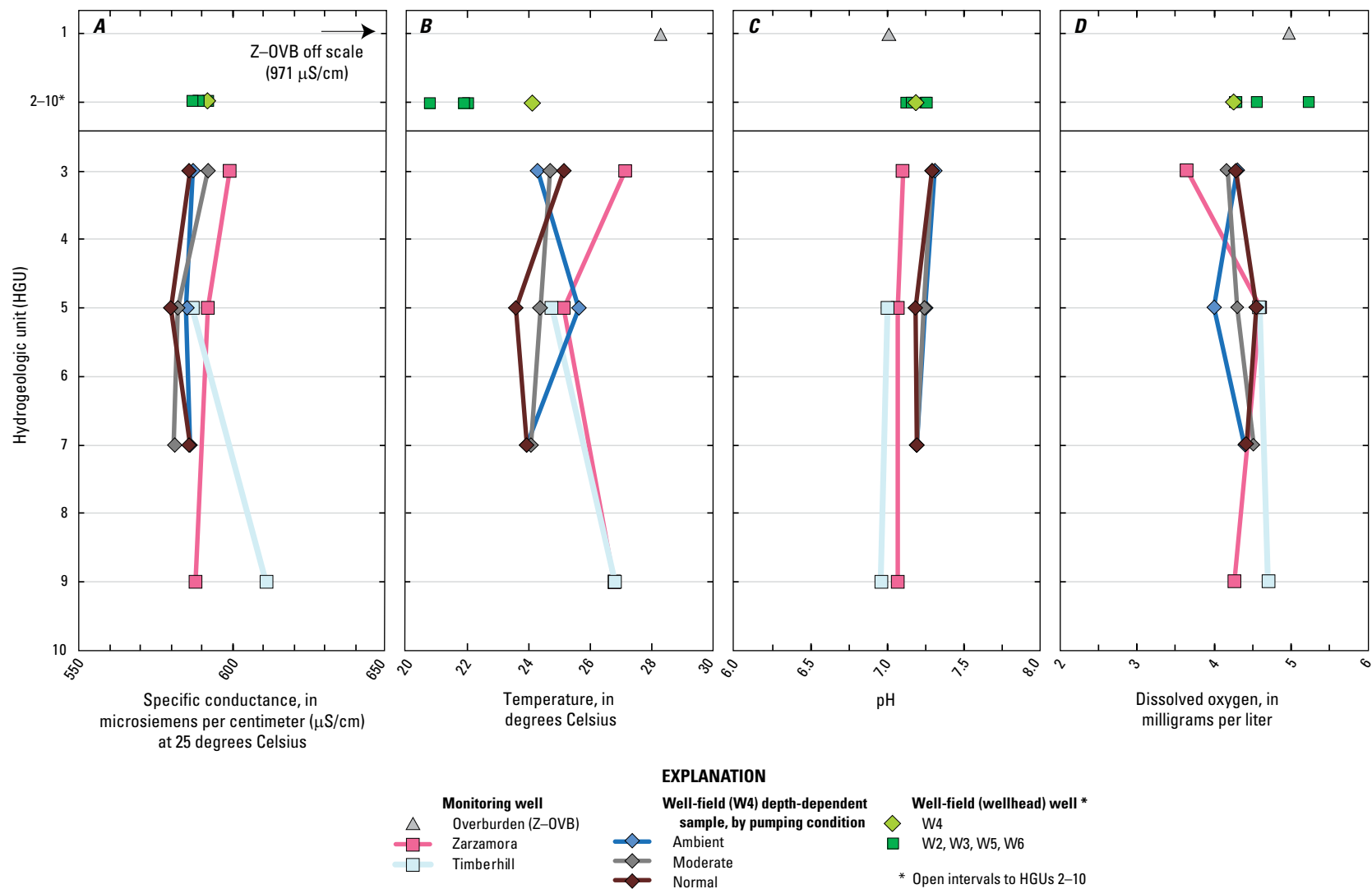


Figure 38. Geochemical results for selected constituents for groundwater samples collected from different depths and under different flow conditions for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas, 2004–9. Results are presented on the basis of associated hydrogeologic unit (HGU) for monitoring-well samples and public-supply well depth-dependent samples; results for well-field samples and the overburden monitoring well are shown for comparison. *A*, Specific conductance. *B*, Temperature. *C*, pH. *D*, Dissolved oxygen concentration.

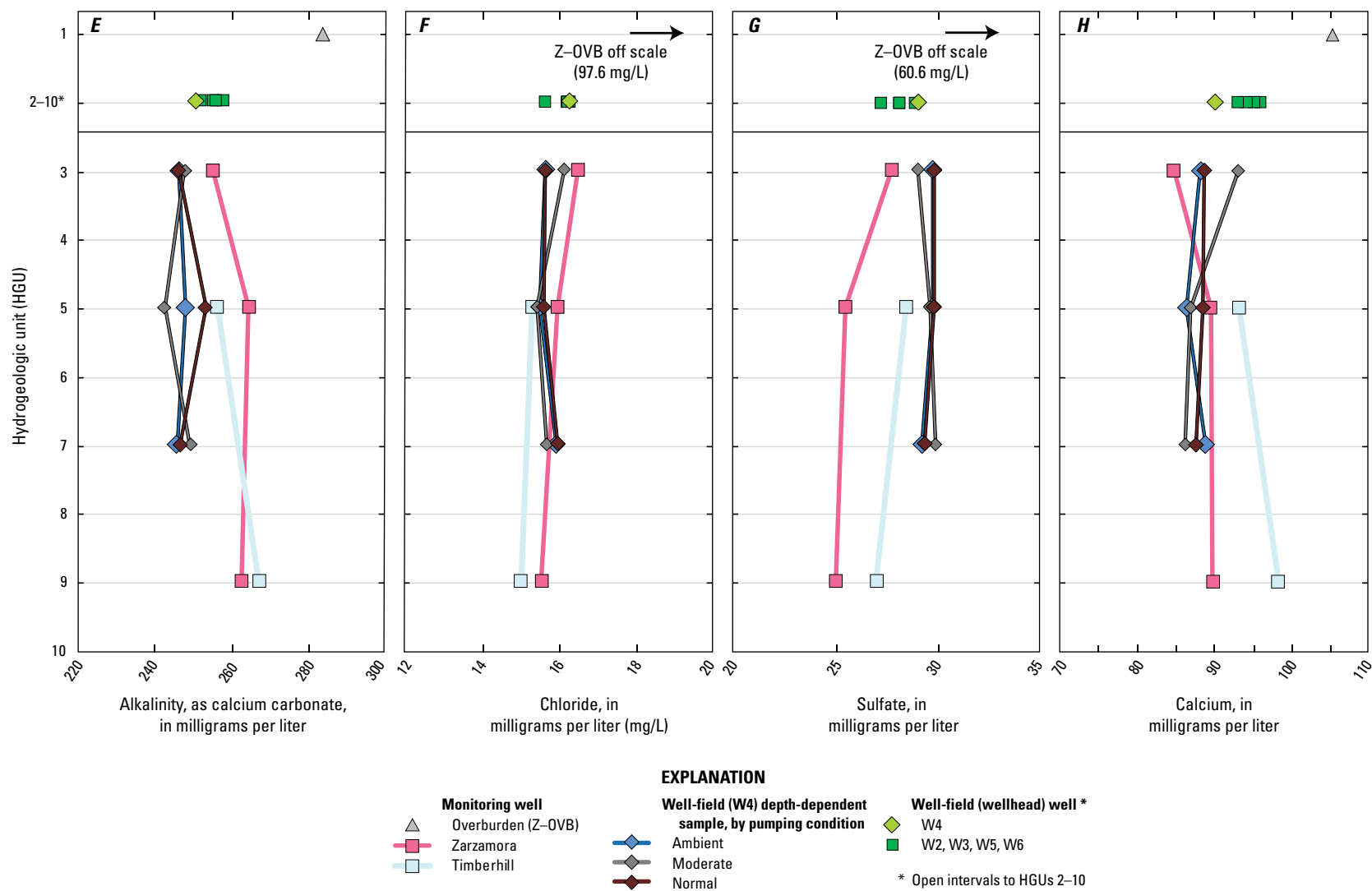


Figure 38. *E*, Alkalinity concentration. *F*, Chloride concentration. *G*, Sulfate concentration. *H*, Calcium concentration.—Continued

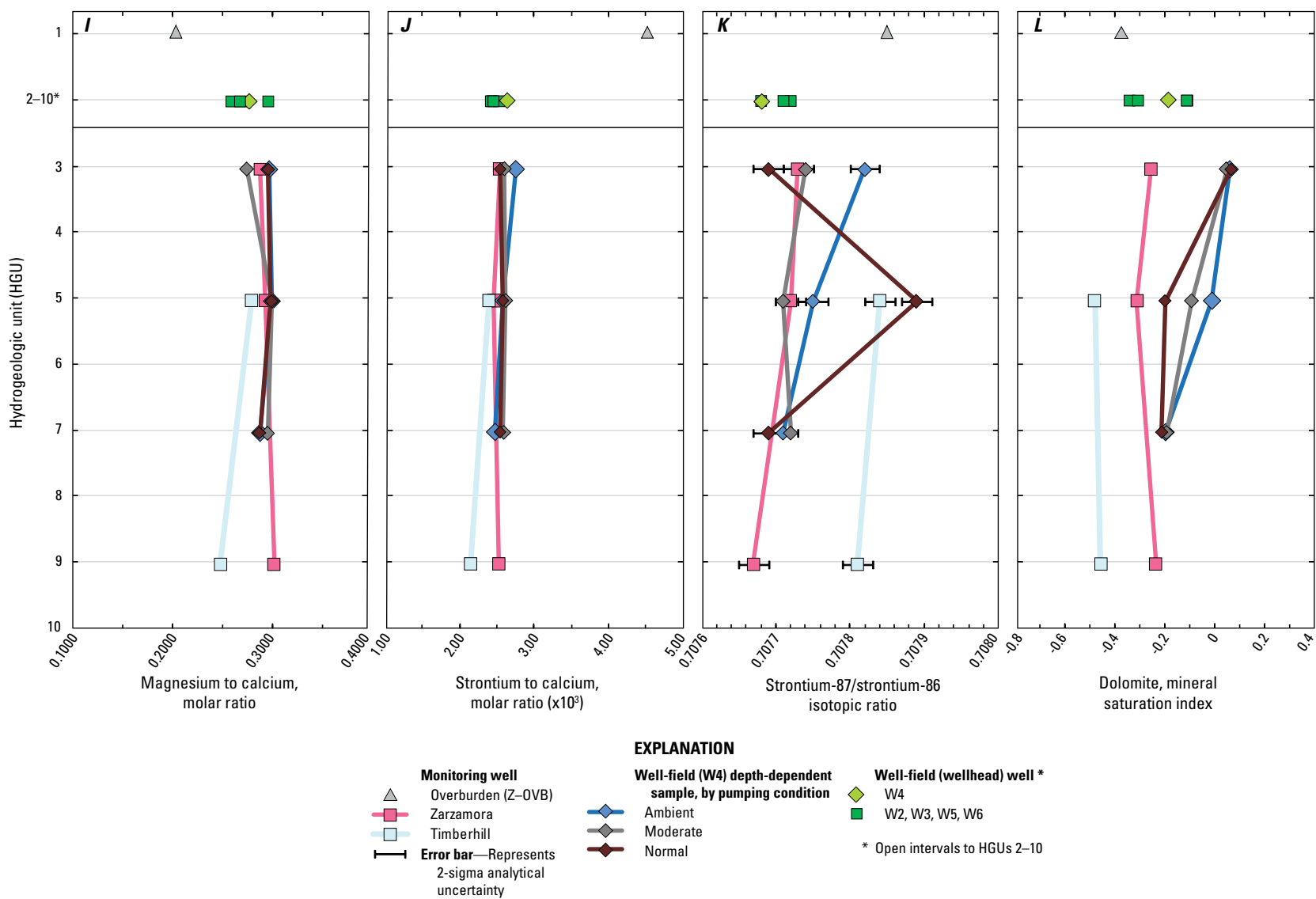


Figure 38. I, Magnesium to calcium molar ratio. J, Strontium to calcium molar ratio ($\times 10^3$). K, Strontium isotope composition. L, Dolomite mineral saturation index.—Continued

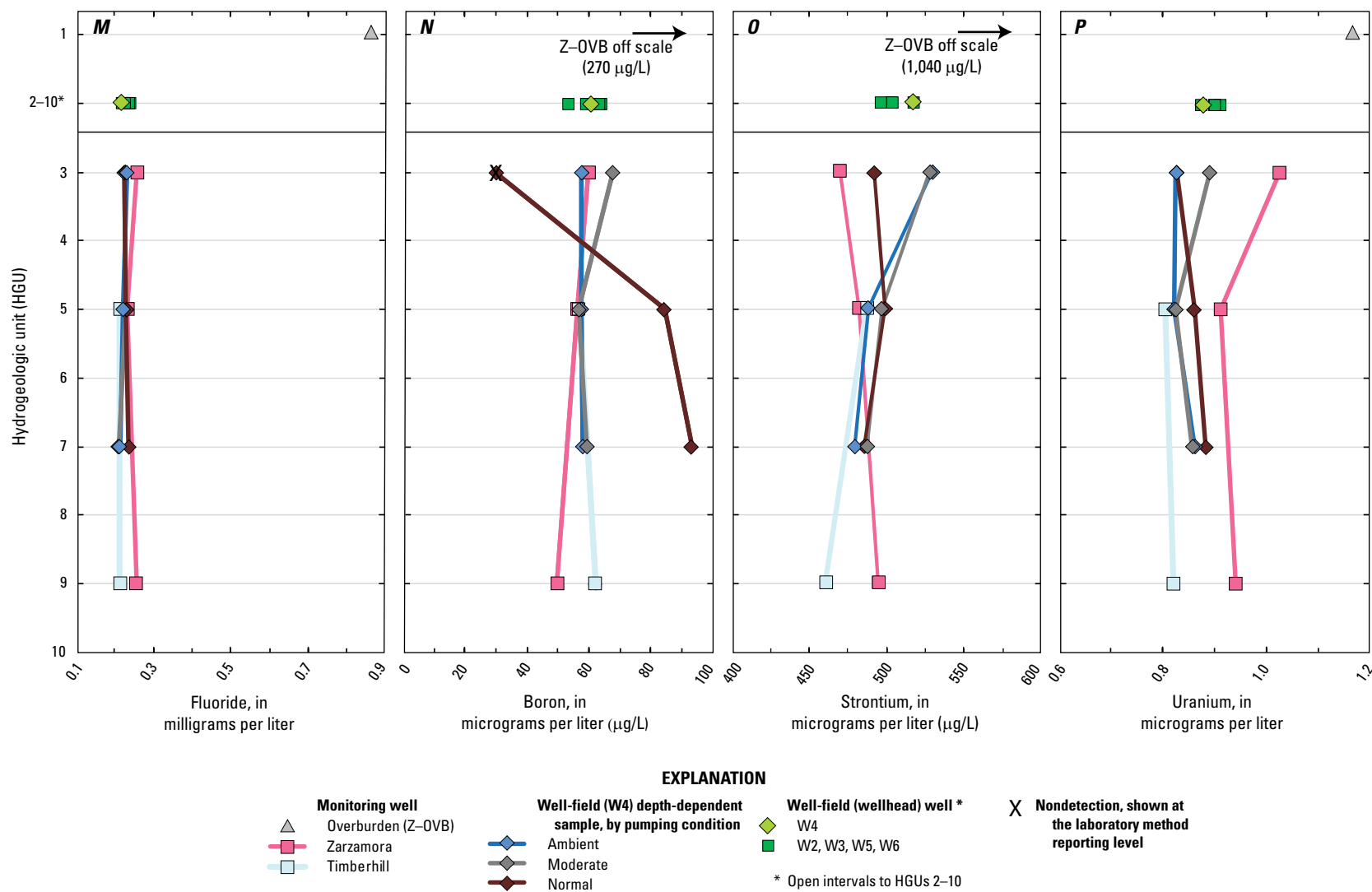


Figure 38. *M*, Fluoride concentration. *N*, Boron concentration. *O*, Strontium concentration. *P*, Uranium concentration.—Continued

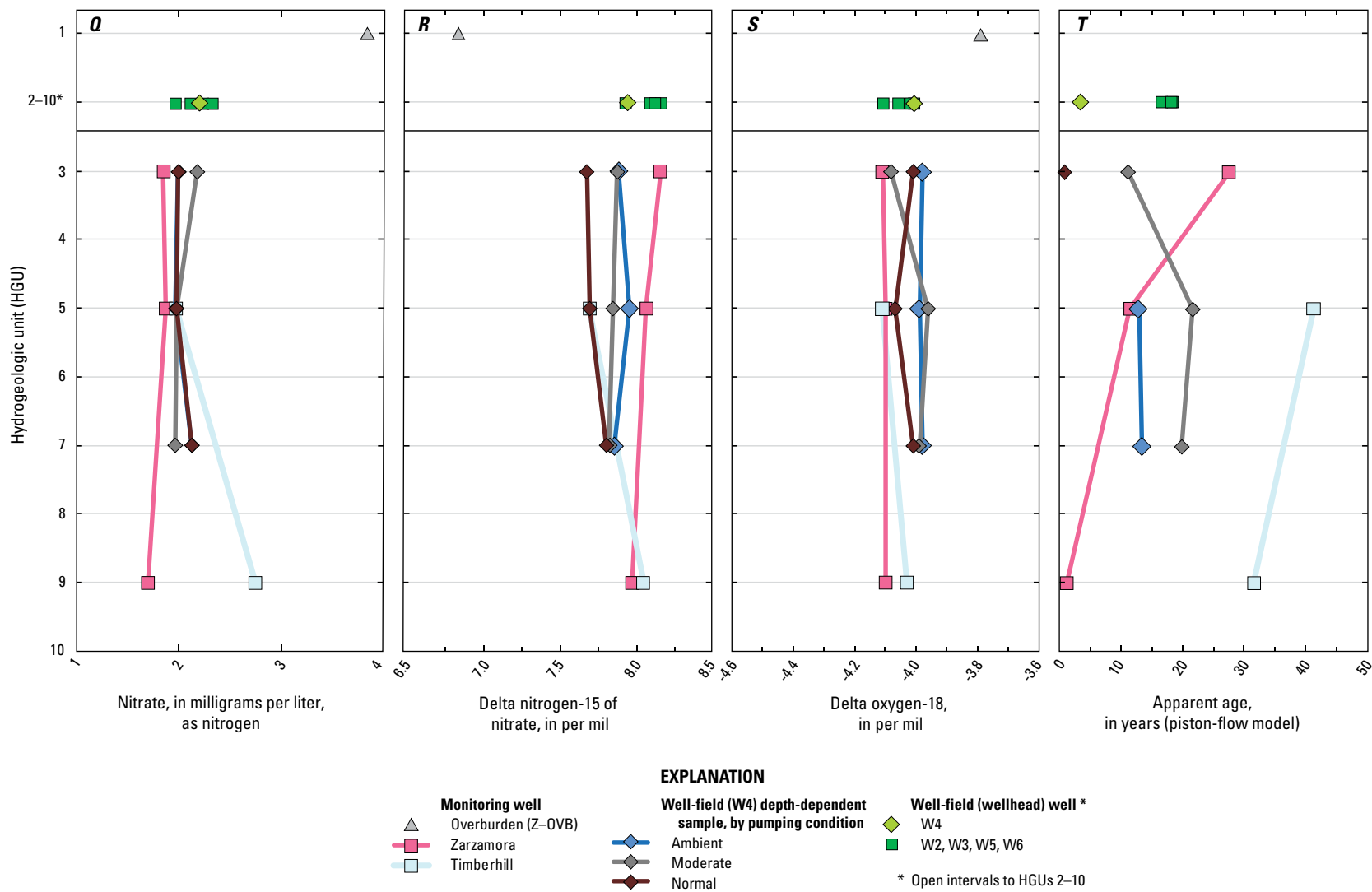


Figure 38. *Q*, Nitrate concentration. *R*, Delta nitrogen-15 of nitrate composition. *S*, Delta oxygen-18 composition. *T*, Groundwater apparent age (based on tritium and helium-3 age tracers).—Continued

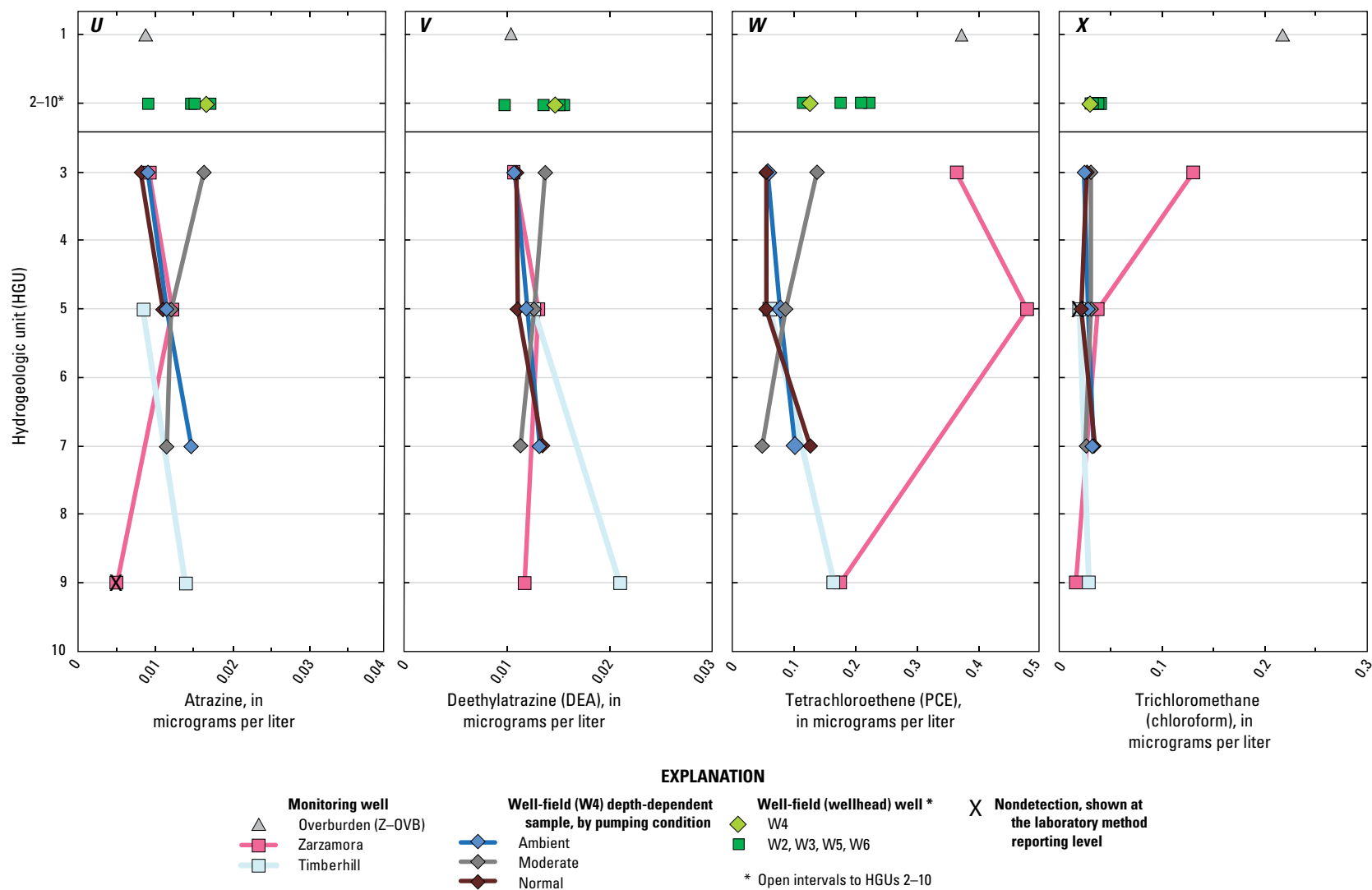


Figure 38. U, Atrazine concentration. V, Deethylatrazine (DEA) concentration. W, Tetrachloroethene (PCE) concentration. X, Trichloromethane (chloroform) concentration.—Continued

altitudes in wells, and springflow (for example, fig. 6). Temporal changes in hydrologic conditions such as water-level altitudes, recharge, groundwater travel times, and flow routing affect hydrogeologic processes, including water-rock interaction processes, and dilution and mixing of recharge, which affect water quality. Rapidly changing hydrologic conditions and a continuum of flow pathways from matrix flow to conduit flow imply that hydrogeologic processes might operate on different temporal and spatial scales in different parts of the aquifer. For example, groundwater dominated by matrix/diffuse flow or less affected by focused recharge might exhibit less temporal variability than would groundwater affected by focused recharge and conduit flow.

Previous work in the Edwards aquifer has demonstrated temporal variability in geochemistry and water quality over multiple time scales in response to changes in hydrologic conditions (Musgrove and Banner, 2004; Mahler and others, 2006; Mahler and Massei, 2007; Mahler, 2008; Musgrove and others, 2010; Wong and others, 2011). In conjunction with samples collected for this study, historical data for samples collected from two wells in the well field extend over an 11-year period, from 1998 through 2009. Historical data collected from these two PSWs (W3 and W6) indicate that geochemistry of the confined groundwater produced by the well field has varied in response to changes in hydrologic conditions over this time period. For example, variability in Mg/Ca molar ratios, indicative of geochemical evolution processes, is consistent with regional-scale changes in hydrologic conditions. Higher Mg/Ca ratios in PSW groundwater samples, indicative of longer groundwater residence times, are correlated with lower water levels in the Bexar County index well (J-17); lower water levels reflect drier regional hydrologic conditions (figs. 6 and 39). The small range of variability for historical samples from the well field relative to regional aquifer variability (Musgrove and others, 2010), as well as similar Mg/Ca ratios relative to the median value for the regional aquifer PSW samples (0.32), is consistent with the hypothesis that the well-field PSWs produce water typical of the confined regional aquifer. In comparison with groundwater from shallow unconfined wells, groundwater from most confined aquifer wells, unless intersecting a fast-flow conduit, is likely dominated by longer residence time matrix flow, which would be accompanied by attenuation of inputs from storm and recharge events. The geochemistry of matrix groundwater is likely less variable and changes more slowly in response to longer term regional hydrologic conditions. Relatively rapid changes in geochemistry have been documented for unconfined wells in the vicinity of the local-scale study area (Musgrove and others, 2010), but the time scales of geochemical variability for the confined aquifer in response to recharge events has, to date, been largely undocumented.

One of the well-field PSWs (W6) and four selected monitoring wells (Zarzamora monitoring wells Z-SED, Z-IED, and Z-DED and an upgradient unconfined monitoring well,

AY-68-27-610) were sampled several times in response to a storm event that supplied recharge to the aquifer in October 2009 (fig. 5; table 4). Event sampling is described earlier in the section "Collection of Water Samples."

Time-series results for water-level altitudes and (or) geochemical constituents from W6 and sampled monitoring wells during the period of event sampling were compared with the regional water-level altitude in the Bexar County index well (J-17) (fig. 40). Water-table altitudes rose in the Bexar County index well (J-17), reflecting regional hydrologic conditions, as well as in the sampled monitoring wells, in response to the recharge event (fig. 40) (water-level altitudes were not measured in W6). The geochemistry in the upgradient monitoring well is notably distinct from that of W6 and the Zarzamora monitoring wells. The upgradient well is characterized by higher specific conductance, Sr isotope composition, and concentrations of NO_3 and chloroform and by lower Mg/Ca molar ratios, dolomite SIs, and PCE concentrations (fig. 40). The lower Mg/Ca ratios and dolomite SIs and higher Sr isotope compositions measured in samples from the upgradient monitoring well are consistent with the hypothesis that groundwater from the upgradient unconfined recharge zone has undergone less water-rock interaction than has groundwater sampled from confined PSWs and Zarzamora monitoring wells. This is consistent with results for shallow unconfined monitoring wells presented in Musgrove and others (2010), which indicated that groundwater from the shallow unconfined aquifer is less geochemically evolved relative to the deeper unconfined and confined part of the aquifer. Higher concentrations of NO_3 (2.62–2.99 mg/L) and chloroform (0.066–0.192 $\mu\text{g/L}$) in samples from the upgradient unconfined monitoring well might be indicative of local sources of these constituents to this well; these concentrations are higher than the median values for shallow/urban unconfined groundwater (1.85 mg/L and 0.032 $\mu\text{g/L}$, respectively) as described by Musgrove and others (2010). Lower PCE concentrations in samples from the upgradient unconfined monitoring well indicate that this well is not affected by local PCE sources, unlike the well-field wells and the Zarzamora monitoring wells.

Geochemical variability in response to the sampled recharge event is generally small in the well-field PSW and the monitoring wells in comparison to the range of geochemical variability observed in the regional aquifer. These results are consistent with the hypothesis that the response to individual recharge events in the confined aquifer, unless intersecting conduit flow paths, might be attenuated by mixing processes along regional flow paths. Nonetheless, numerous geochemical constituents in W6 and the monitoring wells changed in response to the sampled recharge event. Recharging water associated with a rain event would likely be more dilute than groundwater with longer residence time; although specific conductance showed minor variability in response to the recharge event, Cl concentrations showed a dilution curve (an initial decrease, then subsequent increase) in W6 and the monitoring wells as regional water levels rise, consistent with

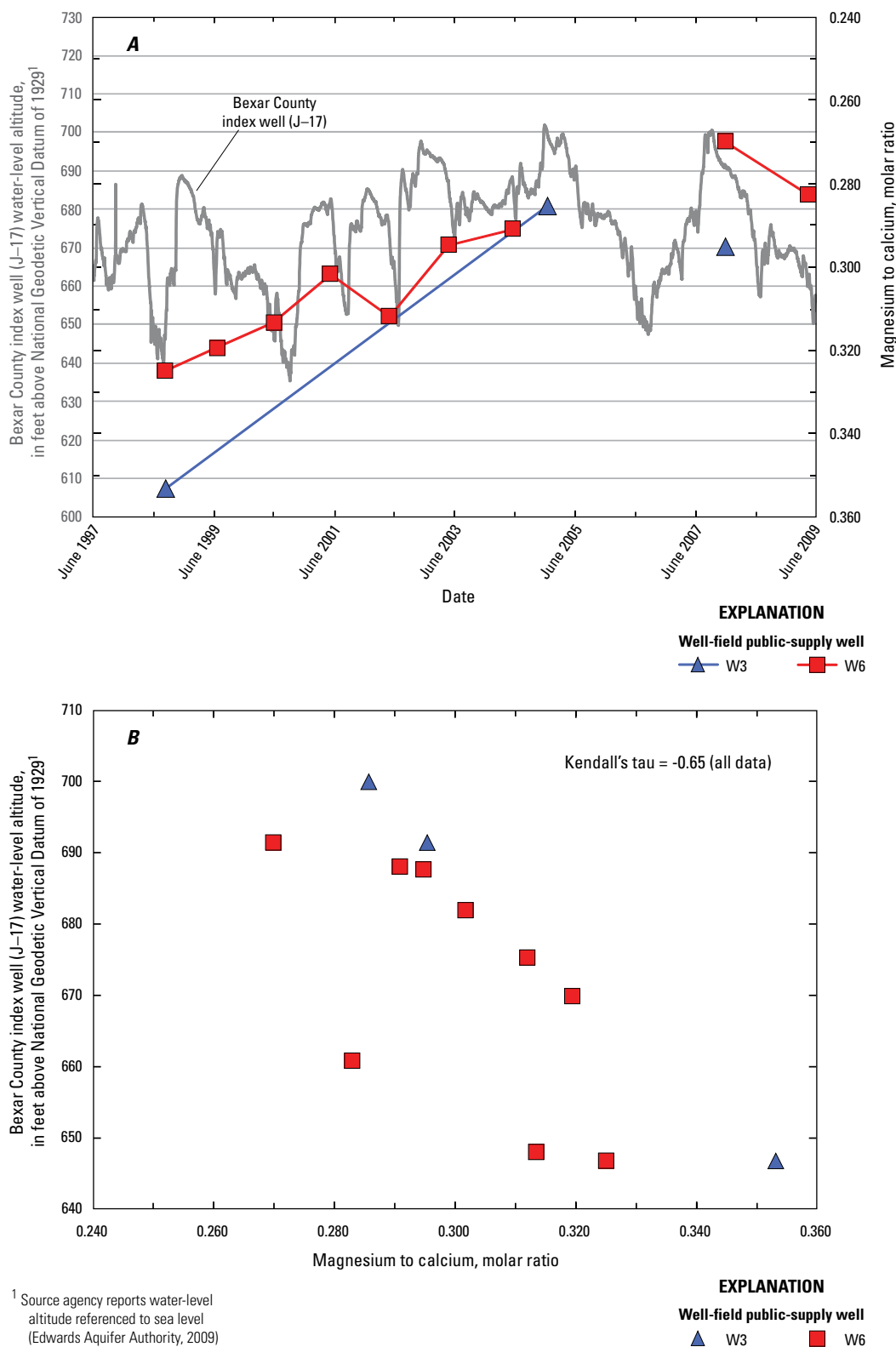


Figure 39. Historical magnesium to calcium molar ratio data for well-field public-supply wells sampled for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas. *A*, Time-series (1997–2009) of Bexar County index well (J-17) water-level altitude (Edwards Aquifer Authority, 2009) and historical magnesium to calcium molar ratios for TANC study well-field wells (W3 and W6). *B*, Relation between Bexar County index well (J-17) water level altitude and historical magnesium to calcium molar ratios for TANC study well-field wells (W3 and W6).

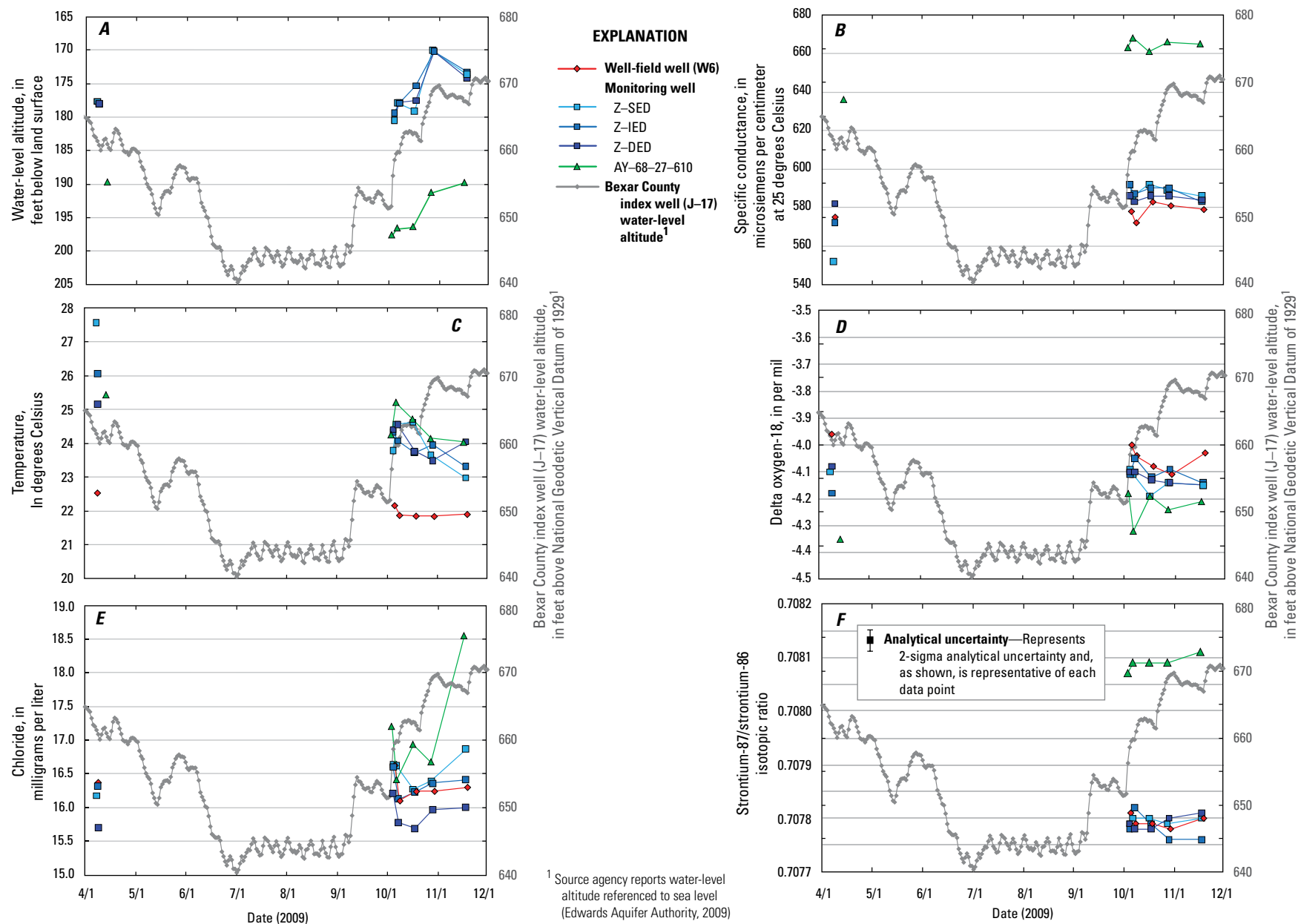


Figure 40. Geochemical results for selected constituents in groundwater samples collected in response to a rain event (October through November, 2009) for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells, south-central Texas. Water-level altitude in Bexar County index well (J-17) is shown for comparison. *A*, Water-level altitude. *B*, Specific conductance. *C*, Temperature. *D*, Delta oxygen-18 composition. *E*, Chloride concentration. *F*, Strontium isotope composition.

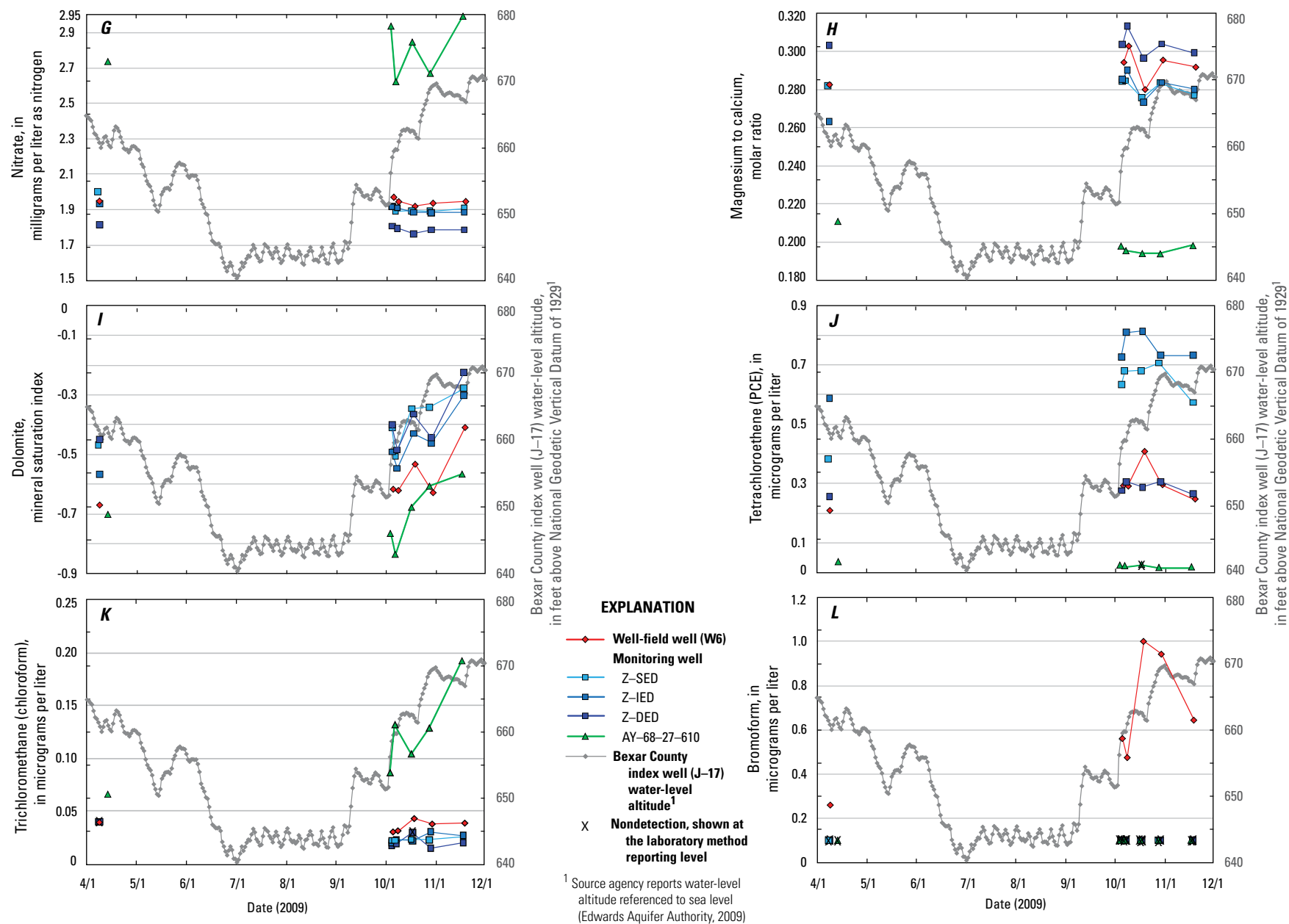


Figure 40. G, Nitrate concentration. H, Magnesium to calcium molar ratio. I, Dolomite mineral saturation index. J, Tetrachloroethene (PCE) concentration. K, Trichloromethane (chloroform) concentration. L, Bromoform concentration.—Continued

temporary dilution during the event following by a return to ambient conditions (fig. 40E). The decrease in Cl concentration was accompanied by decreases in NO_3 concentration and Mg/Ca ratios (fig. 40). Mg/Ca ratios at W6 and the Zarzamora monitoring wells showed a dilution curve, as did dolomite SIs. These results are consistent with previous studies that have proposed that aquifer NO_3 concentrations are diluted by incoming recharge; lower Mg/Ca ratios and dolomite SIs are consistent with shorter residence times that would be associated with recent, less geochemically evolved recharge (Musgrove and others, 2010). The increase in regional water levels associated with the recharge event is also accompanied by an increase and subsequent decrease in PCE concentrations in W6 and the two shallowest Zarzamora monitoring wells (Z-SED and Z-IED), characteristic of a breakthrough curve. This response is consistent with other PSW results, which indicate that the well field and in particular the Zarzamora monitoring wells are affected by a local PCE source and receive recharge with higher PCE concentrations.

Chloroform concentrations were low at W6 and the Zarzamora monitoring wells relative to the upgradient monitoring well, where chloroform concentration increased during the event sampling period (fig. 40K). Bromoform was not detected at any of the monitoring wells during event sampling, but concentrations show a breakthrough curve at W6. This result is consistent with relatively high bromoform concentrations measured in the well field in comparison with regional aquifer PSW and monitoring well samples (fig. 26) and indicates that W6 is affected by a local bromoform source. Chloroform and bromoform are drinking-water chlorination byproducts that are commonly detected nationwide, which is in part attributed to the recycling of chlorinated waters (Zogorski and others, 2006). Backflow to supply wells or the possible use of treated water as a downhole lubricant of well pumps are potential sources of chloroform and bromoform to supply wells.

Processes Affecting Transport of Natural and Anthropogenic Contaminants to the Public-Supply Well

Sources of Water and Solutes and Movement to Supply Wells

Surface water from losing streams is the primary source of recharge in the regional aquifer and the local-scale study area. Stable isotopes of δD and $\delta^{18}\text{O}$ indicate that this water is meteoric in origin. Some evaporation might affect recharge and associated groundwater near specific recharge locations (Musgrove and others, 2010). Groundwater compositions and dissolved-ion sources reflect processes of water-rock

interaction with the host limestone aquifer rocks. Previous studies indicate that groundwater in the unconfined part of the aquifer geochemically evolves to the composition of groundwater in the confined part of the aquifer (Musgrove and others, 2010). Groundwater flow in the San Antonio segment of the aquifer is generally west to east, although model-derived and tracer-inferred flow characteristics can vary markedly in this karst aquifer (Lindgren and others, 2009). The major components of model-derived groundwater flow in the TANC local-scale model area are (1) from the recharge zone southward into the confined zone and (2) from west to east in the confined zone (Lindgren and others, 2011b).

Anthropogenic sources that might influence groundwater quality include septic systems, leakage from municipal water and wastewater systems, and general industrial, commercial, or residential use of fertilizers, pesticides, and VOCs. There are only minor agricultural areas in the region (fig. 10), and these are primarily downgradient of the local-scale study area. There are point sources in the study area and vicinity of the well field for some of these contaminants (figs. 11–14), although nonpoint sources are also likely given the widespread detections of these contaminants in samples from the regional aquifer PSWs. Previous work has indicated that the urban environment is a potential source of organic contaminants to the aquifer (Musgrove and others, 2010).

Results for TANC study groundwater samples had no anthropogenic contaminants at levels of regulation, but they did contain several organic contaminants at low concentrations (mostly less than $1\text{ }\mu\text{g/L}$). Frequent low-level detection of several pesticides and VOCs in samples from the regional aquifer PSWs indicates that sources of pesticides and VOCs are widely distributed in the regional Edwards aquifer. The primary pesticides detected were atrazine, its degradate DEA, and, at less frequent detections, simazine and prometon. The primary VOCs detected were PCE and chloroform, and there were less frequent detections of other THMs (bromoform, dibromochloromethane, and bromodichloromethane). PCE detections and concentrations in samples from the well-field wells and Zarzamora monitoring wells were higher than those for samples from the regional aquifer PSWs. There are point sources in the vicinity of the well field for PCE, including a known plume (Bandera Road Groundwater Plume) to the northwest of the well field. Groundwater flow directions are consistent with the possibility that this might be a possible source of PCE to the well-field wells and Zarzamora monitoring wells, although other sources are also possible. Spatial variations in PCE in the study area indicate that higher PCE concentrations are associated with the urban San Antonio area, with a cluster of higher concentrations in the direct vicinity and north of the well field (fig. 41).

Nitrate concentrations in the aquifer were below levels of regulation for drinking water. Well-field samples had a median NO_3 concentration of 2.2 mg/L . Nitrate concentrations in the regional groundwater wells suggest that NO_3 in the aquifer likely results from nonpoint sources, although a few wells with NO_3 concentrations greater than 3 mg/L might be affected

by local NO_3 sources (fig. 42). Previous studies have proposed that surface-water recharge dilutes background groundwater NO_3 concentrations, and lower NO_3 concentrations are generally associated with groundwater collected from unconfined recharge-zone wells (Mahler and others, 2006; Musgrove and others, 2010). Nitrate concentrations are also generally higher in samples from the urban San Antonio area than outside (Musgrove and others, 2010; fig. 42). Oxidic conditions indicate that denitrification is not an important process in this aquifer, and thus, no natural NO_3 attenuation is likely to occur. Near the downgradient saline-zone boundary, however, deeper Edwards aquifer groundwater can be locally affected by mixing with more saline and less oxidic down-dip water. A small number of groundwater samples from regional wells show evidence for mixing with more saline and less oxidic down-dip groundwater, which might also be associated with down-dip water moving up along faults (Oetting and others, 1996; Musgrove and others, 2010). Groundwater in the local-scale and regional-scale study areas contains no natural contaminants (such as U and As) at levels of concern, and aquifer conditions are not conducive to increases in concentrations of these contaminants.

Results for age tracers indicate that groundwater reaching the well field had a median apparent age of 17 years. The largest fraction of water reaching W4, however, might be younger, with a piston-flow-model age of 3 years. These results are consistent with particle-tracking results from hydrologic modeling for the study area (Lindgren and others, 2011b) indicating that greater than 85 percent of particle tracks to W4 have travel times of less than 5 years and that greater than 95 percent have travel times less than 10 years (fig. 33). Apparent ages are not well correlated with other geochemical tracers of residence time (for example, Mg/Ca ratios, dolomite SIs, and Sr isotopes) (fig. 36). Aquifer units were found to have zones of preferential flow, although samples dominated by groundwater from different stratigraphic units were not geochemically distinct. Aquifer units are composed of Cretaceous-age limestone with minor variations in geochemistry (tables 5 and 6). Groundwater collected from W4 was relatively homogeneous and well mixed under a variety of pumping conditions for numerous geochemical constituents, with a notable exception of age tracers. These results suggest that mixing processes in the aquifer are complex and that hydrogeologic conceptual models used for age interpretations might not adequately account for mixing processes in this aquifer.

Supply-Well Vulnerability to Natural and Anthropogenic Contaminants

Like many karst aquifers, high porosity and permeability and rapid flow of recharging surface water through conduits, with little to no filtration, render the Edwards aquifer highly susceptible to contamination. Constituents of concern in the Edwards aquifer for the long-term sustainability of the groundwater resource include NO_3 and organic contaminants.

Measured concentrations of NO_3 and organic contaminants are currently low relative to levels of regulation. Nonetheless, the dominance of very young groundwater in the TANC study area and, specifically, groundwater that reaches the well field indicates that the aquifer and water supply are vulnerable to contamination.

Other TANC studies have determined that changes in pumping conditions can affect water quality (Katz and others, 2007; Landon and others, 2008; Jurgens and others, 2008; Brown and others, 2009; Bexfield and others, 2011). In contrast, for the Edwards aquifer TANC study area, although changes in pumping conditions might change the distribution of groundwater flow paths contributing water to the well field, under all pumping conditions the majority of the groundwater was young and, as such, is susceptible to anthropogenic contaminants. Borehole imaging of W4 revealed large solution features, consistent with high connectivity throughout the aquifer. Although zones of preferential flow were determined for W4, groundwater samples from different stratigraphic units could not be distinguished geochemically, and no water-quality effects were discernable on the basis of sample stratigraphy. The location of the PSW within an urban area increases the vulnerability of the water supply and its potential susceptibility to contamination.

Groundwater geochemistry and water quality in the Edwards aquifer is affected by changes in hydrologic conditions. Temporal changes in hydrologic conditions such as water-level altitudes, recharge, groundwater travel times, and flow routing affect hydrogeologic processes, including the extent of variability in water-rock interaction processes, dilution of recharge, and associated geochemistry. Specifically for the Edwards aquifer, drier hydrologic conditions are likely associated with lower concentrations of atrazine and higher concentrations of nitrate (Musgrove and others, 2010), but this relation for individual wells is likely affected by specific flow paths. For example, groundwater dominated by matrix/diffuse flow or less affected by focused recharge might be less variable in water quality than groundwater affected by focused recharge and conduit flow.

Water-quality data from the regional aquifer PSWs and well-field monitoring wells, collected over a range of depths in the Edwards aquifer, show little variability. Regional (confined) aquifer PSW samples from the same stratigraphic units where the aquifer is deeper in the subsurface (samples from wells greater than 1,000 ft in depth), however, show notable chemical differences from the rest of the regional aquifer, with lower concentrations of organic contaminants and geochemical evidence for more evolved groundwater compositions.

Understanding how land use affects water quality of PSWs is important because of the strong influence of land use on water quality and the rapid pace at which changes in land use are occurring in the urban part of the study area. This understanding has practical implications with respect to human health (McMahon and others, 2008) but might be complicated by spatial and temporal variations in land use (Brawley and others, 2000; Hiscock and others, 2007) and by

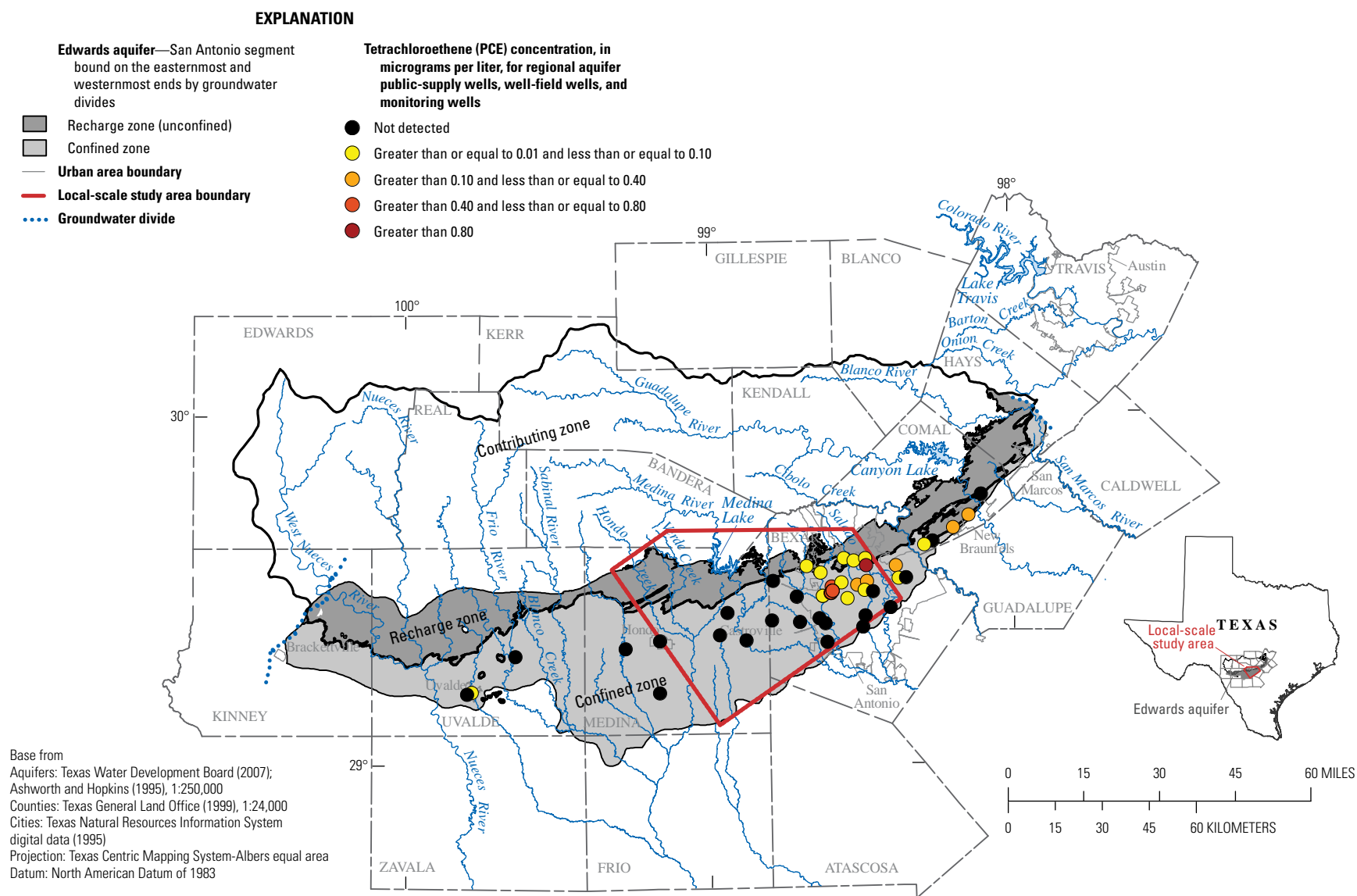


Figure 41. Hydrogeologic setting, location of, and ranges of tetrachloroethene (PCE) concentration (highest value shown where multiple measurements were made during the study period) for regional aquifer public-supply wells sampled for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells in the San Antonio segment of the Edwards aquifer, south-central Texas, 2004–9.

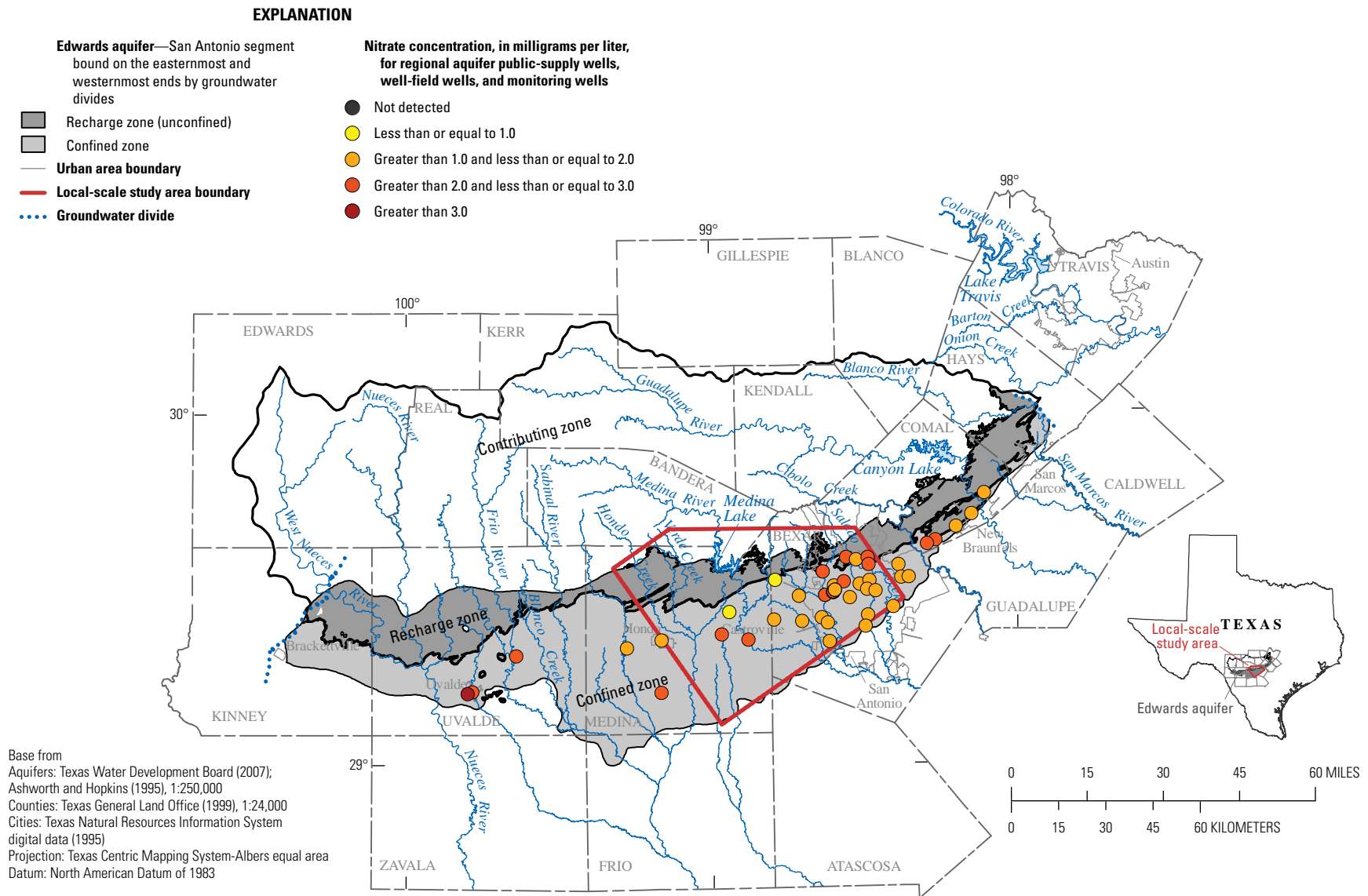


Figure 42. Hydrogeologic setting, location of, and ranges of nitrate concentration (highest value shown where multiple measurements were made during the study period) for regional aquifer public-supply wells sampled for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells in the San Antonio segment of the Edwards aquifer, south-central Texas, 2004–9.

mixing waters of different age and source in long well screens that are typical in PSWs (Kauffman and others, 2001; Böhlke, 2002; Weissmann and others, 2002; Osenbrück and others, 2006; Eberts and others, 2006). Age distributions of water captured by individual long-screen wells in some aquifer settings might span years to millennia or more (Kauffman and others, 2001; Weissmann and others, 2002; McMahon and others, 2004; Plummer and others, 2004; Sturchio and others, 2004; Clark and others, 2007; Starn and Brown, 2007; Crandall and others, 2008; Jurgens and others, 2008). These age distributions in captured water could have positive and negative consequences for the quality of water from a PSW, such as dilution of contaminant concentrations and long lag times in responding to land use change or best management practices. An important implication of long lag times is that, in spite of corrective management actions, contaminant concentrations in water from PSWs could increase in the future (before eventually decreasing) as the fraction of already contaminated young recharge captured by wells increases (Fogg and others, 1999; Böhlke, 2002; Fogg and LaBolle, 2006).

Although TANC study results for the local-scale study area indicate that long lag times are not a likely feature of the Edwards aquifer, understanding the potential response of water quality in the PSW to land-use changes, and the resulting time scales of response in the PSW, is useful for water resource management. A post-MODPATH spreadsheet model (L.J. Kauffman, U.S. Geological Survey, written commun., 2011) that incorporates MODPATH particle-tracking results (Pollock, 1994) and contaminant reaction (degradation) processes was used to analyze the response of water quality in the selected PSW to land-use change. The water-quality response to land-use change was dependent on age distributions of water captured by the PSW. Groundwater flow and particle-tracking simulations are described by Lindgren and others (2011b). Flow and traveltimes associated with the particle tracking to the PSW were used to estimate relative contaminant concentrations in the PSW; a hypothetical conservative (nonreactive) contaminant was modeled. A relative contaminant concentration was assigned to each particle on the basis of the traveltime and by assuming an input concentration history of the contaminant in the contributing area (recharge zone), as described below. Flux-weighted average contaminant concentrations for the PSW were determined by summing the product of concentrations assigned to each particle by the total flow associated with that particle and dividing by the total flow.

Simulated age distribution for contributing-area particles of the PSW ranged from less than 1 to 342 years, with a median of 2.0 years (fig. 33). Particle-track results indicate that 99 percent of the PSW water was less than 50 years old and that 88 percent of the water was less than 5 years old. Younger water (less than 50 years old) is more likely to be affected by human activities, and if present, most anthropogenic contaminants would be associated with younger water. As previously described, these model results concur with the conceptual understanding that a large fraction of water in the

Edwards aquifer, and particularly in the PSW, is young (less than 50 years old). Water samples from the PSW, other wells in the well field, and other PSWs completed in the regional aquifer contained detectable concentrations of a small number of some pesticides and VOCs (fig. 25; tables 14–17). Natural attenuation processes are also likely to be limited in flow systems with young groundwater, in particular for conditions or contaminants where degradation processes are slow (Chapelle and Bradley, 1999; McMahon and others, 2008), such as is the case in this oxic aquifer. In flow systems with a higher fraction of older water (that is, greater than 50 years old), that fraction could dilute anthropogenic contaminant concentrations when waters of different ages mix within the aquifer and the open borehole of a PSW (Osenbrück and others, 2006).

In a model-forecasting exercise, slug inputs were used to examine the range in lag times, dilution, and degradation that might be expected at the PSW. Results of this exercise might be representative of a rapid land-use change overlying the contributing area (recharge area) of a PSW in this aquifer and might represent how quickly the aquifer responds to mitigation. A hypothetical contaminant was introduced over the PSW-model-contributing area continuously for 30 years. Contaminant input concentrations were equal to 1, and all simulated concentrations for the well were relative to this unit value. After 30 years, the contaminant input returned to zero. The hypothetical responses at the PSW for (1) a simulation with no contaminant degradation and (2) a simulation with degradation at a first-order reaction kinetics degradation rate

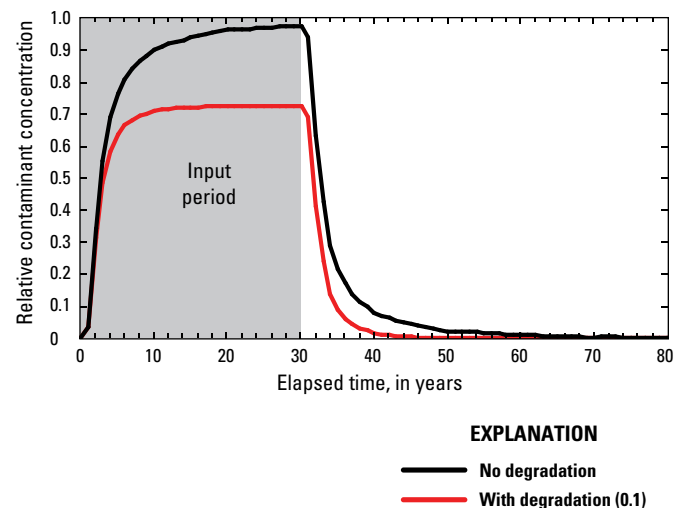


Figure 43. Results of forecasting exercise of time-series of hypothetical contaminant concentrations at the selected public-supply well for the study of the transport of anthropogenic and natural contaminants (TANC) to public-supply wells; results are based on 30 years of contaminant input (at a relative concentration of 1) in the modeled contributing area (based on model and particle-track results described in Lindgren and others, 2011b).

of 0.1/year (10 percent per year) are illustrated in figure 43. For both responses (curves in fig. 43), concentrations at the PSW began to rise within 1 year after loading began and rose quickly to reach 50 percent of peak concentration in 3 and 4 years, respectively. The nondegradation curve continued to rise rapidly until about 5 years, when 77 percent of peak concentration was reached. Concentrations continued to rise for both curves until peak concentrations were reached at the end of the input period of 30 years. Maximum concentrations were 98 percent of input for the nondegradation curve and 73 percent of input for the with-degradation curve. Once the input was stopped after 30 years, both simulations responded with a rapid drop in concentration. Concentration decreased to about 50 percent of input in 3 years in the nondegradation simulation and in a little more than 2 years for the degradation simulation. Concentration decreased to 25 percent of input in about 5 years and in less than 4 years, respectively. Although peak concentrations dropped quickly after input ceased, 32 and 12 years were required, respectively, to flush 99 percent of the hypothetical contaminant and achieve near-background conditions (table 20 at end of report).

Lag times would be expected to be shorter for wells with large fractions of young water (McMahon and others, 2008), such as observed for this study, than for wells with smaller fractions of young water. The effect of short-circuit pathways, for example karst conduits, in the flow system on the movement of young water to the PSW could greatly alter contaminant arrival times compared to what might be expected from advection in a system without short circuiting. This observation is particularly valid for karst such as the Edwards aquifer: numerous voids are present in wells completed in this aquifer, including the PSW. The simulated concentrations show rapid initial response at the beginning and end of chemical input, followed by more gradual response as older water moves through the system. The nature of karst groundwater flow, where flow dominantly occurs via conduit flow paths, could lead to relatively rapid water-quality responses to land-use changes.

This analysis did not consider chemical storage and travel times in the unsaturated zone, dispersion/diffusion, or transient flow conditions in the aquifer, which might yield substantially different results. Storage of large amounts of chemical mass in, and long transit times through, unsaturated zones could result in longer flushing times, as well as longer arrival and lag times, in the aquifer. The impact of not considering dispersion is likely minimized because the focus herein is for large pumping wells and nonpoint sources of contamination. Transient conditions in the flow system could result in more complex areas contributing recharge and more complex groundwater age distributions than were assessed here and should be considered in future studies (Franke and others, 1998; Rock and Kupfersberger, 2002). Spatial variability in input, which would likely have considerable effect on concentrations and movement of contaminants, also was not considered. Short-circuit pathways, such as karst conduits, in some PSW contributing areas can affect age distributions

by allowing relatively rapid movement of young water to the open interval of the PSW. The spatial component of land-use change is important because the complex distribution of particle travel times within the contributing areas strongly influences contaminant arrival times and degradation reaction progress. Results from this and other similar exercises (McMahon and others, 2008) indicate that time scales for change in the quality of water from the PSW could be on the order of a few years to decades for land-use changes that occur over days to decades, which has implications for source-water protection strategies that rely on land-use change to achieve water-quality objectives.

Summary

In 2001, the National Water-Quality Assessment (NAWQA) Program of the U.S. Geological Survey initiated a series of studies on the transport of anthropogenic and natural contaminants (TANC) to public-supply wells (PSWs). The main goal of the TANC project was to better understand the source, transport, and receptor factors that control contaminant movement to PSWs in representative aquifers of the United States. Regional- and local-scale study areas were selected from within existing NAWQA study units, including south-central Texas. This report describes the hydrogeology and geochemistry of the local-scale TANC study near San Antonio, Tex., and discusses factors that might control transport of selected anthropogenic and natural contaminants to a public-supply well within a productive well field. The study area is within the San Antonio segment of the Edwards aquifer, a highly productive karst aquifer developed in Cretaceous-age carbonates. Like many karst aquifers, the Edwards aquifer is subject to rapid changes in flow and discharge rates, water-level altitudes, and water quality and as a result is susceptible to contamination. An understanding of the source, transport, and receptor factors that control contaminant movement to PSWs in the Edwards aquifer is important for water-management practices and the long-term sustainability of the groundwater resource. Results for a variety of geochemical constituents and tracers were compared for groundwater samples from regional aquifer PSWs (mostly from the confined part of the aquifer), from the well field (wellhead samples), from the monitoring wells (Zarzamora and Timberhill monitoring well clusters), from a monitoring well in the overburden, from the PSW depth-dependent sampling, and from temporal sampling in response to a rainfall and recharge event (from the PSW and selected monitoring wells).

Constituents of concern in the Edwards aquifer for the long-term sustainability of the groundwater resource include anthropogenic organic contaminants and the nutrient nitrate. Natural contaminants, such as arsenic or uranium, are not of concern for the Edwards aquifer, and oxic aquifer conditions are not conducive to increases in concentrations of these contaminants. Geochemical and isotopic data are useful tracers of recharge, groundwater flow, fluid mixing, and water-rock

interaction processes that affect water quality. Sources of dissolved constituents to the Edwards aquifer and to the PSW include dissolution of and geochemical interaction with overlying soils and calcite and dolomite minerals that compose the aquifer. The well-field and monitoring well samples had relatively narrow ranges in field parameters and most geochemical constituents were, for the most part, similar in composition to samples from regional aquifer PSWs. Groundwater samples were mostly a dilute calcium bicarbonate or calcium magnesium bicarbonate facies. Mineral saturation indices with respect to calcite, dolomite, and gypsum reflect the carbonate mineralogy of the aquifer host rocks. TANC study results for molar ratios of magnesium to calcium and strontium to calcium, and for strontium isotopes, were used to assess relative extents of water-rock interaction associated with shorter and longer flow paths. Molar ratios of magnesium to calcium and strontium to calcium in carbonate groundwater typically increase along flow paths; strontium isotopes in Edwards aquifer groundwater typically decrease along flow paths, toward values similar to the limestone aquifer rocks. Oxidation-reduction conditions were oxic, which is typical of the regional aquifer, except very close to the downgradient transition, which had more saline, less oxic water.

Nitrate concentrations for samples collected during this study from regional aquifer PSWs (specifically, PSWs in the confined part of the aquifer) were below drinking-water standards and had a similar median value (1.9 milligrams per liter) to regional aquifer samples from the confined part of the aquifer collected in previous studies. Nitrate-isotope results for the well-field and monitoring well samples indicate that soil organic nitrogen is the dominant source of nitrate in the regional aquifer. A comparison with historical nitrate-isotope values, however, suggests that a possible component of nitrate in groundwater from biogenic sources might have increased over the last 30 years. Results for stable isotopes of hydrogen and oxygen for groundwater samples from the well-field wells and monitoring wells indicated that the groundwater is meteoric in origin. A small number of organic contaminants were routinely or frequently detected in groundwater samples collected for this study. These were the pesticides atrazine; its degradate deethylatrazine; the drinking-water disinfection byproduct, trichloromethane (chloroform); and the solvent tetrachloroethene (PCE). Detected concentrations of all organic constituents were low (mostly less than 1 microgram per liter), which is consistent with previous NAWQA findings. The common detection of these anthropogenic contaminants indicates that the aquifer is vulnerable to contamination and effects on water-quality that result from anthropogenic activities. The geochemistry of groundwater from the PSW was similar to that of samples collected from PSWs throughout the extent of the regional confined aquifer. These results indicate that water from the PSW is generally representative of the regional confined aquifer. All samples collected from the well field and specifically those collected from different depths in the PSW under a variety of pumping conditions were relatively homogeneous and well mixed for numerous

geochemical constituents, with the notable exception of age tracers. Although multiple age tracers were sampled, because of contamination effects and age-interpretation complexities, all age interpretations were based on tritium/helium-3 results. Apparent ages based on a piston-flow model for samples from the well field and monitoring wells ranged from less than 1 to 41 years, with a median of 17 years. Apparent ages were inversely stratified for both sets of monitoring wells, with older ages for the deeper monitoring wells. Results for depth-depending samples from the PSW yielded a range of apparent ages (less than 1 to 22 years) but with no consistency with respect to depth or stratigraphic unit contributing to the PSW. Selected geochemical constituents that might provide independent information about groundwater residence time (for example, magnesium to calcium molar ratios, dolomite mineral saturation indices, and strontium isotope ratios) or the extent of impact from anthropogenic activities (for example, the number of organic contaminants detected, concentration of selected organic contaminants) were compared to age-tracer results for groundwater samples. Results of this comparison were not consistent with the range of interpreted apparent ages or age-tracer results, indicating that hydrogeologic conceptual models used in groundwater age interpretations might not adequately account for mixing in this karst system. Mixing processes in this aquifer are complex, and although age tracers provide insight into the young nature of the groundwater, they might not readily allow for distinguishing the mixed water history of water at a PSW or be well suited to distinguish relatively small differences in groundwater age as a result of mixing processes. Nonetheless, age interpretations, indicating that water supplied to the PSW is young, were consistent with particle-tracking results from a companion hydrogeologic modeling study.

One of the well-field PSWs and four selected monitoring wells were sampled several times in response to a storm event that recharged the regional aquifer. Geochemical variability in response to the sampled recharge event was generally small in the well-field PSW and monitoring wells in comparison to the range of geochemical variability observed in the regional aquifer. Several geochemical constituents, however, showed consistent variability in the PSW and monitoring wells that corresponded with associated changes in hydrologic conditions in response to the sampled recharge event. The observed temporal variations in water chemistry were consistent with the hypothesis that the response to individual recharge events in the confined aquifer, unless intersecting conduit flow paths, might be attenuated by mixing processes along regional flow paths.

Anthropogenic sources that might influence groundwater water quality at the PSW include septic systems, leakage from municipal water and wastewater systems, and general industrial, commercial, or residential use of pesticides and volatile organic compounds. Although the urban San Antonio environment is a likely source of anthropogenic contaminants to the PSW, sampling of regional aquifer PSWs indicates that a few specific organic contaminants are widely distributed

throughout the aquifer at low concentrations. There are point sources in the vicinity of the well field for PCE. Higher concentrations of PCE were detected in samples from the well field and Zarzamora monitoring wells relative to the regional aquifer PSWs, which is consistent with the possibility that a local urban source (or sources) might be supplying PCE to these wells. Under all pumping conditions, the majority of water from the PSW was young and, as such, susceptible to contamination. Although zones of preferential flow were determined for the PSW, groundwater samples from different stratigraphic units were not geochemically distinct. Results from groundwater-flow-model particle tracking were used to evaluate a hypothetical contaminant-loading scenario to assess the potential response of the aquifer to contamination. Results indicate that the aquifer responds rapidly (less than 1 year to several years) to contaminant loading; however, it takes a relatively long time (decades) for concentrations to reach the highest values. The aquifer also responds quickly (less than 1 year to several years) to the removal of contaminant loading; again however, it takes a relatively long time (decades) to reach near background concentrations. These results corroborate the existing conceptual understanding of aquifer behavior to contaminant loading.

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Table 1. Site information for wells sampled for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[By sample category and by aquifer category (unconfined, confined). USGS, U.S. Geological Survey; NAD 83, North American Datum of 1983; Y, yes; N, no; --, not measured, known, or applicable; *, sampled 6 times for temporal response to rainfall; psw, public-supply well; **, sampled 9 times for vertically discrete samples; ft, feet; NAVD 1988, North American Vertical Datum of 1988; gal/min, gallons per minute; in., inches; PVC, polyvinyl chloride; †, borehole diameter (diam.), screen is 2.5-in. diam.; ‡, borehole diam., screen is 4.5-in. diam.]

USGS identification number	State well number	Local identifier	Latitude (decimal degrees), NAD 83 datum	Longitude (decimal degrees), NAD 83 datum	County	Aquifer category	Within supply- well scale model area	Pumpage quartile (2004)	Network	Year of first sample	Year of second sample	Year of third sample
Monitoring wells												
292943098354401	AY-68-36-135	Z-OVB	29.49531	-98.59564	Bexar	--	Y	--	discrete interval, overburden	2007	--	--
292943098354402	AY-68-36-133	Z-SED	29.49540	-98.59559	Bexar	confined	Y	--	discrete interval, temporal	2007	2009*	--
292943098354403	AY-68-36-134	Z-IED	29.49519	-98.59558	Bexar	confined	Y	--	discrete interval, temporal	2007	2009*	--
292943098354404	AY-68-36-132	Z-DED	29.49545	-98.59566	Bexar	confined	Y	--	discrete interval, temporal	2007	2009*	--
292851098374401	AY-68-35-315	T-IED	29.48074	-98.62901	Bexar	confined	Y	--	discrete interval	2008	--	--
292851098374402	AY-68-35-314	T-DED	29.48073	-98.62916	Bexar	confined	Y	--	discrete interval	2008	--	--
293252098380801	AY-68-27-610	Parkwood	29.54778	-98.63556	Bexar	unconfined	Y	--	temporal	2009*	--	--
Well-field wells												
292919098360501	AY-68-36-103	W2	--	--	Bexar	confined	Y	1st	wellfield	2007	--	--
292925098360201	AY-68-36-104	W3	--	--	Bexar	confined	Y	1st	wellfield, psw survey	2004	2005	2007
292916098360701	AY-68-36-105	W4	--	--	Bexar	confined	Y	1st	wellfield, depth dependent	2007**	--	--
292920098360601	AY-68-36-130	W5	--	--	Bexar	confined	Y	1st	wellfield	2007	--	--
292923098360301	AY-68-36-131	W6	--	--	Bexar	confined	Y	1st	wellfield, temporal	2007	2009*	--
Regional aquifer public-supply wells												
293359098290301	AY-68-29-414	--	--	--	Bexar	confined	Y	1st	psw survey	2004	2005	--
293358098231101	AY-68-29-610	--	--	--	Bexar	confined	N	1st	psw survey	2004	2005	--
293120098285801	AY-68-29-714	--	--	--	Bexar	confined	N	1st	psw survey	2004	2005	--
293145098224201	AY-68-29-929	--	--	--	Bexar	confined	N	1st	psw survey	2004	--	--
293119098211201	AY-68-30-718	--	--	--	Bexar	confined	N	1st	psw survey	2004	--	--
292843098425101	AY-68-35-106	--	--	--	Bexar	confined	Y	1st	psw survey	2004	--	--
292459098382101	AY-68-35-913	--	--	--	Bexar	confined	Y	1st	psw survey	2004	--	--
292822098325401	AY-68-36-206	--	--	--	Bexar	confined	Y	1st	psw survey	2004	2005	--
292944098292301	AY-68-37-105	--	--	--	Bexar	confined	Y	1st	psw survey	2004	2005	--
292522098291901	AY-68-37-426	--	--	--	Bexar	confined	Y	1st	psw survey	2004	--	--
292643098241801	AY-68-37-601	--	--	--	Bexar	confined	N	1st	psw survey	2004	2005	--
292328098294601	AY-68-37-705	--	--	--	Bexar	confined	N	1st	psw survey	2004	--	--
294225098080301	DX-68-23-601	--	--	--	Comal	confined	N	1st	psw survey	2004	2005	--
291210099475601	YP-69-50-506	--	--	--	Uvalde	confined	N	1st	psw survey	2004	--	--

Table 1. Site information for wells sampled for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category and by aquifer category (unconfined, confined). USGS, U.S. Geological Survey; NAD 83, North American Datum of 1983; Y, yes; N, no; --, not measured, known, or applicable; *, sampled 6 times for temporal response to rainfall; psw, public-supply well; **, sampled 9 times for vertically discrete samples; ft, feet; NAVD 1988, North American Vertical Datum of 1988; gal/min, gallons per minute; in., inches; PVC, polyvinyl chloride; †, borehole diameter (diam.), screen is 2.5-in. diam.; ‡, borehole diam., screen is 4.5-in. diam.]

USGS identification number	State well number	Local identifier	Latitude (decimal degrees), NAD 83 datum	Longitude (decimal degrees), NAD 83 datum	County	Aquifer category	Within supply- well scale model area	Pumpage quartile (2004)	Network	Year of first sample	Year of second sample	Year of third sample
293111098340901	AY-68-28-807	--	--	--	Bexar	confined	Y	2d	psw survey	2005	--	--
293512098291701	AY-68-29-109	--	--	--	Bexar	unconfined	Y	2d	psw survey	2005	--	--
292424098421501	AY-68-35-810	--	--	--	Bexar	confined	Y	2d	psw survey	2005	--	--
292053098365501	AY-68-44-110	--	--	--	Bexar	confined	Y	2d	psw survey	2005	--	--
293807098155301	DX-68-22-901	--	--	--	Comal	confined	N	2d	psw survey	2005	--	--
292604098563201	TD-68-33-501	--	--	--	Medina	confined	Y	2d	psw survey	2005	--	--
292116099095501	TD-69-47-305	--	--	--	Medina	confined	N	2d	psw survey	2005	--	--
291232099470301	YP-69-50-339	--	--	--	Uvalde	confined	N	2d	psw survey	2005	--	--
293128098473101	AY-68-26-814	--	--	--	Bexar	unconfined	Y	3d	psw survey	2005	--	--
293451098313201	AY-68-28-601	--	--	--	Bexar	unconfined	Y	3d	psw survey	2005	--	--
294604098060801	DX-68-16-708	--	--	--	Comal	unconfined	N	3d	psw survey	2005	--	--
294019098114701	DX-68-23-504	--	--	--	Comal	confined	N	3d	psw survey	2005	--	--
292215098580201	TD-68-41-103	--	--	--	Medina	confined	Y	3d	psw survey	2005	--	--
292119098524901	TD-68-41-308	--	--	--	Medina	confined	Y	3d	psw survey	2005	--	--
291943099163301	TD-69-46-601	--	--	--	Medina	confined	N	3d	psw survey	2005	--	--
291840099382601	YP-69-43-606	--	--	--	Uvalde	confined	N	3d	psw survey	2005	--	--
293518098332601	AY-68-28-203	--	--	--	Bexar	unconfined	Y	4th	psw survey	2005	--	--
293023098355401	AY-68-28-702	--	--	--	Bexar	confined	Y	4th	psw survey	2005	--	--
293042098305201	AY-68-28-913	--	--	--	Bexar	confined	Y	4th	psw survey	2005	--	--
292442098474501	AY-68-34-803	--	--	--	Bexar	confined	Y	4th	psw survey	2005	--	--
292405098371201	AY-68-36-704	--	--	--	Bexar	confined	Y	4th	psw survey	2005	--	--
292931098274601	AY-68-37-124	--	--	--	Bexar	confined	Y	4th	psw survey	2005	--	--
293729098173101	DX-68-30-215	--	--	--	Comal	confined	N	4th	psw survey	2005	--	--
291219099095601	TD-69-55-604	--	--	--	Medina	confined	N	4th	psw survey	2005	--	--

Table 1. Site information for wells sampled for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category and by aquifer category (unconfined, confined). USGS, U.S. Geological Survey; NAD 83, North American Datum of 1983; Y, yes; N, no; --, not measured, known, or applicable; *, sampled 6 times for temporal response to rainfall; psw, public-supply well; **, sampled 9 times for vertically discrete samples; ft, feet; NAVD 1988, North American Vertical Datum of 1988; gal/min, gallons per minute; in., inches; PVC, polyvinyl chloride; †, borehole diameter (diam.), screen is 2.5-in. diam.; ‡, borehole diam., screen is 4.5-in. diam.]

USGS identification number	Well type	Water use	Casing material	Land- surface altitude (ft above NAVD 1988)	Rated pump capacity (gal/min)	Well depth (ft below land surface)	Depth to top of open or screened interval (ft from land surface)	Depth to bot- tom of open or screened interval (ft from land surface)	Length of open interval (ft)	Diam- eter of open interval (in.)	Formation(s) of open interval
Monitoring wells											
292943098354401	Observation	Unused	Steel, PVC	848	--	280	203	243	40	11†	Buda Limestone
292943098354402	Observation	Unused	Steel, PVC	848	--	345	335	345	10	5.9†	Cyclic and marine members, undivided
292943098354403	Observation	Unused	Steel, PVC	848	--	480	397	417.5	20.5	5.9†	Leached and collapsed members, undivided
292943098354404	Observation	Unused	Steel, PVC	848	--	616	586	596.5	10.5	5.9†	Kirschberg evaporite member
292851098374401	Observation	Unused	Steel, PVC	795	--	520	480	510	30	5.9†	Leached and collapsed members, undivided
292851098374402	Observation	Unused	Steel, PVC	795	--	660	645	655	10	5.9†	Kirschberg evaporite member
293252098380801	Observation	Unused	Steel, PVC	896	--	229	149	229	80	6.5‡	Leached and collapsed members, undivided to grainstone member
Well-field wells											
292919098360501	Withdrawal	Public	Steel	910	8,750	824	342	824	482	30	All of Edwards
292925098360201	Withdrawal	Public	Steel	890	7,778	808	322	808	486	30	All of Edwards
292916098360701	Withdrawal	Public	Steel	890	7,778	816	321	816	495	30	All of Edwards
292920098360601	Withdrawal	Public	Steel	903	13,958	795	341	795	454	36	All of Edwards
292923098360301	Withdrawal	Public	Steel	906	14,375	820	360	820	460	36	All of Edwards
Regional aquifer public-supply wells											
293359098290301	Withdrawal	Public	Steel	840	3,200	710	380	710	330	16	Most of Edwards
293358098231101	Withdrawal	Public	Steel	828	7,500	811	359	811	452	30	All of Edwards
293120098285801	Withdrawal	Public	Steel	758	8,701	848	408	848	440	30	All of Edwards
293145098224201	Withdrawal	Public	Steel	813	9,722	870	409	870	461	30	All of Edwards
293119098211201	Withdrawal	Public	Steel	895	3,100	877	530	877	347	--	Most of Edwards
292843098425101	Withdrawal	Public	Steel	955	8,264	760	460	750	290	36	Most of Edwards
292459098382101	Withdrawal	Public	Steel	754	8,472	1,040	594	1,040	446	30	All of Edwards
292822098325401	Withdrawal	Public	Steel	756	3,194	748	537	748	211	18	Top of Edwards
292944098292301	Withdrawal	Public	Steel	710	9,722	1,050	574	1,044	470	30	All of Edwards
292522098291901	Withdrawal	Public	Steel	640	8,958	1,114	774	1,114	340	24	Most of Edwards
292643098241801	Withdrawal	Public	Steel	686	3,472	1,150	1,012	1,150	138	20	Top of Edwards
292328098294601	Withdrawal	Public	Steel	601	8,472	1,800	1,333	1,800	467	30	Most of Edwards (from cyclic and marine mem- bers, undivided, to bottom)
294225098080301	Withdrawal	Public	Steel	680	2,100	365	92	365	273	--	One-half of Edwards
291210099475601	Withdrawal	Public	Steel	900	500	525	191	525	334	12	Most of Edwards

Table 1. Site information for wells sampled for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category and by aquifer category (unconfined, confined). USGS, U.S. Geological Survey; NAD 83, North American Datum of 1983; Y, yes; N, no; --, not measured, known, or applicable; *, sampled 6 times for temporal response to rainfall; psw, public-supply well; **, sampled 9 times for vertically discrete samples; ft, feet; NAVD 1988, North American Vertical Datum of 1988; gal/min, gallons per minute; in., inches; PVC, polyvinyl chloride; †, borehole diameter (diam.), screen is 2.5-in. diam.; ‡, borehole diam., screen is 4.5-in. diam.]

USGS identification number	Well type	Water use	Casing material	Land- surface altitude (ft above NAVD 1988)	Rated pump capacity (gal/min)	Well depth (ft below land surface)	Depth to top of open or screened interval (ft from land surface)	Depth to bot- tom of open or screened interval (ft from land surface)	Length of open interval (ft)	Diam- eter of open interval (in.)	Formation(s) of open interval
293111098340901	Withdrawal	Public	Steel	1,000	3,333	685	420	685	265	20	Half of Edwards
293512098291701	Withdrawal	Public	Steel	975	250	460	230	460	230	--	One-half of Edwards
292424098421501	Withdrawal	Public	Steel	835	1,710	735	548	735	187	--	Less than one-half of Edwards
292053098365501	Withdrawal	Public	Steel	668	3,472	1,916	1,380	1,916	536	30	All of Edwards
293807098155301	Withdrawal	Public	Steel	760	1,850	255	148	255	107	--	Top of Edwards
292604098563201	Withdrawal	Public	Steel	1,035	650	1,462	1,190	1,462	272	--	One-half of Edwards
292116099095501	Withdrawal	Public	Steel	910	1,200	1,418	1,206	1,418	212	--	One-half of Edwards
291232099470301	Withdrawal	Public	Steel	904	1,200	400	260	400	140	--	Top of Edwards
293128098473101	Withdrawal	Public	Steel	1,040	153	1,400	900	1,400	500	8	All of Edwards
293451098313201	Withdrawal	Public	Steel	963	160	500	40	500	460	--	All of Edwards
294604098060801	Withdrawal	Public	Steel	802	150	400	200	400	200	6.5	Top of Edwards
294019098114701	Withdrawal	Public	Steel	750	750	215	119	215	96	--	Top of Edwards
292215098580201	Withdrawal	Public	Steel	1,012	750	1,431	1,280	1,431	151	10	Very top of Edwards
292119098524901	Withdrawal	Public	Steel	758	320	715	700	715	15	--	Top of Edwards
291943099163301	Withdrawal	Public	Steel	884	265	1,289	920	1,289	369	--	Most of Edwards
291840099382601	Withdrawal	Public	Steel	1,000	700	698	206	698	492	--	All of Edwards
293518098332601	Withdrawal	Public	Steel	985	250	435	162	435	273	8	One-half of Edwards
293023098355401	Withdrawal	Public	Steel	840	1,320	450	397	450	53	--	Top of Edwards
293042098305201	Withdrawal	Public	Steel	815	1,500	787	448	787	339	20	Most of Edwards
292442098474501	Withdrawal	Public	Steel	900	2,450	680	392	680	288	15	One-half of Edwards
292405098371201	Withdrawal	Public	Steel	720	1,850	1,500	1,108	1,500	392	--	Most of Edwards
292931098274601	Withdrawal	Public	Steel	813	1,000	557	390	557	167	--	Top of Edwards
293729098173101	Withdrawal	Public	Steel	870	2,000	660	185	660	475	--	All of Edwards
291219099095601	Withdrawal	Public	Steel	900	500	2,350	2,100	2,350	250	8	One-half of Edwards

Table 2. Summary of constituents by group for wells sampled for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[By sample category. USGS, U.S. Geological Survey; *, did not analyze in first temporal sample; --, not sampled; **, did not analyze in first three temporal samples; psw, public-supply well; $^{87}\text{Sr}/^{86}\text{Sr}$, strontium-87/strontium-86; δD , delta deuterium; $\delta^{18}\text{O}$, delta oxygen-18; $\delta^{15}\text{N}$, delta nitrogen-15; NO_3 , nitrate]

USGS identification number	State well number	County	Purpose	Local identifier	Year sampled	Physico- chemi- cal constit- uents	Major ions	Trace ele- ments	Nutri- ents
Monitoring wells									
292943098354401	AY-68-36-135	Bexar	discrete interval	Z-OVB	2007	X	X	X	X
292943098354402	AY-68-36-133	Bexar	discrete interval	Z-SED	2007	X	X	X	X
			temporal		2009	X	X	X	X
292943098354403	AY-68-36-134	Bexar	discrete interval	Z-IED	2007	X	X	X	X
			temporal		2009	X	X	X	X
292943098354404	AY-68-36-132	Bexar	discrete interval	Z-DED	2007	X	X	X	X
			temporal		2009	X	X	X	X
292851098374401	AY-68-35-315	Bexar	discrete interval	T-IED	2008	X	X	X	X
292851098374402	AY-68-35-314	Bexar	discrete interval	T-DED	2008	X	X	X	X
293252098380801	AY-68-27-610	Bexar	temporal	Parkwood	2009	X	X	X	X
Well-field wells									
292919098360501	AY-68-36-103	Bexar	wellfield	W2	2007	X	X	X	X
292925098360201	AY-68-36-104	Bexar	psw survey	W3	2004	X	X	X	X
			selected resample		2005	X	X	--	--
			wellfield		2007	X	X	X	X
292916098360701	AY-68-36-105	Bexar	wellfield, depth dependent	W4	2007	X	X	X	X
292920098360601	AY-68-36-130	Bexar	wellfield	W5	2007	X	X	X	X
292923098360301	AY-68-36-131	Bexar	wellfield	W6	2007	X	X	X	X
			temporal		2009	X	X	X	X
Regional aquifer public-supply wells									
293359098290301	AY-68-29-414	Bexar	psw survey	--	2004	X	X	X	X
			selected resample	--	2005	X	--	--	--
293358098231101	AY-68-29-610	Bexar	psw survey	--	2004	X	X	X	X
			selected resample	--	2005	X	--	--	--
293120098285801	AY-68-29-714	Bexar	psw survey	--	2004	X	X	X	X
			selected resample	--	2005	X	--	--	--
293145098224201	AY-68-29-929	Bexar	psw survey	--	2004	X	X	X	X
293119098211201	AY-68-30-718	Bexar	psw survey	--	2004	X	X	X	X
292843098425101	AY-68-35-106	Bexar	psw survey	--	2004	X	X	X	X
292459098382101	AY-68-35-913	Bexar	psw survey	--	2004	X	X	X	X
292822098325401	AY-68-36-206	Bexar	psw survey	--	2004	X	X	X	X
			selected resample	--	2005	X	--	--	--
292944098292301	AY-68-37-105	Bexar	psw survey	--	2004	X	X	X	X
			selected resample	--	2005	X	--	--	--

Table 2. Summary of constituents by group for wells sampled for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

—Continued

[By sample category. USGS, U.S. Geological Survey; *, did not analyze in first temporal sample; --, not sampled; **, did not analyze in first three temporal samples; psw, public-supply well; $^{87}\text{Sr}/^{86}\text{Sr}$, strontium-87/strontium-86; δD , delta deuterium; $\delta^{18}\text{O}$, delta oxygen-18; $\delta^{15}\text{N}$, delta nitrogen-15; NO_3 , nitrate]

USGS identification number	State well number	County	Purpose	Local identifier	Year sampled	Physico- chemi- cal constit- uents	Major ions	Trace ele- ments	Nutri- ents
292522098291901	AY-68-37-426	Bexar	psw survey	--	2004	X	X	X	X
292643098241801	AY-68-37-601	Bexar	psw survey	--	2004	X	X	X	X
			selected resample	--	2005	X	--	--	--
292328098294601	AY-68-37-705	Bexar	psw survey	--	2004	X	X	X	X
294225098080301	DX-68-23-601	Comal	psw survey	--	2004	X	X	X	X
			selected resample	--	2005	X	--	--	--
291210099475601	YP-69-50-506	Uvalde	psw survey	--	2004	X	X	X	X
293111098340901	AY-68-28-807	Bexar	psw survey	--	2005	X	X	X	X
293512098291701	AY-68-29-109	Bexar	psw survey	--	2005	X	X	X	X
292424098421501	AY-68-35-810	Bexar	psw survey	--	2005	X	X	X	X
292053098365501	AY-68-44-110	Bexar	psw survey	--	2005	X	X	X	X
293807098155301	DX-68-22-901	Comal	psw survey	--	2005	X	X	X	X
292604098563201	TD-68-33-501	Medina	psw survey	--	2005	X	X	X	X
292116099095501	TD-69-47-305	Medina	psw survey	--	2005	X	X	X	X
291232099470301	YP-69-50-339	Uvalde	psw survey	--	2005	X	X	X	X
293128098473101	AY-68-26-814	Bexar	psw survey	--	2005	X	X	X	X
293451098313201	AY-68-28-601	Bexar	psw survey	--	2005	X	X	X	X
294604098060801	DX-68-16-708	Comal	psw survey	--	2005	X	X	X	X
294019098114701	DX-68-23-504	Comal	psw survey	--	2005	X	X	X	X
292215098580201	TD-68-41-103	Medina	psw survey	--	2005	X	X	X	X
292119098524901	TD-68-41-308	Medina	psw survey	--	2005	X	X	X	X
291943099163301	TD-69-46-601	Medina	psw survey	--	2005	X	X	X	X
291840099382601	YP-69-43-606	Uvalde	psw survey	--	2005	X	X	X	X
293518098332601	AY-68-28-203	Bexar	psw survey	--	2005	X	X	X	X
293023098355401	AY-68-28-702	Bexar	psw survey	--	2005	X	X	X	X
293042098305201	AY-68-28-913	Bexar	psw survey	--	2005	X	X	X	X
292442098474501	AY-68-34-803	Bexar	psw survey	--	2005	X	X	X	X
292405098371201	AY-68-36-704	Bexar	psw survey	--	2005	X	X	X	X
292931098274601	AY-68-37-124	Bexar	psw survey	--	2005	X	X	X	X
293729098173101	DX-68-30-215	Comal	psw survey	--	2005	X	X	X	X
291219099095601	TD-69-55-604	Medina	psw survey	--	2005	X	X	X	X

Table 2. Summary of constituents by group for wells sampled for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

—Continued

[By sample category. USGS, U.S. Geological Survey; *, did not analyze in first temporal sample; --, not sampled; **, did not analyze in first three temporal samples; psw, public-supply well; $^{87}\text{Sr}/^{86}\text{Sr}$, strontium-87/strontium-86; δD , delta deuterium; $\delta^{18}\text{O}$, delta oxygen-18; $\delta^{15}\text{N}$, delta nitrogen-15; NO_3 , nitrate]

USGS identification number	Dis- solved or- ganic carbon	Vola- tile or- ganic com- pounds	Fuel oxy- gen- ates	Soluble pesti- cide com- pounds	Polar pesti- cide com- pounds	Aceta- mide pesticide com- pounds	Waste- water indicator com- pounds	Age tra- cers	Dis- solved gases	$^{87}\text{Sr}/^{86}\text{Sr}$	δD , $\delta^{18}\text{O}$	$\delta^{15}\text{N}$ in NO_3 , $\delta^{18}\text{O}$ in NO_3	Ra- don	Ra- dium
Monitoring wells														
292943098354401	X	X	--	X	--	--	--	X	X	X	X	X	X	X
292943098354402	X	X	--	X	--	--	--	X	X	X	X	X	X	X
	--	X	--	X**	--	--	--	--	--	X*	X	--	--	--
292943098354403	X	X	--	X	--	--	--	X	X	X	X	X	X	X
	--	X	--	X**	--	--	--	--	--	X*	X	--	--	--
292943098354404	X	X	--	X	--	--	--	X	X	X	X	X	X	X
	--	X	--	X**	--	--	--	--	--	X*	X	--	--	--
292851098374401	X	X	--	X	--	--	--	X	X	X	X	X	X	X
292851098374402	X	X	--	X	--	--	--	X	X	X	X	X	X	X
293252098380801	--	X	--	X**	--	--	--	--	--	X*	X	--	--	--
Well-field wells														
292919098360501	X	X	--	X	--	--	--	X	X	X	X	X	--	X
292925098360201	X	X	X	X	X	--	X	--	--	--	--	--	--	--
	--	X	--	X	X	--	--	--	--	--	--	--	--	--
	X	X	--	X	--	--	X	X	X	X	X	X	--	X
292916098360701	X	X	--	X	--	--	--	X	X	X	X	X	X	X
292920098360601	X	X	--	X	--	--	--	X	X	X	X	X	--	X
292923098360301	X	X	--	X	--	--	--	X	X	X	X	X	--	X
	--	X	--	X**	--	--	--	--	--	X	X*	--	--	--
Regional aquifer public-supply wells														
293359098290301	X	X	X	X	X	--	X	--	--	--	--	--	--	--
	--	X	--	X	X	--	X	--	--	--	--	--	--	--
293358098231101	X	X	X	X	X	--	X	--	--	--	--	--	--	--
	--	X	--	X	X	--	X	--	--	--	--	--	--	--
293120098285801	X	X	X	X	X	--	X	--	--	--	--	--	--	--
	--	X	--	--	X	--	--	--	--	--	--	--	--	--
293145098224201	X	X	X	X	X	--	X	--	--	--	--	--	--	--
293119098211201	X	X	X	X	X	--	X	--	--	--	--	--	--	--
292843098425101	X	X	X	X	X	--	X	--	--	--	--	--	--	--
292459098382101	X	X	X	X	X	--	X	--	--	--	--	--	--	--
292822098325401	X	X	X	X	X	--	X	--	--	--	--	--	--	--
	--	X	--	X	--	--	--	--	--	--	--	--	--	--
292944098292301	X	X	X	X	X	--	X	--	--	--	--	--	--	--
	--	--	--	X	--	--	X	--	--	--	--	--	--	--

Table 2. Summary of constituents by group for wells sampled for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

—Continued

[By sample category. USGS, U.S. Geological Survey; *, did not analyze in first temporal sample; --, not sampled; **, did not analyze in first three temporal samples; psw, public-supply well; $^{87}\text{Sr}/^{86}\text{Sr}$, strontium-87/strontium-86; δD , delta deuterium; $\delta^{18}\text{O}$, delta oxygen-18; $\delta^{15}\text{N}$, delta nitrogen-15; NO_3 , nitrate]

USGS identification number	Dis- solved or- ganic carbon	Vola- tile or- ganic com- pounds	Fuel oxy- gen- ates	Soluble pesti- cide com- pounds	Polar pesti- cide com- pounds	Aceta- mide pesticide com- pounds	Waste- water indicator com- pounds	Age tra- cers	Dis- solved gases	$^{87}\text{Sr}/$ ^{86}Sr	δD , $\delta^{18}\text{O}$	$\delta^{15}\text{N}$ in NO_3 , $\delta^{18}\text{O}$ in NO_3	Ra- don	Ra- dium
292522098291901	X	X	X	X	X	--	X	--	--	--	--	--	--	--
292643098241801	X	X	X	X	X	--	X	--	--	--	--	--	--	--
	--	X	--	--	X	--	--	--	--	--	--	--	--	--
292328098294601	X	X	X	X	X	--	X	--	--	--	--	--	--	--
294225098080301	X	X	X	X	X	--	X	--	--	--	--	--	--	--
	--	X	--	X	--	--	--	--	--	--	--	--	--	--
291210099475601	X	X	X	X	X	--	X	--	--	--	--	--	--	--
293111098340901	X	X	X	X	--	X	--	--	--	--	--	--	--	--
293512098291701	X	X	X	X	--	X	X	--	--	--	--	--	--	--
292424098421501	X	X	X	X	--	X	X	--	--	--	--	--	--	--
292053098365501	X	X	X	X	--	X	--	--	--	--	--	--	--	--
293807098155301	X	X	X	X	--	X	X	--	--	--	--	--	--	--
292604098563201	X	X	X	X	--	X	X	--	--	--	--	--	--	--
292116099095501	X	X	X	X	--	X	X	--	--	--	--	--	--	--
291232099470301	X	X	X	X	--	X	X	--	--	--	--	--	--	--
293128098473101	X	X	X	X	--	X	X	--	--	--	--	--	--	--
293451098313201	X	X	X	X	--	X	X	--	--	--	--	--	--	--
294604098060801	X	X	X	X	--	X	X	--	--	--	--	--	--	--
294019098114701	X	X	X	X	--	X	X	--	--	--	--	--	--	--
292215098580201	X	X	X	X	--	X	X	--	--	--	--	--	--	--
292119098524901	X	X	X	X	--	X	X	--	--	--	--	--	--	--
291943099163301	X	X	X	X	--	X	X	--	--	--	--	--	--	--
291840099382601	X	X	X	X	--	X	X	--	--	--	--	--	--	--
293518098332601	X	X	X	X	--	X	X	--	--	--	--	--	--	--
293023098355401	X	X	X	X	--	X	X	--	--	--	--	--	--	--
293042098305201	X	X	X	X	--	X	X	--	--	--	--	--	--	--
292442098474501	X	X	X	X	--	X	X	--	--	--	--	--	--	--
292405098371201	X	X	X	X	--	X	X	--	--	--	--	--	--	--
292931098274601	X	X	X	X	--	X	X	--	--	--	--	--	--	--
293729098173101	X	X	X	X	--	X	--	--	--	--	--	--	--	--
291219099095601	X	X	X	X	--	X	X	--	--	--	--	--	--	--

Table 3. Summary of depth-dependent sampling for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[By sample category. USGS, U.S. Geological Survey; ft, feet; --, not applicable; DDS, depth-dependent sampling]

USGS identification number	State well number	Local identifier	Sample pump intake depth (ft below land surface)	Well depth (ft below land surface)	Depth to top of open interval (ft below land surface)	Depth to bottom of open interval (ft below land surface)	Length of open interval (ft below land surface)	Formation at pump intake depth	Contributing formation(s) or flow zone(s)	Hydro-geologic unit (HGU)	Comments
Zarzamora monitoring wells											
292943098354402	AY–68–36–133	Z–SED	--	345	335	345	10	--	Georgetown Formation and cyclic and marine members, undivided	2 and 3	--
292943098354403	AY–68–36–134	Z–IED	--	480	397	417.5	20.5	--	Bottom of cyclic and marine members, undivided	3	--
292943098354404	AY–68–36–132	Z–DED	--	616	586	596.5	10.5	--	Top of Kirschberg evaporite member	8	Unable to screen to bottom of open interval because of well collapse
Timberhill monitoring wells											
292851098374401	AY–68–35–315	T– IED	--	520	480	510	30	--	Mostly top of leached and collapsed members, undivided, and possibly some from bottom of cyclic and marine members, undivided	mostly 4 (possibly some 3)	Unable to screen shorter interval because of well collapse
292851098374402	AY–68–35–314	T–DED	--	660	645	655	10	--	Kirschberg evaporite member	8	--
Well-field public-supply well (W4)											
292916098360701	AY–68–36–105	W4 (DDS “well-head”)	320	816	320	320	496	Georgetown Formation	All contributing flow zones	2–10	Represents wellhead sample, above all (4) high-flow zones
		W4 (DDS shallow)	380	816	380	380	--	Near middle of cyclic and marine members	Cyclic and marine members, undivided	3	Targeted middle of cyclic and marine members where flow was being gained or lost to borehole (under different pumping conditions); below high-transmissivity zone at top of cyclic and marine members, undivided
		W4 (DDS intermediate)	430	816	430	430	--	Middle/top of leached and collapsed members	Leached and collapsed members, undivided	4 and 5	Targeted middle of leached and collapsed members, coincident with high flow zone; in middle of high transmissivity zone in leached and collapsed members, undivided
		W4 (DDS deep)	540	816	580	580	--	Grainstone member	Kirschberg evaporite member	8 (also possibly from 9 and 10)	Targeted flow upward from Kirschberg evaporite member or below; above two high-flow zones, one in Kirschberg evaporite member and one in dolomitic member

Table 4. Sample collection for temporal sampling for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2009.

Sample number	Sample collection dates
Pre-storm	4/9/2009–4/14/2009
Storm 1	10/4/2009–10/6/2009
Storm 2	10/7/2009–10/9/2009
Storm 3	10/17/2009–0/19/2009
Storm 4	10/28/2009–10/30/2009
Storm 5	11/17/2009–11/19/2009

Table 5. Summary of organic content and elemental composition from acid extractions for core materials collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[USGS, U.S. Geological Survey; ft, feet; LSD, land surface datum; <2 mm, less than 2 millimeters (size of particulates analyzed); g/kg, grams per kilogram; mg/kg, milligrams per kilogram; --, not analyzed; na, not available; N, normal; HCl, hydrochloric acid; %, percent; HNO₃, nitric acid]

USGS identification number	State well number	Local identi- fier	Sample identifier	Date collected	Geologic or hydrogeologic unit	Depth to top of sample interval (ft from LSD)	Depth to bot- tom of sample interval (ft from LSD)	Total carbon, <2 mm (g/kg)	Inorganic carbon, <2 mm (g/kg)
Carbon analysis									
292943098354403	AY-68-36-134	Z-IED	TX 175A	8/17/2007	Austin Chalk-Eagle Ford Group contact	175.0	176.5	104	103
			TX 175B			175.0	176.5	--	--
			TX 175C			175.0	176.5	--	--
			TX 209	8/18/2007	Eagle Ford Group-Buda Limestone contact	209.0	210.5	129	54
			TX 260	8/18/2007	Buda Limestone-Del Rio Clay contact	260.0	261.5	112	117
			TX 314A	8/18/2007	Del Rio Clay-Georgetown Formation contact	314.0	315.5	59	56
			TX 314B			314.0	315.5	--	--
			TX 314C			314.0	315.5	--	--
			TX 444	8/24/2007	Leached and collapsed members, undivided	444.0	446.0	118	118
292943098354404	AY-68-36-132	Z-DED	TX 526	8/13/2007	Regional dense member-grainstone member contact	525.0	527.0	100	101
			TX 550A	8/14/2007	Grainstone member	550.0	551.0	114	116
			TX 550B			550.0	551.0	--	--
			TX 550C			550.0	551.0	--	--
			TX 589	8/14/2007	Kirschberg evaporite member (fragments)	589.0	591.0	117	116

Table 5. Summary of organic content and elemental composition from acid extractions for core materials collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[USGS, U.S. Geological Survey; ft, feet; LSD, land surface datum; <2 mm, less than 2 millimeters (size of particulates analyzed); g/kg, grams per kilogram; mg/kg, milligrams per kilogram; --, not analyzed; na, not available; N, normal; HCl, hydrochloric acid; %, percent; HNO₃, nitric acid]

Sample identifier	Aluminum (mg/kg)	Barium (mg/kg)	Iron (mg/kg)	Potassium (mg/kg)	Manganese (mg/kg)	Sodium (mg/kg)	Silica (mg/kg)	Magnesium (mg/kg)	Lead (mg/kg)	Uranium (mg/kg)	Arsenic (mg/kg)	Selenium (mg/kg)	Calcium (g/kg)	Calcium carbonate (g/kg)	Magnesium carbonate (g/kg)	Calcium (percent)	Calcium carbonate (percent)	Magnesium carbonate (percent)
*Elemental analysis after hydrochloric acid (6N HCl, cold) leach																		
TX 175A	390	9.2	540	400	462	230	200	3,800	0.19	1.3	--	<0.5	290	724	13.2	29.0	72.4	1.32
TX 175B	330	8.6	490	340	430	190	180	3,300	.15	1.2	--	<.5	303	757	11.5	30.3	75.7	1.15
TX 175C	290	7.9	450	300	390	180	160	2,800	.17	1.4	--	<.5	308	769	9.72	30.8	76.9	.97
TX 209	1,100	17	4,700	880	120	160	450	1,000	.48	2.2	--	<.5	112	280	3.47	11.2	28.0	.35
TX 260	200	4.2	470	210	660	110	170	1,900	<.05	.7	--	<.5	350	874	6.59	35.0	87.4	.66
TX 314A	1,400	56	8,300	690	410	130	360	1,200	7.31	.9	--	<.5	158	395	4.16	15.8	39.5	.42
TX 314B	1,600	48	9,300	830	400	140	410	1,200	7.74	.9	--	<.5	156	390	4.16	15.6	39.0	.42
TX 314C	1,300	46	7,500	710	360	110	410	1,300	7.46	.9	--	<.5	162	405	4.51	16.2	40.5	.45
TX 444	34	1.3	<20	41	17	270	33	40,000	<.05	.3	--	<.5	109	272	139	10.9	27.2	13.9
TX 526	300	4.4	600	330	21	150	200	5,200	.47	1.3	--	<.5	260	649	18.0	26.0	64.9	1.80
TX 550A	35	1.4	110	16	15	47	140	1,100	.13	1.7	--	<.5	345	861	3.82	34.5	86.1	.38
TX 550B	64	1.5	110	13	27	50	130	1,100	.11	1.3	--	<.5	275	687	3.82	27.5	68.7	.38
TX 550C	35	1.5	120	16	15	48	100	1,100	.12	1.6	--	<.5	335	836	3.82	33.5	83.6	.38
TX 589	36	2.4	130	14	13	41	84	15,000	<.05	.8	--	<.5	333	832	52.1	33.3	83.2	5.21
*Elemental analysis after nitric acid (10% HNO ₃) leach																		
TX 175A	54	5.2	140	94	240	120	130	2,200	.05	.2	1.1	.10	150	375	7.63	15.0	37.5	.76
TX 175B	59	4.3	170	79	200	90	120	2,100	.04	.2	1.0	.13	164	410	7.29	16.4	41.0	.73
TX 175C	43	4.3	180	75	190	100	75	2,100	.06	.3	.9	.13	163	407	7.29	16.3	40.7	.73
TX 209	360	12	670	390	100	140	150	820	2.80	1.3	7.9	.75	96	240	2.85	9.60	24.0	.28
TX 260	31	2.4	140	52	340	44	44	1,700	.09	1.5	22	<.1	166	415	5.90	16.6	41.5	.59
TX 314A	380	27	740	330	290	110	160	1,000	1.70	.4	2.4	<.1	157	392	3.47	15.7	39.2	.35
TX 314B	420	33	740	370	330	120	200	1,000	1.60	.4	1.6	.25	153	382	3.47	15.3	38.2	.35
TX 314C	420	34	680	360	330	120	180	1,100	1.60	.4	1.4	<.1	155	387	3.82	15.5	38.7	.38
TX 444	<20	.3	<20	<5	<5	35	<20	5,900	.01	.1	9.9	<.1	15	37	20.5	1.50	3.75	2.05
TX 526	65	2.8	200	98	10	68	180	2,900	.18	.4	2.4	<.1	150	375	10.1	15.0	37.5	1.01
TX 550A	<20	1.0	<20	12	7	12	30	540	.06	.3	1.7	<.1	173	432	1.87	17.3	43.2	.19
TX 550B	<20	1.0	<20	12	8	13	40	540	<.01	.2	1.3	<.1	161	402	1.87	16.1	40.2	.19
TX 550C	<20	1.1	<20	11	8	13	30	560	<.01	.6	1.2	<.1	168	419	1.94	16.8	41.9	.19
TX 589	<20	2.1	<20	13	8	21	29	7,400	.01	<.1	.1	<.1	139	347	25.7	13.9	34.7	2.57

* Elemental analyses are based on leaches, not complete dissolution of samples (residue remaining after leach for most samples).

Table 6. Summary of X-ray diffraction results for core materials collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[USGS, U.S. Geological Survey; <, less than; mm, millimeter; %, percent]

USGS identification number	State well number	Sample identifier	Date collected	Geologic or hydrogeologic unit	Local identifier	Sample top—depth from land surface (feet)	Sample bottom—depth from land surface (feet)	Bulk X-ray diffraction	X-Ray diffraction <1 mm
292943098354403	AY–68–36–134	TX 175A	8/17/2007	Austin Chalk-Eagle Ford Group contact	Z–IED	175.0	176.5	Calcite 85.0%, dolomite 2.2%, quartz 3.8%, total clay 9.0%	Kaolinite 38.1%, illite 22.8%, smectite trace, and chlorite trace. Residual carbonate: calcite 37.8%, dolomite 1.4%
		TX 175B							Dark gray parting/coating: calcite 86.7%, dolomite 0.8%, quartz 4.7%, total clay 7.8%, feldspar trace
292943098354403	AY–68–36–134	TX 209	8/18/2007	Eagle Ford Group-Buda Limestone contact	Z–IED	209.0	210.5	Calcite 52.7%, dolomite 2.0%, quartz 14.0%, total clay 21.8%, pyrite 9.5%	Kaolinite 58.4%, illite 20.4%, smectite 11.3%, chlorite trace. Residual carbonate: calcite 9.9%, dolomite trace
292943098354403	AY–68–36–134	TX 260A	8/18/2007	Buda Limestone-Del Rio Clay contact	Z–IED	260.0	261.5	Light chunks content: calcite 84.0%, dolomite trace, quartz 3.5%, total clay 12.7%	Composite light and dark: kaolinite 45.7%, illite 34.3%, anhydrite or oriented feldspar at trace level. Residual carbonate: calcite 19.2%, dolomite 0.8%
		260B						Dark chunks content: calcite 71.2%, dolomite 2.5%, total clay 26.3%, quartz trace	
292943098354403	AY–68–36–134	TX 314A	8/18/2007	Del Rio Clay-Georgetown Formation contact	Z–IED	314.0	315.5	Bulk yellow: calcite 42.2%, dolomite trace, quartz 44.3%, total clay 13.5%	Kaolinite 81.8%, illite 13.7%, smectite trace, chlorite trace, goethite trace, and hornblende trace. Residual carbonate: calcite 4.5%, dolomite trace
		TX 314B						Dark parting parallel to bedding: calcite 76.0%, possibly dolomite trace, quartz 9.8%, total clay 14.2%	
		TX 314C						Gray waxy nodules: calcite 44.8%, quartz 36.9%, total clay 18.3%, possibly trace goethite and dolomite	
		TX 314D						Bean-like red oxidized nodule with hard goethitic-looking inner core: calcite 3.9%, quartz 8.5%, total clay 4.2%, goethite 83.4%	
		TX 314E						Glassy semi-transparent shell-like nodule: calcite 87.3%, quartz 8.4%, total clay 2.8% (dominantly kaolinite)	

Table 6. Summary of X-ray diffraction results for core materials collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[USGS, U.S. Geological Survey; <, less than; mm, millimeter; %, percent]

USGS identification number	State well number	Sample identifier	Date collected	Geologic or hydrogeologic unit	Local identifier	Sample top—depth from land surface (feet)	Sample bottom—depth from land surface (feet)	Bulk X-ray diffraction	X-Ray diffraction <1 mm
292943098354403	AY-68-36-134	TX 444A TX 444B	8/24/2007	Leached and collapsed members, undivided	Z-IED	444.0	446.0	Calcite 52.4%, dolomite 31.5%, total clay 11.2%, and quartz 4.5%	Re-digestion; residual carbonate: calcite 52.4%, dolomite 31.5%, with traces of kaolinite 9.5%, illite 6.9%, smectite, and chlorite Residual carbonate: dolomite 76.1%, kaolinite 19.2%, illite 5.0%, and smectite and chlorite traces
292943098354404	AY-68-36-132	TX 526	8/13/2007	Regional dense member-grainstone member contact	Z-DED	525.0	526.0	Calcite 73.8%, dolomite 2.1%, quartz 6.5%, total clay 8.3%, pyrite 9.4%	Kaolinite 50.5%, illite 20.6%, smectite (poorly expandable) trace, chlorite trace, and hornblende trace. Residual carbonate: calcite 27.2%, dolomite 1.6%
292943098354404	AY-68-36-132	TX 550A TX 550B	8/14/2007	Grainstone member	Z-DED	550.0	551.0	Calcite 96.5%, dolomite trace, total clay 3.2%	Calcite dilutes trace amounts of clay. Re-digestion: kaolinite 14.1%, illite 2.2%. Residual carbonate: calcite 82.9%, dolomite 0.9% Kaolinite 35.8%, illite 3.5%, smectite trace. Calcite dominates the scan. Residual carbonate: calcite 58.9%, dolomite 0.9%
292943098354404	AY-68-36-132	TX 589A TX 589B TX 589C TX 589D	8/14/2007	Kirschberg evaporite member (fragments)	Z-DED	589.0	591.0	Calcite 94.3%, dolomite 2.5%, total clay 3.2%	Re-digestion (too much calcite). Traces of kaolinite 4.4%, illite 0.4%, and smectite. Residual carbonate: calcite 87.6%, dolomite 7.5% Re-digestion: kaolinite 17.8%, illite 0.7%, smectite trace. Residual carbonate: calcite 2.9%, dolomite 78.6% Dense yellow chunk: calcite 99.9%, traces of quartz, total clay, and possibly goethite. No weight percent calculation because of high calcite content. Principal peaks for trace minerals too close to background. Clear crystals in deep vug: calcite 64.0%, dolomite 25.8%, total clay 10.1% Red crystals in shallow vug: calcite 87.9%, dolomite 0.9%, traces of total clay 11.3% (slightly more clay than sample TX 589C)

Table 7. Physicochemical measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[By sample category. USGS, U.S. Geological Survey; PSW, public-supply well; ft, feet; NTRU, nephelometric turbidity ratiometric unit; mmHg, millimeters of mercury; mg/L, milligrams per liter; $\mu\text{S}/\text{cm}$, microsiemens per centimeter; $^{\circ}\text{C}$, degrees Celsius; CaCO_3 , calcium carbonate; --, not measured; E, estimated; MW, monitoring well; OVB, overburden]

USGS identification number	State well number	Group and description	Sample date	Depth to water (ft below land surface)	Turbid- ity (NTRU)	Air pres- sure (mmHg)	Dis- solved oxygen (mg/L)	pH (stan- dard units)	Specific conduc- tance ($\mu\text{S}/\text{cm}$ at 25°C)	Water temp- era- ture ($^{\circ}\text{C}$)	Alka- linity (mg/L as CaCO_3)
Regional aquifer public-supply wells											
293359098290301	AY-68-29-414	1st (highest) quartile	2004/12/06	--	--	--	2.76	7.0	642	23.0	292
293359098290301	AY-68-29-414	1st quartile	2005/09/01	--	0.26	743	6.63	6.6	661	23.0	289
293358098231101	AY-68-29-610	1st quartile	2004/12/15	--	--	761	3.74	7.0	618	22.0	285
293358098231101	AY-68-29-610	1st quartile	2005/08/29	--	.21	737	3.74	6.9	611	22.5	279
293120098285801	AY-68-29-714	1st quartile	2004/12/14	--	--	765	5.04	7.0	620	21.9	304
293120098285801	AY-68-29-714	1st quartile	2005/08/30	--	.17	740	4.73	6.9	599	22.0	264
293145098224201	AY-68-29-929	1st quartile	2004/12/15	--	--	761	5.52	7.2	519	24.0	260
293119098211201	AY-68-30-718	1st quartile	2004/12/16	--	--	750	5.58	7.2	511	25.3	226
292843098425101	AY-68-35-106	1st quartile	2004/12/13	--	--	756	3.89	7.2	560	22.2	226
292459098382101	AY-68-35-913	1st quartile	2004/12/13	--	--	745	5.84	7.3	498	24.2	201
292925098360201	AY-68-36-104	1st quartile	2004/12/13	--	--	756	4.50	7.1	596	22.3	250
292925098360201	AY-68-36-104	1st quartile	2005/08/31	--	.23	733	4.19	6.9	594	22.5	275
292822098325401	AY-68-36-206	1st quartile	2004/12/13	--	--	756	3.92	7.2	556	23.5	223
292822098325401	AY-68-36-206	1st quartile	2005/08/31	--	.25	742	4.65	6.9	557	23.5	208
292944098292301	AY-68-37-105	1st quartile	2004/12/14	--	--	765	5.07	7.1	552	23.0	250
292944098292301	AY-68-37-105	1st quartile	2005/08/31	--	.15	742	4.96	6.8	673	23.5	226
292522098291901	AY-68-37-426	1st quartile	2004/12/14	--	--	768	6.65	7.2	502	24.8	220
292643098241801	AY-68-37-601	1st quartile	2004/12/15	--	--	760	5.17	7.2	520	26.5	193
292643098241801	AY-68-37-601	1st quartile	2005/08/30	--	.28	742	4.92	7.0	509	26.5	196
292328098294601	AY-68-37-705	1st quartile	2004/12/14	--	--	765	4.57	7.2	531	26.7	207
294225098080301	DX-68-23-601	1st quartile	2004/12/08	--	--	750	2.15	7.1	565	23.4	243
294225098080301	DX-68-23-601	1st quartile	2005/09/01	--	.19	742	5.13	6.8	556	23.5	241
291210099475601	YP-69-50-506	1st quartile	2004/11/30	22.5	--	742	4.68	6.9	557	23.3	**E155
293111098340901	AY-68-28-807	2d quartile	2005/05/26	--	.97	740	1.65	7.0	573	22.5	266
293512098291701	AY-68-29-109	2d quartile	2005/04/20	--	.60	734	7.90	7.3	663	23.0	300
292424098421501	AY-68-35-810	2d quartile	2005/04/21	--	E.32	732	6.95	7.2	494	24.0	214
292053098365501	AY-68-44-110	2d quartile	2005/05/26	--	.78	740	1.53	7.2	492	26.0	209
293807098155301	DX-68-22-901	2d quartile	2005/05/18	--	.52	736	3.69	7.2	518	22.0	232
292604098563201	TD-68-33-501	2d quartile	2005/05/16	--	.27	742	2.51	7.4	494	23.5	189
292116099095501	TD-69-47-305	2d quartile	2005/04/27	--	.34	733	6.74	6.9	479	24.5	211
291232099470301	YP-69-50-339	2d quartile	2005/05/11	--	.39	731	5.72	7.1	548	23.5	211
293128098473101	AY-68-26-814	3d quartile	2005/04/18	--	.32	742	3.59	7.4	544	24.5	213
293451098313201	AY-68-28-601	3d quartile	2005/05/17	--	.52	731	6.28	7.1	593	22.5	268
294604098060801	DX-68-16-708	3d quartile	2005/04/25	--	.36	738	6.04	7.2	613	22.0	286
294019098114701	DX-68-23-504	3d quartile	2005/05/18	--	.36	738	5.09	7.2	586	23.0	258
292215098580201	TD-68-41-103	3d quartile	2005/05/09	--	.49	742	6.38	7.3	501	24.0	207
292119098524901	TD-68-41-308	3d quartile	2005/04/27	--	.44	736	7.46	6.9	502	24.0	209
291943099163301	TD-69-46-601	3d quartile	2005/05/09	--	.46	733	6.85	7.2	480	24.0	205
291840099382601	YP-69-43-606	3d quartile	2005/05/11	--	.31	729	6.69	7.2	508	23.5	199

Table 7. Physicochemical measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; PSW, public-supply well; ft, feet; NTRU, nephelometric turbidity ratiometric unit; mmHg, millimeters of mercury; mg/L, milligrams per liter; $\mu\text{S}/\text{cm}$, microsiemens per centimeter; $^{\circ}\text{C}$, degrees Celsius; CaCO_3 , calcium carbonate; --, not measured; E, estimated; MW, monitoring well; OVB, overburden]

USGS identification number	State well number	Group and description	Sample date	Depth to water (ft below land surface)	Turbidity (NTRU)	Air pressure (mmHg)	Dissolved oxygen (mg/L)	pH (standard units)	Specific conductance ($\mu\text{S}/\text{cm}$ at 25°C)	Water temperature ($^{\circ}\text{C}$)	Alkalinity (mg/L as CaCO_3)
293518098332601	AY-68-28-203	4th quartile	2005/04/19	--	2.60	735	9.39	7.0	740	22.5	286
293023098355401	AY-68-28-702	4th quartile	2005/04/26	--	.38	730	4.78	6.6	593	22.0	264
293042098305201	AY-68-28-913	4th quartile	2005/06/30	--	.79	744	5.34	6.9	617	22.5	258
292442098474501	AY-68-34-803	4th quartile	2005/04/20	--	.67	728	6.28	7.2	481	24.5	211
292405098371201	AY-68-36-704	4th quartile	2005/05/23	--	.37	736	6.52	7.2	495	25.0	207
292931098274601	AY-68-37-124	4th quartile	2005/05/25	--	.29	733	2.79	7.1	517	24.5	223
293729098173101	DX-68-30-215	4th quartile	2005/04/25	--	.42	726	7.14	7.1	633	22.0	281
291219099095601	TD-69-55-604	4th quartile	2005/05/16	--	.55	742	6.75	7.3	531	23.5	207
Monitoring wells											
292943098354401	AY-68-36-135	TANC MW OVB	2007/10/30	127.4	.77	753	4.98	7.0	971	28.3	284
292943098354402	AY-68-36-133	TANC MW	2007/10/31	138.2	.39	749	3.65	7.1	599	27.1	255
292943098354403	AY-68-36-134	TANC MW	2007/10/30	138.2	.39	753	4.60	7.1	592	25.1	264
292943098354404	AY-68-36-132	TANC MW	2007/11/01	138.6	.46	751	4.27	7.1	588	26.8	263
292851098374401	AY-68-35-315	TANC MW	2008/05/06	105.7	.43	758	4.59	7.0	587	24.7	256
292851098374402	AY-68-35-314	TANC MW	2008/05/07	104.1	.34	750	4.71	7.0	611	26.8	267
Well-field depth-dependent sampling (W4)											
292916098360701	AY-68-36-105	W4-shallow (380), normal	2007/11/16	182.6	.32	750	4.29	7.3	586	25.1	246
292916098360701	AY-68-36-105	W4-intermediate (430), normal	2007/11/15	183.1	.29	756	4.55	7.2	580	23.6	253
292916098360701	AY-68-36-105	W4-deep (540), normal	2007/11/15	182.8	.28	756	4.42	7.2	586	23.9	247
292916098360701	AY-68-36-105	W4-shallow (380), ambient	2007/11/16	182.6	.28	750	4.30	7.3	587	24.3	246
292916098360701	AY-68-36-105	W4-intermediate (430), ambient	2007/11/14	181.7	.35	756	4.00	7.3	585	25.6	248
292916098360701	AY-68-36-105	W4-deep (540), ambient	2007/11/15	182.8	.24	756	4.40	7.2	586	23.9	246
292916098360701	AY-68-36-105	W4-shallow (380), moderate	2007/11/16	182.7	.32	750	4.17	7.3	592	24.7	248
292916098360701	AY-68-36-105	W4-intermediate (430), moderate	2007/11/14	181.7	.29	753	4.31	7.2	582	24.4	242
292916098360701	AY-68-36-105	W4-deep (540), moderate	2007/11/15	182.8	.24	756	4.52	7.2	581	24.1	249
Well-field wells (wellhead sampling)											
292919098360501	AY-68-36-103	wellhead	2007/11/29	--	.29	752	4.30	7.3	589	22.0	252
292925098360201	AY-68-36-104	wellhead	2007/11/28	--	.66	750	5.26	7.2	587	20.8	256
292916098360701	AY-68-36-105	*wellhead	2007/11/16	182.7	.35	750	4.29	7.2	592	24.1	251
292920098360601	AY-68-36-130	wellhead	2007/11/29	--	.35	752	4.59	7.3	592	21.9	255
292923098360301	AY-68-36-131	wellhead	2007/11/28	--	.29	750	4.32	7.1	589	21.9	258
Temporal sampling											
292923098360301	AY-68-36-131	TANC PSW pre-storm	2009/04/09	--	.12	732	4.18	7.0	575	22.5	250
292923098360301	AY-68-36-131	TANC PSW storm 1	2009/10/06	--	.17	735	4.40	7.0	578	22.2	248

Table 7. Physicochemical measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; PSW, public-supply well; ft, feet; NTRU, nephelometric turbidity ratiometric unit; mmHg, millimeters of mercury; mg/L, milligrams per liter; $\mu\text{S}/\text{cm}$, microsiemens per centimeter; $^{\circ}\text{C}$, degrees Celsius; CaCO_3 , calcium carbonate; --, not measured; E, estimated; MW, monitoring well; OVB, overburden]

USGS identification number	State well number	Group and description	Sample date	Depth to water (ft below land surface)	Turbid- ity (NTRU)	Air pres- sure (mmHg)	Dis- solved oxygen (mg/L)	pH (stan- dard units)	Specific conduc- tance ($\mu\text{S}/\text{cm}$ at 25°C)	Water temp- era- ture ($^{\circ}\text{C}$)	Alka- linity (mg/L as CaCO_3)
292923098360301	AY-68-36-131	TANC PSW storm 2	2009/10/09	--	0.17	738	4.28	7.0	572	21.9	254
292923098360301	AY-68-36-131	TANC PSW storm 3	2009/10/19	--	.13	744	4.58	7.0	583	21.9	259
292923098360301	AY-68-36-131	TANC PSW storm 4	2009/10/30	--	.14	737	4.64	7.0	581	21.9	252
292923098360301	AY-68-36-131	TANC PSW storm 5	2009/11/19	--	.14	743	4.30	7.1	579	21.9	253
292943098354402	AY-68-36-133	TANC MW pre- storm	2009/04/08	177.7	.39	737	3.90	7.0	552	27.6	254
292943098354402	AY-68-36-133	TANC MW storm 1	2009/10/05	180.4	.21	752	4.15	7.0	592	23.8	274
292943098354402	AY-68-36-133	TANC MW storm 2	2009/10/07	177.9	.23	738	4.40	7.0	586	24.6	262
292943098354402	AY-68-36-133	TANC MW storm 3	2009/10/17	179.1	.24	749	4.07	7.1	592	24.6	269
292943098354402	AY-68-36-133	TANC MW storm 4	2009/10/28	170.0	.23	732	4.18	7.1	589	23.7	258
292943098354402	AY-68-36-133	TANC MW storm 5	2009/11/18	173.6	.15	743	4.29	7.1	586	23.0	275
292943098354403	AY-68-36-134	TANC MW pre- storm	2009/04/09	178.0	.23	732	4.15	6.9	572	26.1	265
292943098354403	AY-68-36-134	TANC MW storm 1	2009/10/05	179.4	.24	752	4.63	7.0	592	24.3	264
292943098354403	AY-68-36-134	TANC MW storm 2	2009/10/08	177.9	.13	737	4.59	7.0	587	24.1	262
292943098354403	AY-68-36-134	TANC MW storm 3	2009/10/18	175.3	.13	750	4.33	7.0	590	23.7	267
292943098354403	AY-68-36-134	TANC MW storm 4	2009/10/29	170.2	.12	733	4.50	7.0	590	24.0	261
292943098354403	AY-68-36-134	TANC MW storm 5	2009/11/18	173.3	.18	743	4.45	7.1	583	23.3	262
292943098354404	AY-68-36-132	TANC MW pre- storm	2009/04/09	178.0	1.67	732	4.12	7.0	582	25.2	262
292943098354404	AY-68-36-132	TANC MW storm 1	2009/10/05	180.5	.30	752	4.44	7.0	586	24.4	263
292943098354404	AY-68-36-132	TANC MW storm 2	2009/10/08	177.9	.18	737	4.81	7.0	583	24.6	266
292943098354404	AY-68-36-132	TANC MW storm 3	2009/10/18	177.5	.14	750	4.34	7.1	586	23.8	260
292943098354404	AY-68-36-132	TANC MW storm 4	2009/10/29	170.2	.15	733	4.55	7.0	586	23.5	259
292943098354404	AY-68-36-132	TANC MW storm 5	2009/11/18	174.2	.18	743	4.27	7.1	584	24.1	261
293252098380801	AY-68-27-610	TANC MW pre- storm	2009/04/14	189.7	.40	753	6.60	6.8	636	25.4	288
293252098380801	AY-68-27-610	TANC MW storm 1	2009/10/04	194.1***	2.63	738	6.35	6.8	663	24.3	308
293252098380801	AY-68-27-610	TANC MW storm 2	2009/10/07	198.5***	1.79	738	6.39	6.7	668	25.2	313
293252098380801	AY-68-27-610	TANC MW storm 3	2009/10/17	200.6***	.27	749	6.45	6.8	661	24.7	307
293252098380801	AY-68-27-610	TANC MW storm 4	2009/10/28	195.1***	.29	732	6.12	6.9	666	24.2	314
293252098380801	AY-68-27-610	TANC MW storm 5	2009/11/17	193.0***	.69	748	6.17	6.9	665	24.0	314

* Sample collected from top of open interval during depth-dependent sampling.

** Alkalinity is low and considered suspect.

*** Data from Edwards Aquifer Authority.

Table 8. Dissolved solids and major ion measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; PSW, public-supply well; mg/L, milligrams per liter; °C, degrees Celsius; --, not measured; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Group and description	Sample date	Dissolved solids (residue on evapora- tion, dried at 180°C) (mg/L)	Calcium (mg/L)	Magnesium (mg/L)	Potassium (mg/L)
Regional aquifer public-supply wells							
293359098290301	AY-68-29-414	1st (highest) quartile	2004/12/06	362	110	12.5	1.25
293359098290301	AY-68-29-414	1st quartile	2005/09/01	--	--	--	--
293358098231101	AY-68-29-610	1st quartile	2004/12/15	360	106	13.3	2.59
293358098231101	AY-68-29-610	1st quartile	2005/08/29	--	--	--	--
293120098285801	AY-68-29-714	1st quartile	2004/12/14	352	105	13.9	1.41
293120098285801	AY-68-29-714	1st quartile	2005/08/30	--	--	--	--
293145098224201	AY-68-29-929	1st quartile	2004/12/15	296	76.6	16.4	1.44
293119098211201	AY-68-30-718	1st quartile	2004/12/16	284	68.8	16.6	1.26
292843098425101	AY-68-35-106	1st quartile	2004/12/13	313	82.0	17.3	1.32
292459098382101	AY-68-35-913	1st quartile	2004/12/13	272	70.1	15.5	1.22
292925098360201	AY-68-36-104	1st quartile	2004/12/13	336	93.0	16.1	1.62
292925098360201	AY-68-36-104	1st quartile	2005/08/31	--	--	--	--
292822098325401	AY-68-36-206	1st quartile	2004/12/13	315	79.9	16.8	1.47
292822098325401	AY-68-36-206	1st quartile	2005/08/31	--	--	--	--
292944098292301	AY-68-37-105	1st quartile	2004/12/14	315	80.8	16.8	1.49
292944098292301	AY-68-37-105	1st quartile	2005/08/31	--	--	--	--
292522098291901	AY-68-37-426	1st quartile	2004/12/14	273	69.5	16.1	1.29
292643098241801	AY-68-37-601	1st quartile	2004/12/15	291	68.1	17.1	1.31
292643098241801	AY-68-37-601	1st quartile	2005/08/30	--	--	--	--
292328098294601	AY-68-37-705	1st quartile	2004/12/14	292	69.6	17.7	1.40
294225098080301	DX-68-23-601	1st quartile	2004/12/08	318	85.1	16.2	1.47
294225098080301	DX-68-23-601	1st quartile	2005/09/01	--	--	--	--
291210099475601	YP-69-50-506	1st quartile	2004/11/30	294	87.5	8.15	1.14
293111098340901	AY-68-28-807	2d quartile	2005/05/26	343	94.2	13.2	1.23
293512098291701	AY-68-29-109	2d quartile	2005/04/20	361	123	9.93	.79
292424098421501	AY-68-35-810	2d quartile	2005/04/21	279	73.4	14.3	1.19
292053098365501	AY-68-44-110	2d quartile	2005/05/26	278	66.4	18.2	1.19
293807098155301	DX-68-22-901	2d quartile	2005/05/18	299	83.6	11.1	.89
292604098563201	TD-68-33-501	2d quartile	2005/05/16	285	69.1	17.3	1.34
292116099095501	TD-69-47-305	2d quartile	2005/04/27	270	70.3	15.2	1.17
291232099470301	YP-69-50-339	2d quartile	2005/05/11	309	89.0	10.1	1.05
293128098473101	AY-68-26-814	3d quartile	2005/04/18	327	80.0	22.4	1.50
293451098313201	AY-68-28-601	3d quartile	2005/05/17	344	107	8.12	1.18
294604098060801	DX-68-16-708	3d quartile	2005/04/25	344	109	16.1	1.18
294019098114701	DX-68-23-504	3d quartile	2005/05/18	332	88.0	15.9	1.40

Table 8. Dissolved solids and major ion measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; PSW, public-supply well; mg/L, milligrams per liter; °C, degrees Celsius; --, not measured; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Group and description	Sample date	Dissolved solids (residue on evapora- tion, dried at 180°C) (mg/L)	Calcium (mg/L)	Magnesium (mg/L)	Potassium (mg/L)
292215098580201	TD-68-41-103	3d quartile	2005/05/09	293	71.6	16.0	1.12
292119098524901	TD-68-41-308	3d quartile	2005/04/27	293	71.8	15.6	1.20
291943099163301	TD-69-46-601	3d quartile	2005/05/09	265	72.5	15.3	1.04
291840099382601	YP-69-43-606	3d quartile	2005/05/11	290	80.5	11.2	1.01
293518098332601	AY-68-28-203	4th quartile	2005/04/19	435	141	7.8	1.20
293023098355401	AY-68-28-702	4th quartile	2005/04/26	346	101	15.5	1.58
293042098305201	AY-68-28-913	4th quartile	2005/06/30	339	100	16.5	1.31
292442098474501	AY-68-34-803	4th quartile	2005/04/20	265	74.5	13.7	1.19
292405098371201	AY-68-36-704	4th quartile	2005/05/23	275	69.3	17.0	1.11
292931098274601	AY-68-37-124	4th quartile	2005/05/25	300	73.4	16.9	1.34
293729098173101	DX-68-30-215	4th quartile	2005/04/25	360	112	11.6	1.38
291219099095601	TD-69-55-604	4th quartile	2005/05/16	292	71.8	15.1	1.10
Monitoring wells							
292943098354401	AY-68-36-135	TANC MW OVB	2007/10/30	600	105	13.0	3.10
292943098354402	AY-68-36-133	TANC MW	2007/10/31	349	84.7	14.8	1.79
292943098354403	AY-68-36-134	TANC MW	2007/10/30	341	89.4	15.9	1.54
292943098354404	AY-68-36-132	TANC MW	2007/11/01	351	89.7	16.4	1.48
292851098374401	AY-68-35-315	TANC MW	2008/05/06	350	93.1	15.8	1.52
292851098374402	AY-68-35-314	TANC MW	2008/05/07	349	98.2	14.8	1.79
Well-field depth-dependent sampling (W4)							
292916098360701	AY-68-36-105	W4-shallow (380), normal	2007/11/16	339	88.6	15.9	1.59
292916098360701	AY-68-36-105	W4-intermediate (430), normal	2007/11/15	343	88.4	16.0	1.56
292916098360701	AY-68-36-105	W4-deep (540), normal	2007/11/15	343	87.5	15.2	1.59
292916098360701	AY-68-36-105	W4-shallow (380), ambient	2007/11/16	343	88.2	15.9	1.56
292916098360701	AY-68-36-105	W4-intermediate (430), ambient	2007/11/14	233	86.4	15.7	1.54
292916098360701	AY-68-36-105	W4-deep (540), ambi- ent	2007/11/15	342	88.7	15.5	1.59
292916098360701	AY-68-36-105	W4-shallow (380), moderate	2007/11/16	341	93.1	15.5	1.66
292916098360701	AY-68-36-105	W4-intermediate (430), moderate	2007/11/14	255	86.9	15.8	1.54
292916098360701	AY-68-36-105	W4-deep (540), mod- erate	2007/11/15	339	86.2	15.4	1.48

Table 8. Dissolved solids and major ion measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; PSW, public-supply well; mg/L, milligrams per liter; °C, degrees Celsius; --, not measured; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Group and description	Sample date	Dissolved solids (residue on evapora- tion, dried at 180°C) (mg/L)	Calcium (mg/L)	Magnesium (mg/L)	Potassium (mg/L)
Well-field wells (wellhead sampling)							
292919098360501	AY-68-36-103	wellhead	2007/11/29	346	95.8	15.5	1.72
292925098360201	AY-68-36-104	wellhead	2007/11/28	349	92.9	16.6	1.59
292916098360701	AY-68-36-105	*wellhead	2007/11/16	343	89.9	15.1	1.57
292920098360601	AY-68-36-130	wellhead	2007/11/29	350	95.0	14.9	1.57
292923098360301	AY-68-36-131	wellhead	2007/11/28	357	94.0	15.4	1.59
Temporal sampling							
292923098360301	AY-68-36-131	TANC PSW pre-storm	2009/04/09	342	87.7	15.0	1.47
292923098360301	AY-68-36-131	TANC PSW storm 1	2009/10/06	316	87.6	15.6	1.54
292923098360301	AY-68-36-131	TANC PSW storm 2	2009/10/09	336	87.1	16.0	1.54
292923098360301	AY-68-36-131	TANC PSW storm 3	2009/10/19	371	92.6	15.7	1.42
292923098360301	AY-68-36-131	TANC PSW storm 4	2009/10/30	340	90.9	16.3	1.47
292923098360301	AY-68-36-131	TANC PSW storm 5	2009/11/19	350	90.9	16.1	1.51
292943098354402	AY-68-36-133	TANC MW pre-storm	2009/04/08	348	93.6	16.0	1.61
292943098354402	AY-68-36-133	TANC MW storm 1	2009/10/05	342	91.5	15.8	1.50
292943098354402	AY-68-36-133	TANC MW storm 2	2009/10/07	338	90.0	15.5	1.45
292943098354402	AY-68-36-133	TANC MW storm 3	2009/10/17	364	93.8	15.7	1.47
292943098354402	AY-68-36-133	TANC MW storm 4	2009/10/28	344	94.7	16.3	1.51
292943098354402	AY-68-36-133	TANC MW storm 5	2009/11/18	351	93.4	15.7	1.47
292943098354403	AY-68-36-134	TANC MW pre-storm	2009/04/09	345	92.9	14.8	1.43
292943098354403	AY-68-36-134	TANC MW storm 1	2009/10/05	338	89.4	15.5	1.43
292943098354403	AY-68-36-134	TANC MW storm 2	2009/10/08	353	88.8	15.6	1.48
292943098354403	AY-68-36-134	TANC MW storm 3	2009/10/18	367	94.0	15.6	1.47
292943098354403	AY-68-36-134	TANC MW storm 4	2009/10/29	344	96.5	16.6	1.47
292943098354403	AY-68-36-134	TANC MW storm 5	2009/11/18	353	92.8	15.8	1.40
292943098354404	AY-68-36-132	TANC MW pre-storm	2009/04/09	342	96.1	17.7	1.60
292943098354404	AY-68-36-132	TANC MW storm 1	2009/10/05	341	87.9	16.2	1.41
292943098354404	AY-68-36-132	TANC MW storm 2	2009/10/08	334	86.5	16.4	1.43
292943098354404	AY-68-36-132	TANC MW storm 3	2009/10/18	412	92.7	16.7	1.41
292943098354404	AY-68-36-132	TANC MW storm 4	2009/10/29	348	93.2	17.2	1.47
292943098354404	AY-68-36-132	TANC MW storm 5	2009/11/18	344	91.2	16.5	1.47
293252098380801	AY-68-27-610	TANC MW pre-storm	2009/04/14	370	104	13.3	.98
293252098380801	AY-68-27-610	TANC MW storm 1	2009/10/04	386	115	13.8	.95
293252098380801	AY-68-27-610	TANC MW storm 2	2009/10/07	379	115	13.6	.91
293252098380801	AY-68-27-610	TANC MW storm 3	2009/10/17	408	115	13.5	.95
293252098380801	AY-68-27-610	TANC MW storm 4	2009/10/28	388	118	13.8	.96
293252098380801	AY-68-27-610	TANC MW storm 5	2009/11/17	386	111	13.4	.92

Table 8. Dissolved solids and major ion measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; PSW, public-supply well; mg/L, milligrams per liter; °C, degrees Celsius; --, not measured; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	Sodium (mg/L)	Bromide (mg/L)	Chloride (mg/L)	Fluoride (mg/L)	Iodide (mg/L)	Sulfate (mg/L)	Silica (mg/L)	Calcite, mineral saturation index	Dolomite, mineral saturation index	Gypsum, mineral saturation index
Regional aquifer public-supply wells										
293359098290301	11.7	E0.20	18.3	0.21	0.003	16.2	14.4	0.11	-0.13	-2.38
293359098290301	--	--	--	--	--	--	--	--	--	--
293358098231101	10.3	.16	14.3	.23	.004	25.5	13.7	-.01	-.61	-2.05
293358098231101	--	--	--	--	--	--	--	--	--	--
293120098285801	9.85	.15	15.0	.22	.002	3.9	12.6	.02	-.53	-1.97
293120098285801	--	--	--	--	--	--	--	--	--	--
293145098224201	10.4	.15	16.6	.24	.002	22.1	14.0	.07	-.19	-2.22
293119098211201	11.0	.14	20.7	.28	.002	19.2	13.1	.02	-.22	-2.31
292843098425101	8.10	.14	12.1	.23	.002	39.1	13.2	.01	-.34	-1.95
292459098382101	9.44	.11	17.9	.20	.002	16.1	13.3	.02	-.28	-2.37
292925098360201	10.4	.15	15.6	.27	E.001	28.1	13.5	.01	-.43	-2.05
292925098360201	--	--	--	--	--	--	--	--	--	--
292822098325401	10.6	.13	16.3	.25	.002	32.3	13.9	.01	-.34	-2.04
292822098325401	--	--	--	--	--	--	--	--	--	--
292944098292301	10.9	.15	16.3	.23	.002	29.2	14.1	-.00	-.36	-2.08
292944098292301	--	--	--	--	--	--	--	--	--	--
292522098291901	10.7	.14	21.0	.23	.002	16.6	13.4	.02	-.28	-2.37
292643098241801	11.3	.15	23.6	.33	.002	24.2	13.5	-.00	-.24	-2.22
292643098241801	--	--	--	--	--	--	--	--	--	--
292328098294601	11.4	.15	24.0	.33	.002	27.7	13.6	-.02	-.26	-2.16
294225098080301	10.4	.14	16.8	.26	.002	22.5	13.6	.03	-.34	-2.17
294225098080301	--	--	--	--	--	--	--	--	--	--
291210099475601	15.2	.15	26.2	.16	.002	2.7	13.6	-.33	-1.36	-2.17
293111098340901	11.6	.12	15.4	.22	E.002	16.6	14.2	-.04	-.62	-2.27
293512098291701	12.4	.10	26.5	.15	<.002	1.7	14.7	.04	-.31	-2.31
292424098421501	9.46	.081	16.8	.22	<.002	16.6	13.4	.04	-.29	-2.34
292053098365501	10.1	.090	20.9	.28	<.002	15.5	13.3	-.07	-.33	-2.42
293807098155301	6.21	.090	10.5	.15	E.002	13.6	11.3	.01	-.54	-2.37
292604098563201	7.10	.085	10.8	.21	<.002	4.8	12.6	.12	-.03	-1.98
292116099095501	8.13	.071	12.7	.24	<.002	16.9	13.6	-.28	-.88	-2.35
291232099470301	13.9	.11	26.9	.12	<.002	14.4	13.6	-.00	-.62	-2.33
293128098473101	7.86	.075	14.0	.37	E.001	56.0	13.1	.24	.26	-1.82
293451098313201	8.42	.10	14.6	.16	E.001	15.2	13.7	.11	-.59	-2.25
294604098060801	7.91	.069	13.3	.20	<.002	15.2	13.3	.25	-.01	-2.27
294019098114701	7.91	.10	15.1	.21	E.001	22.6	13.4	.08	-.26	-2.16

Table 8. Dissolved solids and major ion measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; PSW, public-supply well; mg/L, milligrams per liter; °C, degrees Celsius; --, not measured; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	Sodium (mg/L)	Bromide (mg/L)	Chloride (mg/L)	Fluoride (mg/L)	Iodide (mg/L)	Sulfate (mg/L)	Silica (mg/L)	Calcite, mineral saturation index	Dolomite, mineral saturation index	Gypsum, mineral saturation index
292215098580201	9.56	0.086	18.6	0.21	<0.002	16.2	13.2	0.10	-0.11	-2.36
292119098524901	10.7	.10	20.8	.26	.002	16.5	13.2	-.31	-.94	-2.35
291943099163301	7.45	.070	12.0	.19	<.002	17.3	13.4	-.05	-.44	-2.33
291840099382601	10.9	.10	21.9	.13	<.002	13.1	13.2	.04	-.44	-2.40
293518098332601	20.2	.14	51.8	.12	E.002	15.8	14.2	.17	-.60	-2.17
293023098355401	10.0	.088	15.7	.26	E.001	24.2	13.0	-.40	-1.30	-2.09
293042098305201	10.5	.10	16.1	.22	E.001	37.4	12.4	-.11	-.69	-1.91
292442098474501	8.03	.037	11.9	.21	<.002	17.6	13.5	.04	-.31	-2.31
292405098371201	10.3	.10	20.0	.19	<.002	15.9	13.2	-.02	-.31	-2.39
292931098274601	10.6	.086	16.1	.21	<.002	22.3	13.8	-.05	-.40	-2.22
293729098173101	11.0	.083	20.4	.18	E.001	12.6	14.0	.19	-.29	-2.33
291219099095601	11.6	.12	25.4	.17	<.002	15.7	13.1	.12	-.03	-1.98
Monitoring wells										
292943098354401	78.4	.33	97.6	.86	.006	60.6	26.7	.07	-.37	-1.73
292943098354402	13.2	.10	16.5	.26	.004	27.7	11.3	.06	-.26	-2.09
292943098354403	10.7	.094	15.9	.23	E.001	25.4	11.7	.04	-.31	-2.11
292943098354404	9.89	.10	15.5	.26	E.001	25.0	10.9	.07	-.24	-2.12
292851098374401	10.6	.10	15.3	.22	E.001	28.4	11.8	-.03	-.48	-2.05
292851098374402	10.5	.11	15.0	.22	E.002	26.9	12.5	-.00	-.46	-2.06
Well-field depth-dependent sampling (W4)										
292916098360701	10.7	.095	15.6	.23	<.002	29.7	12.0	.23	.07	-2.05
292916098360701	10.8	.10	15.6	.23	<.002	29.7	11.7	.11	-.20	-2.04
292916098360701	10.9	.10	15.9	.24	E.001	29.3	11.7	.11	-.21	-2.05
292916098360701	10.6	.10	15.6	.23	<.002	29.7	11.9	.23	.06	-2.05
292916098360701	10.7	.10	15.5	.22	<.002	29.7	12.4	.19	-.01	-2.05
292916098360701	11.0	.10	15.9	.21	<.002	29.2	11.9	.11	-.20	-2.05
292916098360701	11.4	.11	16.1	.23	E.001	29.0	12.7	.24	.05	-2.04
292916098360701	10.8	.10	15.4	.23	<.002	29.5	12.4	.15	-.09	-2.05
292916098360701	10.4	.095	15.7	.21	<.002	29.8	11.8	.11	-.19	-2.05

Table 8. Dissolved solids and major ion measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; PSW, public-supply well; mg/L, milligrams per liter; °C, degrees Celsius; --, not measured; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	Sodium (mg/L)	Bromide (mg/L)	Chloride (mg/L)	Fluoride (mg/L)	Iodide (mg/L)	Sulfate (mg/L)	Silica (mg/L)	Calcite, mineral saturation index	Dolomite, mineral saturation index	Gypsum, mineral saturation index
Well-field wells (wellhead sampling)										
292919098360501	11.5	0.10	16.2	0.23	<0.002	28.8	12.4	0.18	-0.11	-2.03
292925098360201	10.8	.083	15.6	.24	<.002	28.1	11.8	.07	-.31	-2.05
292916098360701	11.0	.10	16.2	.22	E.001	29.0	12.3	.13	-.18	-2.05
292920098360601	11.4	.089	16.2	.22	<.002	28.1	12.3	.19	-.11	-2.04
292923098360301	11.2	.091	16.2	.24	<.002	27.1	12.2	.07	-.34	-2.06
Temporal sampling										
292923098360301	10.3	.090	16.4	.23	--	29.2	12.1	-.11	-.67	-2.05
292923098360301	10.4	.094	16.6	.24	--	28.4	12.7	-.09	-.62	-2.06
292923098360301	10.9	.10	16.1	.24	--	28.6	12.0	-.10	-.62	-2.06
292923098360301	10.6	.090	16.2	.26	--	27.3	12.0	-.04	-.53	-2.06
292923098360301	11.0	.091	16.2	.25	--	28.2	12.1	-.10	-.63	-2.05
292923098360301	10.8	.086	16.3	.27	--	29.1	12.0	.02	-.41	-2.04
292943098354402	11.4	.16	16.2	.19	--	25.9	12.4	-.04	-.47	-2.09
292943098354402	10.5	.11	16.6	.23	--	25.9	12.2	.01	-.41	-2.09
292943098354402	10.5	.11	16.6	.22	--	26.2	11.7	-.04	-.51	-2.09
292943098354402	10.4	.11	16.3	.24	--	25.9	11.8	.04	-.35	-2.09
292943098354402	10.9	.11	16.4	.22	--	25.6	11.8	.05	-.34	-2.09
292943098354402	10.6	.11	16.9	.20	--	26.9	11.8	.09	-.28	-2.07
292943098354403	9.86	.10	16.3	.23	--	25.9	11.6	-.07	-.57	-2.09
292943098354403	10.1	.10	16.6	.23	--	25.7	12.1	-.04	-.49	-2.10
292943098354403	10.4	.11	16.1	.23	--	25.7	11.4	-.07	-.55	-2.11
292943098354403	10.2	.11	16.2	.24	--	25.6	11.8	.01	-.43	-2.09
292943098354403	10.9	.10	16.4	.26	--	25.4	11.9	-.02	-.46	-2.09
292943098354403	10.3	.10	16.4	.25	--	25.9	11.6	.07	-.30	-2.09
292943098354404	10.9	.14	15.7	.20	--	27.2	12.5	-.03	-.45	-2.06
292943098354404	9.39	.10	16.2	.24	--	25.8	11.9	-.00	-.40	-2.11
292943098354404	9.90	.10	15.8	.25	--	25.9	11.2	-.05	-.48	-2.11
292943098354404	9.83	.10	15.7	.25	--	25.5	11.5	.02	-.36	-2.10
292943098354404	10.5	.10	16.0	.24	--	25.4	11.5	-.02	-.44	-2.10
292943098354404	9.89	.10	16.0	.26	--	26.2	11.4	.09	-.22	-2.09
293252098380801	8.83	.12	--	--	--	--	13.4	-.08	-.70	--
293252098380801	9.45	.13	17.2	.17	--	13.2	13.5	-.09	-.77	-2.31
293252098380801	9.54	.13	16.4	.16	--	14.7	13.0	-.13	-.84	-2.27
293252098380801	9.33	.13	16.9	.16	--	14.2	12.8	-.05	-.68	-2.29
293252098380801	10.2	.12	16.7	.15	--	14.6	13.0	-.01	-.61	-2.27
293252098380801	9.89	.15	18.5	.13	--	17.9	13.0	.01	-.57	-2.19

* Sample collected from top of open interval during depth-dependent sampling.

Table 9. Trace element measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; PSW, public-supply well; µg/L, micrograms per liter; --, not measured; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Group and description	Sample date	Aluminum (µg/L)	Antimony (µg/L)	Arsenic (µg/L)	Barium (µg/L)	Beryllium (µg/L)	Boron (µg/L)	Cadmium (µg/L)	Chromium (µg/L)	Cobalt (µg/L)
Regional aquifer public-supply wells												
293359098290301	AY-68-29-414	1st (highest) quartile	2004/12/06	E1.13	<0.2	0.46	43.0	<0.06	50.0	<0.04	<0.8	0.48
293359098290301	AY-68-29-414	1st quartile	2005/09/01	--	--	--	--	--	--	--	--	--
293358098231101	AY-68-29-610	1st quartile	2004/12/15	<1.6	<.2	.40	42.5	<.06	57.0	<.04	<.8	.25
293358098231101	AY-68-29-610	1st quartile	2005/08/29	--	--	--	--	--	--	--	--	--
293120098285801	AY-68-29-714	1st quartile	2004/12/14	<1.6	<.2	.39	34.8	<.06	41.8	<.04	<.8	.30
293120098285801	AY-68-29-714	1st quartile	2005/08/30	--	--	--	--	--	--	--	--	--
293145098224201	AY-68-29-929	1st quartile	2004/12/15	E1.26	<.2	.46	49.8	<.06	57.0	<.04	E.49	.18
293119098211201	AY-68-30-718	1st quartile	2004/12/16	<1.6	<.2	.54	103	<.06	50.7	<.04	<.8	.13
292843098425101	AY-68-35-106	1st quartile	2004/12/13	<1.6	<.2	.37	31.0	<.06	59.1	<.04	<.8	.22
292459098382101	AY-68-35-913	1st quartile	2004/12/13	E.80	<.2	.56	47.2	<.06	57.1	<.04	<.8	.18
292925098360201	AY-68-36-104	1st quartile	2004/12/13	<1.6	<.2	.45	37.0	<.06	62.5	<.04	<.8	.35
292925098360201	AY-68-36-104	1st quartile	2005/08/31	--	--	--	--	--	--	--	--	--
292822098325401	AY-68-36-206	1st quartile	2004/12/13	<1.6	<.2	.44	38.4	<.06	55.7	<.04	<.8	.22
292822098325401	AY-68-36-206	1st quartile	2005/08/31	--	--	--	--	--	--	--	--	--
292944098292301	AY-68-37-105	1st quartile	2004/12/14	<1.6	<.2	.36	40.8	<.06	69.0	<.04	<.8	.15
292944098292301	AY-68-37-105	1st quartile	2005/08/31	--	--	--	--	--	--	--	--	--
292522098291901	AY-68-37-426	1st quartile	2004/12/14	E.82	<.2	.57	58.5	<.06	43.2	<.04	<.8	.18
292643098241801	AY-68-37-601	1st quartile	2004/12/15	<1.6	<.2	.48	126	<.06	46.3	<.04	<.8	.14
292643098241801	AY-68-37-601	1st quartile	2005/08/30	--	--	--	--	--	--	--	--	--
292328098294601	AY-68-37-705	1st quartile	2004/12/14	<1.6	<.2	.57	121	<.06	67.0	E.02	<.8	.13
294225098080301	DX-68-23-601	1st quartile	2004/12/08	<1.6	<.2	.51	49.7	<.06	58.3	<.04	3.58	.17
291210099475601	YP-69-50-506	1st quartile	2004/11/30	<1.6	<.2	.68	60.8	<.06	89.5	<.04	<.8	.25
293111098340901	AY-68-28-807	2d quartile	2005/05/26	E1.34	<.2	.45	27.4	<.06	74.5	<.04	<.8	.19
293512098291701	AY-68-29-109	2d quartile	2005/04/20	<1.6	<.2	.75	52.7	<.06	42.6	<.04	<.8	.53
292424098421501	AY-68-35-810	2d quartile	2005/04/21	<1.6	<.2	.38	46.5	<.06	49.8	<.04	<.8	.23
292053098365501	AY-68-44-110	2d quartile	2005/05/26	E1.25	<.2	.53	60.1	<.06	70.8	<.04	<.8	.13
293807098155301	DX-68-22-901	2d quartile	2005/05/18	<1.6	<.2	.41	30.7	<.06	41.6	<.04	<.8	.27
292604098563201	TD-68-33-501	2d quartile	2005/05/16	E.80	<.2	.53	33.1	<.06	59.5	<.04	<.8	.19
292116099095501	TD-69-47-305	2d quartile	2005/04/27	<1.6	<.2	.48	38.6	<.06	52.2	<.04	<.8	.10
291232099470301	YP-69-50-339	2d quartile	2005/05/11	2.04	<.2	.62	53.0	<.06	64.6	<.04	<.8	.14
293128098473101	AY-68-26-814	3d quartile	2005/04/18	E.85	<.2	.36	30.7	<.06	78.7	<.04	<.8	.18
293451098313201	AY-68-28-601	3d quartile	2005/05/17	<1.6	<.2	.48	42.6	<.06	51.1	<.04	<.8	.29
294604098060801	DX-68-16-708	3d quartile	2005/04/25	<1.6	<.2	.38	34.6	<.06	46.3	<.04	<.8	.15
294019098114701	DX-68-23-504	3d quartile	2005/05/18	1.81	<.2	.42	46.9	<.06	65.1	<.04	E.65	.26
292215098580201	TD-68-41-103	3d quartile	2005/05/09	E1.03	<.2	.47	48.3	<.06	69.0	<.04	<.8	.13
292119098524901	TD-68-41-308	3d quartile	2005/04/27	E.84	<.2	.50	44.3	<.06	67.6	<.04	<.8	.11
291943099163301	TD-69-46-601	3d quartile	2005/05/09	<1.6	<.2	.38	37.1	E.03	50.8	<.04	<.8	.13
291840099382601	YP-69-43-606	3d quartile	2005/05/11	<1.6	<.2	.51	51.4	E.04	53.5	<.04	<.8	.13

Table 9. Trace element measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; PSW, public-supply well; µg/L, micrograms per liter; --, not measured; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Group and description	Sample date	Alum- inum (µg/L)	Ant- imony (µg/L)	Ars- enic (µg/L)	Bar- ium (µg/L)	Beryl- lium (µg/L)	Bo- ron (µg/L)	Cad- mium (µg/L)	Chrom- ium (µg/L)	Co- balt (µg/L)
293518098332601	AY-68-28-203	4th quartile	2005/04/19	E.84	<0.2	0.50	54.4	<0.06	58.5	<0.04	<0.8	0.33
293023098355401	AY-68-28-702	4th quartile	2005/04/26	<1.6	<.2	.34	33.0	<.06	60.0	<.04	<.8	.16
293042098305201	AY-68-28-913	4th quartile	2005/06/30	<1.6	<.2	.36	35.1	<.06	54.8	<.04	<.8	.24
292442098474501	AY-68-34-803	4th quartile	2005/04/20	<1.6	<.2	.74	42.5	<.06	52.0	<.04	<.8	.30
292405098371201	AY-68-36-704	4th quartile	2005/05/23	E.81	<.2	.44	51.2	<.06	59.7	<.04	<.8	.21
292931098274601	AY-68-37-124	4th quartile	2005/05/25	E.88	<.2	.47	33.9	<.06	76.5	<.04	<.8	.15
293729098173101	DX-68-30-215	4th quartile	2005/04/25	<1.6	<.2	.44	36.1	<.06	48.2	<.04	<.8	.18
291219099095601	TD-69-55-604	4th quartile	2005/05/16	<1.6	<.2	.57	58.9	<.06	64.6	<.04	<.8	.22
Monitoring wells												
292943098354401	AY-68-36-135	TANC MW OVB	2007/10/30	<1.6	<.14	.49	64.7	<.01	270	<.04	.39	.10
292943098354402	AY-68-36-133	TANC MW	2007/10/31	2.36	<.14	.34	40.9	<.01	60.0	<.04	.32	.14
292943098354403	AY-68-36-134	TANC MW	2007/10/30	<1.6	<.14	.31	36.1	<.01	56.2	<.04	.20	.06
292943098354404	AY-68-36-132	TANC MW	2007/11/01	2.22	.22	.32	35.9	<.01	49.8	E.02	.44	.05
292851098374401	AY-68-35-315	TANC MW	2008/05/06	<1.6	<.14	.35	35.9	<.01	56.8	<.04	.49	.05
292851098374402	AY-68-35-314	TANC MW	2008/05/07	<1.6	<.14	.34	39.5	<.01	62.0	<.04	1.08	.10
Well-field depth-dependent sampling (W4)												
292916098360701	AY-68-36-105	W4-shallow (380), normal	2007/11/16	26.5	<.42	.27	38.0	<.04	<30	<.12	<.36	<.10
292916098360701	AY-68-36-105	W4-intermediate (430), normal	2007/11/15	<1.6	<.14	.37	37.5	<.01	84.3	<.04	E.11	.05
292916098360701	AY-68-36-105	W4-deep (540), normal	2007/11/15	<1.6	<.14	.40	38.9	<.01	93.1	<.04	E.11	.05
292916098360701	AY-68-36-105	W4-shallow (380), ambient	2007/11/16	<1.6	<.14	.36	37.9	<.01	57.7	<.04	E.11	.07
292916098360701	AY-68-36-105	W4-intermediate (430), ambient	2007/11/14	<1.6	<.14	.35	36.9	<.01	57.6	<.04	.12	.05
292916098360701	AY-68-36-105	W4-deep (540), ambient	2007/11/15	<1.6	<.14	.37	38.6	<.01	58.0	<.04	.14	.06
292916098360701	AY-68-36-105	W4-shallow (380), moderate	2007/11/16	<1.6	<.14	.36	40.0	<.01	67.7	<.04	E.10	.05
292916098360701	AY-68-36-105	W4-intermediate (430), moderate	2007/11/14	<1.6	<.14	.34	37.1	<.01	56.9	<.04	.13	.04
292916098360701	AY-68-36-105	W4-deep (540), moderate	2007/11/15	<1.6	<.14	.36	37.2	<.01	59.5	<.04	E.10	.05
Well-field wells (wellhead sampling)												
292919098360501	AY-68-36-103	wellhead	2007/11/29	<1.6	<.14	.37	40.4	<.01	53.2	<.04	.14	.04
292925098360201	AY-68-36-104	wellhead	2007/11/28	<1.6	<.14	.35	36.6	<.01	63.3	<.04	E.10	.04
292916098360701	AY-68-36-105	*wellhead	2007/11/16	1.77	<.14	.38	40.3	<.01	60.6	<.04	E.12	.07
292920098360601	AY-68-36-130	wellhead	2007/11/29	<1.6	<.14	.35	39.7	<.01	63.8	<.04	E.09	.05
292923098360301	AY-68-36-131	wellhead	2007/11/28	E.92	<.14	.37	38.1	<.01	59.2	<.04	.15	.04

Table 9. Trace element measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; PSW, public-supply well; µg/L, micrograms per liter; --, not measured; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Group and description	Sample date	Alum- inum (µg/L)	Ant- imony (µg/L)	Ars- enic (µg/L)	Bar- ium (µg/L)	Beryl- lium (µg/L)	Bo- ron (µg/L)	Cad- mium (µg/L)	Chrom- ium (µg/L)	Co- balt (µg/L)
Temporal sampling												
292923098360301	AY-68-36-131	TANC PSW pre- storm	2009/04/09	<4	E.03	0.38	38.6	<0.02	55.6	<0.02	0.22	0.12
292923098360301	AY-68-36-131	TANC PSW storm 1	2009/10/06	11.8	E.03	.48	38.1	<.01	61.4	<.02	.19	.17
292923098360301	AY-68-36-131	TANC PSW storm 2	2009/10/09	<3.4	E.03	.41	36.4	<.01	58.8	<.02	.26	.15
292923098360301	AY-68-36-131	TANC PSW storm 3	2009/10/19	<3.4	E.03	.36	37.5	<.01	66.1	<.02	.16	.18
292923098360301	AY-68-36-131	TANC PSW storm 4	2009/10/30	<3.4	E.03	.39	36.8	<.01	56.6	<.02	E.11	.15
292923098360301	AY-68-36-131	TANC PSW storm 5	2009/11/19	<3.4	E.03	.43	39.2	<.01	62.6	<.02	E.12	.15
292943098354402	AY-68-36-133	TANC MW pre- storm	2009/04/08	<4	E.03	.34	40.2	<.02	57.5	<.02	.24	.14
292943098354402	AY-68-36-133	TANC MW storm 1	2009/10/05	7.35	E.03	.38	39.2	<.01	58.7	<.02	.31	.19
292943098354402	AY-68-36-133	TANC MW storm 2	2009/10/07	<3.4	E.03	.38	38.8	E.01	58.0	<.02	.22	.20
292943098354402	AY-68-36-133	TANC MW storm 3	2009/10/17	<3.4	E.03	.36	39.0	<.01	60.3	<.02	.29	.21
292943098354402	AY-68-36-133	TANC MW storm 4	2009/10/28	<3.4	E.03	.36	38.2	<.01	56.5	<.02	.14	.17
292943098354402	AY-68-36-133	TANC MW storm 5	2009/11/18	E2.22	E.03	.64	38.3	<.01	60.8	<.02	.16	.21
292943098354403	AY-68-36-134	TANC MW pre- storm	2009/04/09	<4	--	.34	36.1	<.02	52.8	<.02	.27	.13
292943098354403	AY-68-36-134	TANC MW storm 1	2009/10/05	<3.4	--	.39	36.7	<.01	57.0	<.02	.24	.24
292943098354403	AY-68-36-134	TANC MW storm 2	2009/10/08	<3.4	--	.37	35.7	<.01	61.2	<.02	.21	.19
292943098354403	AY-68-36-134	TANC MW storm 3	2009/10/18	<3.4	--	.36	36.0	<.01	61.1	<.02	.29	.20
292943098354403	AY-68-36-134	TANC MW storm 4	2009/10/29	<3.4	--	.38	36.3	<.01	58.5	<.02	E.11	.22
292943098354403	AY-68-36-134	TANC MW storm 5	2009/11/18	<3.4	--	.34	36.4	<.01	60.4	<.02	E.10	.16
292943098354404	AY-68-36-132	TANC MW pre- storm	2009/04/09	<4	--	.32	36.0	<.02	54.2	<.02	.26	.14
292943098354404	AY-68-36-132	TANC MW storm 1	2009/10/05	<3.4	--	.38	35.2	<.01	56.2	<.02	.23	.19
292943098354404	AY-68-36-132	TANC MW storm 2	2009/10/08	<3.4	--	.34	33.9	<.01	54.1	<.02	.24	.16
292943098354404	AY-68-36-132	TANC MW storm 3	2009/10/18	E2.26	--	.33	35.0	<.01	60.7	<.02	.21	.20
292943098354404	AY-68-36-132	TANC MW storm 4	2009/10/29	<3.4	--	.32	34.9	<.01	52.4	<.02	E.11	.16
292943098354404	AY-68-36-132	TANC MW storm 5	2009/11/18	3.68	--	.32	35.5	<.01	58.8	<.02	E.12	.21
293252098380801	AY-68-27-610	TANC MW pre- storm	2009/04/14	<4	<.04	.30	37.2	<.02	41.0	<.02	.20	.50
293252098380801	AY-68-27-610	TANC MW storm 1	2009/10/04	<3.4	<.054	.34	38.9	<.01	44.7	<.02	.57	.23
293252098380801	AY-68-27-610	TANC MW storm 2	2009/10/07	<3.4	<.054	.37	39.5	<.01	45.0	<.02	.42	.25
293252098380801	AY-68-27-610	TANC MW storm 3	2009/10/17	<3.4	E.03	.33	38.9	<.01	48.1	<.02	.33	.22
293252098380801	AY-68-27-610	TANC MW storm 4	2009/10/28	<3.4	<.054	.37	39.7	E.01	51.0	<.02	.18	.26
293252098380801	AY-68-27-610	TANC MW storm 5	2009/11/17	<3.4	<.054	.35	39.5	<.01	47.9	<.02	.17	.28

Table 9. Trace element measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; PSW, public-supply well; µg/L, micrograms per liter; --, not measured; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	Cop- per (µg/L)	Iron (µg/L)	Ferrous (II) iron (µg/L)	Total iron (µg/L)	Lead (µg/L)	Lith- ium (µg/L)	Mang- anese (µg/L)	Molyb- denum (µg/L)	Nick- el (µg/L)	Selen- ium (µg/L)	Sil- ver (µg/L)	Stron- tium (µg/L)	Thal- ium (µg/L)	Uran- ium (µg/L)	Vanad- ium (µg/L)	Zinc (µg/L)
Regional aquifer public-supply wells																
293359098290301	3.04	<6	--	--	0.30	3.07	<0.2	E0.36	2.78	E0.36	<0.20	232	<0.04	0.97	3.62	2.17
293359098290301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293358098231101	3.52	<6	--	--	2.16	7.92	<.2	.66	3.60	.52	<.20	353	.04	.95	2.73	5.59
293358098231101	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293120098285801	1.97	<6	--	--	1.03	3.67	<.2	.65	5.61	.70	<.20	421	<.04	.93	2.47	5.29
293120098285801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293145098224201	3.13	<6	--	--	1.29	4.44	<.2	1.28	5.12	.56	<.20	556	.08	.91	3.18	10.6
293119098211201	3.11	<6	--	--	.35	7.66	<.2	.90	.83	.69	<.20	1,100	.05	.77	3.30	2.34
292843098425101	2.06	<6	--	--	.26	4.31	<.2	.85	2.50	.63	<.20	598	<.04	.84	2.15	6.13
292459098382101	9.41	<6	--	--	4.45	4.00	<.2	.78	1.59	.68	<.20	609	E.02	.83	3.60	2.30
292925098360201	3.14	<6	--	--	1.28	4.65	<.2	.80	4.52	.69	<.20	538	<.04	.97	2.49	5.84
292925098360201	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292822098325401	4.59	<6	--	--	.92	4.91	<.2	.77	3.26	.66	<.20	555	<.04	.94	2.87	8.86
292822098325401	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292944098292301	5.43	<6	--	--	2.21	5.10	<.2	.73	1.41	.50	<.20	509	E.03	.90	2.81	23.6
292944098292301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292522098291901	1.58	<6	--	--	5.03	5.31	<.2	.87	2.48	.82	<.20	955	.04	.85	3.91	9.95
292643098241801	6.84	<6	--	--	1.80	10.6	<.2	2.52	1.49	.76	<.20	1,540	.07	.94	3.49	4.92
292643098241801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292328098294601	1.14	<6	--	--	3.47	12.5	<.2	7.79	8.74	.88	<.20	1,710	.14	1.16	4.41	2.00
294225098080301	2.88	<6	--	--	1.28	6.24	<.2	.67	.77	.79	<.20	598	E.03	.84	2.85	1.42
291210099475601	4.33	<6	--	--	.83	3.08	.23	.93	1.16	.80	<.20	324	.07	.88	6.08	5.08
293111098340901	6.32	<6	--	0.02	.98	4.52	<.2	.59	4.09	1.03	<.20	436	<.04	.78	2.96	8.16
293512098291701	3.90	<6	--	.04	.32	2.19	<.2	E.21	14.2	1.00	<.20	214	<.04	.91	5.31	2.77
292424098421501	2.00	<6	--	.03	.85	2.78	E.12	.64	4.47	E.34	<.20	477	E.03	.83	3.91	.95
292053098365501	1.45	<6	--	.05	.13	3.90	<.2	1.16	3.08	.91	<.20	1,470	<.04	.82	3.88	1.19
293807098155301	2.77	<6	--	.00	.16	2.82	<.2	E.40	3.95	.51	<.20	154	<.04	.56	2.53	4.19
292604098563201	2.01	<6	--	.00	.68	4.85	<.2	.77	2.77	.65	<.20	542	<.04	.72	2.51	.71
292116099095501	4.26	<6	--	.01	.83	2.68	<.2	.70	.28	.73	<.20	350	E.03	.77	2.88	5.43
291232099470301	2.68	<6	--	.01	.20	2.42	<.2	.62	1.13	.59	<.20	261	E.04	.79	5.81	1.89
293128098473101	1.63	6.78	--	<.01	.14	5.66	.34	1.11	3.40	.67	<.20	945	E.02	.92	3.60	4.18
293451098313201	2.56	<6	--	.01	.09	3.45	E.16	E.35	4.38	.61	<.20	235	<.04	.72	3.01	29.2
294604098060801	3.46	<6	--	.05	.55	3.19	<.2	.42	.37	.74	<.20	426	<.04	.86	2.64	2.69
294019098114701	2.48	<6	--	.02	.42	6.64	<.2	.63	4.40	.77	<.20	524	<.08	.74	2.73	2.73
292215098580201	5.54	<6	--	.03	.74	3.84	E.11	.65	1.51	.82	<.20	659	E.03	.72	4.69	11.3
292119098524901	0.41	<6	--	.01	.12	3.82	<.2	.60	.37	.68	<.20	549	E.03	.73	3.17	4.88
291943099163301	3.14	<6	--	.01	.26	2.57	<.2	.61	1.03	.54	<.20	343	<.04	.73	3.40	6.28
291840099382601	2.13	<6	--	.01	.20	2.14	<.2	.54	1.02	.66	<.20	402	E.02	.71	4.15	2.50

Table 9. Trace element measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

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USGS identification number	Cop- per (µg/L)	Iron (µg/L)	Ferrous (II) iron (µg/L)	Total iron (µg/L)	Lead (µg/L)	Lith- ium (µg/L)	Mang- anese (µg/L)	Molyb- denum (µg/L)	Nick- el (µg/L)	Selen- ium (µg/L)	Sil- ver (µg/L)	Stron- tium (µg/L)	Thal- ium (µg/L)	Uran- ium (µg/L)	Vanad- ium (µg/L)	Zinc (µg/L)
293518098332601	3.88	E3.61	--	0.10	0.30	2.75	0.58	E0.25	5.90	0.74	<0.20	228	<0.04	0.77	4.06	3.52
293023098355401	2.88	<6	--	.06	.26	3.55	<2	.76	.75	.63	<.20	499	<.04	.89	2.95	5.96
293042098305201	0.74	24.0	--	.03	.21	3.68	1.17	.67	3.24	.77	<.20	470	E.02	.84	2.41	1.06
292442098474501	4.09	<6	--	.05	.87	2.90	<2	.62	8.36	.87	<.20	366	<.04	.77	4.05	5.61
292405098371201	4.73	<6	--	.10	.88	4.32	<2	.65	1.65	.71	<.20	677	.04	.76	3.48	10.2
292931098274601	1.68	<6	--	.05	.26	3.81	<2	.68	3.38	.94	<.20	475	E.02	.79	3.12	1.25
293729098173101	3.35	E5.20	--	.11	.35	3.32	.95	E.34	.60	.76	<.20	236	E.03	.81	3.35	2.91
291219099095601	1.87	<6	--	.02	1.45	3.85	<2	.75	3.02	.91	<.20	890	E.04	.65	4.05	.64
Monitoring wells																
292943098354401	<1	<8	--	--	<.08	35.1	1.03	.89	.89	3.89	<.1	1,040	<.04	1.17	1.52	E1.25
292943098354402	<1	<8	--	--	<.08	4.86	34.9	1.06	2.89	.60	<.1	470	0.06	1.03	2.19	3.30
292943098354403	E.54	<8	--	--	<.08	3.82	E.19	.85	.66	.52	<.1	482	<.04	.91	2.17	1.99
292943098354404	<1	<8	--	--	<.08	3.50	.46	.93	1.03	.50	<.1	495	<.04	.94	2.12	2.60
292851098374401	<1	<8	--	--	<.08	4.32	.38	.81	1.57	.50	<.1	487	<.04	.80	2.12	<1.8
292851098374402	E.60	<8	--	--	<.08	4.89	1.83	1.01	2.30	.55	<.1	461	<.04	.82	2.12	E1.14
Well-field depth-dependent sampling (W4)																
292916098360701	<3	<8	--	--	<.24	<5	<.6	.67	<1.0	.54	<.3	492	<.12	.83	2.29	<5.4
292916098360701	<1	<8	--	--	<.08	5.05	<2	.74	.77	.57	<.1	499	<.04	.86	2.45	<1.8
292916098360701	<1	<8	--	--	<.08	5.34	E.16	.77	.67	.58	<.1	486	<.04	.88	2.47	<1.8
292916098360701	<1	<8	--	--	<.08	3.90	E.14	.72	.83	.58	<.1	530	<.04	.83	2.41	<1.8
292916098360701	<1	<8	--	--	<.08	4.01	.85	.73	.61	.53	<.1	489	<.04	.82	2.32	<1.8
292916098360701	<1	<8	--	--	<.08	3.91	<2	.76	.76	.58	<.1	480	<.04	.86	2.51	<1.8
292916098360701	<1	<8	--	--	<.08	3.86	<2	.69	.49	.58	<.1	528	E.02	.89	2.47	<1.8
292916098360701	<1	<8	--	--	<.08	3.93	.99	.72	.59	.50	<.1	497	<.04	.83	2.36	<1.8
292916098360701	<1	<8	--	--	<.08	4.16	<2	.73	.69	.57	<.1	488	<.04	.86	2.44	<1.8
Well-field wells (wellhead sampling)																
292919098360501	1.39	<8	--	--	.16	3.75	<2	.68	.71	.57	<.1	504	<.04	.90	2.23	3.84
292925098360201	4.26	<8	--	--	1.53	4.48	<2	.75	.69	.57	<.1	496	<.04	.87	2.31	2.93
292916098360701	<1	<8	--	--	<.08	3.98	E.10	.73	.83	.61	<.1	517	<.04	.88	2.54	E.92
292920098360601	4.38	<8	--	--	.90	4.26	<2	.74	.81	.59	<.1	503	<.04	.91	2.44	2.03
292923098360301	3.71	<8	--	--	.39	3.99	.25	.72	.59	.59	<.1	518	<.04	.90	2.34	E1.52

Table 9. Trace element measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

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Temporal sampling																
292923098360301	3.36	<4	--	--	0.42	3.64	E0.11	0.73	0.67	0.60	<0.01	487	<.04	0.89	2.46	<2
292923098360301	4.85	<6	--	--	.56	4.20	<.26	.74	1.16	.67	.01	467	E.02	.89	2.57	E2.40
292923098360301	3.67	<6	--	--	.44	3.83	<.26	.66	1.14	.64	<.01	456	E.02	.82	2.29	E2.00
292923098360301	4.07	<6	--	--	.47	4.45	<.26	.70	1.07	.63	<.01	495	E.02	.88	2.39	E1.88
292923098360301	3.33	<6	--	--	.34	4.17	<.26	.71	1.02	.59	<.01	457	E.02	.87	2.36	E2.69
292923098360301	3.99	<6	--	--	.50	3.76	<.26	.72	1.12	.61	<.01	498	E.02	.88	2.39	E2.29
292943098354402	<1	E3.03	--	--	<.06	3.87	1.01	.85	1.30	.57	<.01	496	E.03	.96	2.22	<2
292943098354402	<1	<6	--	--	<.03	4.06	.51	.80	1.67	.64	E.01	463	.04	.93	2.16	<2.8
292943098354402	1.03	<6	--	--	<.03	3.93	.36	.79	1.84	.66	<.01	471	.04	.90	2.22	<2.8
292943098354402	<1	<6	--	--	<.03	4.11	.40	.78	1.68	.57	<.01	477	.04	.89	2.27	E1.89
292943098354402	<1	<6	--	--	<.03	4.15	.31	.76	1.46	.61	<.01	452	.03	.93	2.21	E1.43
292943098354402	<1	<6	--	--	<.03	3.88	.36	.79	1.57	.65	<.01	480	.03	.95	2.55	<2.8
292943098354403	<1	<4	--	--	<.06	3.32	E.14	.75	.81	.58	<.01	467	<.04	.90	2.25	<2
292943098354403	<1	<6	--	--	<.03	3.84	<.26	.79	1.55	.70	<.01	461	E.02	.90	2.20	<2.8
292943098354403	<1	<6	--	--	<.03	4.05	<.26	.78	1.38	.65	<.01	465	E.02	.88	2.23	<2.8
292943098354403	<1	<6	--	--	<.03	4.02	<.26	.72	1.15	.59	<.01	470	E.02	.87	2.30	<2.8
292943098354403	<1	<6	--	--	<.03	4.12	<.26	.79	1.25	.61	<.01	470	E.02	.92	2.34	<2.8
292943098354403	<1	<6	--	--	E.02	3.82	<.26	.79	1.18	.56	<.01	478	E.02	.92	2.17	<2.8
292943098354404	<1	<4	--	--	<.06	3.61	E.19	.83	1.30	.54	<.01	511	<.04	.93	2.16	<2
292943098354404	<1	<6	--	--	<.03	3.73	<.26	.77	1.32	.63	<.01	473	E.02	.89	2.21	<2.8
292943098354404	<1	<6	--	--	<.03	3.47	<.26	.76	1.27	.60	<.01	467	E.01	.86	2.08	<2.8
292943098354404	<1	<6	--	--	<.03	4.00	<.26	.78	1.17	.56	<.01	488	E.02	.87	2.25	<2.8
292943098354404	<1	<6	--	--	<.03	3.76	<.26	.76	1.10	.57	<.01	461	E.02	.90	2.17	<2.8
292943098354404	<1	<6	--	--	<.03	3.69	<.26	.82	1.21	.55	E.01	494	E.02	.91	2.12	<2.8
293252098380801	<1	<4	--	--	<.06	2.96	E.12	.51	.90	.44	<.01	318	<.04	.83	2.41	<2
293252098380801	<1	E4.28	--	--	<.03	3.21	E.16	.56	1.92	.55	<.01	310	.02	.87	2.36	<2.8
293252098380801	<1	<6	--	--	<.03	3.11	<.26	.48	1.97	.53	<.01	329	.02	.83	2.49	<2.8
293252098380801	<1	<6	--	--	<.03	3.36	<.26	.41	1.34	.44	<.01	323	.02	.85	2.46	<2.8
293252098380801	<1	<6	--	--	<.03	3.72	<.26	.42	1.49	.48	<.01	325	.02	.89	2.52	<2.8
293252098380801	<1	E3.17	--	--	<.03	3.16	<.26	.43	1.67	.47	<.01	323	.02	.91	2.36	<2.8

* Sample collected from top of open interval during depth-dependent sampling.

Table 10. Dissolved gas and associated data for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[By sample category. mg/L, milligrams per liter; cm³ STP/kg, cubic centimeters per kilogram at standard temperature (20°C) and pressure (1 atmosphere); --, not measured or calculated]

USGS identification number	State well number	Local identifier and description	Sample date	Methane (mg/L)	Carbon dioxide (mg/L)	Nitrogen (mg/L)	*Oxygen (mg/L)	Argon (mg/L)	Calcu- lated excess nitrogen (mg/L)	Calculated excess air (cm ³ STP/ kg)
Monitoring wells										
292943098354401	AY-68-36-135	Z-OVB	10/30/2007	<0.0005	51.3	23.4	3.40	0.76	0	6.2
292943098354402	AY-68-36-133	Z-SED	10/31/2007	<.0005	37.9	22.8	1.60	.71	0	7.3
292943098354403	AY-68-36-134	Z-IED	10/30/2007	<.0005	43.7	22.6	.21	.70	0	7.3
292943098354404	AY-68-36-132	Z-DED	11/1/2007	<.0005	37.6	21.3	2.16	.68	0	5.9
292851098374401	AY-68-35-315	T-IED	5/6/2008	<.0005	41.5	26.8	1.25	.79	0	11.0
292851098374402	AY-68-35-314	T-DED	5/7/2008	<.0005	37.7	21.2	1.40	.68	0	5.4
Well-field wells (wellhead sampling)										
292919098360501	AY-68-36-103	W2	11/29/2007	<.0005	27.6	16.6	6.05	.58	0	1.9
292925098360201	AY-68-36-104	W3	11/28/2007	<.0005	34.6	19.8	5.00	.65	0	4.8
292916098360701	AY-68-36-105	**W4	11/16/2007	<.0005	39.1	20.9	2.00	.67	0	5.8
292920098360601	AY-68-36-130	W5	11/29/2007	<.0005	37.7	20.5	4.91	.67	0	5.1
292923098360301	AY-68-36-131	W6	11/28/2007	<.0005	37.3	20.3	4.82	.66	0	5.0
Well-field depth-dependent sampling (W4)										
292916098360701	AY-68-36-105	W4-shallow (380), ambient	11/16/2007	<.0005	36.3	21.2	3.17	.68	0	5.9
292916098360701	AY-68-36-105	W4-intermediate (430), ambient	11/14/2007	--	--	--	--	--	--	--
292916098360701	AY-68-36-105	W4-deep (540), ambient	11/15/2007	<.0005	36.2	20.8	2.78	.67	0	5.3
292916098360701	AY-68-36-105	W4-shallow (380), moder- ate	11/16/2007	<.0005	38.5	20.9	2.42	.67	0	5.7
292916098360701	AY-68-36-105	W4 intermediate (430), moder- ate	11/14/2007	<.0005	34.6	21.3	3.97	.68	0	5.9
292916098360701	AY-68-36-105	W4-deep (540), moderate	11/15/2007	<.0005	34.6	20.8	3.66	.67	0	5.9
292916098360701	AY-68-36-105	W4-shallow (380), normal	11/16/2007	<.0005	36.2	21.1	3.01	.68	0	5.6
292916098360701	AY-68-36-105	W4-intermediate (430), normal	11/15/2007	<.0005	35.4	21.4	3.95	.68	0	6.0
292916098360701	AY-68-36-105	W4-deep (540), normal	11/15/2007	<.0005	36.5	20.5	2.71	.67	0	5.3

* Similar to dissolved oxygen in table 9; however, determined by laboratory rather than field method.

** Sample collected from top of open interval during depth-dependent sampling.

Table 11. Nutrient measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; PSW, public-supply well; mg/L, milligrams per liter; N, nitrogen; P, phosphorus; --, not measured; E, estimated; V, contamination; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Group and description	Sample date	Nitrate plus nitrite (mg/L as N)	Ammonia (mg/L as N)	Nitrite (mg/L as N)	Orthophos- phate (mg/L as P)	Organic carbon (mg/L)
Regional aquifer public-supply wells								
293359098290301	AY–68–29–414	1st (highest) quartile	2004/12/06	2.16	<0.04	<0.008	0.011	V1.13
293359098290301	AY–68–29–414	1st quartile	2005/09/01	--	--	--	--	--
293358098231101	AY–68–29–610	1st quartile	2004/12/15	1.73	E.02	<.008	.007	V.43
293358098231101	AY–68–29–610	1st quartile	2005/08/29	--	--	--	--	--
293120098285801	AY–68–29–714	1st quartile	2004/12/14	1.59	<.04	<.008	.006	V.47
293120098285801	AY–68–29–714	1st quartile	2005/08/30	--	--	--	--	--
293145098224201	AY–68–29–929	1st quartile	2004/12/15	1.86	<.04	<.008	E.004	V.28
293119098211201	AY–68–30–718	1st quartile	2004/12/16	1.85	<.04	<.008	<.006	V.23
292843098425101	AY–68–35–106	1st quartile	2004/12/13	1.17	<.04	<.008	E.005	V.46
292459098382101	AY–68–35–913	1st quartile	2004/12/13	1.98	<.04	<.008	.006	V9.90
292925098360201	AY–68–36–104	1st quartile	2004/12/13	2.08	<.04	<.008	.006	V.48
292925098360201	AY–68–36–104	1st quartile	2005/08/31	--	--	--	--	--
292822098325401	AY–68–36–206	1st quartile	2004/12/13	1.69	<.04	<.008	E.005	V3.27
292822098325401	AY–68–36–206	1st quartile	2005/08/31	--	--	--	--	--
292944098292301	AY–68–37–105	1st quartile	2004/12/14	1.75	<.04	<.008	E.004	V.35
292944098292301	AY–68–37–105	1st quartile	2005/08/31	--	--	--	--	--
292522098291901	AY–68–37–426	1st quartile	2004/12/14	1.98	<.04	<.008	E.003	V.30
292643098241801	AY–68–37–601	1st quartile	2004/12/15	1.72	<.04	<.008	<.006	V.25
292643098241801	AY–68–37–601	1st quartile	2005/08/30	--	--	--	--	--
292328098294601	AY–68–37–705	1st quartile	2004/12/14	1.50	<.04	<.008	E.003	V.27
294225098080301	DX–68–23–601	1st quartile	2004/12/08	1.88	<.04	<.008	.006	V.47
294225098080301	DX–68–23–601	1st quartile	2005/09/01	--	--	--	--	--
291210099475601	YP–69–50–506	1st quartile	2004/11/30	3.12	<.04	<.008	.007	V12.8
293111098340901	AY–68–28–807	2d quartile	2005/05/26	2.34	<.04	<.008	E.003	V.47
293512098291701	AY–68–29–109	2d quartile	2005/04/20	2.32	<.04	<.008	.008	V.48
292424098421501	AY–68–35–810	2d quartile	2005/04/21	1.97	<.04	<.008	E.004	V.34
292053098365501	AY–68–44–110	2d quartile	2005/05/26	1.91	<.04	<.008	<.006	V.45
293807098155301	DX–68–22–901	2d quartile	2005/05/18	2.46	<.04	<.008	.012	V24.3
292604098563201	TD–68–33–501	2d quartile	2005/05/16	.73	<.04	<.008	<.006	V84.0
292116099095501	TD–69–47–305	2d quartile	2005/04/27	1.64	<.04	<.008	<.006	V.30
291232099470301	YP–69–50–339	2d quartile	2005/05/11	2.82	<.04	<.008	E.003	V.31
293128098473101	AY–68–26–814	3d quartile	2005/04/18	.95	<.04	<.008	<.006	V248
293451098313201	AY–68–28–601	3d quartile	2005/05/17	1.39	<.04	<.008	E.004	V58.7
294604098060801	DX–68–16–708	3d quartile	2005/04/25	1.85	<.04	<.008	<.006	V.46
294019098114701	DX–68–23–504	3d quartile	2005/05/18	1.80	<.04	<.008	E.004	V18.3
292215098580201	TD–68–41–103	3d quartile	2005/05/09	2.01	<.04	<.008	<.006	V.26

Table 11. Nutrient measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; PSW, public-supply well; mg/L, milligrams per liter; N, nitrogen; P, phosphorus; --, not measured; E, estimated; V, contamination; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Group and description	Sample date	Nitrate plus nitrite (mg/L as N)	Ammonia (mg/L as N)	Nitrite (mg/L as N)	Orthophosphate (mg/L as P)	Organic carbon (mg/L)
292119098524901	TD-68-41-308	3d quartile	2005/04/27	2.02	<0.04	<0.008	<0.006	V0.34
291943099163301	TD-69-46-601	3d quartile	2005/05/09	1.44	<.04	<.008	E.005	V.28
291840099382601	YP-69-43-606	3d quartile	2005/05/11	2.77	<.04	<.008	<.006	V.32
293518098332601	AY-68-28-203	4th quartile	2005/04/19	2.36	<.04	<.008	.006	V2.43
293023098355401	AY-68-28-702	4th quartile	2005/04/26	1.85	<.04	<.008	<.006	V.57
293042098305201	AY-68-28-913	4th quartile	2005/06/30	1.82	<.04	<.008	E.003	V15.6
292442098474501	AY-68-34-803	4th quartile	2005/04/20	1.81	<.04	<.008	<.006	V.37
292405098371201	AY-68-36-704	4th quartile	2005/05/23	1.98	<.04	<.008	<.006	V2.02
292931098274601	AY-68-37-124	4th quartile	2005/05/25	1.79	<.04	<.008	<.006	V.49
293729098173101	DX-68-30-215	4th quartile	2005/04/25	2.50	<.04	<.008	E.003	V.44
291219099095601	TD-69-55-604	4th quartile	2005/05/16	2.58	<.04	<.008	E.003	V92.5
Monitoring wells								
292943098354401	AY-68-36-135	TANC MW OVB	2007/10/30	3.83	<.02	E.001	.010	7.23
292943098354402	AY-68-36-133	TANC MW	2007/10/31	1.86	<.02	.013	.006	.53
292943098354403	AY-68-36-134	TANC MW	2007/10/30	1.87	<.02	<.002	.008	1.03
292943098354404	AY-68-36-132	TANC MW	2007/11/01	1.69	<.02	.005	.007	.44
292851098374401	AY-68-35-315	TANC MW	2008/05/06	1.97	E.01	<.002	.011	.41
292851098374402	AY-68-35-314	TANC MW	2008/05/07	2.74	<.02	<.002	.010	.60
Well-field depth-dependent sampling (W4)								
292916098360701	AY-68-36-105	W4-shallow (380), normal	2007/11/16	1.99	<.02	<.002	.008	E.36
292916098360701	AY-68-36-105	W4-intermediate (430), normal	2007/11/15	1.97	<.02	<.002	.008	E.36
292916098360701	AY-68-36-105	W4-deep (540), normal	2007/11/15	2.12	<.02	<.002	.009	E.38
292916098360701	AY-68-36-105	W4-shallow (380), ambient	2007/11/16	1.99	<.02	<.002	.008	E.35
292916098360701	AY-68-36-105	W4-intermediate (430), ambient	2007/11/14	1.97	<.02	<.002	E.006	.44
292916098360701	AY-68-36-105	W4-deep (540), ambient	2007/11/15	2.12	<.02	<.002	.009	.45
292916098360701	AY-68-36-105	W4-shallow (380), moderate	2007/11/16	2.18	<.02	<.002	.009	.43
292916098360701	AY-68-36-105	W4-intermediate (430), moderate	2007/11/14	1.97	<.02	<.002	.008	.56
292916098360701	AY-68-36-105	W4-deep (540), moderate	2007/11/15	1.97	<.02	<.002	.008	E.39
Well-field wells (wellhead sampling)								
292919098360501	AY-68-36-103	wellhead	2007/11/29	2.24	<.02	<.002	.008	E.40
292925098360201	AY-68-36-104	wellhead	2007/11/28	1.97	<.02	<.002	.008	.42

Table 11. Nutrient measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; PSW, public-supply well; mg/L, milligrams per liter; N, nitrogen; P, phosphorus; --, not measured; E, estimated; V, contamination; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Group and description	Sample date	Nitrate plus nitrite (mg/L as N)	Ammonia (mg/L as N)	Nitrite (mg/L as N)	Orthophos- phate (mg/L as P)	Organic carbon (mg/L)
292916098360701	AY-68-36-105	*wellhead	2007/11/16	2.21	<0.02	<0.002	0.009	0.43
292920098360601	AY-68-36-130	wellhead	2007/11/29	2.33	<.02	<.002	.008	E.35
292923098360301	AY-68-36-131	wellhead	2007/11/28	2.11	<.02	<.002	.009	E.39
Temporal sampling								
292923098360301	AY-68-36-131	TANC PSW pre- storm	2009/04/09	1.95	<.02	<.002	.013	--
292923098360301	AY-68-36-131	TANC PSW storm 1	2009/10/06	1.97	<.02	<.002	.010	--
292923098360301	AY-68-36-131	TANC PSW storm 2	2009/10/09	1.95	<.02	<.002	.011	--
292923098360301	AY-68-36-131	TANC PSW storm 3	2009/10/19	1.92	<.02	<.002	.009	--
292923098360301	AY-68-36-131	TANC PSW storm 4	2009/10/30	1.94	<.02	<.002	.009	--
292923098360301	AY-68-36-131	TANC PSW storm 5	2009/11/19	1.95	<.02	<.002	<.008	--
292943098354402	AY-68-36-133	TANC MW pre- storm	2009/04/08	2.00	<.02	<.002	.011	--
292943098354402	AY-68-36-133	TANC MW storm 1	2009/10/05	1.92	<.02	<.002	.010	--
292943098354402	AY-68-36-133	TANC MW storm 2	2009/10/07	1.89	<.02	<.002	.010	--
292943098354402	AY-68-36-133	TANC MW storm 3	2009/10/17	1.89	<.02	<.002	.009	--
292943098354402	AY-68-36-133	TANC MW storm 4	2009/10/28	1.89	<.02	<.002	.010	--
292943098354402	AY-68-36-133	TANC MW storm 5	2009/11/18	1.91	<.02	<.002	<.008	--
292943098354403	AY-68-36-134	TANC MW pre- storm	2009/04/09	1.93	<.02	<.002	.012	--
292943098354403	AY-68-36-134	TANC MW storm 1	2009/10/05	1.92	<.02	<.002	.011	--
292943098354403	AY-68-36-134	TANC MW storm 2	2009/10/08	1.91	<.02	<.002	.010	--
292943098354403	AY-68-36-134	TANC MW storm 3	2009/10/18	1.89	<.02	<.002	.012	--
292943098354403	AY-68-36-134	TANC MW storm 4	2009/10/29	1.88	<.02	<.002	.009	--
292943098354403	AY-68-36-134	TANC MW storm 5	2009/11/18	1.89	<.02	<.002	.009	--
292943098354404	AY-68-36-132	TANC MW pre- storm	2009/04/09	1.82	<.02	<.002	.011	--
292943098354404	AY-68-36-132	TANC MW storm 1	2009/10/05	1.81	<.02	<.002	.010	--
292943098354404	AY-68-36-132	TANC MW storm 2	2009/10/08	1.80	<.02	<.002	.010	--
292943098354404	AY-68-36-132	TANC MW storm 3	2009/10/18	1.76	<.02	<.002	.009	--
292943098354404	AY-68-36-132	TANC MW storm 4	2009/10/29	1.79	<.02	<.002	.010	--
292943098354404	AY-68-36-132	TANC MW storm 5	2009/11/18	1.79	<.02	<.002	.008	--
293252098380801	AY-68-27-610	TANC MW pre- storm	2009/04/14	2.74	E.01	<.002	.012	--
293252098380801	AY-68-27-610	TANC MW storm 1	2009/10/04	2.94	<.02	<.002	.010	--
293252098380801	AY-68-27-610	TANC MW storm 2	2009/10/07	2.62	<.02	E.002	.084	--
293252098380801	AY-68-27-610	TANC MW storm 3	2009/10/17	2.85	<.02	<.002	.010	--
293252098380801	AY-68-27-610	TANC MW storm 4	2009/10/28	2.67	<.02	<.002	.011	--
293252098380801	AY-68-27-610	TANC MW storm 5	2009/11/17	2.99	<.02	<.002	.010	--

* Sample collected from top of open interval during depth-dependent sampling.

Table 12. Isotope and radionuclide measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; PSW, public-supply well; pCi/L, picoCuries per liter; $\delta^{18}\text{O}$, delta oxygen-18; δD , delta deuterium; $^{87}\text{Sr}/^{86}\text{Sr}$, strontium-87/strontium-86; $\delta^{15}\text{N}$, delta nitrogen-15; NO_3^- , nitrate; --, not measured; R, result less than sample-specific critical level]

USGS identifica- tion number	State well number	Group and description	Sample date	$\delta^{18}\text{O}$ (per mil)	δD (per mil)	$^{87}\text{Sr}/^{86}\text{Sr}$	Radon- 222 (pCi/L)	Ra- dium- 226 (pCi/L)	Ra- dium- 228 (pCi/L)	$\delta^{15}\text{N}$ in NO_3^- (per mil)	$\delta^{18}\text{O}$ in NO_3^- (per mil)
Monitoring wells											
292943098354401	AY-68-36-135	TANC MW OVB	2007/10/30	-3.79	-22.7	.70785	173	0.46	R.18	6.84	5.32
292943098354402	AY-68-36-133	TANC MW	2007/10/31	-4.11	-23.5	.70773	95	.24	.36	8.15	5.12
292943098354403	AY-68-36-134	TANC MW	2007/10/30	-4.10	-21.8	.70772	73	.11	R-.14	8.06	5.27
292943098354404	AY-68-36-132	TANC MW	2007/11/01	-4.10	-22.4	.70767	40	.19	R-.27	7.97	5.17
292851098374401	AY-68-35-315	TANC MW	2008/05/06	-4.11	-23.4	.70784	18	.18	R-.01	7.69	4.59
292851098374402	AY-68-35-314	TANC MW	2008/05/07	-4.03	-23.6	.70781	53	.17	R.02	8.04	4.82
Well-field depth-dependent sampling (W4)											
292916098360701	AY-68-36-105	W4-shallow (380), normal	2007/11/16	-4.01	-22.7	.70769	35	.16	R.02	7.67	4.98
292916098360701	AY-68-36-105	W4-intermediate (430), normal	2007/11/15	-4.07	-22.8	.70789	30	.23	R.03	7.69	5.08
292916098360701	AY-68-36-105	W4-deep (540), normal	2007/11/15	-4.01	-22.7	.70769	46	.20	R.12	7.80	5.35
292916098360701	AY-68-36-105	W4-shallow (380), ambient	2007/11/16	-3.98	-22.0	.70782	34	.15	R.05	7.88	5.16
292916098360701	AY-68-36-105	W4-intermediate (430), ambient	2007/11/14	-3.99	-22.5	.70775	35	.16	R-.15	7.95	5.24
292916098360701	AY-68-36-105	W4-deep (540), ambient	2007/11/15	-3.98	-23.2	.70771	40	.18	R-.12	7.85	5.28
292916098360701	AY-68-36-105	W4-shallow (380), moderate	2007/11/16	-4.08	-22.6	.70774	67	.16	R.01	7.87	5.30
292916098360701	AY-68-36-105	W4-intermediate (430), moderate	2007/11/14	-3.96	-23.5	.70771	36	.18	R-.19	7.84	5.51
292916098360701	AY-68-36-105	W4-deep (540), moderate	2007/11/15	-3.99	-23.2	.70772	40	.07	R.03	7.82	5.22
Well-field wells (wellhead sampling)											
292919098360501	AY-68-36-103	wellhead	2007/11/29	-4.06	-22.9	.70772	--	.20	R.08	8.16	5.45
292925098360201	AY-68-36-104	wellhead	2007/11/28	-4.01	-22.4	.70768	--	.16	R.02	7.93	5.15
292916098360701	AY-68-36-105	*wellhead	2007/11/16	-4.01	-21.6	.70768	57	.12	R-.04	7.94	5.06
292920098360601	AY-68-36-130	wellhead	2007/11/29	-4.11	-22.5	.70772	--	.20	R.05	8.09	5.35
292923098360301	AY-68-36-131	wellhead	2007/11/28	-4.02	-23.4	.70771	--	.18	R.20	8.12	5.52
Temporal sampling											
292923098360301	AY-68-36-131	TANC PSW pre- storm	2009/04/09	-3.96	-22.7	--	--	--	--	--	--
292923098360301	AY-68-36-131	TANC PSW storm 1	2009/10/06	-4.00	-21.8	.70781	--	--	--	--	--
292923098360301	AY-68-36-131	TANC PSW storm 2	2009/10/09	-4.04	-21.3	.70779	--	--	--	--	--
292923098360301	AY-68-36-131	TANC PSW storm 3	2009/10/19	-4.08	-21.8	.70779	--	--	--	--	--
292923098360301	AY-68-36-131	TANC PSW storm 4	2009/10/30	-4.11	-22.8	.70778	--	--	--	--	--
292923098360301	AY-68-36-131	TANC PSW storm 5	2009/11/19	-4.03	-21.7	.70780	--	--	--	--	--

Table 12. Isotope and radionuclide measurements for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; PSW, public-supply well; pCi/L, picoCuries per liter; $\delta^{18}\text{O}$, delta oxygen-18; δD , delta deuterium; $^{87}\text{Sr}/^{86}\text{Sr}$, strontium-87/strontium-86; $\delta^{15}\text{N}$, delta nitrogen-15; NO_3^- , nitrate; --, not measured; R, result less than sample-specific critical level]

USGS identifica- tion number	State well number	Group and description	Sample date	$\delta^{18}\text{O}$ (per mil)	δD (per mil)	$^{87}\text{Sr}/^{86}\text{Sr}$	Radon- 222 (pCi/L)	Ra- dium- 226 (pCi/L)	Ra- dium- 228 (pCi/L)	$\delta^{15}\text{N}$ in NO_3^- (per mil)	$\delta^{18}\text{O}$ in NO_3^- (per mil)
292943098354402	AY-68-36-133	TANC MW pre- storm	2009/04/08	-4.10	-22.1	--	--	--	--	--	--
292943098354402	AY-68-36-133	TANC MW storm 1	2009/10/05	-4.09	-22.9	0.70779	--	--	--	--	--
292943098354402	AY-68-36-133	TANC MW storm 2	2009/10/07	-4.11	-21.4	.70780	--	--	--	--	--
292943098354402	AY-68-36-133	TANC MW storm 3	2009/10/17	-4.19	-22.5	.70780	--	--	--	--	--
292943098354402	AY-68-36-133	TANC MW storm 4	2009/10/28	-4.14	-21.8	.70779	--	--	--	--	--
292943098354402	AY-68-36-133	TANC MW storm 5	2009/11/18	-4.15	-22.7	.70780	--	--	--	--	--
292943098354403	AY-68-36-134	TANC MW pre- storm	2009/04/09	-4.18	-21.8	--	--	--	--	--	--
292943098354403	AY-68-36-134	TANC MW storm 1	2009/10/05	-4.11	-23.2	.70778	--	--	--	--	--
292943098354403	AY-68-36-134	TANC MW storm 2	2009/10/08	-4.05	-23.6	.70782	--	--	--	--	--
292943098354403	AY-68-36-134	TANC MW storm 3	2009/10/18	-4.12	-23.8	.70779	--	--	--	--	--
292943098354403	AY-68-36-134	TANC MW storm 4	2009/10/29	-4.09	-23.8	.70776	--	--	--	--	--
292943098354403	AY-68-36-134	TANC MW storm 5	2009/11/18	-4.14	-22.6	.70776	--	--	--	--	--
292943098354404	AY-68-36-132	TANC MW pre- storm	2009/04/09	-4.08	-22.8	--	--	--	--	--	--
292943098354404	AY-68-36-132	TANC MW storm 1	2009/10/05	-4.10	-23.8	.70779	--	--	--	--	--
292943098354404	AY-68-36-132	TANC MW storm 2	2009/10/08	-4.10	-24.2	.70778	--	--	--	--	--
292943098354404	AY-68-36-132	TANC MW storm 3	2009/10/18	-4.13	-21.7	.70778	--	--	--	--	--
292943098354404	AY-68-36-132	TANC MW storm 4	2009/10/29	-4.14	-23.7	.70780	--	--	--	--	--
292943098354404	AY-68-36-132	TANC MW storm 5	2009/11/18	-4.15	-23.5	.70781	--	--	--	--	--
293252098380801	AY-68-27-610	TANC MW pre- storm	2009/04/14	-4.35	-23.8	--	--	--	--	--	--
293252098380801	AY-68-27-610	TANC MW storm 1	2009/10/04	-4.18	-23.3	.70807	--	--	--	--	--
293252098380801	AY-68-27-610	TANC MW storm 2	2009/10/07	-4.32	-23.7	.70809	--	--	--	--	--
293252098380801	AY-68-27-610	TANC MW storm 3	2009/10/17	-4.19	-24.0	.70809	--	--	--	--	--
293252098380801	AY-68-27-610	TANC MW storm 4	2009/10/28	-4.24	-23.1	.70809	--	--	--	--	--
293252098380801	AY-68-27-610	TANC MW storm 5	2009/11/17	-4.21	-23.6	.70811	--	--	--	--	--

* Sample collected from top of open interval during depth-dependent sampling.

Table 13. Wastewater indicator compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; <, nondetection less than laboratory method reporting level; --, not measured; V, contaminated]

USGS identification number	State well number	Group and description	Sample date	1,4-Di-chloro-benzene (µg/L)	1-Methyl-naphth-alene (µg/L)	2,6-Di-methyl-naphth-alene (µg/L)	2-Methyl-naphth-alene (µg/L)	3-beta-Copro-standol (µg/L)	3-Methyl-1H-indole (µg/L)	3-tert-Butyl-4-hydroxy-anisole (µg/L)	4-Cumyl-phenol (µg/L)	4-n-Octyl-phenol (µg/L)	4-Nonyl-phenol (sum of isomers) (µg/L)
Regional aquifer public-supply wells													
293359098290301	AY-68-29-414	1st (highest) quartile	2004/12/06	<0.5	<0.5	<0.5	<0.5	<2.0	<1.0	<5.0	<1.0	<1.0	V1.10
293359098290301	AY-68-29-414	1st quartile	2005/09/01	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
293358098231101	AY-68-29-610	1st quartile	2004/12/15	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	V.470
293358098231101	AY-68-29-610	1st quartile	2005/08/29	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	V1.10
293120098285801	AY-68-29-714	1st quartile	2004/12/14	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
293120098285801	AY-68-29-714	1st quartile	2005/08/30	--	--	--	--	--	--	--	--	--	--
293145098224201	AY-68-29-929	1st quartile	2004/12/15	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	V.510
293119098211201	AY-68-30-718	1st quartile	2004/12/16	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
292843098425101	AY-68-35-106	1st quartile	2004/12/13	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
292459098382101	AY-68-35-913	1st quartile	2004/12/13	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
292925098360201	AY-68-36-104	1st quartile	2004/12/13	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
292925098360201	AY-68-36-104	1st quartile	2005/08/31	--	--	--	--	--	--	--	--	--	--
292822098325401	AY-68-36-206	1st quartile	2004/12/13	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
292822098325401	AY-68-36-206	1st quartile	2005/08/31	--	--	--	--	--	--	--	--	--	--
292944098292301	AY-68-37-105	1st quartile	2004/12/14	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	V.470
292944098292301	AY-68-37-105	1st quartile	2005/08/31	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	V2.00
292522098291901	AY-68-37-426	1st quartile	2004/12/14	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
292643098241801	AY-68-37-601	1st quartile	2004/12/15	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	V.670
292643098241801	AY-68-37-601	1st quartile	2005/08/30	--	--	--	--	--	--	--	--	--	--
292328098294601	AY-68-37-705	1st quartile	2004/12/14	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
294225098080301	DX-68-23-601	1st quartile	2004/12/08	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
294225098080301	DX-68-23-601	1st quartile	2005/09/01	--	--	--	--	--	--	--	--	--	--
291210099475601	YP-69-50-506	1st quartile	2004/11/30	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0

Table 13. Wastewater indicator compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; <, nondetection less than laboratory method reporting level; --, not measured; V, contaminated]

USGS identification number	State well number	Group and description	Sample date	1,4-Di- chloro- benzene (µg/L)	1-Methyl- naphth- alene (µg/L)	2,6-Di- methyl- naphth- alene (µg/L)	2-Methyl- naphth- alene (µg/L)	3-beta- Copro- stanol (µg/L)	3-Methyl- 1H-indole (µg/L)	3-tert- Butyl- 4-hydroxy- anisole (µg/L)	4-Cumyl- phenol (µg/L)	4-n- Octyl- phenol (µg/L)	4-Nonyl- phenol (sum of isomers) (µg/L)
293111098340901	AY-68-28-807	2d quartile	2005/05/26	--	--	--	--	--	--	--	--	--	--
293512098291701	AY-68-29-109	2d quartile	2005/04/20	<0.5	<0.5	<0.5	<0.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
292424098421501	AY-68-35-810	2d quartile	2005/04/21	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	V1.10
292053098365501	AY-68-44-110	2d quartile	2005/05/26	--	--	--	--	--	--	--	--	--	--
293807098155301	DX-68-22-901	2d quartile	2005/05/18	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
292604098563201	TD-68-33-501	2d quartile	2005/05/16	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
292116099095501	TD-69-47-305	2d quartile	2005/04/27	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	V.520
291232099470301	YP-69-50-339	2d quartile	2005/05/11	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
293128098473101	AY-68-26-814	3d quartile	2005/04/18	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
293451098313201	AY-68-28-601	3d quartile	2005/05/17	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	V.980
294604098060801	DX-68-16-708	3d quartile	2005/04/25	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	V3.20
294019098114701	DX-68-23-504	3d quartile	2005/05/18	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
292215098580201	TD-68-41-103	3d quartile	2005/05/09	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
292119098524901	TD-68-41-308	3d quartile	2005/04/27	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	V.760
291943099163301	TD-69-46-601	3d quartile	2005/05/09	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	V1.20
291840099382601	YP-69-43-606	3d quartile	2005/05/11	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
293518098332601	AY-68-28-203	4th quartile	2005/04/19	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
293023098355401	AY-68-28-702	4th quartile	2005/04/26	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	V.990
293042098305201	AY-68-28-913	4th quartile	2005/06/30	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
292442098474501	AY-68-34-803	4th quartile	2005/04/20	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
292405098371201	AY-68-36-704	4th quartile	2005/05/23	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
292931098274601	AY-68-37-124	4th quartile	2005/05/25	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
293729098173101	DX-68-30-215	4th quartile	2005/04/25	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	<5.0
291219099095601	TD-69-55-604	4th quartile	2005/05/16	<.5	<.5	<.5	<.5	<2.0	<1.0	<5.0	<1.0	<1.0	V.830

Table 13. Wastewater indicator compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; <, nondetection less than laboratory method reporting level; --, not measured; V, contaminated]

USGS identification number	4-Nonyl- phenol dieth- oxylate (sum of isomers) (µg/L)	4-tert- Octyl- phenol dieth- oxylate (µg/L)	4-tert- Octyl- phenol mono- ethoxylate (µg/L)	4-tert- Octyl- phenol (µg/L)	5-Methyl- 1H- benzo- triazole (µg/L)	9,10- Anthra- quinone (µg/L)	Aceto- phe- none (µg/L)	Acetyl hexa- methyl tetrahydro naph- thalene (AHTN) (µg/L)	Anth- racene (µg/L)	Benzo- [a]- pyrene (µg/L)	Benzo- phe- none (µg/L)	beta- Sito- sterol (µg/L)	beta- Stigma- stanol (µg/L)	Brom- acil (µg/L)	Cam- phor (µg/L)	Caffeine (µg/L)	Carba- zole (µg/L)
Regional aquifer public-supply wells																	
293359098290301	<5.0	<1.0	<1.0	<1.0	<2.0	<0.5	<0.5	<0.5	<0.5	<0.5	V0.068	<2.0	<2.0	<0.018	<0.5	<0.018	<0.5
293359098290301	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.018	<.5	<.018	<.5
293358098231101	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.018	<.5	<.018	<.5
293358098231101	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.018	<.5	<.018	<.5
293120098285801	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.018	<.5	V.007	<.5
293120098285801	--	--	--	--	--	--	--	--	--	--	--	--	--	<.018	--	<.018	--
293145098224201	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.018	<.5	<.018	<.5
293119098211201	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.018	<.5	<.018	<.5
292843098425101	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.018	<.5	<.018	<.5
292459098382101	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.018	<.5	<.018	<.5
292925098360201	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.018	<.5	V.009	<.5
292925098360201	--	--	--	--	--	--	--	--	--	--	--	--	--	<.018	--	<.018	--
292822098325401	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.018	<.5	<.018	<.5
292822098325401	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292944098292301	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.018	<.5	<.018	<.5
292944098292301	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.5	V.004	<.5	<.5
292522098291901	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.018	<.5	<.018	<.5
292643098241801	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.018	<.5	V.005	<.5
292643098241801	--	--	--	--	--	--	--	--	--	--	--	--	--	<.018	--	<.018	--
292328098294601	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.018	<.5	<.018	<.5
294225098080301	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.018	<.5	V.009	<.5
294225098080301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291210099475601	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.018	<.5	<.018	<.5

Table 13. Wastewater indicator compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; <, nondetection less than laboratory method reporting level; --, not measured; V, contaminated]

USGS identification number	4-Nonyl- phenol dieth- oxylate (sum of isomers) (µg/L)	4-tert- Octyl- phenol dieth- oxylate (µg/L)	4-tert- Octyl- phenol mono- ethoxylate (µg/L)	4-tert- Octyl- phenol (µg/L)	5-Methyl- 1H- benzo- triazole (µg/L)	9,10- Anthra- quinone (µg/L)	Aceto- phe- none (µg/L)	Acetyl hexa- methyl tetrahydro naph- thalene (AHTN) (µg/L)	Anth- racene (µg/L)	Benzo- [a]- pyrene (µg/L)	Benzo- phe- none (µg/L)	beta- Sito- sterol (µg/L)	beta- Stigma- stanol (µg/L)	Brom- acil (µg/L)	Cam- phor (µg/L)	Caffeine (µg/L)	Carba- zole (µg/L)
293111098340901	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293512098291701	<5.0	<1.0	<1.0	<1.0	<2.0	<0.5	<0.5	<0.5	<0.5	<0.5	<0.5	<2.0	<2.0	<0.5	<0.5	<0.5	<0.5
292424098421501	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	V.048	<2.0	<2.0	<.5	<.5	<.5	<.5
292053098365501	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293807098155301	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	V.066	<2.0	<2.0	<.5	<.5	<.5	<.5
292604098563201	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.5	<.5	<.5	<.5
292116099095501	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.5	<.5	<.5	<.5
291232099470301	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	V.030	<2.0	<2.0	<.5	<.5	<.5	<.5
293128098473101	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.5	<.5	<.5	<.5
293451098313201	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	V.067	<2.0	<2.0	<.5	<.5	<.5	<.5
294604098060801	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	V.220	<2.0	<2.0	<.5	<.5	<.5	<.5
294019098114701	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.5	<.5	<.5	<.5
292215098580201	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	V.027	<2.0	<2.0	<.5	<.5	<.5	<.5
292119098524901	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.5	<.5	<.5	<.5
291943099163301	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	V.013	<2.0	<2.0	<.5	<.5	<.5	<.5
291840099382601	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	V.016	<2.0	<2.0	<.5	<.5	<.5	<.5
293518098332601	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.5	<.5	<.5	<.5
293023098355401	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	V.027	<2.0	<2.0	<.5	<.5	<.5	<.5
293042098305201	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	V.023	<2.0	<2.0	<.5	V.006	<.5	<.5
292442098474501	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.5	<.5	<.5	<.5
292405098371201	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.5	<.5	<.5	<.5
292931098274601	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293729098173101	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	<.5	<2.0	<2.0	<.5	<.5	<.5	<.5
291219099095601	<5.0	<1.0	<1.0	<1.0	<2.0	<.5	<.5	<.5	<.5	<.5	V.017	<2.0	<2.0	<.5	<.5	<.5	<.5

Table 13. Wastewater indicator compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; <, nondetection less than laboratory method reporting level; --, not measured; V, contaminated]

[illegible]

Table 13. Wastewater indicator compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; <, nondetection less than laboratory method reporting level; --, not measured; V, contaminated]

USGS identification number	Chol- esterol (µg/L)	Cotinine (µg/L)	N,N- diethyl- meta- tolua- mide (µg/L)	D-Limo- nene (µg/L)	Fluor- anthene (µg/L)	Hexahydro- hexamethyl cyclopenta- benzopyran (HHCb) (µg/L)	Indole (µg/L)	Iso- borneol (µg/L)	Iso- phorone (µg/L)	Iso- propyl- benzene (µg/L)	Iso- quino- line (µg/L)	Men- thol (µg/L)	Meta- laxyl (µg/L)	Methyl sali- cylate (µg/L)
293111098340901	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293512098291701	<2.0	<1.0	<0.5	<0.5	<0.5	<0.5	<0.5	<0.5	<0.5	<0.5	<0.5	<0.5	<0.5	<0.5
292424098421501	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
292053098365501	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293807098155301	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
292604098563201	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
292116099095501	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
291232099470301	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	V.023	<.5	<.5	<.5	<.5
293128098473101	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
293451098313201	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
294604098060801	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
294019098114701	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
292215098580201	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
292119098524901	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
291943099163301	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
291840099382601	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
293518098332601	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
293023098355401	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
293042098305201	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
292442098474501	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
292405098371201	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
292931098274601	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293729098173101	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5
291219099095601	<2.0	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5	<.5

Table 13. Wastewater indicator compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; <, nondetection less than laboratory method reporting level; --, not measured; V, contaminated]

USGS identification number	Naphth- alene (µg/L)	p-Cresol (µg/L)	Phen- anthrene (µg/L)	Phenol (µg/L)	Pyrene (µg/L)	Tetra- chloro- ethene (PCE) (µg/L)	Tribromo- methane (bromo- form) (µg/L)	Tributyl phos- phate (µg/L)	Tri- closan (µg/L)	Tri- ethyl citrate (µg/L)	Tri- phenyl phos- phate (µg/L)	Tris (2-but- oxyethyl) phos- phate (µg/L)	Tris (2-chloro- ethyl) phosphate (FYROL CEF) (µg/L)	Tris (dichloro- isopropyl) phosphate (FYROL FR 2) (µg/L)
Regional aquifer public-supply wells														
293359098290301	<0.5	<1.0	<0.5	V0.820	<.5	V0.280	<0.5	<0.5	<1.0	<0.5	<0.5	<0.5	<0.5	<0.5
293359098290301	<.5	<1.0	<.5	<.5	<.5	V.250	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
293358098231101	<.5	<1.0	<.5	V.320	<.5	V.130	V.023	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
293358098231101	<.5	<1.0	<.5	<.5	<.5	V.093	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
293120098285801	<.5	<1.0	<.5	V.890	<.5	V.086	V1.90	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
293120098285801	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293145098224201	<.5	<1.0	<.5	V1.50	<.5	V.037	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
293119098211201	<.5	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292843098425101	<.5	<1.0	<.5	V.210	<.5	<.5	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292459098382101	<.5	<1.0	<.5	V.480	<.5	<.5	V.054	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292925098360201	<.5	<1.0	<.5	<.5	<.5	V.055	V1.10	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292925098360201	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292822098325401	<.5	<1.0	<.5	V.210	<.5	V.022	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292822098325401	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292944098292301	<.5	<1.0	<.5	V1.10	<.5	V.032	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292944098292301	<.5	<1.0	<.5	<.5	<.5	V.015	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292522098291901	<.5	<1.0	<.5	V.460	<.5	<.5	V.024	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292643098241801	<.5	<1.0	<.5	V.690	<.5	<.5	V.660	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292643098241801	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292328098294601	<.5	<1.0	<.5	V.140	<.5	V.020	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
294225098080301	<.5	<1.0	<.5	V.400	<.5	V.072	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
294225098080301	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291210099475601	<.5	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5

Table 13. Wastewater indicator compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; <, nondetection less than laboratory method reporting level; --, not measured; V, contaminated]

USGS identification number	Naphth- alene (µg/L)	p-Cresol (µg/L)	Phen- anthrene (µg/L)	Phenol (µg/L)	Pyrene (µg/L)	Tetra- chloro- ethene (PCE) (µg/L)	Tribromo- methane (bromo- form) (µg/L)	Tributyl phos- phate (µg/L)	Tri- closan (µg/L)	Tri- ethyl citrate (µg/L)	Tri- phenyl phos- phate (µg/L)	Tris (2-but- oxyethyl) phos- phate (µg/L)	Tris (2-chloro- ethyl) phosphate (FYROL CEF) (µg/L)	Tris (dichloro- isopropyl) phosphate (FYROL FR 2) (µg/L)
293111098340901	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293512098291701	<0.5	<1.0	<0.5	V1.90	<0.5	<0.5	<0.5	<0.5	<1.0	<0.5	<0.5	<0.5	<0.5	<0.5
292424098421501	<.5	<1.0	<.5	V1.70	<.5	<.5	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292053098365501	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293807098155301	<.5	<1.0	<.5	V1.50	<.5	<.5	V.500	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292604098563201	<.5	<1.0	<.5	V2.20	<.5	<.5	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292116099095501	<.5	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
291232099470301	<.5	<1.0	<.5	V.500	<.5	V.040	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
293128098473101	<.5	<1.0	<.5	V1.30	<.5	<.5	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
293451098313201	<.5	<1.0	<.5	V1.00	<.5	<.5	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
294604098060801	<.5	<1.0	<.5	V.850	<.5	<.5	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
294019098114701	<.5	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292215098580201	<.5	V.024	<.5	V4.0	<.5	<.5	V.078	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292119098524901	<.5	<1.0	<.5	<.5	<.5	<.5	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
291943099163301	<.5	<1.0	<.5	V2.40	<.5	<.5	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
291840099382601	<.5	<1.0	<.5	V.280	<.5	<.5	V.026	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
293518098332601	<.5	<1.0	<.5	V1.40	<.5	V.042	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
293023098355401	<.5	<1.0	<.5	V1.40	<.5	V.140	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
293042098305201	<.5	<1.0	<.5	V.180	<.5	V.130	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292442098474501	<.5	<1.0	<.5	<.5	<.5	<.5	V1.80	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292405098371201	<.5	<1.0	<.5	V.110	<.5	<.5	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
292931098274601	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293729098173101	<.5	<1.0	<.5	V.850	<.5	<.5	<.5	<.5	<1.0	<.5	<.5	<.5	<.5	<.5
291219099095601	<.5	<1.0	<.5	V1.90	<.5	<.5	<.5	<.5	<1.0	<.5	<.5	V.098	<.5	<.5

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Group and description	Sample date	1,1,1,2- Tetra- chloro- ethane (µg/L)	1,1,1-Tri- chloro- ethane (µg/L)	1,1,2,2- Tetra- chloro- ethane (µg/L)	1,1,2-Tri- chloro- 1,2,2-tri- fluoro- ethane (CFC-113) (µg/L)	1,1,2-Tri- chloro- ethane (µg/L)	1,1-Di- chloro- ethane (µg/L)	1,1-Di- chloro- ethene (µg/L)	1,1-Di- chloro- propene (µg/L)	1,2,3,4- Tetra- methyl- benzene (prehnite- tene) (µg/L)	1,2,3,5- Tetra- methyl- benzene (iso- durene) (µg/L)
Regional aquifer public-supply wells													
293359098290301	AY-68-29-414	1st (highest) quartile	2004/12/06	<0.03	<0.032	<0.08	E0.038	<0.04	<0.035	<0.024	<0.026	<0.14	<0.14
293359098290301	AY-68-29-414	1st quartile	2005/09/01	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
293358098231101	AY-68-29-610	1st quartile	2004/12/15	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
293358098231101	AY-68-29-610	1st quartile	2005/08/29	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
293120098285801	AY-68-29-714	1st quartile	2004/12/14	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
293120098285801	AY-68-29-714	1st quartile	2005/08/30	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
293145098224201	AY-68-29-929	1st quartile	2004/12/15	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
293119098211201	AY-68-30-718	1st quartile	2004/12/16	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292843098425101	AY-68-35-106	1st quartile	2004/12/13	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292459098382101	AY-68-35-913	1st quartile	2004/12/13	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292925098360201	AY-68-36-104	1st quartile	2004/12/13	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292925098360201	AY-68-36-104	1st quartile	2005/08/31	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292822098325401	AY-68-36-206	1st quartile	2004/12/13	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292822098325401	AY-68-36-206	1st quartile	2005/08/31	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292944098292301	AY-68-37-105	1st quartile	2004/12/14	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292944098292301	AY-68-37-105	1st quartile	2005/08/31	--	--	--	--	--	--	--	--	--	--
292522098291901	AY-68-37-426	1st quartile	2004/12/14	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292643098241801	AY-68-37-601	1st quartile	2004/12/15	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292643098241801	AY-68-37-601	1st quartile	2005/08/30	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292328098294601	AY-68-37-705	1st quartile	2004/12/14	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
294225098080301	DX-68-23-601	1st quartile	2004/12/08	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
294225098080301	DX-68-23-601	1st quartile	2005/09/01	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
291210099475601	YP-69-50-506	1st quartile	2004/11/30	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
293111098340901	AY-68-28-807	2d quartile	2005/05/26	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
293512098291701	AY-68-29-109	2d quartile	2005/04/20	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292424098421501	AY-68-35-810	2d quartile	2005/04/21	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Group and description	Sample date	1,1,1,2- Tetra- chloro- ethane (µg/L)	1,1,1-Tr- chloro- ethane (µg/L)	1,1,2,2- Tetra- chloro- ethane (µg/L)	1,1,2-Tri- chloro- 1,2,2-tri- fluoro- ethane (CFC-113) (µg/L)	1,1,2-Tri- chloro- ethane (µg/L)	1,1-Di- chloro- ethane (µg/L)	1,1-Di- chloro- ethene (µg/L)	1,1-Di- chloro- propene (µg/L)	1,2,3,4- Tetra- methyl- benzene (prehnite- tene) (µg/L)	1,2,3,5- Tetra- methyl- benzene (iso- durene) (µg/L)
292053098365501	AY-68-44-110	2d quartile	2005/05/26	<0.03	<0.032	<0.08	<0.038	<0.04	<0.035	<0.024	<0.026	<0.14	<0.14
293807098155301	DX-68-22-901	2d quartile	2005/05/18	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292604098563201	TD-68-33-501	2d quartile	2005/05/16	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292116099095501	TD-69-47-305	2d quartile	2005/04/27	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
291232099470301	YP-69-50-339	2d quartile	2005/05/11	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
293128098473101	AY-68-26-814	3d quartile	2005/04/18	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
293451098313201	AY-68-28-601	3d quartile	2005/05/17	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
294604098060801	DX-68-16-708	3d quartile	2005/04/25	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
294019098114701	DX-68-23-504	3d quartile	2005/05/18	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292215098580201	TD-68-41-103	3d quartile	2005/05/09	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292119098524901	TD-68-41-308	3d quartile	2005/04/27	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
291943099163301	TD-69-46-601	3d quartile	2005/05/09	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
291840099382601	YP-69-43-606	3d quartile	2005/05/11	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
293518098332601	AY-68-28-203	4th quartile	2005/04/19	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
293023098355401	AY-68-28-702	4th quartile	2005/04/26	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
293042098305201	AY-68-28-913	4th quartile	2005/06/30	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292442098474501	AY-68-34-803	4th quartile	2005/04/20	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292405098371201	AY-68-36-704	4th quartile	2005/05/23	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
292931098274601	AY-68-37-124	4th quartile	2005/05/25	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
293729098173101	DX-68-30-215	4th quartile	2005/04/25	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
291219099095601	TD-69-55-604	4th quartile	2005/05/16	<.03	<.032	<.08	<.038	<.04	<.035	<.024	<.026	<.14	<.14
Monitoring wells													
292943098354401	AY-68-36-135	TANC MW OVB	2007/10/30	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
292943098354402	AY-68-36-133	TANC MW	2007/10/31	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
292943098354403	AY-68-36-134	TANC MW	2007/10/30	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
292943098354404	AY-68-36-132	TANC MW	2007/11/01	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Group and description	Sample date	1,1,1,2- Tetra- chloro- ethane (µg/L)	1,1,1-Trichloro- ethane (µg/L)	1,1,2,2- Tetra- chloro- ethane (µg/L)	1,1,2-Tri- chloro- 1,2,2-tri- fluoro- ethane (CFC-113) (µg/L)	1,1,2-Trichloro- ethane (µg/L)	1,1-Di- chloro- ethane (µg/L)	1,1-Di- chloro- ethene (µg/L)	1,1-Di- chloro- propene (µg/L)	1,2,3,4- Tetra- methyl- benzene (prehnite- tene) (µg/L)	1,2,3,5- Tetra- methyl- benzene (isodurene) (µg/L)
292851098374401	AY-68-35-315	TANC MW	2008/05/06	<0.04	<0.02	<0.1	<0.04	<0.06	<0.04	<0.02	<0.04	<0.14	<0.12
292851098374402	AY-68-35-314	TANC MW	2008/05/07	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
Well-field depth-dependent sampling (W4)													
292916098360701	AY-68-36-105	W4-shallow (380), normal	2007/11/16	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
292916098360701	AY-68-36-105	W4-intermediate (430), normal	2007/11/15	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
292916098360701	AY-68-36-105	W4-deep (540), normal	2007/11/15	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
292916098360701	AY-68-36-105	W4-shallow (380), ambient	2007/11/16	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
292916098360701	AY-68-36-105	W4-intermediate (430), ambient	2007/11/14	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
292916098360701	AY-68-36-105	W4-deep (540), ambient	2007/11/15	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
292916098360701	AY-68-36-105	W4-shallow (380), moderate	2007/11/16	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
292916098360701	AY-68-36-105	W4-intermediate (430), moderate	2007/11/14	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
292916098360701	AY-68-36-105	W4-deep (540), moderate	2007/11/15	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
Well-field wells (wellhead sampling)													
292919098360501	AY-68-36-103	wellhead	2007/11/29	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
292925098360201	AY-68-36-104	wellhead	2007/11/28	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
292916098360701	AY-68-36-105	*wellhead	2007/11/16	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
292920098360601	AY-68-36-130	wellhead	2007/11/29	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
292923098360301	AY-68-36-131	wellhead	2007/11/28	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.14	<.12
Temporal sampling													
292923098360301	AY-68-36-131	TANC PSW pre- storm	2009/04/09	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.08	<.08
292923098360301	AY-68-36-131	TANC PSW storm 1	2009/10/06	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Group and description	Sample date	1,1,1,2- Tetra- chloro- ethane (µg/L)	1,1,1-Tr- chloro- ethane (µg/L)	1,1,2,2- Tetra- chloro- ethane (µg/L)	1,1,2-Tri- chloro- 1,2,2-tri- fluoro- ethane (CFC-113) (µg/L)	1,1,2-Tri- chloro- ethane (µg/L)	1,1-Di- chloro- ethane (µg/L)	1,1-Di- chloro- ethene (µg/L)	1,1-Di- chloro- propene (µg/L)	1,2,3,4- Tetra- methyl- benzene (prehnite- tene) (µg/L)	1,2,3,5- Tetra- methyl- benzene (iso- durene) (µg/L)
292923098360301	AY-68-36-131	TANC PSW storm 2	2009/10/09	<0.04	<0.03	<0.14	<0.034	<0.046	<0.044	<0.022	<0.03	<0.08	<0.08
292923098360301	AY-68-36-131	TANC PSW storm 3	2009/10/19	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292923098360301	AY-68-36-131	TANC PSW storm 4	2009/10/30	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292923098360301	AY-68-36-131	TANC PSW storm 5	2009/11/19	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292943098354402	AY-68-36-133	TANC MW pre- storm	2009/04/08	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.08	<.08
292943098354402	AY-68-36-133	TANC MW storm 1	2009/10/05	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292943098354402	AY-68-36-133	TANC MW storm 2	2009/10/07	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292943098354402	AY-68-36-133	TANC MW storm 3	2009/10/17	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292943098354402	AY-68-36-133	TANC MW storm 4	2009/10/28	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292943098354402	AY-68-36-133	TANC MW storm 5	2009/11/18	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292943098354403	AY-68-36-134	TANC MW pre- storm	2009/04/09	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.08	<.08
292943098354403	AY-68-36-134	TANC MW storm 1	2009/10/05	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292943098354403	AY-68-36-134	TANC MW storm 2	2009/10/08	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292943098354403	AY-68-36-134	TANC MW storm 3	2009/10/18	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292943098354403	AY-68-36-134	TANC MW storm 4	2009/10/29	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292943098354403	AY-68-36-134	TANC MW storm 5	2009/11/18	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292943098354404	AY-68-36-132	TANC MW pre- storm	2009/04/09	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.08	<.08
292943098354404	AY-68-36-132	TANC MW storm 1	2009/10/05	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292943098354404	AY-68-36-132	TANC MW storm 2	2009/10/08	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292943098354404	AY-68-36-132	TANC MW storm 3	2009/10/18	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292943098354404	AY-68-36-132	TANC MW storm 4	2009/10/29	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
292943098354404	AY-68-36-132	TANC MW storm 5	2009/11/18	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
293252098380801	AY-68-27-610	TANC MW pre- storm	2009/04/14	<.04	<.02	<.1	<.04	<.06	<.04	<.02	<.04	<.08	<.08
293252098380801	AY-68-27-610	TANC MW storm 1	2009/10/04	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
293252098380801	AY-68-27-610	TANC MW storm 2	2009/10/07	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
293252098380801	AY-68-27-610	TANC MW storm 3	2009/10/17	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
293252098380801	AY-68-27-610	TANC MW storm 4	2009/10/28	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08
293252098380801	AY-68-27-610	TANC MW storm 5	2009/11/17	<.04	<.03	<.14	<.034	<.046	<.044	<.022	<.03	<.08	<.08

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	1,2,3-Tri- chloro- benzene (µg/L)	1,2,3-Tri- chloro- propane (µg/L)	1,2,3-Tri- methyl- benzene (µg/L)	1,2,4-Tri- chloro- benzene (µg/L)	1,2,4-Tri- methyl- benzene (µg/L)	1,2-Di- bromo- 3-chloro- propane (µg/L)	1,2-Di- bromo- ethane (µg/L)	1,2-Di- chloro- benzene (µg/L)	1,2-Di- chloro- ethane (µg/L)	1,2-Di- chloro- propane (µg/L)	1,3,5-Tri- methyl- benzene (µg/L)	1,3-Di- chloro- benzene (µg/L)	1,3-Di- chloro- propane (µg/L)	1,4-Di- chloro- benzene (µg/L)
Regional aquifer public-supply wells														
293359098290301	<0.18	<0.18	<0.06	<0.12	<0.056	<0.51	<0.036	<0.048	<0.13	<0.029	<0.044	<0.03	<0.06	<0.034
293359098290301	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
293358098231101	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
293358098231101	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
293120098285801	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
293120098285801	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
293145098224201	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
293119098211201	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292843098425101	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292459098382101	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292925098360201	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292925098360201	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292822098325401	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292822098325401	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292944098292301	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292944098292301	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292522098291901	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292643098241801	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292643098241801	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292328098294601	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
294225098080301	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
294225098080301	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
291210099475601	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
293111098340901	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
293512098291701	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292424098421501	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	1,2,3-Tri- chloro- benzene (µg/L)	1,2,3-Tri- chloro- propane (µg/L)	1,2,3-Tri- methyl- benzene (µg/L)	1,2,4-Tri- chloro- benzene (µg/L)	1,2,4-Tri- methyl- benzene (µg/L)	1,2-Di- bromo- 3-chloro- propane (µg/L)	1,2-Di- bromo- ethane (µg/L)	1,2-Di- chloro- benzene (µg/L)	1,2-Di- chloro- ethane (µg/L)	1,2-Di- chloro- propane (µg/L)	1,3,5-Tri- methyl- benzene (µg/L)	1,3-Di- chloro- benzene (µg/L)	1,3-Di- chloro- propane (µg/L)	1,4-Di- chloro- benzene (µg/L)
292053098365501	<0.18	<0.18	<0.06	<0.12	<0.056	<0.51	<0.036	<0.048	<0.13	<0.029	<0.044	<0.03	<0.06	<0.034
293807098155301	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292604098563201	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292116099095501	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
291232099470301	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
293128098473101	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
293451098313201	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
294604098060801	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
294019098114701	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292215098580201	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292119098524901	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
291943099163301	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
291840099382601	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
293518098332601	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
293023098355401	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
293042098305201	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292442098474501	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292405098371201	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
292931098274601	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
293729098173101	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
291219099095601	<.18	<.18	<.06	<.12	<.056	<.51	<.036	<.048	<.13	<.029	<.044	<.03	<.06	<.034
Monitoring wells														
292943098354401	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
292943098354402	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
292943098354403	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
292943098354404	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	1,2,3-Tri- chloro- benzene (µg/L)	1,2,3-Tri- chloro- propane (µg/L)	1,2,3-Tri- methyl- benzene (µg/L)	1,2,4-Tri- chloro- benzene (µg/L)	1,2,4-Tri- methyl- benzene (µg/L)	1,2-Di- bromo- 3-chloro- propane (µg/L)	1,2-Di- bromo- ethane (µg/L)	1,2-Di- chloro- benzene (µg/L)	1,2-Di- chloro- ethane (µg/L)	1,2-Di- chloro- propane (µg/L)	1,3,5-Tri- methyl- benzene (µg/L)	1,3-Di- chloro- benzene (µg/L)	1,3-Di- chloro- propane (µg/L)	1,4-Di- chloro- benzene (µg/L)
292851098374401	<0.08	<0.12	<0.08	<0.08	<0.04	<0.5	<0.04	<0.02	<0.06	<0.02	<0.04	<0.04	<0.06	<0.02
292851098374402	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
Well-field depth-dependent sampling (W4)														
292916098360701	<.08	<.12	<.08	<.08	E.040	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
292916098360701	<.08	<.12	<.08	<.08	E.046	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
292916098360701	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
292916098360701	<.08	<.12	<.08	<.08	E.026	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
292916098360701	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
292916098360701	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
292916098360701	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
292916098360701	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
292916098360701	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
292916098360701	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
Well-field wells (wellhead sampling)														
292919098360501	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
292925098360201	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
292916098360701	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
292920098360601	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
292923098360301	<.08	<.12	<.08	<.08	<.04	<.5	<.04	<.02	<.06	<.02	<.04	<.04	<.06	<.02
Temporal sampling														
292923098360301	<.06	<.12	<.08	<.04	<.04	<1	<.04	<.02	<.06	<.02	<.04	<.02	<.06	<.02
292923098360301	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	1,2,3-Tri- chloro- benzene (µg/L)	1,2,3-Tri- chloro- propane (µg/L)	1,2,3-Tri- methyl- benzene (µg/L)	1,2,4-Tri- chloro- benzene (µg/L)	1,2,4-Tri- methyl- benzene (µg/L)	1,2-Di- bromo- 3-chloro- propane (µg/L)	1,2-Di- bromo- ethane (µg/L)	1,2-Di- chloro- benzene (µg/L)	1,2-Di- chloro- ethane (µg/L)	1,2-Di- chloro- propane (µg/L)	1,3,5-Tri- methyl- benzene (µg/L)	1,3-Di- chloro- benzene (µg/L)	1,3-Di- chloro- propane (µg/L)	1,4-Di- chloro- benzene (µg/L)
292923098360301	<0.06	<0.12	<0.06	<0.08	<0.032	<0.34	<0.05	<0.028	<0.08	<0.026	<0.032	<0.024	<0.06	<0.026
292923098360301	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292923098360301	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292923098360301	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354402	<.06	<.12	<.08	<.04	<.04	<1	<.04	<.02	<.06	<.02	<.04	<.02	<.06	<.02
292943098354402	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354402	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354402	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354402	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354402	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354402	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354403	<.06	<.12	<.08	<.04	<.04	<1	<.04	<.02	<.06	<.02	<.04	<.02	<.06	<.02
292943098354403	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354403	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354403	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354403	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354403	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354403	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354404	<.06	<.12	<.08	<.04	<.04	<1	<.04	<.02	<.06	<.02	<.04	<.02	<.06	<.02
292943098354404	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354404	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354404	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354404	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354404	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
292943098354404	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
293252098380801	<.06	<.12	<.08	<.04	<.04	<1	<.04	<.02	<.06	<.02	<.04	<.02	<.06	<.02
293252098380801	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
293252098380801	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
293252098380801	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
293252098380801	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026
293252098380801	<.06	<.12	<.06	<.08	<.032	<.34	<.05	<.028	<.08	<.026	<.032	<.024	<.06	<.026

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	2,2-Di- chloro- propane (µg/L)	2- Chloro- toluene (µg/L)	2-Ethyl- toluene (µg/L)	3-Chloro- propene (µg/L)	4-Chloro- toluene (µg/L)	4-Iso- propyl- toluene (µg/L)	Acetone (µg/L)	Acrylo- nitrile (µg/L)	Benzene (µg/L)	Bromo- benzene (µg/L)	Bromo- chloro- methane (µg/L)	Bromo- dichloro- methane (µg/L)	Bromo- ethene (µg/L)	Bromo- meth- ane (µg/L)	Carbon di- sulfide (µg/L)
Regional aquifer public-supply wells															
293359098290301	<0.05	<0.04	<0.06	<0.5	<0.05	<0.08	<6	<0.8	<0.021	<0.028	<0.12	<0.028	<0.1	<0.26	<0.038
293359098290301	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
293358098231101	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
293358098231101	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
293120098285801	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	.163	<.1	<.26	<.038
293120098285801	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	E.035	<.1	<.26	<.038
293145098224201	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
293119098211201	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
292843098425101	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
292459098382101	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
292925098360201	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
292925098360201	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
292822098325401	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
292822098325401	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
292944098292301	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
292944098292301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292522098291901	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
292643098241801	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	.788	<.1	<.26	<.038
292643098241801	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
292328098294601	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
294225098080301	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
294225098080301	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
291210099475601	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
293111098340901	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
293512098291701	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	E.027	<.1	<.26	<.038
292424098421501	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	2,2-Di- chloro- propane (µg/L)	2- Chloro- toluene (µg/L)	2-Ethyl- toluene (µg/L)	3-Chloro- propene (µg/L)	4-Chloro- toluene (µg/L)	4-Iso- propyl- toluene (µg/L)	Acetone (µg/L)	Acrylo- nitrile (µg/L)	Benzene (µg/L)	Bromo- benzene (µg/L)	Bromo- chloro- methane (µg/L)	Bromo- dichloro- methane (µg/L)	Bromo- ethene (µg/L)	Bromo- meth- ane (µg/L)	Carbon di- sulfide (µg/L)
292053098365501	<0.05	<0.04	<0.06	<0.5	<0.05	<0.08	<6	<0.8	<0.021	<0.028	<0.12	E0.046	<0.1	<0.26	<0.038
293807098155301	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	E.056	<.1	<.26	<.038
292604098563201	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
292116099095501	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
291232099470301	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
293128098473101	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
293451098313201	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	E.029	<.1	<.26	<.038
294604098060801	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
294019098114701	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
292215098580201	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
292119098524901	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
291943099163301	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
291840099382601	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
293518098332601	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
293023098355401	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
293042098305201	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
292442098474501	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	.123	<.1	<.26	<.038
292405098371201	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
292931098274601	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
293729098173101	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
291219099095601	<.05	<.04	<.06	<.5	<.05	<.08	<6	<.8	<.021	<.028	<.12	<.028	<.1	<.26	<.038
Monitoring wells															
292943098354401	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	E.026
292943098354402	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	E.074
292943098354403	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	<.06
292943098354404	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	<.06

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	2,2-Di- chloro- propane (µg/L)	2- Chloro- toluene (µg/L)	2-Ethyl- toluene (µg/L)	3-Chloro- propene (µg/L)	4-Chloro- toluene (µg/L)	4-Iso- propyl- toluene (µg/L)	Acetone (µg/L)	Acrylo- nitrile (µg/L)	Benzene (µg/L)	Bromo- benzene (µg/L)	Bromo- chloro- methane (µg/L)	Bromo- dichloro- methane (µg/L)	Bromo- ethene (µg/L)	Bromo- meth- ane (µg/L)	Carbon di- sulfide (µg/L)
292851098374401	<0.06	<0.04	<0.04	<0.08	<0.04	<0.08	<4	<0.4	<0.02	<0.02	<0.06	<0.04	<0.12	<0.4	<0.06
292851098374402	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	<.06
Well-field depth-dependent sampling (W4)															
292916098360701	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	<.06
292916098360701	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	<.06
292916098360701	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	<.06
292916098360701	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	<.06
292916098360701	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	<.06
292916098360701	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	<.06
292916098360701	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	<.06
292916098360701	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	<.06
292916098360701	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	<.06
292916098360701	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	<.06
Well-field wells (wellhead sampling)															
292919098360501	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	<.06
292925098360201	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	E.071	<.12	<.4	<.06
292916098360701	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	<.04	<.12	<.4	<.06
292920098360601	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	E.032	<.12	<.4	<.06
292923098360301	<.06	<.04	<.04	<.08	<.04	<.08	<4	<.4	<.02	<.02	<.06	E.037	<.12	<.4	<.06
Temporal sampling															
292923098360301	<.06	<.02	<.02	<.08	<.02	<.06	<4	<.4	<.016	<.02	<.06	<.04	<.12	<.4	<.04
292923098360301	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	E.032	<.12	<.2	<.04

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	2,2-Di- chloro- propane (µg/L)	2- Chloro- toluene (µg/L)	2-Ethyl- toluene (µg/L)	3-Chloro- propene (µg/L)	4-Chloro- toluene (µg/L)	4-Iso- propyl- toluene (µg/L)	Acetone (µg/L)	Acrylo- nitrile (µg/L)	Benzene (µg/L)	Bromo- benzene (µg/L)	Bromo- chloro- methane (µg/L)	Bromo- dichloro- methane (µg/L)	Bromo- ethene (µg/L)	Bromo- meth- ane (µg/L)	Carbon di- sulfide (µg/L)
292923098360301	<0.06	<0.028	<0.032	<0.08	<0.042	<0.06	<3.4	<0.8	<0.026	<0.022	<0.06	<0.034	<0.12	<0.2	<0.04
292923098360301	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	E.096	<.12	<.2	<.04
292923098360301	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	E.066	<.12	<.2	<.04
292923098360301	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354402	<.06	<.02	<.02	<.08	<.02	<.06	<4	<.4	<.016	<.02	<.06	<.04	<.12	<.4	<.04
292943098354402	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354402	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354402	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354402	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354402	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354402	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354403	<.06	<.02	<.02	<.08	<.02	<.06	<4	<.4	<.016	<.02	<.06	<.04	<.12	<.4	<.04
292943098354403	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354403	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354403	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354403	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354403	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354403	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354404	<.06	<.02	<.02	<.08	<.02	<.06	<4	<.4	<.016	<.02	<.06	<.04	<.12	<.4	<.04
292943098354404	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354404	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354404	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354404	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354404	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
292943098354404	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
293252098380801	<.06	<.02	<.02	<.08	<.02	<.06	<4	<.4	<.016	<.02	<.06	<.04	<.12	<.4	<.04
293252098380801	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	E.053	<.12	<.2	<.04
293252098380801	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	<.034	<.12	<.2	<.04
293252098380801	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	E.077	<.12	<.2	<.04
293252098380801	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	E.071	<.12	<.2	<.04
293252098380801	<.06	<.028	<.032	<.08	<.042	<.06	<3.4	<.8	<.026	<.022	<.06	.155	<.12	<.2	<.04

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	Chloro- benzene (µg/L)	Chloro- ethane (µg/L)	Chloro- methane (µg/L)	cis-1,2- Di- chloro- ethene (µg/L)	cis-1,3- Di- chloro- propene (µg/L)	Di- bromo- chloro- methane (µg/L)	Di- bromo- methane (µg/L)	Dichloro- difluoro- methane (CFC-12) (µg/L)	Di- chloro- meth- ane (µg/L)	Di- ethyl ether (µg/L)	Di- iso- propyl ether (µg/L)	Ethyl meth- acrylate (µg/L)	Ethyl methyl ketone (µg/L)	Ethyl- ben- zene (µg/L)	Hexa- chloro- buta- diene (µg/L)	Hexa- chloro- ethane (µg/L)
Regional aquifer public-supply wells																
293359098290301	<0.028	<0.12	<0.17	<0.024	<0.05	<0.1	<0.05	<0.18	<0.06	<0.08	<0.1	<0.18	<2	<0.03	<0.14	<0.14
293359098290301	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
293358098231101	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
293358098231101	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
293120098285801	<.028	<.12	<.17	<.024	<.05	.519	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
293120098285801	<.028	<.12	<.17	<.024	<.05	E.099	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
293145098224201	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
293119098211201	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292843098425101	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292459098382101	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292925098360201	<.028	<.12	<.17	<.024	<.05	.154	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292925098360201	<.028	<.12	<.17	<.024	<.05	E.057	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292822098325401	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292822098325401	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292944098292301	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292944098292301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292522098291901	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292643098241801	<.028	<.12	<.17	<.024	<.05	1.01	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292643098241801	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292328098294601	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
294225098080301	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
294225098080301	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
291210099475601	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
293111098340901	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
293512098291701	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292424098421501	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	Chloro- benzene (µg/L)	Chloro- ethane (µg/L)	Chloro- methane (µg/L)	cis-1,2- Di- chloro- ethene (µg/L)	cis-1,3- Di- chloro- propene (µg/L)	Di- bromo- chloro- methane (µg/L)	Di- bromo- methane (µg/L)	Dichloro- difluoro- methane (CFC-12) (µg/L)	Di- chloro- meth- ane (µg/L)	Di- ethyl ether (µg/L)	Di- iso- propyl ether (µg/L)	Ethyl meth- acrylate (µg/L)	Ethyl methyl ketone (µg/L)	Ethyl- ben- zene (µg/L)	Hexa- chloro- buta- diene (µg/L)	Hexa- chloro- ethane (µg/L)
292053098365501	<0.028	<0.12	<0.17	<0.024	<0.05	E0.081	<0.05	<0.18	<0.06	<0.08	<0.1	<0.18	<2	<0.03	<0.14	<0.14
293807098155301	<.028	<.12	<.17	<.024	<.05	.166	<.05	<.18	<.06	<.08	<.1	<.18	<2	E.019	<.14	<.14
292604098563201	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292116099095501	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
291232099470301	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
293128098473101	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
293451098313201	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
294604098060801	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
294019098114701	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292215098580201	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292119098524901	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
291943099163301	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
291840099382601	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
293518098332601	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
293023098355401	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
293042098305201	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292442098474501	<.028	<.12	<.17	<.024	<.05	.263	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292405098371201	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
292931098274601	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
293729098173101	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
291219099095601	<.028	<.12	<.17	<.024	<.05	<.1	<.05	<.18	<.06	<.08	<.1	<.18	<2	<.03	<.14	<.14
Monitoring wells																
292943098354401	<.02	<.1	<.1	<.02	<.1	<.12	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292943098354402	<.02	<.1	<.1	<.02	<.1	<.12	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292943098354403	<.02	<.1	<.1	<.02	<.1	<.12	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292943098354404	<.02	<.1	<.1	<.02	<.1	<.12	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	Chloro- benzene (µg/L)	Chloro- ethane (µg/L)	Chloro- methane (µg/L)	cis-1,2- Di- chloro- ethene (µg/L)	cis-1,3- Di- chloro- propene (µg/L)	Di- bromo- chloro- methane (µg/L)	Di- bromo- methane (µg/L)	Dichloro- difluoro- methane (CFC-12) (µg/L)	Di- chloro- meth- ane (µg/L)	Di- ethyl ether (µg/L)	Di- iso- propyl ether (µg/L)	Ethyl meth- acrylate (µg/L)	Ethyl methyl ketone (µg/L)	Ethyl- ben- zene (µg/L)	Hexa- chloro- buta- diene (µg/L)	Hexa- chloro- ethane (µg/L)
292851098374401	<0.02	<0.1	<0.1	<0.02	<0.1	<0.12	<0.004	<0.14	<0.04	<0.12	<0.06	<0.14	<1.6	<0.04	<0.06	<0.14
292851098374402	<.02	<.1	<.1	<.02	<.1	<.12	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
Well-field depth-dependent sampling (W4)																
292916098360701	<.02	<.1	<.1	<.02	<.1	<.12	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292916098360701	<.02	<.1	<.1	<.02	<.1	<.12	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292916098360701	<.02	<.1	<.1	<.02	<.1	<.12	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292916098360701	<.02	<.1	<.1	<.02	<.1	<.12	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292916098360701	<.02	<.1	<.1	<.02	<.1	<.12	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292916098360701	<.02	<.1	<.1	<.02	<.1	<.12	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292916098360701	<.02	<.1	<.1	<.02	<.1	<.12	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292916098360701	<.02	<.1	<.1	<.02	<.1	<.12	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292916098360701	<.02	<.1	<.1	<.02	<.1	<.12	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
Well-field wells (wellhead sampling)																
292919098360501	<.02	<.1	<.1	<.02	<.1	<.12	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292925098360201	<.02	<.1	<.1	<.02	<.1	.299	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292916098360701	<.02	<.1	<.1	<.02	<.1	<.12	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292920098360601	<.02	<.1	<.1	<.02	<.1	.123	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292923098360301	<.02	<.1	<.1	<.02	<.1	.203	<.04	<.14	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
Temporal sampling																
292923098360301	<.02	<.1	<.14	<.02	<.1	<.12	<.04	<.1	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292923098360301	<.016	<.06	<.14	<.022	<.1	E.114	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	Chloro- benzene (µg/L)	Chloro- ethane (µg/L)	Chloro- methane (µg/L)	cis-1,2- Di- chloro- ethene (µg/L)	cis-1,3- Di- chloro- propene (µg/L)	Di- bromo- chloro- methane (µg/L)	Di- bromo- methane (µg/L)	Dichloro- difluoro- methane (CFC-12) (µg/L)	Di- chloro- meth- ane (µg/L)	Di- ethyl ether (µg/L)	Di- iso- propyl ether (µg/L)	Ethyl meth- acrylate (µg/L)	Ethyl methyl ketone (µg/L)	Ethyl- ben- zene (µg/L)	Hexa- chloro- buta- diene (µg/L)	Hexa- chloro- ethane (µg/L)
292923098360301	<0.016	<0.06	<0.14	<0.022	<0.1	<0.12	<0.05	<0.1	<0.038	<0.08	<0.06	<0.14	<1.6	<0.036	<0.06	<0.14
292923098360301	<.016	<.06	<.14	<.022	<.1	.310	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292923098360301	<.016	<.06	<.14	<.022	<.1	.225	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292923098360301	<.016	<.06	<.14	<.022	<.1	E.115	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354402	<.02	<.1	<.14	<.02	<.1	<.12	<.04	<.1	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292943098354402	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354402	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354402	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354402	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354402	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354402	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354403	<.02	<.1	<.14	<.02	<.1	<.12	<.04	<.1	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292943098354403	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354403	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354403	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354403	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354403	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354403	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354404	<.02	<.1	<.14	<.02	<.1	<.12	<.04	<.1	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
292943098354404	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354404	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354404	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354404	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354404	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354404	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
292943098354404	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
293252098380801	<.02	<.1	<.14	<.02	<.1	<.12	<.04	<.1	<.04	<.12	<.06	<.14	<1.6	<.04	<.06	<.14
293252098380801	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
293252098380801	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
293252098380801	<.016	<.06	<.14	<.022	<.1	<.12	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14
293252098380801	<.016	<.06	<.14	<.022	<.1	E.048	<.05	<.1	<.038	<.08	<.06	<.14	<1.6	<.036	<.06	<.14

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	Iodo- methane (µg/L)	Iso- butyl methyl ketone (µg/L)	Iso- propyl- ben- zene (µg/L)	Methyl arylo- nitrile (µg/L)	Methyl acrylate (µg/L)	Methyl meth- acrylate (µg/L)	Methyl tert- pentyl ether (µg/L)	m- plus p- Xylene (µg/L)	Naphth- alene (µg/L)	n-Butyl methyl ketone (µg/L)	n-Butyl- ben- zene (µg/L)	n- Propyl- benzene (µg/L)	o- Xylene (µg/L)	sec- Butyl- benzene (µg/L)	Styrene (µg/L)	tert- Amyl alcohol (µg/L)
Regional aquifer public-supply wells																
293359098290301	<0.5	<0.37	<0.038	<0.4	<1	<0.2	<0.04	<0.06	<0.52	<0.4	<0.12	<0.042	<0.038	<0.06	<0.042	<1
293359098290301	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	--
293358098231101	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
293358098231101	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	--
293120098285801	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
293120098285801	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	--
293145098224201	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
293119098211201	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
292843098425101	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
292459098382101	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
292925098360201	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
292925098360201	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	--
292822098325401	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
292822098325401	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	--
292944098292301	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
292944098292301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292522098291901	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
292643098241801	<.5	<.37	<.038	<.4	<1	<.2	<.04	E.021	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
292643098241801	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	--
292328098294601	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
294225098080301	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
294225098080301	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	--
291210099475601	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
293111098340901	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
293512098291701	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
292424098421501	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	Iodo- methane (µg/L)	Iso- butyl methyl ketone (µg/L)	Iso- propyl- ben- zene (µg/L)	Methyl arylo- nitrile (µg/L)	Methyl acrylate (µg/L)	Methyl meth- acrylate (µg/L)	Methyl tert- pentyl ether (µg/L)	m- plus p- Xylene (µg/L)	Naphth- alene (µg/L)	n-Butyl methyl ketone (µg/L)	n-Butyl- ben- zene (µg/L)	n- Propyl- benzene (µg/L)	o- Xylene (µg/L)	sec- Butyl- benzene (µg/L)	Styrene (µg/L)	tert- Amyl alcohol (µg/L)
292053098365501	<0.5	<0.37	<0.038	<0.4	<1	<0.2	<0.04	<0.06	<0.52	<0.4	<0.12	<0.042	<0.038	<0.06	<0.042	<1
293807098155301	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
292604098563201	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
292116099095501	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
291232099470301	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
293128098473101	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
293451098313201	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
294604098060801	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
294019098114701	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
292215098580201	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
292119098524901	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
291943099163301	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
291840099382601	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
293518098332601	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
293023098355401	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
293042098305201	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
292442098474501	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
292405098371201	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
292931098274601	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
293729098173101	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
291219099095601	<.5	<.37	<.038	<.4	<1	<.2	<.04	<.06	<.52	<.4	<.12	<.042	<.038	<.06	<.042	<1
Monitoring wells																
292943098354401	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
292943098354402	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
292943098354403	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
292943098354404	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	Iodo- methane (µg/L)	Iso- butyl methyl ketone (µg/L)	Iso- propyl- ben- zene (µg/L)	Methyl arylo- nitrile (µg/L)	Methyl acrylate (µg/L)	Methyl meth- acrylate (µg/L)	Methyl tert- pentyl ether (µg/L)	m- plus p- Xylene (µg/L)	Naphth- alene (µg/L)	n-Butyl methyl ketone (µg/L)	n-Butyl- ben- zene (µg/L)	n- Propyl- benzene (µg/L)	o- Xylene (µg/L)	sec- Butyl- benzene (µg/L)	Styrene (µg/L)	tert- Amyl alcohol (µg/L)
292851098374401	<0.4	<0.4	<0.04	<0.2	<0.6	<0.2	<0.06	<0.08	<0.2	<0.6	<0.14	<0.04	<0.04	<0.04	<0.04	--
292851098374402	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
Well-field depth-dependent sampling (W4)																
292916098360701	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
292916098360701	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
292916098360701	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
292916098360701	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
292916098360701	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
292916098360701	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
292916098360701	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
292916098360701	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
292916098360701	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
Well-field wells (wellhead sampling)																
292919098360501	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
292925098360201	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
292916098360701	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
292920098360601	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
292923098360301	<.4	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.14	<.04	<.04	<.04	<.04	--
Temporal sampling																
292923098360301	<.8	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.08	<.04	<.04	<.02	<.04	--
292923098360301	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	Iodo- methane (µg/L)	Iso- butyl methyl ketone (µg/L)	Iso- propyl- ben- zene (µg/L)	Methyl arylo- nitrile (µg/L)	Methyl acrylate (µg/L)	Methyl meth- acrylate (µg/L)	Methyl tert- pentyl ether (µg/L)	m- plus p- Xylene (µg/L)	Naphth- alene (µg/L)	n-Butyl methyl ketone (µg/L)	n-Butyl- ben- zene (µg/L)	n- Propyl- benzene (µg/L)	o- Xylene (µg/L)	sec- Butyl- benzene (µg/L)	Styrene (µg/L)	tert- Amyl alcohol (µg/L)
292923098360301	<0.26	<0.32	<0.042	<0.26	<0.56	<0.22	<0.06	<0.08	<0.18	<0.46	<0.08	<0.036	<0.032	<0.034	<0.03	--
292923098360301	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292923098360301	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292923098360301	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354402	<.8	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.08	<.04	<.04	<.02	<.04	--
292943098354402	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354402	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354402	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354402	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354402	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354402	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354403	<.8	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.08	<.04	<.04	<.02	<.04	--
292943098354403	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354403	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354403	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354403	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354403	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354403	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354404	<.8	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.08	<.04	<.04	<.02	<.04	--
292943098354404	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354404	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354404	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354404	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
292943098354404	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
293252098380801	<.8	<.4	<.04	<.2	<.6	<.2	<.06	<.08	<.2	<.6	<.08	<.04	<.04	<.02	<.04	--
293252098380801	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
293252098380801	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
293252098380801	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
293252098380801	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--
293252098380801	<.26	<.32	<.042	<.26	<.56	<.22	<.06	<.08	<.18	<.46	<.08	<.036	<.032	<.034	<.03	--

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	tert- Butyl alcohol (µg/L)	tert- Butyl ethyl ether (µg/L)	Methyl tert-butyl ether (MTBE) (µg/L)	tert- Butyl- benzene (µg/L)	Tetra- chloro- ethene (PCE) (µg/L)	Tetra- chloro- methane (µg/L)	Tetra- hydro- furan (µg/L)	Toluene (µg/L)	trans- 1,2-Di- chloro- ethene (µg/L)	trans- 1,3-Di- chloro- propene (µg/L)	trans- 1,4-Di- chloro- 2-butene (µg/L)	Tri- bromo- meth- ane (bromo- form) (µg/L)	Tri- chloro- ethene (TCE) (µg/L)	Tri- chloro- fluoro- methane (CFC-11) (µg/L)	Tri- chloro- methane (chloro- form) (µg/L)	Vinyl chloro- ride (µg/L)
Regional aquifer public-supply wells																
293359098290301	<2	<0.03	<0.1	<0.06	0.549	<0.06	<1	<0.02	<0.032	<0.09	<0.7	0.1	0.038	0.275	0.094	0.08
293359098290301	--	<.03	<.1	<.06	.830	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	.109	E.085	<.08
293358098231101	<2	<.03	<.1	<.06	.158	<.06	<1	<.02	<.032	<.09	<.7	<.1	E.016	<.08	E.035	<.08
293358098231101	--	<.03	<.1	<.06	.183	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	E.035	<.08
293120098285801	<2	<.03	<.1	<.06	.130	<.06	<1	<.02	<.032	<.09	<.7	1.90	E.023	<.08	E.089	<.08
293120098285801	--	<.03	<.1	<.06	.244	<.06	<1	<.02	<.032	<.09	<.7	.386	E.021	<.08	E.056	<.08
293145098224201	<2	<.03	<.1	<.06	E.046	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	E.021	<.08
293119098211201	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	<.024	<.08
292843098425101	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	<.024	<.08
292459098382101	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	<.024	<.08
292925098360201	<2	<.03	<.1	<.06	.107	<.06	<1	<.02	<.032	<.09	<.7	.477	<.038	<.08	E.020	<.08
292925098360201	--	<.03	<.1	<.06	.140	<.06	<1	<.02	<.032	<.09	<.7	.352	<.038	<.08	E.025	<.08
292822098325401	<2	<.03	<.1	<.06	E.017	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	E.022	<.08
292822098325401	--	<.03	<.1	<.06	E.019	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	E.030	<.08
292944098292301	<2	<.03	<.1	<.06	E.030	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	<.024	<.08
292944098292301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292522098291901	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	<.024	<.08
292643098241801	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	1.02	<.038	<.08	.360	<.08
292643098241801	--	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	E.017	<.08
292328098294601	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	<.024	<.08
294225098080301	<2	<.03	<.1	<.06	.110	<.06	<1	E.011	<.032	<.09	<.7	<.1	<.038	<.08	E.026	<.08
294225098080301	--	<.03	<.1	<.06	E.087	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	E.024	<.08
291210099475601	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	E.010	<.08
293111098340901	<2	<.03	<.1	<.06	E.046	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	E.040	<.08
293512098291701	<2	<.03	<.1	<.06	E.037	<.06	<1	E.009	<.032	<.09	<.7	E.057	<.038	<.08	E.077	<.08
292424098421501	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	<.024	<.08

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	tert- Butyl alcohol (µg/L)	tert- Butyl ethyl ether (µg/L)	Methyl tert-butyl ether (MTBE) (µg/L)	tert- Butyl- benzene (µg/L)	Tetra- chloro- ethene (PCE) (µg/L)	Tetra- chloro- methane (µg/L)	Tetra- hydro- furan (µg/L)	Toluene (µg/L)	trans- 1,2-Di- chloro- ethene (µg/L)	trans- 1,3-Di- chloro- propene (µg/L)	trans- 1,4-Di- chloro- 2-butene (µg/L)	Tri- bromo- meth- ane (bromo- form) (µg/L)	Tri- chloro- ethene (TCE) (µg/L)	Tri- chloro- fluoro- methane (CFC-11) (µg/L)	Tri- chloro- methane (chloro- form) (µg/L)	Vinyl chloro- ride (µg/L)
292053098365501	<2	<0.03	<0.1	<0.06	<0.03	<0.06	<1	<0.02	<0.032	<0.09	<0.7	0.347	<0.038	<0.08	E.025	<0.08
293807098155301	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	.494	<.038	<.08	E.019	<.08
292604098563201	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	<.024	<.08
292116099095501	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	<.024	<.08
291232099470301	<2	<.03	<.1	<.06	E.069	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	E.015	<.08
293128098473101	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	<.024	<.08
293451098313201	<2	<.03	<.1	<.06	E.026	<.06	<1	<.02	<.032	<.09	<.7	<.1	E.01502	<.08	.101	<.08
294604098060801	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	<.024	<.08
294019098114701	<2	<.03	<.1	<.06	.120	<.06	<1	<.02	<.032	<.09	<.7	E.099	<.038	<.08	E.021	<.08
292215098580201	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	.101	<.038	<.08	E.014	<.08
292119098524901	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	E.011	<.08
291943099163301	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	E.011	<.08
291840099382601	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	E.015	<.08
293518098332601	<2	<.03	<.1	<.06	E.059	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	.100	<.08
293023098355401	<2	<.03	<.1	<.06	.223	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	E.016	<.08
293042098305201	<2	<.03	<.1	<.06	.197	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	E.044	<.08
292442098474501	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	.345	<.038	<.08	E.056	<.08
292405098371201	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	<.024	<.08
292931098274601	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	.172	<.038	<.08	E.020	<.08
291219099095601	<2	<.03	<.1	<.06	<.03	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	<.024	<.08
293729098173101	<2	<.03	<.1	<.06	E.016	<.06	<1	<.02	<.032	<.09	<.7	<.1	<.038	<.08	E.019	<.08
Monitoring wells																
292943098354401	--	<.04	<.1	<.06	.372	<.08	<1.4	<.02	<.02	<.1	<.6	<.08	<.02	<.08	.218	<.08
292943098354402	--	<.04	<.1	<.06	.365	<.08	<1.4	<.02	<.02	<.1	<.6	<.08	<.02	<.08	.131	<.08
292943098354403	--	<.04	<.1	<.06	.479	<.08	<1.4	<.02	<.02	<.1	<.6	<.08	<.02	<.08	E.037	<.08
292943098354404	--	<.04	<.1	<.06	.174	<.08	<1.4	<.02	<.02	<.1	<.6	<.08	<.02	<.08	E.015	<.08

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	tert- Butyl alcohol (µg/L)	tert- Butyl ethyl ether (µg/L)	Methyl tert-butyl ether (MTBE) (µg/L)	tert- Butyl- benzene (µg/L)	Tetra- chloro- ethene (PCE) (µg/L)	Tetra- chloro- methane (µg/L)	Tetra- hydro- furan (µg/L)	Toluene (µg/L)	trans- 1,2-Di- chloro- ethene (µg/L)	trans- 1,3-Di- chloro- propene (µg/L)	trans- 1,4-Di- chloro- 2-butene (µg/L)	Tri- bromo- meth- ane (bromo- form) (µg/L)	Tri- chloro- ethene (TCE) (µg/L)	Tri- chloro- fluoro- methane (CFC-11) (µg/L)	Tri- chloro- methane (chloro- form) (µg/L)	Vinyl chloro- ride (µg/L)
292851098374401	--	<0.04	<0.1	<0.06	E0.060	<0.08	<1.4	<0.02	<0.02	<0.1	<0.6	<0.08	<0.02	<0.08	<0.02	<0.08
292851098374402	--	<.04	<.1	<.06	.164	<.08	<1.4	<.02	<.02	<.1	<.6	<.08	E.066	E.032	E.028	<.08
Well-field depth-dependent sampling (W4)																
292916098360701	--	<.04	<.1	<.06	E.055	<.08	<1.4	<.02	<.02	<.1	<.6	<.08	<.02	<.08	E.028	<.08
292916098360701	--	<.04	<.1	<.06	E.055	<.08	<1.4	<.02	<.02	<.1	<.6	<.08	<.02	<.08	E.021	<.08
292916098360701	--	<.04	<.1	<.06	.127	<.08	<1.4	<.02	<.02	<.1	<.6	<.08	<.02	<.08	E.034	<.08
292916098360701	--	<.04	<.1	<.06	E.057	<.08	<1.4	<.02	<.02	<.1	<.6	<.08	<.02	<.08	E.024	<.08
292916098360701	--	<.04	<.1	<.06	E.078	<.08	<1.4	<.02	<.02	<.1	<.6	<.08	<.02	<.08	E.028	<.08
292916098360701	--	<.04	<.1	<.06	.102	<.08	<1.4	<.02	<.02	<.1	<.6	<.08	<.02	<.08	E.033	<.08
292916098360701	--	<.04	<.1	<.06	.137	<.08	<1.4	<.02	<.02	<.1	<.6	<.08	<.02	<.08	E.030	<.08
292916098360701	--	<.04	<.1	<.06	E.086	<.08	<1.4	<.02	<.02	<.1	<.6	<.08	<.02	<.08	E.031	<.08
292916098360701	--	<.04	<.1	<.06	E.049	<.08	<1.4	<.02	<.02	<.1	<.6	<.08	<.02	<.08	E.026	<.08
Well-field wells (wellhead sampling)																
292919098360501	--	<.04	<.1	<.06	.176	<.08	<1.4	<.02	<.02	<.1	<.6	.327	<.02	<.08	E.038	<.08
292925098360201	--	<.04	<.1	<.06	.115	<.08	<1.4	<.02	<.02	<.1	<.6	2.01	<.02	<.08	E.029	<.08
292916098360701	--	<.04	<.1	<.06	.126	<.08	<1.4	<.02	<.02	<.1	<.6	<.08	<.02	<.08	E.029	<.08
292920098360601	--	<.04	<.1	<.06	.223	<.08	<1.4	<.02	<.02	<.1	<.6	.756	<.02	<.08	E.032	<.08
292923098360301	--	<.04	<.1	<.06	.210	<.08	<1.4	<.02	<.02	<.1	<.6	1.15	<.02	<.08	E.037	<.08
Temporal sampling																
292923098360301	--	<.04	<.1	<.06	.208	<.06	<1.4	<.018	<.018	<.1	<.4	.260	<.02	<.08	E.040	<.08
292923098360301	--	<.032	<.1	<.06	.292	<.052	<1.4	<.018	<.018	<.14	<.36	.561	<.022	<.08	E.031	<.06

Table 14. Volatile organic and fuel oxygenate compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured or not applicable; E, estimated; <, nondetection less than laboratory method reporting level]

USGS identification number	tert- Butyl alcohol (µg/L)	tert- Butyl ethyl ether (µg/L)	Methyl tert-butyl ether (MTBE) (µg/L)	tert- Butyl- benzene (µg/L)	Tetra- chloro- ethene (PCE) (µg/L)	Tetra- chloro- methane (µg/L)	Tetra- hydro- furan (µg/L)	Toluene (µg/L)	trans- 1,2-Di- chloro- ethene (µg/L)	trans- 1,3-Di- chloro- propene (µg/L)	trans- 1,4-Di- chloro- 2-butene (µg/L)	Tri- bromo- meth- ane (bromo- form) (µg/L)	Tri- chloro- ethene (TCE) (µg/L)	Tri- chloro- fluoro- methane (CFC-11) (µg/L)	Tri- chloro- methane (chloro- form) (µg/L)	Vinyl chloro- ride (µg/L)
292923098360301	--	<0.032	<0.1	<0.06	0.290	<0.052	<1.4	<0.018	<0.018	<0.14	<0.36	0.475	<0.022	<0.08	E0.032	<0.06
292923098360301	--	<.032	<.1	<.06	.408	<.052	<1.4	<.018	<.018	<.14	<.36	1.00	<.022	<.08	E.044	<.06
292923098360301	--	<.032	<.1	<.06	.295	<.052	<1.4	<.018	<.018	<.14	<.36	.943	<.022	<.08	E.038	<.06
292923098360301	--	<.032	<.1	<.06	.247	<.052	<1.4	<.018	<.018	<.14	<.36	.644	<.022	<.08	E.039	<.06
292943098354402	--	<.04	<.1	<.06	.383	<.06	<1.4	<.018	<.018	<.1	<.4	<.1	<.02	<.08	<.04	<.08
292943098354402	--	<.032	<.1	<.06	.634	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	E.011	E.023	<.06
292943098354402	--	<.032	<.1	<.06	.680	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	E.023	<.06
292943098354402	--	<.032	<.1	<.06	.681	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	E.024	<.06
292943098354402	--	<.032	<.1	<.06	.706	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	E.024	<.06
292943098354402	--	<.032	<.1	<.06	.574	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	E.026	<.06
292943098354403	--	<.04	<.1	<.06	.587	<.06	<1.4	<.018	<.018	<.1	<.4	<.1	<.02	<.08	<.04	<.08
292943098354403	--	<.032	<.1	<.06	.727	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	E.021	<.06
292943098354403	--	<.032	<.1	<.06	.810	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	E.024	<.06
292943098354403	--	<.032	<.1	<.06	.814	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	E.020	<.06
292943098354403	--	<.032	<.1	<.06	.733	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	E.031	<.06
292943098354403	--	<.032	<.1	<.06	.733	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	E.028	<.06
292943098354404	--	<.04	<.1	<.06	.257	<.06	<1.4	<.018	<.018	<.1	<.4	<.1	<.02	<.08	<.04	<.08
292943098354404	--	<.032	<.1	<.06	.277	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	E.018	<.06
292943098354404	--	<.032	<.1	<.06	.306	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	E.020	<.06
292943098354404	--	<.032	<.1	<.06	.288	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	<.03	<.06
292943098354404	--	<.032	<.1	<.06	.307	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	E.016	<.06
292943098354404	--	<.032	<.1	<.06	.265	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	E.021	<.06
293252098380801	--	<.04	<.1	<.06	E.037	<.06	<1.4	<.018	<.018	<.1	<.4	<.1	<.02	<.08	E.066	<.08
293252098380801	--	<.032	<.1	<.06	E.025	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	E.087	<.06
293252098380801	--	<.032	<.1	<.06	E.024	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	.132	<.06
293252098380801	--	<.032	<.1	<.06	<.026	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	.104	<.06
293252098380801	--	<.032	<.1	<.06	E.018	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	.129	<.06
293252098380801	--	<.032	<.1	<.06	E.020	<.052	<1.4	<.018	<.018	<.14	<.36	<.1	<.022	<.08	.192	<.06

*Sample collected from top of open interval during depth-dependent sampling.

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Group and description	Sample date	1-Naph- thol (µg/L)	2,6-Di- ethyl- aniline (µg/L)	2-Chloro- 4-isopropyl- amino-6- triazine (CIAT) (µg/L)	2-Chloro- N-(2,6-di- ethylphenyl)- acetamide (alachlor 2nd amide) (µg/L)	2-Ethyl- 6-methyl- aniline (µg/L)	3,4-Di- chloro- aniline (µg/L)	3,5-Di- chloro- aniline (µg/L)	4-Chloro- 2-methyl- phenol (µg/L)	Aceto- chlor (µg/L)	Ala- chlor (µg/L)
Regional aquifer public-supply wells													
293359098290301	AY-68-29-414	1st (highest) quartile	2004/12/06	<0.088	<0.006	E0.011	<0.005	<0.005	<0.005	--	<0.006	<0.006	<0.005
293359098290301	AY-68-29-414	1st quartile	2005/09/01	<.088	<.006	E.010	<.005	<.005	<.005	<0.004	<.006	<.006	<.005
293358098231101	AY-68-29-610	1st quartile	2004/12/15	<.088	<.006	E.012	<.005	<.005	<.005	--	<.006	<.006	<.005
293358098231101	AY-68-29-610	1st quartile	2005/08/29	<.088	<.006	E.013	<.005	<.005	<.005	<.004	<.006	<.006	<.005
293120098285801	AY-68-29-714	1st quartile	2004/12/14	<.088	<.006	E.011	<.005	<.005	<.005	--	<.006	<.006	<.005
293120098285801	AY-68-29-714	1st quartile	2005/08/30	--	--	E.012	--	--	--	--	--	--	--
293145098224201	AY-68-29-929	1st quartile	2004/12/15	<.088	<.006	<.01	<.005	<.005	<.005	--	<.006	<.006	<.005
293119098211201	AY-68-30-718	1st quartile	2004/12/16	<.088	<.006	<.006	<.005	<.005	<.005	--	<.006	<.006	<.005
292843098425101	AY-68-35-106	1st quartile	2004/12/13	<.088	<.006	<.006	<.005	<.005	<.005	--	<.006	<.006	<.005
292459098382101	AY-68-35-913	1st quartile	2004/12/13	<.088	<.006	<.006	<.005	<.005	<.005	--	<.006	<.006	<.005
292925098360201	AY-68-36-104	1st quartile	2004/12/13	<.088	<.006	E.009	<.005	<.005	<.005	--	<.006	<.006	<.005
292925098360201	AY-68-36-104	1st quartile	2005/08/31	<.088	<.006	E.008	<.005	<.005	<.005	<.004	<.006	<.006	<.005
292822098325401	AY-68-36-206	1st quartile	2004/12/13	<.088	<.006	E.009	<.005	<.005	<.005	--	<.006	<.006	<.005
292822098325401	AY-68-36-206	1st quartile	2005/08/31	<.088	<.006	E.012	<.005	<.005	<.005	<.004	<.006	<.006	<.005
292944098292301	AY-68-37-105	1st quartile	2004/12/14	<.088	<.006	E.010	<.005	<.005	<.005	--	<.006	<.006	<.005
292944098292301	AY-68-37-105	1st quartile	2005/08/31	<.088	<.006	E.011	<.005	<.005	<.005	<.004	<.006	<.006	<.005
292522098291901	AY-68-37-426	1st quartile	2004/12/14	<.088	<.006	<.006	<.005	<.005	<.005	--	<.006	<.006	<.005
292643098241801	AY-68-37-601	1st quartile	2004/12/15	<.088	<.006	<.006	<.005	<.005	<.005	--	<.006	<.006	<.005
292643098241801	AY-68-37-601	1st quartile	2005/08/30	--	--	<.028	--	--	--	--	--	--	--
292328098294601	AY-68-37-705	1st quartile	2004/12/14	<.088	<.006	<.006	<.005	<.005	<.005	--	<.006	<.006	<.005
294225098080301	DX-68-23-601	1st quartile	2004/12/08	<.088	<.006	E.010	<.005	<.005	<.005	--	<.006	<.006	<.005
294225098080301	DX-68-23-601	1st quartile	2005/09/01	<.088	<.006	E.008	<.005	<.005	<.005	<.004	<.006	<.006	<.005
291210099475601	YP-69-50-506	1st quartile	2004/11/30	<.088	<.006	E.006	<.005	<.005	<.005	--	<.006	<.006	<.005
293111098340901	AY-68-28-807	2d quartile	2005/05/26	<.088	<.006	E.007	<.005	<.005	<.005	--	<.006	<.006	<.005
293512098291701	AY-68-29-109	2d quartile	2005/04/20	<.088	<.006	E.010	<.005	<.005	<.005	--	<.006	<.006	<.005
292424098421501	AY-68-35-810	2d quartile	2005/04/21	<.088	<.006	E.008	<.005	<.005	<.005	--	<.006	<.006	<.005

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Group and description	Sample date	1-Naph- thol (µg/L)	2,6-Di- ethyl- aniline (µg/L)	2-Chloro- 4-isopropyl- amino-6- triazine (CIAT) (µg/L)	2-Chloro- N-(2,6-di- ethylphenyl)- acetamide (alachlor 2nd amide) (µg/L)	2-Ethyl- 6-methyl- aniline (µg/L)	3,4-Di- chloro- aniline (µg/L)	3,5-Di- chloro- aniline (µg/L)	4-Chloro- 2-methyl- phenol (µg/L)	Aceto- chlor (µg/L)	Ala- chlor (µg/L)
292053098365501	AY-68-44-110	2d quartile	2005/05/26	<0.088	<0.006	E.004	<0.005	<0.005	<0.005	--	<0.006	<0.006	<0.005
293807098155301	DX-68-22-901	2d quartile	2005/05/18	<.088	<.006	E.005	<.005	<.005	<.005	--	<.006	<.006	<.005
292604098563201	TD-68-33-501	2d quartile	2005/05/16	<.088	<.006	<.006	<.005	<.005	<.005	--	<.006	<.006	<.005
292116099095501	TD-69-47-305	2d quartile	2005/04/27	<.088	<.006	E.006	<.005	<.005	<.005	--	<.006	<.006	<.005
291232099470301	YP-69-50-339	2d quartile	2005/05/11	<.088	<.006	E.005	<.005	<.005	--	--	<.006	<.006	<.005
										--			
293128098473101	AY-68-26-814	3d quartile	2005/04/18	<.088	<.006	<.006	<.005	<.005	<.005	--	<.006	<.006	<.005
293451098313201	AY-68-28-601	3d quartile	2005/05/17	<.088	<.006	E.015	<.005	<.005	E.003	--	<.006	<.006	<.005
294604098060801	DX-68-16-708	3d quartile	2005/04/25	<.088	<.006	<.006	<.005	<.005	<.005	--	<.006	<.006	<.005
294019098114701	DX-68-23-504	3d quartile	2005/05/18	<.088	<.006	E.007	<.005	<.005	<.005	--	<.006	<.006	<.005
292215098580201	TD-68-41-103	3d quartile	2005/05/09	<.088	<.006	E.006	<.005	<.005	--	--	<.006	<.006	<.005
292119098524901	TD-68-41-308	3d quartile	2005/04/27	<.088	<.006	E.005	<.005	<.005	<.005	--	<.006	<.006	<.005
291943099163301	TD-69-46-601	3d quartile	2005/05/09	<.088	<.006	<.006	<.005	<.005	--	--	<.006	<.006	<.005
291840099382601	YP-69-43-606	3d quartile	2005/05/11	<.088	<.006	E.006	<.005	<.005	--	--	<.006	<.006	<.005
293518098332601	AY-68-28-203	4th quartile	2005/04/19	<.088	<.006	E.005	<.005	<.005	<.005	--	<.006	<.006	<.005
293023098355401	AY-68-28-702	4th quartile	2005/04/26	<.088	<.006	E.011	<.005	<.005	<.005	--	<.006	<.006	<.005
293042098305201	AY-68-28-913	4th quartile	2005/06/30	<.088	<.006	E.010	<.005	<.005	<.005	<.004	<.006	<.006	<.005
292442098474501	AY-68-34-803	4th quartile	2005/04/20	<.088	<.006	E.013	<.005	<.005	<.005	--	<.006	<.006	<.005
292405098371201	AY-68-36-704	4th quartile	2005/05/23	<.088	<.006	E.006	<.005	<.005	<.005	--	<.006	<.006	<.005
292931098274601	AY-68-37-124	4th quartile	2005/05/25	<.088	<.006	E.009	<.005	<.005	<.005	--	<.006	<.006	<.005
293729098173101	DX-68-30-215	4th quartile	2005/04/25	<.088	<.006	E.006	<.005	<.005	<.005	--	<.006	<.006	<.005
291219099095601	TD-69-55-604	4th quartile	2005/05/16	<.088	<.006	E.006	<.005	<.005	<.005	--	<.006	<.006	<.005
Monitoring wells													
292943098354401	AY-68-36-135	MW OVB	2007/10/30	<.04	<.006	E.010	<.01	<.010	<.006	<.008	<.005	<.006	<.006
292943098354402	AY-68-36-133	TANC MW	2007/10/31	<.04	<.006	E.011	<.01	<.010	<.006	<.008	<.005	<.0100	<.006
292943098354403	AY-68-36-134	TANC MW	2007/10/30	<.04	<.006	E.013	<.01	<.010	<.006	<.008	<.005	<.006	<.006
292943098354404	AY-68-36-132	TANC MW	2007/11/01	<.04	<.006	E.012	<.01	<.010	<.006	<.008	<.005	<.006	<.006

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

[illegible]

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Group and description	Sample date	1-Naph- thol (µg/L)	2,6-Di- ethyl- aniline (µg/L)	2-Chloro- 4-isopropyl- amino-6- triazine (CIAT) (µg/L)	2-Chloro- N-(2,6-di- ethylphenyl)- acetamide (alachlor 2nd amide) (µg/L)	2-Ethyl- 6-methyl- aniline (µg/L)	3,4-Di- chloro- aniline (µg/L)	3,5-Di- chloro- aniline (µg/L)	4-Chloro- 2-methyl- phenol (µg/L)	Aceto- chlor (µg/L)	Ala- chlor (µg/L)
292923098360301	AY-68-36-131	TANC PSW storm 3	2009/10/19	<0.036	<0.006	E.015	<0.010	<0.010	<0.004	<0.003	<0.003	<0.010	<0.008
292923098360301	AY-68-36-131	TANC PSW storm 4	2009/10/30	<.036	<.006	E.016	<.010	<.010	<.004	<.003	<.003	<.010	<.008
292923098360301	AY-68-36-131	TANC PSW storm 5	2009/11/19	<.036	<.006	E.015	<.010	<.010	<.004	<.003	<.003	<.010	<.008
292943098354402	AY-68-36-133	TANC MW pre-storm	2009/04/08	--	--	--	--	--	--	--	--	--	--
292943098354402	AY-68-36-133	TANC MW storm 1	2009/10/05	--	--	--	--	--	--	--	--	--	--
292943098354402	AY-68-36-133	TANC MW storm 2	2009/10/07	--	--	--	--	--	--	--	--	--	--
292943098354402	AY-68-36-133	TANC MW storm 3	2009/10/17	<.036	<.006	E.015	<.010	<.010	<.004	<.003	<.003	<.010	<.008
292943098354402	AY-68-36-133	TANC MW storm 4	2009/10/28	<.036	<.006	E.015	<.010	<.010	<.004	<.003	<.003	<.010	<.008
292943098354402	AY-68-36-133	TANC MW storm 5	2009/11/18	<.036	<.006	E.015	<.010	<.010	<.004	<.003	<.003	<.010	<.008
292943098354403	AY-68-36-134	TANC MW pre-storm	2009/04/09	--	--	--	--	--	--	--	--	--	--
292943098354403	AY-68-36-134	TANC MW storm 1	2009/10/05	--	--	--	--	--	--	--	--	--	--
292943098354403	AY-68-36-134	TANC MW storm 2	2009/10/08	--	--	--	--	--	--	--	--	--	--
292943098354403	AY-68-36-134	TANC MW storm 3	2009/10/18	<.036	<.006	E.016	<.010	<.010	<.004	<.003	<.003	<.010	<.008
292943098354403	AY-68-36-134	TANC MW storm 4	2009/10/29	<.036	<.006	E.019	<.010	<.010	<.004	<.003	<.003	<.010	<.008
292943098354403	AY-68-36-134	TANC MW storm 5	2009/11/18	<.036	<.006	E.016	<.010	<.010	<.004	<.003	<.003	<.010	<.008
292943098354404	AY-68-36-132	TANC MW pre-storm	2009/04/09	--	--	--	--	--	--	--	--	--	--
292943098354404	AY-68-36-132	TANC MW storm 1	2009/10/05	--	--	--	--	--	--	--	--	--	--
292943098354404	AY-68-36-132	TANC MW storm 2	2009/10/08	--	--	--	--	--	--	--	--	--	--
292943098354404	AY-68-36-132	TANC MW storm 3	2009/10/18	<.036	<.006	E.017	<.010	<.010	<.004	<.003	<.003	<.010	<.008
292943098354404	AY-68-36-132	TANC MW storm 4	2009/10/29	<.036	<.006	E.018	<.010	<.010	<.004	<.003	<.003	<.010	<.008
292943098354404	AY-68-36-132	TANC MW storm 5	2009/11/18	<.036	<.006	E.020	<.010	<.010	<.004	<.003	<.003	<.010	<.008
293252098380801	AY-68-27-610	TANC MW pre-storm	2009/04/14	--	--	--	--	--	--	--	--	--	--
293252098380801	AY-68-27-610	TANC MW storm 1	2009/10/04	--	--	--	--	--	--	--	--	--	--
293252098380801	AY-68-27-610	TANC MW storm 2	2009/10/07	--	--	--	--	--	--	--	--	--	--
293252098380801	AY-68-27-610	TANC MW storm 3	2009/10/17	<.036	<.006	E.028	<.010	<.010	<.004	<.003	<.003	<.010	<.008
293252098380801	AY-68-27-610	TANC MW storm 4	2009/10/28	<.036	<.006	E.032	<.010	<.010	<.004	<.003	<.003	<.010	<.008
293252098380801	AY-68-27-610	TANC MW storm 5	2009/11/17	<.036	<.006	E.036	<.010	<.010	<.004	<.003	<.003	<.010	<.008

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	alpha- Endo- sulfan (µg/L)	Atra- zine (µg/L)	Azin- phos- methyl oxygen analog (µg/L)	Azin- phos- methyl (µg/L)	Ben- fluralin (µg/L)	Carb- aryl (µg/L)	Carbo- furan (µg/L)	Chlor- pyrifos (µg/L)	cis- Perm- ethrin (µg/L)	cis- Pro- picon- azole (µg/L)	Cyana- zine (µg/L)	Cyflu- thrin (µg/L)	lambda- Cyhalo- thrin (µg/L)	Cy- perm- ethrin (µg/L)	Dacthal (DCPA) (µg/L)	Chlor- pyrifos oxygen analog (µg/L)	Desulf- inyl- fipronil amide (µg/L)
Regional aquifer public-supply wells																	
293359098290301	--	0.01	<0.070	<0.05	<0.010	<0.041	--	<0.005	<0.006	--	--	<0.008	--	<0.009	<0.003	<0.056	<0.029
293359098290301	<0.005	.008	<.070	<.05	<.01	<.041	<.02	<.005	<.006	<0.008	<0.018	<.027	<0.009	<.009	<.003	<.056	<.029
293358098231101	--	.014	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.008	--	<.009	<.003	<.056	<.029
293358098231101	<.005	.011	<.070	<.05	<.01	<.041	<.02	<.005	<.006	<.008	<.018	<.027	<.009	<.009	<.003	<.056	<.029
293120098285801	--	<.01	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.008	--	<.009	<.003	<.056	<.029
293120098285801	--	E.005	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293145098224201	--	<.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.008	--	<.009	<.003	<.056	<.029
293119098211201	--	<.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.008	--	<.009	<.003	<.056	<.029
292843098425101	--	<.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.008	--	<.009	<.003	<.056	<.029
292459098382101	--	<.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.008	--	<.009	<.003	<.056	<.029
292925098360201	--	.008	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.008	--	<.009	<.003	<.056	<.029
292925098360201	<.005	E.007	<.070	<.05	<.01	<.041	<.02	<.005	<.006	<.008	<.018	<.027	<.009	<.009	<.003	<.056	<.029
292822098325401	--	.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.008	--	<.009	<.003	<.056	<.029
292822098325401	<.005	E.007	<.070	<.05	<.01	<.041	<.02	<.005	<.006	<.008	<.018	<.027	<.009	<.009	<.003	<.056	<.029
292944098292301	--	.008	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.008	--	<.009	<.003	<.056	<.029
292944098292301	<.005	E.007	<.070	<.05	<.01	<.041	<.02	<.005	<.006	<.008	<.018	<.027	<.009	<.009	<.003	<.056	<.029
292522098291901	--	<.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.008	--	<.009	<.003	<.056	<.029
292643098241801	--	<.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.008	--	<.009	<.003	<.056	<.029
292643098241801	--	<.008	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292328098294601	--	<.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.008	--	<.009	<.003	<.056	<.029
294225098080301	--	.008	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.008	--	<.009	<.003	<.056	<.029
294225098080301	<.005	E.006	<.070	<.05	<.01	<.041	<.02	<.005	<.006	<.008	<.018	<.027	<.009	<.009	<.003	<.056	<.029
291210099475601	--	E.004	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.008	--	<.009	<.003	<.056	<.029
293111098340901	--	.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
293512098291701	--	.009	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
292424098421501	--	.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	alpha- Endo- sulfan (µg/L)	Atra- zine (µg/L)	Azin- phos- methyl oxygen analog (µg/L)	Azin- phos- methyl (µg/L)	Ben- fluralin (µg/L)	Carb- aryl (µg/L)	Carbo- furan (µg/L)	Chlor- pyrifos (µg/L)	cis- Perm- ethrin (µg/L)	cis- Pro- picon- azole (µg/L)	Cyana- zine (µg/L)	Cyflu- thrin (µg/L)	lambda- Cyhalo- thrin (µg/L)	Cy- perm- ethrin (µg/L)	Dacthal (DCPA) (µg/L)	Chlor- pyrifos oxygen analog (µg/L)	Desulf- inyl- fipronil amide (µg/L)
292053098365501	--	<0.007	<0.070	<0.05	<0.010	<0.041	--	<0.005	<0.006	--	--	<0.027	--	<0.009	<0.003	<0.056	<0.029
293807098155301	--	E.004	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
292604098563201	--	<.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
292116099095501	--	<.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
291232099470301	--	E.004	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
	--						--			--	--						
293128098473101	--	<.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
293451098313201	--	.018	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
294604098060801	--	<.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
294019098114701	--	E.005	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
292215098580201	--	<.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
292119098524901	--	<.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
291943099163301	--	<.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
291840099382601	--	E.004	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
293518098332601	--	E.006	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
293023098355401	--	.012	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
293042098305201	<0.005	E.007	<.070	<.05	<.01	<.041	<0.02	<.005	<.006	<0.008	<0.018	<.027	<0.009	<.009	<.003	<.056	<.029
292442098474501	--	E.005	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
292405098371201	--	<.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
292931098274601	--	.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
293729098173101	--	E.005	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
291219099095601	--	<.007	<.070	<.05	<.010	<.041	--	<.005	<.006	--	--	<.027	--	<.009	<.003	<.056	<.029
Monitoring wells																	
292943098354401	<.006	.009	<.042	<.12	<.0059	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
292943098354402	<.006	.009	<.042	<.12	<.004	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
292943098354403	<.006	.012	<.042	<.12	<.004	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
292943098354404	<.006	<.008	<.042	<.12	<.004	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
292851098374401	<0.006	0.008	<0.06	<0.12	<0.01	<0.06	<0.02	<0.005	<0.01	<0.006	<0.02	<0.016	<0.004	<0.014	<0.003	<0.056	<0.029

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	alpha- Endo- sulfan (µg/L)	Atra- zine (µg/L)	Azin- phos- methyl oxygen analog (µg/L)	Azin- phos- methyl (µg/L)	Ben- fluralin (µg/L)	Carb- aryl (µg/L)	Carbo- furan (µg/L)	Chlor- pyrifos (µg/L)	cis- Perm- ethrin (µg/L)	cis- Pro- picon- azole (µg/L)	Cyana- zine (µg/L)	Cyflu- thrin (µg/L)	lambda- Cyhalo- thrin (µg/L)	Cy- perm- ethrin (µg/L)	Dacthal (DCPA) (µg/L)	Chlor- pyrifos oxygen analog (µg/L)	Desulf- inyl- fipronil amide (µg/L)
292851098374402	<.006	E.014	<.042	<.12	<.01	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
Well-field depth-dependent sampling (W4)																	
292916098360701	<.006	E.008	<.042	<.12	<.004	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
292916098360701	<.006	.011	<.042	<.12	<.004	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
292916098360701	<.006	u	<.042	<.12	<.004	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
292916098360701	<.006	.009	<.042	<.12	<.004	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
292916098360701	<.006	E.011	<.042	<.12	<.004	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
292916098360701	<.006	.015	<.042	<.12	<.004	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
292916098360701	<.006	.016	<.042	<.12	<.004	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
292916098360701	<.006	.012	<.042	<.12	<.004	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
292916098360701	<.006	.011	<.042	<.12	<.004	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
Well-field wells (wellhead sampling)																	
292919098360501	<.006	.017	<.042	<.12	<.004	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
292925098360201	<.006	.009	<.042	<.12	<.004	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
292916098360701	<.006	.017	<.042	<.12	<.004	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
292920098360601	<.006	.015	<.042	<.12	<.004	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
292923098360301	<.006	.015	<.042	<.12	<.0058	<.06	<.02	<.005	<.01	<.006	<.02	<.016	<.004	<.014	<.003	<.056	<.029
Temporal sampling																	
292923098360301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292923098360301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292923098360301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292923098360301	<0.006	0.008	<0.042	<0.12	<0.014	<0.06	<0.060	<0.010	<0.014	<0.006	<0.022	<0.016	<0.010	<0.020	<0.008	<0.05	<0.029
292923098360301	<.006	E.006	<.042	<.12	<.014	<.06	<.060	<.010	<.014	<.006	<.022	<.016	<.010	<.020	<.008	<.05	<.029

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	alpha- Endo- sulfan (µg/L)	Atra- zine (µg/L)	Azin- phos- methyl oxygen analog (µg/L)	Azin- phos- methyl (µg/L)	Ben- fluralin (µg/L)	Carb- aryl (µg/L)	Carbo- furan (µg/L)	Chlor- pyrifos (µg/L)	cis- Perm- ethrin (µg/L)	cis- Pro- picon- azole (µg/L)	Cyana- zine (µg/L)	Cyflu- thrin (µg/L)	lambda- Cyhalo- thrin (µg/L)	Cy- perm- ethrin (µg/L)	Dacthal (DCPA) (µg/L)	Chlor- pyrifos oxygen analog (µg/L)	Desulf- inyl- fipronil amide (µg/L)
292923098360301	<.006	E.008	<.042	<.12	<.014	<.06	<.060	<.010	<.014	<.006	<.022	<.016	<.010	<.020	<.008	u	<.029
292943098354402	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354402	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354402	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354402	<.006	.009	<.042	<.12	<.014	<.06	<.060	<.010	<.014	<.006	<.022	<.016	<.010	<.020	<.008	<.05	<.029
292943098354402	<.006	.009	<.042	<.12	<.014	<.06	<.060	<.010	<.014	<.006	<.022	<.016	<.010	<.020	<.008	<.05	<.029
292943098354402	<.006	.008	<.042	<.12	<.014	<.06	<.060	<.010	<.014	<.006	<.022	<.016	<.010	<.020	<.008	<.05	<.029
292943098354403	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354403	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354403	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354403	<.006	.009	<.042	<.12	<.014	<.06	<.060	<.010	<.014	<.006	<.022	<.016	<.010	<.020	<.008	<.05	<.029
292943098354403	<.006	.009	<.042	<.12	<.014	<.06	<.060	<.010	<.014	<.006	<.022	<.016	<.010	<.020	<.008	<.05	<.029
292943098354403	<.006	.008	<.042	<.12	<.014	<.06	<.060	<.010	<.014	<.006	<.022	<.016	<.010	<.020	<.008	u	<.029
292943098354404	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354404	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354404	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354404	<.006	.009	<.042	<.12	<.014	<.06	<.060	<.010	<.014	<.006	<.022	<.016	<.010	<.020	<.008	<.05	<.029
292943098354404	<.006	.009	<.042	<.12	<.014	<.06	<.060	<.010	<.014	<.006	<.022	<.016	<.010	<.020	<.008	u	<.029
292943098354404	<.006	.009	<.042	<.12	<.014	<.06	<.060	<.010	<.014	<.006	<.022	<.016	<.010	<.020	<.008	u	<.029
293252098380801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293252098380801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293252098380801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293252098380801	<.006	.017	<.042	<.12	<.014	<.06	<.060	<.010	<.014	<.006	<.022	<.016	<.010	<.020	<.008	<.05	<.029
293252098380801	<.006	.015	<.042	<.12	<.014	<.06	<.060	<.010	<.014	<.006	<.022	<.016	<.010	<.020	<.008	u	<.029
293252098380801	<.006	.017	<.042	<.12	<.014	<.06	<.060	<.010	<.014	<.006	<.022	<.016	<.010	<.020	<.008	<.05	<.029

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	Diaz- inon (µg/L)	Diaz- oxon (µg/L)	Di- chlor- vos (µg/L)	Di- croto- phos (µg/L)	Diel- drin (µg/L)	Di- meth- oate (µg/L)	Di- sulf- oton sulfone (µg/L)	Di- sulf- oton (µg/L)	Endo- sulfan sulfate (µg/L)	EPTC (µg/L)	Ethion mon- oxon (µg/L)	Ethion (µg/L)	Etho- prop (µg/L)	Fen- amiphos sulfone (µg/L)	Fen- ami- phos sulf- oxide (µg/L)	Fen- ami- phos (µg/L)	Desulf- inyl- fipronil (µg/L)
Regional aquifer public-supply wells																	
293359098290301	<0.005	<0.006	<0.012	<0.084	<0.009	<0.006	--	--	--	--	<0.002	<0.004	--	<0.049	<0.039	<0.029	<0.012
293359098290301	<.005	<.006	<.012	<.084	<.009	<.006	<0.006	<0.021	<0.014	<0.004	<.002	<.004	<0.005	<.049	<.039	<.029	<.012
293358098231101	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	u	<.029	<.012
293358098231101	<.005	<.006	<.012	<.084	<.009	<.006	<.006	<.021	<.014	<.004	<.002	<.004	<0.005	<.049	<.039	<.029	<.012
293120098285801	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	u	<.029	<.012
293120098285801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293145098224201	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	u	<.029	<.012
293119098211201	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	u	<.029	<.012
292843098425101	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
292459098382101	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	u	<.029	<.012
292925098360201	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
292925098360201	<.005	<.006	<.012	<.084	<.009	<.006	<.006	<.021	<.014	<.004	<.002	<.004	<0.005	<.049	<.039	<.029	<.012
292822098325401	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
292822098325401	<.005	<.006	<.012	<.084	<.009	<.006	<.006	<.021	<.014	<.004	<.002	<.004	<0.005	<.049	<.039	<.029	<.012
292944098292301	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	u	<.029	<.012
292944098292301	<.005	<.006	<.012	<.084	<.009	<.006	<.006	<.021	<.014	<.004	<.002	<.004	<0.005	<.049	<.039	<.029	<.012
292522098291901	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	u	<.029	<.012
292643098241801	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	u	<.029	<.012
292643098241801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292328098294601	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	u	<.029	<.012
294225098080301	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
294225098080301	<.005	<.006	<.012	<.084	<.009	<.006	<.006	<.021	<.014	<.004	<.002	<.004	<0.005	<.049	<.039	<.029	<.012
291210099475601	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
293111098340901	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
293512098291701	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
292424098421501	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	Diaz- inon (µg/L)	Diaz- oxon (µg/L)	Di- chlor- vos (µg/L)	Di- croto- phos (µg/L)	Diel- drin (µg/L)	Di- meth- oate (µg/L)	Di- sulf- oton sulfone (µg/L)	Di- sulf- oton (µg/L)	Endo- sulfan sulfate (µg/L)	EPTC (µg/L)	Ethion mon- oxon (µg/L)	Ethion (µg/L)	Etho- prop (µg/L)	Fen- amiphos sulfone (µg/L)	Fen- ami- phos sulf- oxide (µg/L)	Fen- ami- phos (µg/L)	Desulf- inyl- fipronil (µg/L)
292053098365501	<0.005	<0.006	<0.012	<0.084	<0.009	<0.006	--	--	--	--	<0.002	<0.004	--	<0.049	<0.039	<0.029	<0.012
293807098155301	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
292604098563201	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
292116099095501	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
291232099470301	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
293128098473101	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
293451098313201	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
294604098060801	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
294019098114701	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
292215098580201	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
292119098524901	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
291943099163301	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
291840099382601	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
293518098332601	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
293023098355401	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
293042098305201	<.005	<.006	<.012	<.084	<.009	<.006	<0.006	<0.021	<0.014	<0.004	<.002	<.004	<0.005	<.049	<.039	<.029	<.012
292442098474501	<.005	<.006	<.012	<.084	E.003	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
292405098371201	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
292931098274601	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
293729098173101	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
291219099095601	<.005	<.006	<.012	<.084	<.009	<.006	--	--	--	--	<.002	<.004	--	<.049	<.039	<.029	<.012
Monitoring wells																	
292943098354401	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.06	<.029	<.012
292943098354402	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.060	<.022	<.002	<.021	<.006	<.012	<.053	<.040	<.029	<.012
292943098354403	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.06	<.029	<.012
292943098354404	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.040	<.029	<.012
292851098374401	<0.005	<0.006	<0.013	<0.084	<0.009	<0.006	<0.014	<0.04	<0.022	<0.002	<0.021	<0.006	<0.012	<0.053	<0.2	<0.029	<0.012

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	Diaz- inon (µg/L)	Diaz- oxon (µg/L)	Di- chlor- vos (µg/L)	Di- croto- phos (µg/L)	Diel- drin (µg/L)	Di- meth- oate (µg/L)	Di- sulf- oton sulfone (µg/L)	Di- sulf- oton (µg/L)	Endo- sulfan sulfate (µg/L)	EPTC (µg/L)	Ethion mon- oxon (µg/L)	Ethion (µg/L)	Etho- prop (µg/L)	Fen- amiphos sulfone (µg/L)	Fen- ami- phos sulf- oxide (µg/L)	Fen- ami- phos (µg/L)	Desulf- inyl- fipronil (µg/L)
292851098374402	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.2	<.029	<.012
Well-field depth-dependent sampling (W4)																	
292916098360701	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.040	<.029	<.012
292916098360701	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.040	<.029	<.012
292916098360701	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.040	<.029	<.012
292916098360701	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.040	<.029	<.012
292916098360701	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.040	<.029	<.012
292916098360701	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.040	<.029	<.012
292916098360701	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.06	<.029	<.012
292916098360701	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.040	<.029	<.012
292916098360701	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.040	<.029	<.012
Well-field wells (wellhead sampling)																	
292919098360501	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.040	<.029	<.012
292925098360201	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.040	<.029	<.012
292916098360701	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.040	<.029	<.012
292920098360601	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.040	<.029	<.012
292923098360301	<.005	<.006	<.013	<.084	<.009	<.006	<.014	<.04	<.022	<.002	<.021	<.006	<.012	<.053	<.06	<.029	<.012
Temporal sampling																	
292923098360301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292923098360301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292923098360301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292923098360301	<0.005	--	<0.02	<0.08	<0.009	<0.006	<0.014	<0.04	<0.014	<0.002	<0.021	<0.008	<0.016	<0.053	<0.08	<0.030	<0.012

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	Diaz- inon (µg/L)	Diaz- oxon (µg/L)	Di- chlor- vos (µg/L)	Di- cro- phos (µg/L)	Diel- drin (µg/L)	Di- meth- oate (µg/L)	Di- sulf- oton sulfone (µg/L)	Di- sulf- oton (µg/L)	Endo- sulfan sulfate (µg/L)	EPTC (µg/L)	Ethion mon- oxon (µg/L)	Ethion (µg/L)	Etho- prop (µg/L)	Fen- amiphos sulfone (µg/L)	Fen- ami- phos sulf- oxide (µg/L)	Fen- ami- phos (µg/L)	Desulf- inyl- fipronil (µg/L)
292923098360301	<.005	--	<.02	<.08	<.009	<.006	<.014	<.04	<.014	<.002	<.021	<.008	<.016	<.053	<.08	<.030	<.012
292923098360301	<.005	--	<.02	<.08	<.009	<.006	<.014	<.04	<.014	<.002	<.021	<.008	<.016	<.053	<.08	<.030	<.012
292943098354402	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354402	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354402	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354402	<.005	--	<.02	<.08	<.009	<.006	<.014	<.04	<.014	<.002	<.021	<.008	<.016	<.053	<.08	<.030	<.012
292943098354402	<.005	--	<.02	<.08	<.009	<.006	<.014	<.04	<.014	<.002	<.021	<.008	<.016	<.053	<.08	<.030	<.012
292943098354402	<.005	--	<.02	<.08	<.009	<.006	<.014	<.04	<.014	<.002	<.021	<.008	<.016	<.053	<.08	<.030	<.012
292943098354403	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354403	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354403	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354403	<.005	--	<.02	<.08	<.009	<.006	<.014	<.04	<.014	<.002	<.021	<.008	<.016	<.053	<.08	<.030	<.012
292943098354403	<.005	--	<.02	<.08	<.009	<.006	<.014	<.04	<.014	<.002	<.021	<.008	<.016	<.053	<.08	<.030	<.012
292943098354403	<.005	--	<.02	<.08	<.009	<.006	<.014	<.04	<.014	<.002	<.021	<.008	<.016	<.053	<.08	<.030	<.012
292943098354404	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354404	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354404	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354404	<.005	--	<.02	<.08	<.009	<.006	<.014	<.04	<.014	<.002	<.021	<.008	<.016	<.053	<.08	<.030	<.012
292943098354404	<.005	--	<.02	<.08	<.009	<.006	<.014	<.04	<.014	<.002	<.021	<.008	<.016	<.053	<.08	<.030	<.012
292943098354404	<.005	--	<.02	<.08	<.009	<.006	<.014	<.04	<.014	<.002	<.021	<.008	<.016	<.053	<.08	<.030	<.012
293252098380801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293252098380801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293252098380801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293252098380801	<.005	--	<.02	<.08	<.009	<.006	<.014	<.04	<.014	<.002	<.021	<.008	<.016	<.053	<.08	<.030	<.012
293252098380801	<.005	--	<.02	<.08	<.009	<.006	<.014	<.04	<.014	<.002	<.021	<.008	<.016	<.053	<.08	<.030	<.012
293252098380801	<.005	--	<.02	<.08	<.009	<.006	<.014	<.04	<.014	<.002	<.021	<.008	<.016	<.053	<.08	<.030	<.012

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	Fip- ronil sulfide (µg/L)	Fip- ronil sulfone (µg/L)	Fip- ronil (µg/L)	Fonofos oxygen analog (µg/L)	Fon- ofos (µg/L)	Hexa- zinone (µg/L)	Ipro- dione (µg/L)	Isofen- phos (µg/L)	Mala- oxon (µg/L)	Mala- thion (µg/L)	Meta- laxyl (µg/L)	Meth- ida- thion (µg/L)	Methyl para- oxon (µg/L)	Methyl para- thion (µg/L)	Meto- la- chlor (µg/L)	Metri- buzin (µg/L)	Mol- inate (µg/L)	Myclo- but- anil (µg/L)
Regional aquifer public-supply wells																		
293359098290301	<0.013	<0.024	<0.016	<0.003	<0.003	<0.013	<0.387	<0.003	<0.030	<0.027	<0.005	<0.006	<0.030	<0.015	<0.006	<0.006	--	<0.008
293359098290301	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	<0.003	<.008
293358098231101	<.013	<.024	<.016	<.003	<.003	<.013	<.387	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
293358098231101	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	<.003	<.008
293120098285801	<.013	<.024	<.016	<.003	<.003	<.013	<.387	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
293120098285801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293145098224201	<.013	<.024	<.016	<.003	<.003	<.013	<.387	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
293119098211201	<.013	<.024	<.016	<.003	<.003	<.013	<.387	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
292843098425101	<.013	<.024	<.016	<.003	<.003	<.013	<.387	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
292459098382101	<.013	<.024	<.016	<.003	<.003	<.013	<.387	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
292925098360201	<.013	<.024	<.016	<.003	<.003	<.013	<.387	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
292925098360201	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	<.003	<.008
292822098325401	<.013	<.024	<.016	<.003	<.003	<.013	<.387	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
292822098325401	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	<.003	<.008
292944098292301	<.013	<.024	<.016	<.003	<.003	<.013	<.387	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
292944098292301	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	<.003	<.008
292522098291901	<.013	<.024	<.016	<.003	<.003	<.013	<.387	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
292643098241801	<.013	<.024	<.016	<.003	<.003	<.013	<.387	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
292643098241801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292328098294601	<.013	<.024	<.016	<.003	<.003	<.013	<.387	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
294225098080301	<.013	<.024	<.016	<.003	<.003	<.013	<.387	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
294225098080301	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	<.003	<.008
291210099475601	<.013	<.024	<.016	<.003	<.003	<.013	<.387	<.003	<.030	<.027	<.005	<.006	<.030	<.015	E.003	<.006	--	<.008
293111098340901	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
293512098291701	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
292424098421501	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	Fip- ronil sulfide (µg/L)	Fip- ronil sulfone (µg/L)	Fip- ronil (µg/L)	Fonofos oxygen analog (µg/L)	Fon- ofos (µg/L)	Hexa- zinone (µg/L)	Ipro- dione (µg/L)	Isofen- phos (µg/L)	Mala- oxon (µg/L)	Mala- thion (µg/L)	Meta- laxyl (µg/L)	Meth- ida- thion (µg/L)	Methyl para- oxon (µg/L)	Methyl para- thion (µg/L)	Meto- la- chlor (µg/L)	Metri- buzin (µg/L)	Mol- inate (µg/L)	Myclo- but- anil (µg/L)
292053098365501	<0.013	<0.024	<0.016	--	<0.003	<0.013	<0.538	<0.003	<0.030	<0.027	<0.005	<0.006	<0.030	<0.015	<0.006	<0.006	--	<0.008
293807098155301	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
292604098563201	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
292116099095501	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
291232099470301	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
293128098473101	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
293451098313201	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
294604098060801	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
294019098114701	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
292215098580201	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
292119098524901	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
291943099163301	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
291840099382601	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
293518098332601	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
293023098355401	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
293042098305201	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	<0.003	<.008
292442098474501	E.008	E.009	E.008	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
292405098371201	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
292931098274601	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
293729098173101	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
291219099095601	<.013	<.024	<.016	--	<.003	<.013	<.538	<.003	<.030	<.027	<.005	<.006	<.030	<.015	<.006	<.006	--	<.008
Monitoring wells																		
292943098354401	<.013	<.024	<.02	--	<.01	<.008	<.01	<.006	<.02	<.016	<.007	<.004	<.01	<.008	<.010	<.012	<.003	<.01
292943098354402	<.013	<.024	<.02	--	<.01	<.008	<.01	<.006	<.02	<.016	<.007	<.004	<.01	<.008	<.010	<.012	<.003	<.01
292943098354403	<.013	<.024	<.02	--	<.01	<.008	<.01	<.006	<.02	<.016	<.007	<.004	<.01	<.008	<.010	<.012	<.003	<.01
292943098354404	<.013	<.024	<.02	--	<.01	<.008	<.01	<.006	<.02	<.016	<.007	<.004	<.01	<.008	<.010	<.012	<.003	<.01

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

[illegible]

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	Fip- ronil sulfide (µg/L)	Fip- ronil sulfone (µg/L)	Fip- ronil (µg/L)	Fonofos oxygen analog (µg/L)	Fon- ofos (µg/L)	Hexa- zinone (µg/L)	Ipro- dione (µg/L)	Isofen- phos (µg/L)	Mala- oxon (µg/L)	Mala- thion (µg/L)	Meta- laxyl (µg/L)	Meth- ida- thion (µg/L)	Methyl para- oxon (µg/L)	Methyl para- thion (µg/L)	Meto- la- chlor (µg/L)	Metri- buzin (µg/L)	Mol- inate (µg/L)	Myclo- but- anil (µg/L)
292923098360301	<0.013	<0.024	<0.018	--	<0.004	<0.008	<0.014	<0.006	<0.08	<0.016	<0.007	<0.006	<0.010	<0.008	<0.014	<0.012	<0.003	<0.010
292923098360301	<.013	<.024	<.018	--	<.004	<.008	<.014	<.006	<.08	<.016	<.007	<.006	<.010	<.008	<.014	<.012	<.003	<.010
292923098360301	<.013	<.024	<.018	--	<.004	<.008	<.014	<.006	<.08	<.016	<.007	<.006	<.010	<.008	<.014	<.012	<.003	<.010
292943098354402	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354402	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354402	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354402	<.013	<.024	<.018	--	<.004	<.008	<.014	<.006	<.08	<.016	<.007	<.006	<.010	<.008	<.014	<.012	<.003	<.010
292943098354402	<.013	<.024	<.018	--	<.004	<.008	<.014	<.006	<.08	<.016	<.007	<.006	<.010	<.008	<.014	<.012	<.003	<.010
292943098354402	<.013	<.024	<.018	--	<.004	<.008	<.014	<.006	<.08	<.016	<.007	<.006	<.010	<.008	<.014	<.012	<.003	<.010
292943098354403	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354403	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354403	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354403	<.013	<.024	<.018	--	<.004	<.008	<.014	<.006	<.08	<.016	<.007	<.006	<.010	<.008	<.014	<.012	<.003	<.010
292943098354403	<.013	<.024	<.018	--	<.004	<.008	<.014	<.006	<.08	<.016	<.007	<.006	<.010	<.008	<.014	<.012	<.003	<.010
292943098354403	<.013	<.024	<.018	--	<.004	<.008	<.014	<.006	<.08	<.016	<.007	<.006	<.010	<.008	<.014	<.012	<.003	<.010
292943098354404	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354404	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354404	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354404	<.013	<.024	<.018	--	<.004	<.008	<.014	<.006	<.08	<.016	<.007	<.006	<.010	<.008	<.014	<.012	<.003	<.010
292943098354404	<.013	<.024	<.018	--	<.004	<.008	<.014	<.006	<.08	<.016	<.007	<.006	<.010	<.008	<.014	<.012	<.003	<.010
292943098354404	<.013	<.024	<.018	--	<.004	<.008	<.014	<.006	<.08	<.016	<.007	<.006	<.010	<.008	<.014	<.012	<.003	<.010
293252098380801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293252098380801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293252098380801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293252098380801	<.013	<.024	<.018	--	<.004	<.008	<.014	<.006	<.08	<.016	<.007	<.006	<.010	<.008	<.014	<.012	<.003	<.010
293252098380801	<.013	<.024	<.018	--	<.004	<.008	<.014	<.006	<.08	<.016	<.007	<.006	<.010	<.008	<.014	<.012	<.003	<.010
293252098380801	<.013	<.024	<.018	--	<.004	<.008	<.014	<.006	<.08	<.016	<.007	<.006	<.010	<.008	<.014	<.012	<.003	<.010

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	Oxy- fluor- fen (µg/L)	Pendi- meth- alin (µg/L)	Phor- ate oxy- gen analog (µg/L)	Phor- ate (µg/L)	Phos- met oxy- gen analog (µg/L)	Phos- met (µg/L)	Pro- meton (µg/L)	Pro- metryn (µg/L)	Pro- pyz- amide (µg/L)	Prop- anil (µg/L)	Prop- argite (µg/L)	Sim- azine (µg/L)	Tebu- con- azole (µg/L)	Tebu- thi- uron (µg/L)	Teflu- thrin (µg/L)	Terb- ufos oxy- gen analog sul- fone (µg/L)	Terb- ufos (µg/L)	Ter- buthyl- azine (µg/L)	Thio- ben- carb (µg/L)	trans- Propi- con- azole (µg/L)	Tri- bu- phos (µg/L)	Tri- flur- alin (µg/L)
Regional aquifer public-supply wells																						
293359098290301	--	<0.022	<0.105	<0.011	<0.051	<0.008	0.013	<0.005	<0.004	--	--	0.008	--	<0.016	--	<0.068	<0.017	<0.010	--	--	--	<0.009
293359098290301	<0.007	<.022	<.105	<.011	<.051	<.008	.011	<.005	<.004	<0.011	<0.023	.009	<0.014	<.016	<0.008	<.068	<.017	<.010	<0.010	<0.013	<0.004	<.009
293358098231101	--	<.022	<.105	<.011	<.051	<.008	E.009	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
293358098231101	<.007	<.022	<.105	<.011	<.051	<.008	E.008	<.005	<.004	<.011	<.023	E.007	<.014	<.016	<.008	<.068	<.017	<.010	<.010	<.013	<.004	<.009
293120098285801	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
293120098285801	--	--	--	--	--	--	--	--	--	--	--	--	--	<.026	--	--	--	--	--	--	--	--
293145098224201	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
293119098211201	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
292843098425101	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
292459098382101	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
292925098360201	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
292925098360201	<.007	<.022	<.105	<.011	u	u	E.004	<.005	<.004	<.011	<.023	E.006	<.014	<.016	<.008	<.068	<.017	<.010	<.010	<.013	<.004	<.009
292822098325401	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
292822098325401	<.016	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	<.011	<.023	0.008	<.014	<.016	<.010	<.068	<.017	<.010	<.010	<.013	<.004	<.009
292944098292301	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
292944098292301	<.007	<.022	<.105	<.011	u	u	E.004	<.005	<.004	<.011	<.023	E.007	<.014	<.016	<.008	<.068	<.017	<.010	<.010	<.013	<.004	<.009
292522098291901	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
292643098241801	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
292643098241801	--	--	--	--	--	--	--	--	--	--	--	--	--	<.026	--	--	--	--	--	--	--	--
292328098294601	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
294225098080301	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	.008	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
294225098080301	<.007	<.022	<.105	<.011	<.051	<.008	E.007	<.005	<.004	<.011	<.023	.008	<.014	<.016	<.008	<.068	<.017	<.010	<.010	<.013	<.004	<.009
291210099475601	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
293111098340901	--	<.022	<.105	<.011	--	u	E.008	<.005	<.004	--	--	.007	--	<.016	--	<.068	<.017	<.0102	--	--	--	<.009
293512098291701	--	<.022	<.105	<.011	<.051	<.008	0.010	<.005	<.004	--	--	.010	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
292424098421501	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	Oxy- fluor- fen (µg/L)	Pendi- meth- alin (µg/L)	Phor- ate oxy- gen analog (µg/L)	Phor- ate (µg/L)	Phos- met oxy- gen analog (µg/L)	Phos- met (µg/L)	Pro- meton (µg/L)	Pro- metryn (µg/L)	Pro- pyz- amide (µg/L)	Prop- anil (µg/L)	Prop- argite (µg/L)	Sim- azine (µg/L)	Tebu- con- azole (µg/L)	Tebu- thi- uron (µg/L)	Teflu- thrin (µg/L)	Terb- ufos oxy- gen analog sul- fone (µg/L)	Terb- ufos (µg/L)	Ter- buthyl- azine (µg/L)	Thio- ben- carb (µg/L)	trans- Propi- con- azole (µg/L)	Tri- bu- phos (µg/L)	Tri- flur- alin (µg/L)
292053098365501	--	<0.022	<0.105	<0.011	u	u	<0.010	<0.005	<0.004	--	--	<0.005	--	<0.016	--	<0.068	<0.017	<0.010	--	--	--	<0.009
293807098155301	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	.006	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
292604098563201	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
292116099095501	--	<.022	<.105	<.011	u	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
291232099470301	--	<.022	<.105	<.011	<.051	<.008	E.005	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
293128098473101	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
293451098313201	--	<.022	<.105	<.011	<.051	<.008	E.007	<.005	<.004	--	--	.010	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
294604098060801	--	<.022	<.105	<.011	u	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
294019098114701	--	<.022	<.105	<.011	<.051	<.008	E.005	<.005	<.004	--	--	.006	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
292215098580201	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
292119098524901	--	<.022	<.105	<.011	u	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
291943099163301	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
291840099382601	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
293518098332601	--	<.022	<.105	<.011	<.051	<.008	E.005	<.005	<.004	--	--	E.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
293023098355401	--	<.022	<.105	<.011	u	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
293042098305201	<.007	<.022	<.105	<.011	<.051	<.008	E.005	<.005	<.004	<0.011	<0.023	E.007	<.014	<.016	<0.008	<.068	<.017	<.010	<0.010	<0.013	<0.004	<.009
292442098474501	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
292405098371201	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
292931098274601	--	<.022	<.105	<.011	u	u	<.010	<.005	<.004	--	--	.006	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
293729098173101	--	<.022	<.105	<.011	u	<.008	E.008	<.005	<.004	--	--	.007	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
291219099095601	--	<.022	<.105	<.011	<.051	<.008	<.010	<.005	<.004	--	--	<.005	--	<.016	--	<.068	<.017	<.010	--	--	--	<.009
Monitoring wells																						
292943098354401	<.006	<.012	<.027	<.04	u	<.008	<.01	<.006	<.004	<.006	<.04	E.006	<.018	<.016	<.003	<.045	<.018	<.008	<.010	<.02	<.035	<.006
292943098354402	<.006	<.012	<.027	<.04	<.051	<.008	<.01	<.006	<.004	<.006	<.04	E.005	<.018	<.016	<.003	<.045	<.018	<.008	<.010	<.02	<.035	<.006
292943098354403	<.006	<.012	<.027	<.04	u	<.008	<.01	<.006	<.004	<.006	<.04	<.006	<.018	<.016	<.003	<.045	<.018	<.008	<.010	<.02	<.035	<.006
292943098354404	<.006	<.012	<.027	<.04	<.051	<.008	<.01	<.006	<.004	<.006	<.04	<.006	<.018	<.016	<.003	<.045	<.018	<.008	<.010	<.02	<.035	<.006

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	Oxy- fluor- fen (µg/L)	Pendi- meth- alin (µg/L)	Phor- ate oxy- gen analog (µg/L)	Phor- ate (µg/L)	Phos- met oxy- gen analog (µg/L)	Phos- met (µg/L)	Pro- meton (µg/L)	Pro- metryn (µg/L)	Pro- pyz- amide (µg/L)	Prop- anil (µg/L)	Prop- argite (µg/L)	Sim- azine (µg/L)	Tebu- con- azole (µg/L)	Tebu- thi- uron (µg/L)	Teflu- thrin (µg/L)	Terb- ufos oxy- gen analog sul- fone (µg/L)	Terb- ufos (µg/L)	Ter- buthyl- azine (µg/L)	Thio- ben- carb (µg/L)	trans- Propi- con- azole (µg/L)	Tri- bu- phos (µg/L)	Tri- flur- alin (µg/L)
292851098374401	<0.006	<0.012	<0.027	<0.04	<0.051	<0.008	<0.01	<0.006	<0.004	<0.006	<0.04	E0.006	<0.018	<0.016	<0.003	<0.045	<0.018	<0.008	<0.010	<0.02	<0.035	<0.009
292851098374402	<.006	<.012	<.027	<.04	<.051	<.008	<.01	<.006	<.004	<.006	<.04	E.006	<.018	<.016	<.003	<.045	<.018	<.008	<.010	<.02	<.035	<.009
Well-field depth-dependent sampling (W4)																						
292916098360701	<.006	<.012	<.027	<.04	<.051	<.008	<.01	<.006	<.004	<.006	<.04	<.006	<.018	<.016	<.003	<.045	<.018	<.008	<.010	<.02	<.035	<.006
292916098360701	<.006	<.012	<.027	<.04	<.051	<.008	<.01	<.006	<.004	<.006	<.04	.008	<.018	<.016	<.003	<.045	<.018	<.008	<.010	<.02	<.035	<.006
292916098360701	<.006	<.012	<.027	<.04	<.051	<.008	<.01	<.006	u	<.006	<.04	<.006	<.018	<.016	<.003	<.045	<.018	u	<.010	<.02	<.035	<.006
292916098360701	<.006	<.012	<.027	<.04	<.051	<.008	<.01	<.006	<.004	<.006	<.04	E.005	<.018	<.016	<.003	<.045	<.018	<.008	<.010	<.02	<.035	<.006
292916098360701	<.006	<.012	<.027	<.04	<.051	<.008	<.01	<.006	<.004	<.006	<.04	<.006	<.018	<.016	<.003	<.045	<.018	u	<.010	<.02	<.035	<.006
292916098360701	<.006	<.012	<.027	<.04	<.051	<.008	<.01	<.006	<.004	<.006	<.04	.009	<.018	<.016	<.003	<.045	<.018	E.007	<.010	<.02	<.035	<.006
292916098360701	<.006	<.012	<.027	<.04	<.051	<.008	<.01	<.006	<.004	<.006	<.04	E.006	<.018	<.016	<.003	<.045	<.018	<.008	<.010	<.02	<.035	<.006
292916098360701	<.006	<.012	<.027	<.04	<.051	<.008	<.01	<.006	<.004	<.006	<.04	.009	<.018	<.016	<.003	<.045	<.018	<.008	<.010	<.02	<.035	<.006
292916098360701	<.006	<.012	<.027	<.04	<.051	<.008	<.01	<.006	<.004	<.006	<.04	.008	<.018	<.016	<.003	<.045	<.018	<.008	<.010	<.02	<.035	<.006
Well-field wells (wellhead sampling)																						
292919098360501	<.006	<.012	<.027	<.04	<.051	<.008	<.01	<.006	<.004	<.006	<.04	E.005	<.018	<.016	<.003	<.045	<.018	<.008	<.010	<.02	<.035	<.006
292925098360201	<.006	<.012	<.027	<.04	<.051	<.008	<.01	<.006	<.004	<.006	<.04	<.006	<.018	<.016	<.003	<.045	<.018	<.008	<.010	<.02	<.035	<.006
292916098360701	<.006	<.012	<.027	<.04	<.051	<.008	<.01	<.006	<.004	<.006	<.04	E.005	<.018	<.016	<.003	<.045	<.018	<.008	<.010	<.02	<.035	<.006
292920098360601	<.006	<.012	<.027	<.04	<.051	<.008	<.01	<.006	<.004	<.006	<.04	E.005	<.018	<.016	<.003	<.045	<.018	<.008	<.010	<.02	<.035	<.006
292923098360301	<.006	<.012	<.027	<.04	u	<.008	<.01	<.006	<.004	<.006	<.04	<.006	<.018	<.016	<.003	<.045	<.018	<.008	<.010	<.02	<.035	<.006
Temporal sampling																						
292923098360301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292923098360301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292923098360301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--

Table 15. Soluble pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; MW, monitoring well; OVB, overburden; µg/L, micrograms per liter; --, not measured; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level]

USGS identification number	Oxy- fluor- fen (µg/L)	Pendi- meth- alin (µg/L)	Phor- ate oxy- gen analog (µg/L)	Phor- ate (µg/L)	Phos- met oxy- gen analog (µg/L)	Phos- met (µg/L)	Pro- meton (µg/L)	Pro- metryn (µg/L)	Pro- pyz- amide (µg/L)	Prop- anil (µg/L)	Prop- argite (µg/L)	Sim- azine (µg/L)	Tebu- con- azole (µg/L)	Tebu- thi- uron (µg/L)	Teflu- thrin (µg/L)	Terb- ufos oxy- gen analog sul- fone (µg/L)	Terb- ufos (µg/L)	Ter- buthyl- azine (µg/L)	Thio- ben- carb (µg/L)	trans- Propi- con- azole (µg/L)	Tri- bu- phos (µg/L)	Tri- flur- alin (µg/L)
292923098360301	<0.010	<0.012	<0.027	<0.02	<0.051	<0.034	E.008	<0.006	<0.004	<0.010	<0.020	<0.006	--	<0.028	<0.010	<0.045	<0.018	<0.006	<0.016	<0.02	<0.018	<0.018
292923098360301	<.010	<.012	<.027	<.02	<.051	<.034	<.012	<.006	<.004	<.010	<.020	<.006	--	<.028	<.010	<.045	<.018	<.006	<.016	<.02	<.018	<.018
292923098360301	<.010	<.012	<.027	<.02	u	<.034	<.012	<.006	<.004	<.010	<.020	E.006	--	<.028	<.010	<.040	<.018	<.006	<.016	<.02	<.018	<.018
292943098354402	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354402	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354402	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354402	<.010	<.012	<.027	<.02	<.051	<.034	<.012	<.006	<.007	<.010	<.020	.008	--	<.028	<.010	<.045	<.018	<.006	<.016	<.02	<.018	<.018
292943098354402	<.010	<.012	<.027	<.02	<.051	<.034	<.012	<.006	u	<.010	<.020	<.006	--	<.028	<.010	<.045	<.018	<.011	<.016	<.02	<.018	<.018
292943098354402	<.010	<.012	<.027	<.02	<.051	<.034	<.012	<.006	<.004	<.010	<.020	E.007	--	<.028	<.010	<.045	<.018	<.007	<.016	<.02	<.018	<.018
292943098354403	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354403	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354403	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354403	<.010	<.012	<.027	<.02	<.051	<.034	<.012	<.006	u	<.010	<.020	<.008	--	<.028	<.010	<.045	<.018	<.006	<.016	<.02	<.018	<.018
292943098354403	<.010	<.012	<.027	<.02	<.051	<.034	<.012	<.006	<.004	<.010	<.020	E.003	--	<.028	<.010	<.045	<.018	<.006	<.016	<.02	<.018	<.018
292943098354403	<.010	<.012	<.027	<.02	u	<.034	<.012	<.006	u	<.010	<.020	E.003	--	<.028	<.010	<.045	<.018	<.006	<.016	<.02	<.018	<.018
292943098354404	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354404	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354404	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292943098354404	<.010	<.012	<.027	<.02	<.051	<.034	<.012	<.006	<.009	<.010	<.020	<.008	--	<.028	<.010	<.045	<.018	<.006	<.016	<.02	<.018	<.018
292943098354404	<.010	<.012	<.027	<.02	<.051	<.034	<.012	<.006	<.004	<.010	<.020	<.007	--	<.028	<.010	<.045	<.018	<.006	<.016	<.02	<.018	<.018
292943098354404	<.010	<.012	<.027	<.02	<.051	<.034	<.012	<.006	<.004	<.010	<.020	<.007	--	<.028	<.010	<.045	<.018	<.006	<.016	<.02	<.018	<.018
293252098380801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293252098380801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293252098380801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293252098380801	<.010	<.012	<.027	<.02	<.051	<.034	<.012	<.006	u	<.010	<.020	.008	--	<.028	<.010	<.045	<.018	<.006	<.016	<.02	<.018	<.018
293252098380801	<.010	<.012	<.027	<.02	<.051	<.034	E.004	<.006	<.004	<.010	<.020	E.007	--	<.028	<.010	<.045	<.018	<.006	<.016	<.02	<.018	<.018
293252098380801	<.010	<.012	<.027	<.02	<.051	<.034	<.012	<.006	<.004	<.010	<.020	E.007	--	<.028	<.010	<.045	<.018	<.006	<.016	<.02	<.018	<.018

* Sample collected from top of open interval during depth-dependent sampling.

Table 16. Polar pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level; --, not measured]

USGS identification number	State well number	Group and description	Sample date	2,4-D methyl ester (µg/L)	2,4-D (µg/L)	2,4-DB (µg/L)	2-Chloro- 6-ethyl- 4-amino- s-triazine (CEAT) (µg/L)	2-Hydroxy- 4-isopropyl- amino- 6-ethyl- amino-s- triazine (OIET) (µg/L)	3-Ke- to- carbo- furan (µg/L)	3-Hy- droxy carbo- furan (µg/L)	Acifluor- fen (µg/L)	Aldi- carb sulf- one (µg/L)	Aldi- carb sulf- oxide (µg/L)	Aldi- carb (µg/L)	Bendio- carb (µg/L)
Regional aquifer public-supply wells															
293359098290301	AY-68-29-414	1st (highest) quartile	2004/12/06	<0.016	<0.038	<0.02	E0.006	<0.032	<0.02	<0.008	<0.028	<0.018	<0.022	<0.04	<0.02
293359098290301	AY-68-29-414	1st quartile	2005/09/01	<.016	<.038	<.02	<.08	<.032	<.02	<.008	<.028	<.018	<.022	<.04	<.02
293358098231101	AY-68-29-610	1st quartile	2004/12/15	<.016	<.038	<.02	E.005	E.007	<.02	<.008	<.028	<.018	<.022	<.04	<.02
293358098231101	AY-68-29-610	1st quartile	2005/08/29	<.016	<.038	<.020	E.009	<.032	<.02	<.008	<.028	<.018	<.022	<.04	<.020
293120098285801	AY-68-29-714	1st quartile	2004/12/14	<.016	<.038	<.02	E.005	E.004	<.02	<.008	<.028	<.018	<.022	<.04	<.02
293120098285801	AY-68-29-714	1st quartile	2005/08/30	<.016	<.038	<.020	<.08	<.032	<.02	<.008	<.028	<.018	<.022	<.04	<.020
293145098224201	AY-68-29-929	1st quartile	2004/12/15	<.016	<.038	<.02	<.08	<.032	<.02	<.008	<.028	<.018	<.022	<.04	<.02
293119098211201	AY-68-30-718	1st quartile	2004/12/16	<.016	<.038	<.02	<.08	<.032	<.02	<.008	<.028	<.018	<.022	<.04	<.02
292843098425101	AY-68-35-106	1st quartile	2004/12/13	<.016	<.038	<.02	<.08	<.032	<.02	<.008	<.028	<.018	<.022	<.04	<.02
292459098382101	AY-68-35-913	1st quartile	2004/12/13	<.016	<.038	<.02	<.08	<.032	<.02	<.008	<.028	<.018	<.022	<.04	<.02
292925098360201	AY-68-36-104	1st quartile	2004/12/13	<.016	<.038	<.02	E.005	E.003	<.02	<.008	<.028	<.018	<.022	<.04	<.02
292925098360201	AY-68-36-104	1st quartile	2005/08/31	<.016	<.038	<.020	<.08	<.032	<.02	<.008	<.028	<.018	<.022	<.04	<.020
292822098325401	AY-68-36-206	1st quartile	2004/12/13	<.016	<.038	<.02	E.003	<.032	<.02	<.008	<.028	<.018	<.022	<.04	<.02
292822098325401	AY-68-36-206	1st quartile	2005/08/31	--	--	--	--	--	--	--	--	--	--	--	--
292944098292301	AY-68-37-105	1st quartile	2004/12/14	<.016	<.038	<.02	<.08	<.032	<.02	<.008	<.028	<.018	<.022	<.04	<.02
292944098292301	AY-68-37-105	1st quartile	2005/08/31	--	--	--	--	--	--	--	--	--	--	--	--
292522098291901	AY-68-37-426	1st quartile	2004/12/14	<.016	<.038	<.02	<.08	<.032	<.02	<.008	<.028	<.018	<.022	<.04	<.02
292643098241801	AY-68-37-601	1st quartile	2004/12/15	<.016	E.009	<.02	<.08	<.032	<.02	<.008	<.028	<.018	<.022	<.04	<.02
292643098241801	AY-68-37-601	1st quartile	2005/08/30	<.016	<.038	<.020	<.08	<.032	<.02	<.008	<.028	<.018	<.022	<.04	<.020
292328098294601	AY-68-37-705	1st quartile	2004/12/14	<.016	<.038	<.02	<.08	<.032	<.02	<.008	<.028	<.018	<.022	<.04	<.02
294225098080301	DX-68-23-601	1st quartile	2004/12/08	<.016	<.038	<.02	<.08	<.032	<.02	<.008	<.028	<.018	<.022	<.04	<.02
294225098080301	DX-68-23-601	1st quartile	2005/09/01	--	--	--	--	--	--	--	--	--	--	--	--
291210099475601	YP-69-50-506	1st quartile	2004/11/30	<.016	<.038	<.02	<.08	<.032	<.02	<.008	<.028	<.018	<.022	<.04	<.02

Table 16. Polar pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level; --, not measured]

USGS identification number	State well number	Group and description	Sample date	2,4-D methyl ester (µg/L)	2,4-D (µg/L)	2,4-DB (µg/L)	2-Chloro- 6-ethyl- 4-amino- s-triazine (CEAT) (µg/L)	2-Hydroxy- 4-isopropyl- 6-ethyl- amino-s- triazine (OIET) (µg/L)	3-Ke- to- carbo- furan (µg/L)	3-Hy- droxy carbo- furan (µg/L)	Acifluor- fen (µg/L)	Aldi- carb sulf- one (µg/L)	Aldi- carb sulf- oxide (µg/L)	Aldi- carb (µg/L)	Bendio- carb (µg/L)
293111098340901	AY-68-28-807	2d quartile	2005/05/26	--	--	--	--	--	--	--	--	--	--	--	--
293512098291701	AY-68-29-109	2d quartile	2005/04/20	--	--	--	--	--	--	--	--	--	--	--	--
292424098421501	AY-68-35-810	2d quartile	2005/04/21	--	--	--	--	--	--	--	--	--	--	--	--
292053098365501	AY-68-44-110	2d quartile	2005/05/26	--	--	--	--	--	--	--	--	--	--	--	--
293807098155301	DX-68-22-901	2d quartile	2005/05/18	--	--	--	--	--	--	--	--	--	--	--	--
292604098563201	TD-68-33-501	2d quartile	2005/05/16	--	--	--	--	--	--	--	--	--	--	--	--
292116099095501	TD-69-47-305	2d quartile	2005/04/27	--	--	--	--	--	--	--	--	--	--	--	--
291232099470301	YP-69-50-339	2d quartile	2005/05/11	--	--	--	--	--	--	--	--	--	--	--	--
293128098473101	AY-68-26-814	3d quartile	2005/04/18	--	--	--	--	--	--	--	--	--	--	--	--
293451098313201	AY-68-28-601	3d quartile	2005/05/17	--	--	--	--	--	--	--	--	--	--	--	--
294604098060801	DX-68-16-708	3d quartile	2005/04/25	--	--	--	--	--	--	--	--	--	--	--	--
294019098114701	DX-68-23-504	3d quartile	2005/05/18	--	--	--	--	--	--	--	--	--	--	--	--
292215098580201	TD-68-41-103	3d quartile	2005/05/09	--	--	--	--	--	--	--	--	--	--	--	--
292119098524901	TD-68-41-308	3d quartile	2005/04/27	--	--	--	--	--	--	--	--	--	--	--	--
291943099163301	TD-69-46-601	3d quartile	2005/05/09	--	--	--	--	--	--	--	--	--	--	--	--
291840099382601	YP-69-43-606	3d quartile	2005/05/11	--	--	--	--	--	--	--	--	--	--	--	--
293518098332601	AY-68-28-203	4th quartile	2005/04/19	--	--	--	--	--	--	--	--	--	--	--	--
293023098355401	AY-68-28-702	4th quartile	2005/04/26	--	--	--	--	--	--	--	--	--	--	--	--
293042098305201	AY-68-28-913	4th quartile	2005/06/30	--	--	--	--	--	--	--	--	--	--	--	--
292442098474501	AY-68-34-803	4th quartile	2005/04/20	--	--	--	--	--	--	--	--	--	--	--	--
292405098371201	AY-68-36-704	4th quartile	2005/05/23	--	--	--	--	--	--	--	--	--	--	--	--
292931098274601	AY-68-37-124	4th quartile	2005/05/25	--	--	--	--	--	--	--	--	--	--	--	--
293729098173101	DX-68-30-215	4th quartile	2005/04/25	--	--	--	--	--	--	--	--	--	--	--	--
291219099095601	TD-69-55-604	4th quartile	2005/05/16	--	--	--	--	--	--	--	--	--	--	--	--

Table 16. Polar pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level; --, not measured]

USGS identification number	Ben- omyl (µg/L)	Bensulf- uron- methyl (µg/L)	Bentazon (µg/L)	Brom- oxynil (µg/L)	Carb- aryl (µg/L)	Carbo- furan (µg/L)	Chlor- amben methyl ester (µg/L)	Chlor- im- uron- ethyl (µg/L)	Chloro- diamino- s-triazine (CAAT) (µg/L)	Chloro- thal- onil (µg/L)	Clopyr- alid (µg/L)	Cyclo- ate (µg/L)	Dacthal- mono- acid (µg/L)	Di- camba (µg/L)	Di- chlor- prop (µg/L)	Dino- seb (µg/L)	Di- phen- amid (µg/L)
Regional aquifer public-supply wells																	
293359098290301	<0.022	<0.018	<0.012	<0.028	<0.018	<0.016	<0.024	<0.032	<0.04	<0.035	<0.024	<0.014	<0.028	<0.036	<0.028	<0.038	<0.01
293359098290301	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.01
293358098231101	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.01
293358098231101	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.010
293120098285801	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.01
293120098285801	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.010
293145098224201	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.01
293119098211201	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.01
292843098425101	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.01
292459098382101	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.01
292925098360201	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.01
292925098360201	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.010
292822098325401	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.01
292822098325401	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292944098292301	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.01
292944098292301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292522098291901	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.01
292643098241801	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.01
292643098241801	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.010
292328098294601	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.01
294225098080301	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.01
294225098080301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291210099475601	<.022	<.018	<.012	<.028	<.018	<.016	<.024	<.032	<.04	<.035	<.024	<.014	<.028	<.036	<.028	<.038	<.01

Table 16. Polar pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level; --, not measured]

USGS identification number	Ben- omyl (µg/L)	Bensulf- uron- methyl (µg/L)	Bentazon (µg/L)	Brom- oxynil (µg/L)	Carb- aryl (µg/L)	Carbo- furan (µg/L)	Chlor- amben methyl ester (µg/L)	Chlor- im- uron- ethyl (µg/L)	Chloro- diamino- s-triazine (CAAT) (µg/L)	Chloro- thal- onil (µg/L)	Clopyr- alid (µg/L)	Cyclo- ate (µg/L)	Dacthal- mono- acid (µg/L)	Di- camba (µg/L)	Di- chlor- prop (µg/L)	Dino- seb (µg/L)	Di- phen- amid (µg/L)
293111098340901	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293512098291701	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292424098421501	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292053098365501	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293807098155301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292604098563201	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292116099095501	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291232099470301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293128098473101	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293451098313201	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
294604098060801	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
294019098114701	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292215098580201	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292119098524901	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291943099163301	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291840099382601	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293518098332601	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293023098355401	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293042098305201	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292442098474501	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292405098371201	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292931098274601	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293729098173101	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291219099095601	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--	--

Table 16. Polar pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level; --, not measured]

USGS identification number	Diuron (µg/L)	Fenuron (µg/L)	Flumet- sulam (µg/L)	Fluo- meturon (µg/L)	Imaz- aquin (µg/L)	Imaz- etha- pyr (µg/L)	Imida- cloprid (µg/L)	Linuron (µg/L)	MCPA (µg/L)	MCPB (µg/L)	Methio- carb (µg/L)	Meth- omyl (µg/L)	Met- sulfuron (µg/L)	N-(4-chloro- phenyl)-N'- methyl-urea (µg/L)
Regional aquifer public-supply wells														
293359098290301	<0.015	<0.019	<0.04	<0.016	<0.036	<0.038	<0.02	<0.014	<0.03	<0.01	<0.02	<0.025	<0.036	<0.012
293359098290301	<.015	<.019	<.04	<.016	<.036	<.038	<.02	<.014	<.03	<.01	<.01	<.02	<.025	<.036
293358098231101	<.015	<.019	<.04	<.016	<.036	<.038	<.02	<.014	u	<.01	<.01	<.02	<.025	<.036
293358098231101	<.015	<.019	<.040	<.016	<.036	<.038	<.020	<.014	<.030	<.010	<.010	<.020	<.025	<.036
293120098285801	<.015	<.019	<.04	<.016	<.036	<.038	<.02	<.014	u	<.01	<.01	<.02	<.025	<.036
293120098285801	<.015	<.019	<.040	<.016	<.036	<.038	<.020	<.014	<.030	<.010	<.010	<.020	<.025	<.036
293145098224201	<.015	<.019	<.04	<.016	<.036	<.038	<.02	<.014	u	<.01	<.01	<.02	<.025	<.036
293119098211201	<.015	<.019	<.04	<.016	<.036	<.038	<.02	<.014	u	<.01	<.01	<.02	<.025	<.036
292843098425101	<.015	<.019	<.04	<.016	<.036	<.038	<.02	<.014	<.03	<.01	<.01	<.02	<.025	<.036
292459098382101	<.015	<.019	<.04	<.016	<.036	<.038	<.02	<.014	<.03	<.01	<.01	<.02	E.035	<.036
292925098360201	<.015	<.019	<.04	<.016	<.036	<.038	<.02	<.014	<.03	<.01	<.01	<.02	<.025	<.036
292925098360201	<.015	<.019	<.040	<.016	<.036	<.038	<.020	<.014	<.030	<.010	<.010	<.020	<.025	<.036
292822098325401	<.015	<.019	<.04	<.016	<.036	<.038	<.02	<.014	<.03	<.01	<.01	<.02	<.025	<.036
292822098325401	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292944098292301	<.015	<.019	<.04	<.016	<.036	<.038	<.02	<.014	u	<.01	<.01	<.02	<.025	<.036
292944098292301	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292522098291901	<.015	<.019	<.04	<.016	<.036	<.038	<.02	<.014	u	<.01	<.01	<.02	<.025	<.036
292643098241801	<.015	<.019	<.04	<.016	<.036	<.038	<.02	<.014	u	<.01	<.01	<.02	<.025	<.036
292643098241801	<.015	<.019	<.040	<.016	<.036	<.038	<.020	<.014	<.030	<.010	<.010	<.020	<.025	<.036
292328098294601	<.015	<.019	<.04	<.016	<.036	<.038	<.02	<.014	u	<.01	<.01	<.02	<.025	<.036
294225098080301	<.015	<.019	<.04	<.016	<.036	<.038	<.02	<.014	<.03	<.01	<.01	<.02	<.025	<.036
294225098080301	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291210099475601	<.015	<.019	<.04	<.016	<.036	<.038	<.02	<.014	<.03	<.01	<.01	<.02	<.025	<.036

Table 16. Polar pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level; --, not measured]

USGS identification number	Diuron (µg/L)	Fenuron (µg/L)	Flumet- sulam (µg/L)	Fluo- meturon (µg/L)	Imaz- aquin (µg/L)	Imaz- etha- pyr (µg/L)	Imida- cloprid (µg/L)	Linuron (µg/L)	MCPA (µg/L)	MCPB (µg/L)	Methio- carb (µg/L)	Meth- omyl (µg/L)	Met- sulfuron (µg/L)	N-(4-chloro- phenyl)-N'- methyl-urea (µg/L)
293111098340901	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293512098291701	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292424098421501	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292053098365501	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293807098155301	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292604098563201	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292116099095501	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291232099470301	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293128098473101	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293451098313201	--	--	--	--	--	--	--	--	--	--	--	--	--	--
294604098060801	--	--	--	--	--	--	--	--	--	--	--	--	--	--
294019098114701	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292215098580201	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292119098524901	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291943099163301	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291840099382601	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293518098332601	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293023098355401	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293042098305201	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292442098474501	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292405098371201	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292931098274601	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293729098173101	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291219099095601	--	--	--	--	--	--	--	--	--	--	--	--	--	--

Table 16. Polar pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level; --, not measured]

USGS identification number	Neburon (µg/L)	Nico- sulfuron (µg/L)	Nor- flurazon (µg/L)	Oryz- alin (µg/L)	Oxa- myl (µg/L)	Pic- loram (µg/L)	Prop- ham (µg/L)	Propi- con- azole (µg/L)	Prop- oxur (µg/L)	Siduron (µg/L)	Sulfo- meturon (µg/L)	Terbacil (µg/L)	Triben- uron- methyl (µg/L)	Triclopyr (µg/L)
Regional aquifer public-supply wells														
293359098290301	<0.04	<0.02	<0.012	<0.03	<0.032	<0.03	<0.01	<0.008	<0.02	<0.038	<0.016	u	<0.026	<0.026
293359098290301	<.012	<.04	<.02	<.012	<.03	<.032	<.03	<.01	<.008	<.02	<.038	<0.016	--	<.026
293358098231101	<.012	<.04	<.02	<.012	<.03	<.032	<.03	<.01	<.008	<.02	<.038	<0.016	u	<.026
293358098231101	<.012	<.04	<.020	<.012	<.030	<.032	<.030	<.010	<.008	<.020	<.038	<0.016	--	<.026
293120098285801	<.012	<.04	<.02	<.012	<.03	<.032	<.03	<.01	<.008	<.02	<.038	<0.016	u	<.026
293120098285801	<.012	<.04	<.020	<.012	<.030	<.032	<.030	<.010	<.008	<.020	<.038	<0.016	--	<.026
293145098224201	<.012	<.04	<.02	<.012	<.03	<.032	<.03	<.01	<.008	<.02	<.038	<0.016	u	<.026
293119098211201	<.012	<.04	<.02	<.012	<.03	<.032	<.03	<.01	<.008	<.02	<.038	<0.016	u	<.026
292843098425101	<.012	<.04	<.02	<.012	<.03	<.032	<.03	<.01	<.008	<.02	<.038	<0.016	u	<.026
292459098382101	<.012	<.04	<.02	<.012	<.03	<.032	<.03	<.01	<.008	<.02	<.038	<0.016	u	<.026
292925098360201	<.012	<.04	<.02	<.012	<.03	<.032	<.03	<.01	<.008	<.02	<.038	<0.016	u	<.026
292925098360201	<.012	<.04	<.020	<.012	<.030	<.032	<.030	<.010	<.008	<.020	<.038	<0.016	--	<.026
292822098325401	<.012	<.04	<.02	<.012	<.03	<.032	<.03	<.01	<.008	<.02	<.038	<0.016	u	<.026
292822098325401	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292944098292301	<.012	<.04	<.02	<.012	<.03	<.032	<.03	<.01	<.008	<.02	<.038	<0.016	u	<.026
292944098292301	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292522098291901	<.012	<.04	<.02	<.012	<.03	<.032	<.03	<.01	<.008	<.02	<.038	<0.016	u	<.026
292643098241801	<.012	<.04	<.02	<.012	<.03	<.032	<.03	<.01	<.008	<.02	<.038	<0.016	u	<.026
292643098241801	<.012	<.04	<.020	<.012	<.030	<.032	<.030	<.010	<.008	<.020	<.038	<0.016	--	<.026
292328098294601	<.012	<.04	<.02	<.012	<.03	<.032	<.03	<.01	<.008	<.02	<.038	<0.016	u	<.026
294225098080301	<.012	<.04	<.02	<.012	<.03	<.032	<.03	<.01	<.008	<.02	<.038	<0.016	u	<.026
294225098080301	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291210099475601	<.012	<.04	<.02	<.012	<.03	<.032	<.03	<.01	<.008	<.02	<.038	<0.016	u	<.026

Table 16. Polar pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; E, estimated; u, unable to determine; <, nondetection less than laboratory method reporting level; --, not measured]

USGS identification number	Neburon (µg/L)	Nico- sulfuron (µg/L)	Nor- flurazon (µg/L)	Oryz- alin (µg/L)	Oxa- myl (µg/L)	Pic- loram (µg/L)	Prop- ham (µg/L)	Propi- con- azole (µg/L)	Prop- oxur (µg/L)	Siduron (µg/L)	Sulfo- meturon (µg/L)	Terbacil (µg/L)	Triben- uron- methyl (µg/L)	Triclopyr (µg/L)
293111098340901	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293512098291701	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292424098421501	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292053098365501	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293807098155301	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292604098563201	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292116099095501	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291232099470301	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293128098473101	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293451098313201	--	--	--	--	--	--	--	--	--	--	--	--	--	--
294604098060801	--	--	--	--	--	--	--	--	--	--	--	--	--	--
294019098114701	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292215098580201	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292119098524901	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291943099163301	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291840099382601	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293518098332601	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293023098355401	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293042098305201	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292442098474501	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292405098371201	--	--	--	--	--	--	--	--	--	--	--	--	--	--
292931098274601	--	--	--	--	--	--	--	--	--	--	--	--	--	--
293729098173101	--	--	--	--	--	--	--	--	--	--	--	--	--	--
291219099095601	--	--	--	--	--	--	--	--	--	--	--	--	--	--

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; --, not measured; <, nondetection less than laboratory method reporting level]

[illegible]

Table 17. Acetamide pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; --, not measured; <, nondetection less than laboratory method reporting level]

USGS identification number	State well number	Sample date	Group and description	2-[(2-Ethyl- 6-methyl- phenyl) amino]- 2-oxo- ethane- sulfonic acid (N-(eth- methphen) ox ESA) (µg/L)	Ala- chlor 2nd amide, (µg/L)	2-Chloro- N-(2-ethyl- 6-methyl- phenyl)- acetamid (aceto- chlor 2nd amide) (µg/L)	Aceto- chlor ethane- sulf- onic acid (aceto- chlor ESA) (µg/L)	Aceto- chlor oxa- nilic acid (aceto- chlor OA) (µg/L)	Aceto- chlor sulf- ynil- acetic acid (aceto- chlor SAA) (µg/L)	Alachlor ethane- sulfonic acid sec- ondary amide (ala- chlor ESA SA) (µg/L)	Ala- chlor ethane- sulfonic acid (ala- chlor ESA) (µg/L)	Ala- chlor oxanilic acid (ala- chlor OA) (µg/L)	Alachlor sulfynil- acetic acid (ala- chlor SAA) (µg/L)	De- chloro- aceto- chlor (µg/L)
293111098340901	AY-68-28-807	2005/05/26	2d quartile	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	0.02	<0.02	<0.02	<0.02
293512098291701	AY-68-29-109	2005/04/20	2d quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
292424098421501	AY-68-35-810	2005/04/21	2d quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
292053098365501	AY-68-44-110	2005/05/26	2d quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
293807098155301	DX-68-22-901	2005/05/18	2d quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
292604098563201	TD-68-33-501	2005/05/16	2d quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
292116099095501	TD-69-47-305	2005/04/27	2d quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
291232099470301	YP-69-50-339	2005/05/11	2d quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
293128098473101	AY-68-26-814	2005/04/18	3d quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
293451098313201	AY-68-28-601	2005/05/17	3d quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
294604098060801	DX-68-16-708	2005/04/25	3d quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
294019098114701	DX-68-23-504	2005/05/18	3d quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
292215098580201	TD-68-41-103	2005/05/09	3d quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
292119098524901	TD-68-41-308	2005/04/27	3d quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
291943099163301	TD-69-46-601	2005/05/09	3d quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
291840099382601	YP-69-43-606	2005/05/11	3d quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
293518098332601	AY-68-28-203	2005/04/19	4th quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
293023098355401	AY-68-28-702	2005/04/26	4th quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
293042098305201	AY-68-28-913	2005/06/30	4th quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
292442098474501	AY-68-34-803	2005/04/20	4th quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
292405098371201	AY-68-36-704	2005/05/23	4th quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
292931098274601	AY-68-37-124	2005/05/25	4th quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	0.02	<.02	<.02	<.02
293729098173101	DX-68-30-215	2005/04/25	4th quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02
291219099095601	TD-69-55-604	2005/05/16	4th quartile	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02

Table 17. Acetamide pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; --, not measured; <, nondetection less than laboratory method reporting level]

[illegible]

Table 17. Acetamide pesticide compounds for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. USGS, U.S. Geological Survey; µg/L, micrograms per liter; --, not measured; <, nondetection less than laboratory method reporting level]

USGS identification number	De- chloro- ala- chlor (µg/L)	De- chloro- di- methen- amid (µg/L)	De- chloro- metola- chlor (µg/L)	Di- methen- amid ethane- sulfonic acid (µg/L)	Di- methen- amid oxa- nilic acid (µg/L)	Di- methen- amid (µg/L)	Flufen- acet ethane- sulfonic acid (µg/L)	Flufen- acet oxa- nilic acid (µg/L)	Flufen- acet (µg/L)	Hy- droxy- aceto- chlor (µg/L)	Hy- droxy- ala- chlor (µg/L)	Hy- droxy- di- methen- amid (µg/L)	Hy- droxy- metola- chlor (µg/L)	Metola- chlor ethane- sulf- onic acid (µg/L)	Metola- chlor oxa- nilic acid (µg/L)	Prop- achlor ethane- sulf- onic acid (µg/L)	Prop- achlor oxa- nilic acid (µg/L)
293111098340901	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<0.05	<0.02
293512098291701	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
292424098421501	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
292053098365501	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
293807098155301	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
292604098563201	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
292116099095501	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
291232099470301	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
293128098473101	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
293451098313201	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
294604098060801	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
294019098114701	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
292215098580201	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
292119098524901	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
291943099163301	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
291840099382601	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
293518098332601	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
293023098355401	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
293042098305201	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
292442098474501	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
292405098371201	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
292931098274601	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
293729098173101	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02
291219099095601	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.02	<.05	<.02

Table 18. Summary of sulfur hexafluoride and chlorofluorocarbon age-tracer data and piston-flow model results for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[By sample category. SF₆, sulfur hexafluoride; CFC, chlorofluorocarbon; °C, degrees Celsius; m, meter; pptv, parts per trillion by volume; C, contaminated]

USGS identification number	State well number	Local identifier and description	Sample date	Recharge temperature (°C)	Recharge elevation (m)	SF ₆ atmospheric mixing ratio (pptv)*	SF ₆ apparent age, (years)**	CFC-11, atmospheric mixing ratio (pptv)*	CFC-12, atmospheric mixing ratio (pptv)*	CFC-113, atmospheric mixing ratio (pptv)*	CFC-11 piston-flow apparent age (years)**	CFC-12 piston-flow apparent age (years)**	CFC-113 piston-flow apparent age (years)**
Monitoring wells													
292943098354401	AY-68-36-135	Z-OVB	10/30/2007	28.3	258.5	56.13, 57.32	C	4515.5, 4607.1	1972.9, 1930.3	4040.7, 3852.0	C	C	C
292943098354402	AY-68-36-133	Z-SED	10/31/2007	27.1	258.5	20.92, 39.23	C	4541.5, 4402.1	1487.8, 1449.3	295.4, 285.8	C	C	C
292943098354403	AY-68-36-134	Z-IED	10/30/2007	25.1	258.5	17.89, 18.78	C	1615.9, 4949.0, 4488.3	1519.1, 1628.3, 1403.0	244.2, 281.7, 281.3	C	C	C
292943098354404	AY-68-36-132	Z-DED	11/1/2007	26.8	258.5	12.77, 8.58	C	3563.5, 3021.7	1375.4, 1332.1	238.9, 242.1	C	C	C
292851098374401	AY-68-35-315	T-IED	5/6/2008	16.7	242.3	7.37, 7.11	C	5588.6, 5511.1	24893.7, 23877.0	493.4, 493.4	C	C	C
292851098374402	AY-68-35-314	T-DED	5/7/2008	15.8	242.3	11.84, 11.59	C	2385.8, 2482.6, 2524.1	1279.6, 1366.7, 1385.5	456.9, 485.9, 493.5	C	C	C
Well-field wells (wellhead sampling)													
292919098360501	AY-68-36-103	W2	11/29/2007	22.0	276.5	15.76, 15.70	C	4548.4, 4676.9	2566.0, 2758.6	345.0, 361.5	C	C	C
292925098360201	AY-68-36-104	W3	11/28/2007	20.8	269.8	11.44, 16.94	C	4952.2, 4967.3, 4479.9	2033.1, 2074.9, 2210.3	563.8, 485.7, 526.7	C	C	C
292916098360701	AY-68-36-105	W4	11/16/2007	24.1	270.4	27.33, 18.54	C	4952.2, 4967.3, 4479.9	4497.9, 3999.8, 3337.9	258.6, 449.6, 449.6	C	C	C
292920098360601	AY-68-36-130	W5	11/29/2007	21.9	275.2	17.54, 24.67	C	5229.2, 5592.4	3141.1, 3072.6	441.0, 404.9	C	C	C
292920098360601	AY-68-36-130	W5 replicate						58.4, 47.8	248.8, 229.7	28.0, 24.0	38.4, 39.4	31.9, 32.9	26.4, 27.9
292920098360601	AY-68-36-130	W5 replicate						53.3, 49.1, 63.1	223.7, 204.6, 406.1	47.3, 21.7, 26.1	38.9, 39.4, 37.4	33.4, 33.9, 22.4	22.4, 28.4, 26.9
292923098360301	AY-68-36-131	W6	11/28/2007	21.9	276.2	16.11, 21.46	C	4160.1, 3616.1, 4925.8	13871.5, 16355.0, 5380.3	455.5, 373.5, 404.6	C	C	C
292923098360301	AY-68-36-131	W6 replicate						4166.9, 1400.9, 5039.2	4762.2, 5377.8, 2641.0	383.4, 332.8, 371.7	C	C	C

Table 18. Summary of sulfur hexafluoride and chlorofluorocarbon age-tracer data and piston-flow model results for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. SF₆, sulfur hexafluoride; CFC, chlorofluorocarbon; °C, degrees Celsius; m, meter; pptv, parts per trillion by volume; C, contaminated]

USGS identification number	State well number	Local identifier and description	Sample date	Recharge temperature (°C)	Recharge elevation (m)	SF ₆ , atmospheric mixing ratio (pptv)*	SF ₆ , apparent age, (years)**	CFC-11, atmospheric mixing ratio (pptv)*	CFC-12, atmospheric mixing ratio (pptv)*	CFC-113, atmospheric mixing ratio (pptv)*	CFC-11 piston-flow apparent age (years)**	CFC-12 piston-flow apparent age (years)**	CFC-113 piston-flow apparent age (years)**
Well-field depth-dependent sampling (W4)													
292916098360701	AY-68-36-105	W4-shallow (380), ambient	11/16/2007	24.3	270.4	39.62, 40.24	C	3778.9, 3757.7	3596.5, 6780.4	549.3, 542.1	C	C	C
292916098360701	AY-68-36-105	W4-intermediate (430), ambient	11/14/2007	25.6	270.4	12.86, 9.52	C	3047.8, 2311.2, 1879.9, 3080.8	3508.5, 2477.3, 2076.0, 2626.5	527.0, 517.6, 444.0, 585.0	C	C	C
292916098360701	AY-68-36-105	W4-deep (540), ambient	11/15/2007	23.9	270.4	24.60, 32.54	C	4272.3, 2249.2, 3794.6	3045.7, 2169.3, 3277.3	464.7, 342.5, 497.4	C	C	C
292916098360701	AY-68-36-105	W4-shallow (380), moderate	11/16/2007	24.7	270.4	13.60, 12.38	C	4497.8, 4170.3, 4498.5	3687.3, 3110.6, 4210.5	420.3, 457.8, 420.2	C	C	C
292916098360701	AY-68-36-105	W4-intermediate (430), moderate	11/14/2007	24.4	270.4	41.21, 30.57	C	3049.0, 2437.7, 1257.2	2656.7, 2647.3, 2445.3	569.8, 263.6, 437.8	C	C	C
292916098360701	AY-68-36-105	W4-deep (540), moderate	11/15/2007	24.1	270.4	15.73, 11.08	C	3205.6, 3399.0, 3236.0	2480.6, 2558.7, 4272.7	595.1, 554.2, 537.4	C	C	C
292916098360701	AY-68-36-105	W4-shallow (380), normal	11/16/2007	25.1	270.4	23.62, 15.13	C	3394.5, 3032.0, 2284.9	3919.6, 3640.5, 3251.3	566.4, 553.3, 506.2	C	C	C
292916098360701	AY-68-36-105	W4-intermediate (430), normal	11/15/2007	23.6	270.4	14.81, 46.17	C	3031.0, 3176.5, 2919.2	3043.9, 2766.6, 2319.1	541.8, 551.4, 577.6	C	C	C
292916098360701	AY-68-36-105	W4-deep (540), normal	11/15/2007	23.9	270.4	25.91, 47.14	C	4341.1, 4033.8	3512.0, 3430.5	260.6, 261.8	C	C	C

* Corrected for excess air and recharge temperature.

** C, contaminated, age not resolvable (tracer concentration greater than would result from atmospheric sources, indicating contamination for the purposes of apparent-age determination).

Table 19. Summary of tritium/helium-3 age-tracer data and piston-flow model results for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

[By sample category. °C, degrees Celsius; m, meter; ³H, tritium; ³H₀, tritium naught; ³He, helium-3; ³He_{int}, helium derived from tritium decay; ⁴He, helium-4; ⁴He_{terr}, helium derived from terrigenic sources; cm³ STP/g H₂O, cubic centimeters per gram at standard temperature (20°C) and pressure (1 atmosphere); --, not determined]

USGS identification number	State well number	Local identifier and description	Sample date	Re- charge temper- ature (°C)	Re- charge eleva- tion (m)	³ H (tritium units)	³ H error (tritium units)	Delta ³ He (Δ ³ He), measured (uncor- rected) (percent)	Δ ³ He error (1 sigma) (percent)	Helium-4 (⁴ He), measured (cm ³ STP/g H ₂ O)	⁴ He error (cm ³ STP/g H ₂ O)	Excess ⁴ He (Δ ⁴ He) (percent)
Monitoring wells												
292943098354401	AY-68-36-135	Z-OVB	10/30/2007	28.3	258.5	1.30	0.048	-8.94	0.23	8.34937E-08	1.57E-10	96.8
292943098354402	AY-68-36-133	Z-SED	10/31/2007	27.1	258.5	2.17	.061	-.44	.18	1.06614E-07	3.58E-10	151
292943098354403	AY-68-36-134	Z-IED	10/30/2007	25.1	258.5	2.08	.056	-.21	.23	8.78542E-08	1.67E-10	105
292943098354404	AY-68-36-132	Z-DED	11/1/2007	26.8	258.5	2.13	.065	-1.88	.30	8.18674E-08	2.01E-10	92.3
292851098374401	AY-68-35-315	T-IED	5/6/2008	16.7	242.3	1.77	.185	.66	.17	1.39E-07	2.50E-10	217
292851098374402	AY-68-35-314	T-DED	5/7/2008	15.8	242.3	1.72	.185	-1.85	.23	9.49E-08	1.00E-10	115
Well-field wells (wellhead samples)												
292919098360501	AY-68-36-103	W2	11/29/2007	22	276.5	2.08	.060	-.04	.23	8.01809E-08	1.52E-10	86.3
292925098360201	AY-68-36-104	W3	11/28/2007	20.8	269.8	2.05	.057	-3.02	.23	8.19502E-08	1.54E-10	89.5
292916098360701	AY-68-36-105	W4	11/16/2007	24.1	270.4	1.88	.075	.09	.29	7.8893E-08	1.87E-10	84.2
292920098360601	AY-68-36-130	W5	11/29/2007	21.9	275.2	2.18	.070	3.87	.25	4.37746E-08	8.10E-11	1.60
292923098360301	AY-68-36-131	W6	11/28/2007	21.9	276.2	1.91	.058	-1.03	.19	8.25234E-08	2.72E-10	91.6
Well-field depth-dependent sampling (W4)												
292916098360701	AY-68-36-105	W4-shallow (380), ambient	11/16/2007	24.3	270.4	<1	--	-.40	.30	7.99629E-08	1.97E-10	86.8
292916098360701	AY-68-36-105	W4-intermediate (430), ambient	11/14/2007	25.6	270.4	2.20	.074	-1.30	.19	7.96141E-08	2.61E-10	86.7
292916098360701	AY-68-36-105	W4-deep (540), ambient	11/15/2007	23.9	270.4	1.97	.046	-.17	.30	8.04951E-08	1.99E-10	87.9
292916098360701	AY-68-36-105	W4-shallow (380), moderate	11/16/2007	24.7	270.4	1.82	.066	-.93	.19	7.81165E-08	2.62E-10	82.7
292916098360701	AY-68-36-105	W4-intermediate (430), moderate	11/14/2007	24.4	270.4	1.98	.161	-.06	.19	8.28429E-08	2.81E-10	93.6
292916098360701	AY-68-36-105	W4-deep (540), moderate	11/15/2007	24.1	270.4	1.63	.073	-.77	.19	8.34208E-08	2.80E-10	94.8
292916098360701	AY-68-36-105	W4-shallow (380), normal	11/16/2007	25.1	270.4	2.23	.071	-.54	.29	8.14245E-08	1.94E-10	90.7
292916098360701	AY-68-36-105	W4-intermediate (430), normal	11/15/2007	23.6	270.4	1.78	.053	67.81	1.49	5.61E-11	7.57E-12	-99.9
292916098360701	AY-68-36-105	W4-deep (540), normal	11/15/2007	23.9	270.4	1.90	.058	3.58	.21	4.58499E-08	1.53E-10	7.00

Table 19. Summary of tritium/helium-3 age-tracer data and piston-flow model results for groundwater samples collected for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.—Continued

[By sample category. °C, degrees Celsius; m, meter; ³H, tritium; ³H₀, tritium naught; ³He, helium-3; ³He_{trit}, helium derived from tritium decay; ⁴He, helium-4; ⁴He_{terr}, helium derived from terrigenic sources; cm³ STP/g H₂O, cubic centimeters per gram at standard temperature (20°C) and pressure (1 atmosphere); --, not determined]

USGS identification number	Excess Neon (ΔNe) (per- cent)	Tritio- genic ³ He (³ He _{trit}) (tritium units)	³ He _{trit} error (tritium units)	Terrigenic helium-4 (⁴ He _{terr}) (cm ³ STP/g H ₂ O)	⁴ He _{terr} (Δ ⁴ He _{terr}), percent of measured helium	“Tritium naught” (³ H ₀) (³ H+ ³ He _{trit}) (tritium units)	³ H/ ³ H ₀	³ H/ ³ He apparent age, years (piston- flow model)	³ H/ ³ He appar- ent- age error (years)	Comment
Monitoring wells										
292943098354401	79.0	-2.38**	0.16	2.5942E ⁻⁰⁹	3.11	--	--	--	--	Age not resolvable because of low ³ H
292943098354402	101	8.03	.24	1.4519E ⁻⁰⁸	13.6	10.2	0.21	27.5	0.6	
292943098354403	84.5	1.90	.18	3.009E ⁻⁰⁹	3.43	3.97	.52	11.5	.8	
292943098354404	77.4	0.16	.64	1.222E ⁻⁰⁹	1.49	2.28	.93	1.3	5.0	
292851098374401	126	16.3	1.22	2.822E ⁻⁰⁸	20.3	18.1	.10	41.3	2.1	
292851098374402	63.9	8.43	.61	1.656E ⁻⁰⁸	17.5	10.1	.17	31.6	1.9	
Well-field wells (wellhead samples)										
292919098360501	61.1	3.71	.16	6.1193E ⁻⁰⁹	7.63	5.79	.36	18.2	.6	
292925098360201	58.6	3.73	.16	8.6435E ⁻⁰⁹	10.5	5.79	.35	18.4	.6	
292916098360701	73.1	.12	.64	-5.1124E ⁻¹⁰	-.65**	2.00	.94	3.4*	1.0*	Age calculation uncorrected for Δ ⁴ He _{terr} because of negative Δ ⁴ He _{terr} value
292920098360601	-26.3	9.04	.10	1.4099E ⁻⁰⁸	32.2	11.2	.19	--	--	Age not resolvable because of negative ΔNe value
292923098360301	66.7	2.95	.18	5.5689E ⁻⁰⁹	6.75	4.86	.39	16.6	.8	
Well-field depth-dependent sampling (W4)										
292916098360701	70.6	1.23	.63	1.8748E ⁻⁰⁹	2.34	--	--	--	--	Age not resolvable because of low ³ H
292916098360701	65.2	2.33	.18	4.6682E ⁻⁰⁹	5.86	4.53	.48	12.9	.8	
292916098360701	68.1	2.24	.62	3.5324E ⁻⁰⁹	4.39	4.21	.47	13.5	2.6	
292916098360701	65.0	1.58	.17	2.9632E ⁻⁰⁹	3.79	3.40	.53	11.1	1.0	
292916098360701	64.1	4.73	.19	8.0402E ⁻⁰⁹	9.71	6.72	.30	21.7	1.1	
292916098360701	68.9	3.36	.19	6.1269E ⁻⁰⁹	7.34	4.99	.33	19.9	.9	
292916098360701	80.5	-.61**	.67	-1.3155E ⁻⁰⁹	-1.62**	2.23	1.00	.8*	1.0*	Age calculation uncorrected for ⁴ He _{terr} because of negative ⁴ He _{terr} value
292916098360701	--	4.44	.03	7.4103E ⁻⁰⁹	13,212	6.23	.29	--	--	Age not resolvable because of negative Δ ⁴ He value and because of ΔNe value
292916098360701	-21.1	8.70	.12	1.3561E ⁻⁰⁸	29.6	10.6	.18	--	--	Age not resolvable because of negative Δ ⁴ He value and because of low ΔNe value

* Age calculation not corrected for ⁴He_{terr} because of negative ⁴He_{terr} value.

** Negative values result from errors associated with measurements values near 0.

Table 20. Results of forecasting analysis for the study of the transport of anthropogenic and natural contaminants (TANC) to public supply wells in the San Antonio segment of the Edwards aquifer near San Antonio, south-central Texas, 2004–9.

Forecast model scenario	Peak concentration (or lag time, in years) relative to time zero (start of input)	Dilution at peak concentration (concentration at peak/concentration of input at time zero (1))	Arrival of 1 percent of total mass (years)	Flush of 99 percent of total mass (years) relative to year of end of input
Uniform input across area contributing recharge to well, no degradation	30	0.98	1	32
Uniform input across area contributing recharge to well, first-order reaction kinetics degradation rate constant = 0.1/year	30	.73	1	12

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