

Prepared in cooperation with McHenry County, Illinois

Water Quality of Sand and Gravel Aquifers in McHenry County, Illinois, 2020 and Comparisons to Conditions in 2010



Scientific Investigations Report 2022–5110



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By Amy M. Gahala, Lance R. Gruhn, Jennifer C. Murphy, and Lisa A. Matson
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Contents

Acknow	rledgments	iii
Abstrac	t	1
Introduc	ction	1
Pu	rpose and Scope	3
De	scription of Study Area	3
Pre	evious Water-Quality Investigations	8
Method	s of Study	8
Wa	ater-Quality Sampling	8
Sta	itistical Analysis	9
2020 Wa	ater Quality	9
Fie	ld Properties	10
Ma	ijor lons	12
Tra	ice Metals	15
Nu	trients	17
Co	ntaminants of Emerging Concern	17
	itability of Water for Drinking	
	ison to 2010 Water-Quality Results	
•	al Comparison	
	ison of Chloride to Bromide Ratios	
	гу	
	ces Cited	
	ix 1. Quality Assurance and Quality Control of Water-Quality Results	
Figure	es ·	
1.	Map showing groundwater monitoring wells of the McHenry County Groundwater Monitoring Network and U.S. Geological Survey National Water-Quality Assessment Project and three surface-water-quality monitoring sites sampled in June–July 2020 in McHenry County, Illinois	2
2.	Map showing distribution of sodium concentrations in samples from the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment Project monitoring wells, McHenry County, Illinois, 2020	
3.	Graph showing dissolved solids concentrations by well depth in the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment Project monitoring wells, McHenry County, Illinois 2020	17
4.	Map showing distribution of dissolved solids concentrations in the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment Project monitoring wells, McHenry County, Illinois, 2020	18
5.	Map showing distribution of chloride concentrations in the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment Project monitoring wells, McHenry County, Illinois, 2020	
6.	Map showing distribution of arsenic concentrations in the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment monitoring wells, McHenry County, Illinois, 2020	

7.	Map showing distribution of iron concentrations in the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment monitoring wells, McHenry County, Illinois, 2020	25
8.	Map showing distribution of manganese concentrations in the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment monitoring wells, McHenry County, Illinois, 2020	
9.	Map showing number of detections for pharmaceuticals, pesticides, and wastewater indicator compounds in the McHenry County Groundwater Monitoring Network wells and four National Water-Quality Assessment monitoring wells, McHenry County, Illinois, 2020	32
10.	Graphs showing differences in the shallow, intermediate, and deep parts of the sand and gravel aquifers for concentrations of selected constituents between 2010 and 2020 for water-quality samples in McHenry County, Illinois	
11.	Diagrams showing major ions in wells from 2010 and 2020 in McHenry County,	37
12.	Graph showing chloride-bromide ratio plotted against chloride concentrations in the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment monitoring wells, McHenry County, Illinois, 2020	
13.	Graph showing chloride-bromide ratio plotted against chloride concentration for wells that show shifts between 2010 and 2020 in the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment monitoring wells, McHenry County, Illinois, 2020	40
Tables		
1.	Site information and water-quality sampling plan for field properties, major ions, trace metals, nutrients, and contaminants of emerging concern analyzed in the McHenry County Groundwater Monitoring Network, U.S. Geological Survey National Water-Quality Assessment monitoring wells, and three surface-water locations in Illinois, 2020	4
2.	Water use by category, volume of groundwater withdrawals, and population served, McHenry County, Illinois, 2010 and 2015	
3.	U.S. Environmental Protection Agency water-quality thresholds, number of wells exceeding water-quality thresholds in samples collected in 2020, and statistical description	
4.	Results for field properties for the McHenry County Groundwater Monitoring Network wells, 4 National Water-Quality Assessment Project monitoring wells, and 3 surface-water sites, McHenry County, Illinois, 2020	11
5.	Results for concentrations of major ions from samples collected from the McHenry County Groundwater Monitoring Network wells, 4 National Water-Quality Assessment wells, and 3 surface-water sites in McHenry County, Illinois, 2020	
6.	Results for concentrations of trace metals from samples collected from the McHenry County Groundwater Monitoring Network wells, 4 National Water-Quality Assessment wells, and 2 surface-water sites in McHenry	

7.	Results for concentrations of nutrients from samples collected from the McHenry County Groundwater Monitoring Network wells, 4 National Water-Quality Assessment wells, and 2 surface-water sites in McHenry County, Illinois, 2020	27
8.	Concentrations of pharmaceuticals and wastewater indicator compounds detected in samples collected from selected wells within the McHenry County Groundwater Monitoring Network and three surface-water sites in McHenry County, Illinois, 2020	30
9.	Summary statistics for selected constituents, results of the Wilcoxon signed-rank test for groundwater-quality changes between 2010 and 2020 for uncensored data and the paired Prentice-Wilcoxon test for uncensored data, and evaluation of differences between the 2010 and 2020 aquifer-depth groups of shallow, intermediate, and deep parts of the sand and gravel aquifers in McHenry County, Illinois	34
	Monething County, Illinois	04

Conversion Factors

U.S. customary units to International System of Units

Multiply	Ву	To obtain		
	Length			
inch (in.)	2.54	centimeter (cm)		
inch (in.)	25.4	millimeter (mm)		
foot (ft)	0.3048	meter (m)		
mile (mi)	1.609	kilometer (km)		
	Area			
square mile (mi ²)	259.0	hectare (ha)		
square mile (mi ²)	2.590	square kilometer (km²)		

International System of Units to U.S. customary units

Multiply	Ву	To obtain
	Mass	
milligram (mg)	3.527×10 ⁻⁵	ounce, avoirdupois (oz)

Temperature in degrees Celsius (°C) may be converted to degrees Fahrenheit (°F) as follows:

$$^{\circ}F = (1.8 \times ^{\circ}C) + 32.$$

Datum

Vertical coordinate information is referenced to the North American Vertical Datum of 1988 (NAVD 88).

Horizontal coordinate information is referenced to the North American Datum of 1983 (NAD 83).

Elevation, as used in this report, refers to distance above the vertical datum.

Supplemental Information

Specific conductance is given in microsiemens per centimeter at 25 degrees Celsius (μ S/cm at 25 °C).

Concentrations of chemical constituents in water are given in either milligrams per liter (mg/L), micrograms per liter (μ g/L), or nanograms per liter (μ g/L).

Turbidity is given in nephelometric turbidity units (NTU).

Abbreviations

CEC contaminant of emerging concern

DEET N,N-Diethyl-*m*-toluamide
DWA drinking water advisory

EPA U.S. Environmental Protection Agency

MCGMN McHenry County Groundwater Monitoring Network

MCL maximum contaminant level

NAWQA National Water-Quality Assessment NWQL National Water Quality Laboratory

p probability

PPW paired Prentice-Wilcoxon [test]

QAQC quality assurance and quality control SMCL secondary maximum contaminant level

USGS U.S. Geological Survey

WIC wastewater indicator compound

WSR Wilcoxon signed rank [test]

< less than

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Abstract

McHenry County, Illinois, obtains most of its drinking water from shallow sand and gravel aquifers (groundwater). To evaluate this groundwater resource, the U.S. Geological Survey, in cooperation with McHenry County, Illinois, collected water-quality samples from 41 of 42 monitoring wells in the McHenry County Groundwater Monitoring Network and 4 monitoring wells from the U.S. Geological Survey National Water-Quality Assessment Project. Additionally, a subset of 12 monitoring wells was sampled and analyzed for pharmaceuticals and wastewater indicator compounds (WICs), collectively referred to as "contaminants of emerging concern" (CECs). Results from this 2020 study were compared to the 2010 results to assess changes in groundwater quality. Statistical analyses and chloride-bromide ratio analyses also were completed to assess changes in water quality.

Health-based benchmarks were exceeded for arsenic (about 24 percent; 11 of 45 monitoring wells), sodium (40 percent, 18 of 45), and manganese (about 2 percent, 1 of 45). Aesthetically based benchmarks were exceeded for dissolved solids (about 29 percent, 13 of 45), chloride (about 4 percent, 2 of 45), iron (about 87 percent, 39 of 45), and manganese (about 29 percent, 13 of 45). CECs were detected at low or estimated concentrations in 8 of the 12 (about 67 percent) monitoring wells analyzed.

In addition to sampling the groundwater monitoring wells, three surface-water-quality monitoring sites also were sampled and analyzed for pharmaceuticals and WICs to provide a preliminary assessment of the presence of CECs in the surface waters. CECs were detected in all three of the surface-water-quality monitoring samples collected, and WICs were more prevalent and more frequently detected than pharmaceutical compounds. These results provided a cursory understanding of the presence of CECs in surface waters and do not constitute a robust analysis of sources, seasonality, range of concentrations, persistence, or effects.

The 2020 groundwater-quality results had measurements of field properties, and concentrations of major ions, trace metals, and nutrients that were consistent with 2010 results with statistically significant increases for calcium, magnesium,

and silica, and decreases for aluminum, ammonia, arsenic, barium, bromide, calcium, molybdenum, phosphate, specific conductance, sulfate, and dissolved solids. Increases generally were detected in the intermediate and deep parts of the sand and gravel aquifer, and decreases were detected in the shallow parts of the sand and gravel aquifer. The mixed distribution of increases and decreases among the various constituents and aquifer-depth groups could be reflecting dissolution and mobility of some of the redox sensitive constituents and dilution of some constituents in the shallow aquifer depths. These changes may be attributed to a combination of stable population of the past decade (2010–20), land-use management practices, and the recent wet years of 2017 through 2019 causing a dilution of the major ions in the shallow parts of the aquifer.

Introduction

McHenry County, Illinois, sources their drinking water primarily from the shallow (less than 300 feet [ft]) sand and gravel aquifers (groundwater). To evaluate this groundwater resource, a network of 42 groundwater monitoring wells (referred to as the "McHenry County Groundwater Monitoring Network" [MCGMN]) is used to monitor the water levels and water quality in the sand and gravel aquifers (fig. 1). In 2010, water-quality samples were obtained at 41 monitoring wells from the MCGMN, plus 5 additional monitoring wells available from the U.S. Geological Survey (USGS) National Water-Quality Assessment (NAWQA) Project were sampled and analyzed for field properties, major ions, trace metals, and nutrients. In 2010, samples from a subset of the monitoring wells also were analyzed for volatile organic compounds, pesticides, and herbicides. The results of that study are presented in Gahala (2017). In short, the findings of that report indicated health-based standards were exceeded for arsenic in 22 percent of the monitoring wells sampled, sodium in 37 percent of the monitoring wells sampled, and nitrates in 2 percent of the monitoring wells sampled. Aesthetically based standards were exceeded for total dissolved solids in 33 percent of the wells sampled, chloride in 4 percent of the wells sampled, iron in 85 percent of the wells sampled, and manganese in

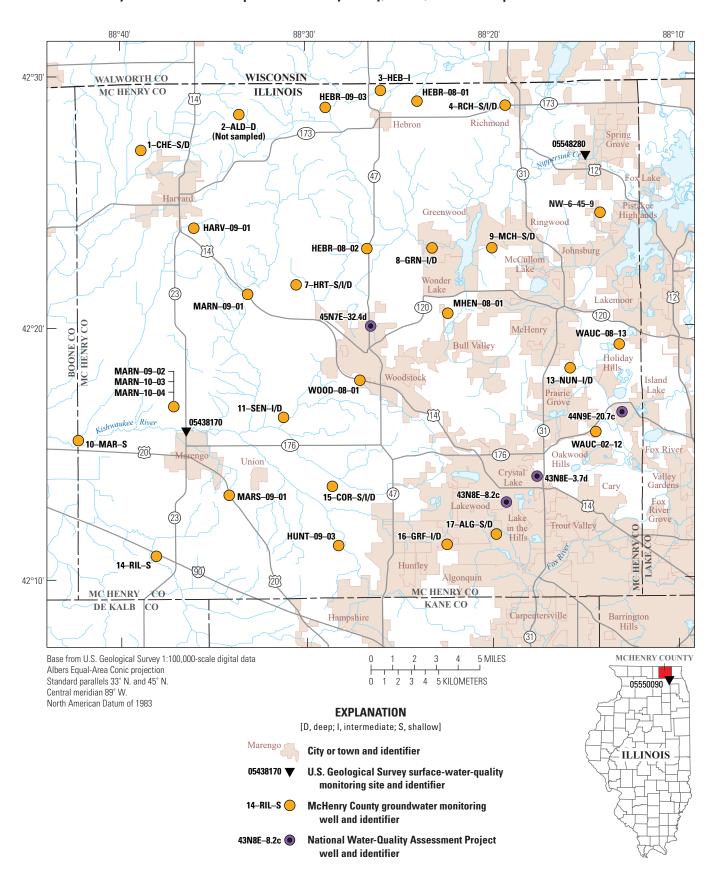


Figure 1. Groundwater monitoring wells of the McHenry County Groundwater Monitoring Network and U.S. Geological Survey National Water-Quality Assessment Project and three surface-water-quality monitoring sites sampled in June–July 2020 in McHenry County, Illinois.

30 percent of the wells sampled. Volatile organic compounds and pesticides were detected at trace levels in shallow wells (less than 46 ft below land surface) near urban settings. To identify changes in groundwater conditions in the shallow sand and gravel aquifer since 2010, the USGS, in cooperation with McHenry County, Illinois, and in collaboration with the McHenry County Planning and Development Committee, completed a decadal water-quality assessment of the sand and gravel aquifers in June–July of 2020.

The MCGMN monitoring wells, along with four USGS NAWQA Project shallow (less than about 46 ft) monitoring wells in urban areas, were sampled and analyzed for field properties, major ions, trace metals, and nutrients (table 1). Similar to the 2010 water-quality assessment, targeted monitoring wells (12 total) were sampled for contaminants of emerging concern (CECs). CECs can refer to many kinds of chemicals, including medicines, personal care products, cleaning products, and lawn care and agricultural products (U.S. Geological Survey, 2022). These CECs can enter our groundwaters and surface waters through runoff, septic systems, discharges from wastewater-treatment plants, and various industrial processes (U.S. Geological Survey, 2022). To identify if any CECs are present in the sand and gravel aquifers, pharmaceuticals and wastewater indicator compounds (WICs) were selected for analysis in the groundwater at targeted monitoring wells. In addition to the groundwater monitoring wells, three surface-water sites also were sampled and analyzed for CECs, and two of the sites included analyses for major ions, trace metals, and nutrients (table 1) to assess general stream-water quality downstream from wastewatertreatment plants and industries. Surface waters are the primary drainage ways for groundwater, stormwater runoff, and other land-use discharges. This preliminary assessment of the presence of CECs provides a cursory understanding of concentrations detected but does not constitute a robust analysis of sources, seasonality, range of concentrations, persistence, or effects in surface water.

Purpose and Scope

The primary objectives of this study were to assess changes in the water quality as it pertains to human health and the environment and to identify potential effects of land use on water quality. This investigation provides an update on the water quality of the sand and gravel aquifers from the countywide network of groundwater monitoring wells and selected surface-water sites. The purpose of this report is to summarize and describe the findings from the analysis of water-quality samples collected during June and July of 2020 from the MCGMN, plus four NAWQA monitoring wells and three surface-water sites, and to complete the following:

• Describe the status of groundwater quality in 2020.

- Identify statistically significant differences in groundwater quality by applying the Wilcoxon signed-rank (WSR) test and paired Prentice-Wilcoxon (PPW) test.
- Provide an update on the sources of chloride (chloridebromide analysis) and compare results to the 2010 data.
- Summarize the results of the water-quality analysis for CECs (pharmaceuticals and WICs) identified in samples from targeted groundwater monitoring sites and surface-water locations.
- Identify exceedances of drinking water standards and statistically significant differences in concentrations and potentially relate these results to land-use or other effects near the well.

Description of Study Area

McHenry County, Ill. (611 square miles), is 35 miles northwest of the Chicago metropolitan area. As described in Gahala (2017), the county is generally geographically split by type of land use. The western half of the county is dominated by agricultural land use, and the eastern half is primarily urban land use. Geologically, the land is described as hilly with moraines, kames, eskers, and outwash plains resulting from several glacial advances and retreats during the Quaternary Period (Hackett and McComas, 1969; Berg and others, 1997; Stiff and Hansel, 2004). McHenry County is divided by two hydrologic drainage basins, the Fox River drainage basin, which drains towards the east, and the Kishwaukee River drainage basin, which drains towards the west (Nicholas and Krohelski, 1984; Berg and others, 1997; Illinois State Water Survey, 2022; Gahala, 2017, fig. 3).

Precipitation ranged from 27.07 inches (in.) in 2012 to 49.9 in. in 2018, with an average of 44.51 in. over the decade (2010-20) according to the Global Historical Climate Network observation in Woodstock, Ill. (Woodstock 5NW) (National Centers for Environmental Information, 2022a). The legacy climatological period from 1981 to 2010 from weather observation station Woodstock 5NW, in McHenry County, Ill., indicates an average annual precipitation of 36.62 in. (National Centers for Environmental Information, 2022b). The most recent standard climatological period (1991–2020) indicates an average annual precipitation of 38.68 in. (National Centers for Environmental Information, 2022c), and the latest 15-year period from 2006 to 2020 indicates an average annual precipitation of 41.18 in. (National Centers for Environmental Information, 2022d). The average annual precipitation, calculated from the daily values from 2010 to 2020 from data downloaded for the observation station Woodstock 5NW is slightly greater than the 2006–20 average annual precipitation (44.51 in.) for the decade. Increases in precipitation can affect water quality by increasing the amount of runoff and infiltration of various land-use inputs through mobilization of

4 Water Quality of Sand and Gravel Aquifers in McHenry County, Illinois, 2020 and Comparisons to Conditions in 2010

Table 1. Site information and water-quality sampling plan for field properties, major ions, trace metals, nutrients, and contaminants of emerging concern analyzed in the McHenry County Groundwater Monitoring Network, U.S. Geological Survey National Water-Quality Assessment monitoring wells, and three surface-water locations in Illinois, 2020.

[Data collected for sites in this table can be accessed in the U.S. Geological Survey (USGS) National Water Information System database (U.S. Geological Survey, 2020b) using the USGS station identifiers. ft, foot; NAVD 88, North American Vertical Datum of 1988; X, analyzed; —, no detection; Co, County; CRN, Climate Response Network; IL, Illinois]

USGS station name	USGS station identifier	Field identifier	Latitude	Longitude	Land-surface elevation (ft above NAVD 88)	Well depth (ft)
43N5E-27.4h1 (14-RIL-S)	421056088380801	14-RIL-S	42.182	-88.636	806.75	20.4
43N7E-19.8d (HUNT-09-03)	421120088281801	HUNT-09-03	42.189	-88.472	878	150.7
43N7E-23.1d1 (16-GRF-I)	421122088222701	16-GRF-I	42.189	-88.374	879.51	99
43N7E-23.1d2 (16-GRF-D)	421122088222702	16-GRF-D	42.189	-88.374	879.51	139.1
43N8E-20.6h1 (17-ALG-S)	421145088194801	17-ALG-S	42.196	-88.330	880.03	47.3
43N8E-20.6h2 (17-ALG-D)	421145088194802	17-ALG-D	42.196	-88.330	880.03	187.8
43N6E-07.1g (MARS-09-01)	421321088341101	MARS-09-01	42.223	-88.570	928	190.3
43N6E-01.3b1 (15-COR-S)	421341088283701	15-COR-S	42.228	-88.477	851.23	55.1
43N6E-01.3b2 (15-COR-I)	421341088283702	15-COR-I	42.228	-88.477	851.48	103.3
43N6E-01.3b3 (15-COR-D)	421341088283703	15-COR-D	42.228	-88.477	851.23	116.1
44N5E-30.8c1 (10-MAR-S) (McHenry Co CRN)	421533088421801	10-MAR-S	42.259	-88.705	780.93	20.3
44N9E-25.1d (WAUC-02-12)	421547088142301	WAUC-02-12	42.263	-88.240	835	192.3
44N6E-22.4c1 (11-SEN-I)	421626088311401	11-SEN-I	42.274	-88.521	830.63	75.4
44N6E-22.4c2 (11-SEN-D)	421626088311402	11-SEN-D	42.274	-88.521	830.63	153.2
44N5E-23.5g (MARN-09-02)	421653088370901	MARN-09-02	42.281	-88.619	827	110.6
44N7E-17.8h1 (WOOD-08-01)	421747088270701	WOOD-08-01	42.296	-88.452	943	202.3
44N8E-11.3d1 (13-NUN-I)	421820088154501	13-NUN-I	42.306	-88.263	785.32	113
44N8E-11.3d2 (13-NUN-D)	421820088154502	13-NUN-D	42.306	-88.263	785.32	152.2
44N9E-05.7d1 (WAUC-08-13)	421914088125301	WAUC-08-13	42.321	-88.215	766	105.3
45N7E-25.7a (MHEN-08-01)	422032088222001	MHEN-08-01	42.342	-88.372	860	103.3
45N6E-29.1h (MARN-09-01)	422120088330901	MARN-09-01	42.356	-88.552	909	100.7
45N6E-23.7d1 (7-HRT-S)	422142088303101	7-HRT-S	42.362	-88.509	924.14	62.3
45N6E-23.7d2 (7-HRT-I)	422142088303102	7-HRT-I	42.362	-88.509	924.14	114.9
45N6E-23.7d3 (7-HRT-D)	422142088303103	7-HRT-D	42.362	-88.509	924.14	165.7
45N8E-17.7h1 (9-MCH-S)	422308088195601	9-MCH-S	42.386	-88.332	863.18	25.9
45N8E-17.7h2 (9-MCH-D)	422308088195602	9-MCH-D	42.386	-88.332	863.18	180
45N7E-11.5a1 (8-GRN-I)	422308088231001	8-GRN-I	42.386	-88.386	855.97	70.3
45N7E-11.5a2 (8-GRN-D)	422308088231002	8-GRN-D	42.386	-88.386	855.97	153.1
45N7E-08.5a (HEBR-08-02)	422308088264201	HEBR-08-02	42.385	-88.445	898	100.3
45N5E-12.5h (HARV-09-01)	422358088360201	HARV-09-01	42.400	-88.601	942	120.1
45N9E-06.7e (NW-6-45-9)	422433088140601	NW-6-45-9	42.409	-88.235	855	73
46N5E-21.1d1 (1-CHE-S)	422704088385301	1-CHE-S	42.451	-88.648	896.24	40.3
46N5E-21.1d2 (1-CHE-D)	422704088385302	1-CHE-D	42.451	-88.648	896.24	110.8
46N6E-12.5d (HEBR-09-03)	422845088285401	HEBR-09-03	42.480	-88.481	949	120.6
46N8E-08.2e1 (4-RCH-S)	422848088191001	4-RCH-S	42.480	-88.320	844.15	24
46N8E-08.2e2 (4-RCH-I)	422848088191002	4–RCH–I	42.480	-88.320	844.15	98.3
46N8E-08.2e3 (4-RCH-D)	422848088191003	4-RCH-D	42.480	-88.320	844.15	176
46N7E-10.2f (HEBR-08-01)	422858088235601	HEBR-08-01	42.483	-88.399	898	145.3
46N7E-04.8b1 (3-HEB-I)	422925088255401	3-HEB-I	42.490	-88.432	867.93	66.3
44N5E-23.5g2 (MARN-10-03)	421653088370902	MARN-10-03	42.281	-88.619	828	160

Table 1. Site information and water-quality sampling plan for field properties, major ions, trace metals, nutrients, and contaminants of emerging concern analyzed in the McHenry County Groundwater Monitoring Network, U.S. Geological Survey National Water-Quality Assessment monitoring wells, and three surface-water locations in Illinois, 2020.—Continued

[Data collected for sites in this table can be accessed in the U.S. Geological Survey (USGS) National Water Information System database (U.S. Geological Survey, 2020b) using the USGS station identifiers. ft, foot; NAVD 88, North American Vertical Datum of 1988; X, analyzed; —, no detection; Co, County; CRN, Climate Response Network; IL, Illinois]

verage ground-		Water-quality analytes					
water level for period of record (ft below land surface)	Aquifer unit ¹	Field properties	Major ions	Trace metals	Nutrients	Contaminants o	
7.78	Haeger-Beverly Unit	X	X	X	X	_	
26.39	Ashmore Sand Unit	X	X	X	X	_	
18.91	Ashmore Sand Unit	X	X	X	X	_	
21.27	Lower Glasford Sand Unit	X	X	X	X	_	
0.66	Yorkville-Batestown Unit	X	X	X	X	_	
99.32	Shallow Bedrock Aquifer ²	X	X	X	X	_	
72.15	Upper Glasford Sand Unit	X	X	X	X	_	
9.08	Haeger-Beverly Unit	X	X	X	X	_	
9.64	Ashmore Sand Unit	X	X	X	X	X	
9.08	Ashmore Sand Unit	X	X	X	X	_	
4.96	Haeger-Beverly Unit	X	X	X	X	X	
97.27	Lower Glasford Sand Unit	X	X	X	X	_	
3.85	Ashmore Sand Unit	X	X	X	X	_	
5.11	Lower Glasford Sand Unit	X	X	X	X	_	
18.27	Upper Glasford Sand Unit	X	X	X	X	_	
78.11	Ashmore Sand Unit	X	X	X	X	X	
47.28	Yorkville-Batestown Unit	X	X	X	X	_	
47.38	Shallow Bedrock Aquifer and Lower Glasford Sand Unit	X	X	X	X	_	
20.93	Haeger-Beverly Unit	X	X	X	X	_	
34.01	Haeger-Beverly Unit	X	X	X	X	_	
32.95	Upper Glasford Sand Unit	X	X	X	X	X	
35.83	Haeger-Beverly Unit	X	X	X	X	X	
36.10	Upper Glasford Sand Unit	X	X	X	X	_	
37.61	Shallow Bedrock Aquifer and Lower Glasford Sand Unit	X	X	X	X	_	
10.53	Haeger-Beverly Unit	X	X	X	X	X	
53.96	Ashmore Sand Unit	X	X	X	X	_	
4.99	Haeger-Beverly Unit	X	X	X	X	X	
16.95	Shallow Bedrock Aquifer and Lower Glasford Sand Unit	X	X	X	X	X	
11.04	Haeger-Beverly Unit	X	X	X	X	X	
31.81	Ashmore Sand Unit	X	X	X	X	X	
31.27	Haeger-Beverly Unit	X	X	X	X	_	
7.40	Ashmore Sand Unit	X	X	X	X	_	
7.34	Shallow Bedrock Aquifer ²	X	X	X	X	_	
24.54	Ashmore Sand Unit	X	X	X	X	X	
7.13	Haeger-Beverly Unit	X	X	X	X	X	
11.70	Upper Glasford Sand Unit	X	X	X	X	_	
11.00	Shallow Bedrock Aquifer and Lower Glasford Sand Unit	X	X	X	X	_	
27.85	Ashmore Sand Unit	X	X	X	X	_	
-12.95	Upper Glasford Sand Unit	X	X	X	X	_	
27.82	Lower Glasford Sand Unit	X	X	X	X		

Table 1. Site information and water-quality sampling plan for field properties, major ions, trace metals, nutrients, and contaminants of emerging concern analyzed in the McHenry County Groundwater Monitoring Network, U.S. Geological Survey National Water-Quality Assessment monitoring wells, and three surface-water locations in Illinois, 2020.—Continued

[Data collected for sites in this table can be accessed in the U.S. Geological Survey (USGS) National Water Information System database (U.S. Geological Survey, 2020b) using the USGS station identifiers. ft, foot; NAVD 88, North American Vertical Datum of 1988; X, analyzed; —, no detection; Co, County; CRN, Climate Response Network; IL, Illinois]

USGS station name	USGS station identifier	Field identifier	Latitude	Longitude	Land-surface elevation (ft above NAVD 88)	Well depth (ft)
44N5E-23.5g3 (MARN-10-04)	421653088370903	MARN-10-04	42.281	-88.619	827	82
46N6E-08.46l (2-ALD-D)	422828088333301	2-ALD-D	42.474	-88.559	1,176.35	344.4
43N8E-8.2c	421301088191501	43N8E-8.2c	42.217	-88.321	900	46.1
45N7E-32.4d	422002088263001	45N7E-32.4d	42.334	-88.442	900	30.4
43N8E-3.7d	421402088173501	43N8E-3.7d	42.234	-88.293	920	58.7
44N9E-20.7c	421633088125801	44N9E-20.7c	42.276	-88.216	750	26.1
		Surface-water site	es			
Nippersink Creek Near Spring Grove, IL	05548280	Nippersink Creek	42.443	-88.248	_	_
Fox River at Huntley Road at Carpentersville, IL	05550090	Fox River	42.108	-88.290	_	_
Kishwaukee River at Marengo, IL	05438170	Kishwaukee River	42.265	-88.608	_	_

constituents into nearby surface waters (Hatfield and Prueger, 2004) and potentially imparting geochemical changes and transport within an aquifer.

Aguifers used to supply drinking water in McHenry County were mapped and described in detail by Meyer and others (2013). Drinking water supplies for McHenry County are pumped from deeper bedrock aquifers such as the Cambrian-Ordovician, Silurian dolomite, and the shallow sand and gravel aquifers within glacial drift deposits (Woller and Sanderson, 1976). Directions of groundwater flow were determined by Meyer and others (2013) and Gahala (2017) and are briefly summarized in this section. Groundwater is replenished (recharge) from precipitation that falls onto the land surface and infiltrates through the soil and sediment to the water table, where it flows downward and laterally through hydraulically connected sediments over time. Some of the groundwater flows laterally and discharges into local surface-water bodies. Deeper parts of the sand and gravel aquifers that are overlain by a low-permeability clay (till) can receive recharge through downward transport through saturated clays or through preferential, permeable flow paths made by thin or discontinuous clay deposits.

This study focuses on the water quality within the shallow sand and gravel aquifers. The aquifer units within the shallow sand and gravel aquifers that the monitoring wells represent are described in detail in Meyer and others (2013) and Gahala (2017) (table 1). Two wells are screened in the upper 10 ft of the Shallow Bedrock Aquifer. The source of the water is recharged from the sand and gravel aquifers.

The MCGMN and NAWQA wells are used to monitor for the effects of various land uses and changes. Samples collected from the MCGMN and NAWQA wells give a snapshot of water quality at shallow, intermediate, and deep depths. The depths of the MCGMN and NAQWA monitoring wells range from 20.3 to 344.4 ft and are generally comparable to the depth range of production and private (residential) wells extracting drinking water from the shallow sand and gravel aquifers. According to the Illinois State Water Survey, estimates from the ILWATER database indicate about 58 percent of production wells are extracting groundwater from the sand and gravel aquifers at depths of 16 to 272 ft. About 69 percent of private wells are extracting groundwater from sand and gravel aquifers at depths from 8 to 372 ft (Daniel Abrams, Illinois State Water Survey, written commun., March 23, 2022). It is important to note that these estimates from the

Table 1. Site information and water-quality sampling plan for field properties, major ions, trace metals, nutrients, and contaminants of emerging concern analyzed in the McHenry County Groundwater Monitoring Network, U.S. Geological Survey National Water-Quality Assessment monitoring wells, and three surface-water locations in Illinois, 2020.—Continued

[Data collected for sites in this table can be accessed in the U.S. Geological Survey (USGS) National Water Information System database (U.S. Geological Survey, 2020b) using the USGS station identifiers. ft, foot; NAVD 88, North American Vertical Datum of 1988; X, analyzed; —, no detection; Co, County; CRN, Climate Response Network; IL, Illinois]

Average ground-			Wa	ter-quality analyte	-quality analytes		
water level for period of record (ft below land surface)	Aquifer unit ¹	Field properties	Major ions	Trace metals	Nutrients	Contaminants of emerging concern	
19.43	Upper Glasford Sand Unit	X	X	X	X	_	
218.31	Lower Glasford Sand Unit	_	_	_	_	_	
_	Haeger-Beverly Unit	X	X	X	X	_	
_	Haeger-Beverly Unit	X	X	X	X	_	
_	Haeger-Beverly Unit	X	X	X	X	_	
12.60	Haeger-Beverly Unit	X	X	X	X	_	
	Su	rface-water sites					
_	_	X	X	X	X	X	
_	_	X	X	X	X	X	
_	_	X	_	_	_	X	

¹Aquifer unit names originated from Illinois State Water Survey (Meyer and others, 2013) and are used in this report for consistency purposes. See Gahala (2017) for further details on aquifer unit and monitoring well designations.

ILWATER database are uncertain because the database does not account for wells that have not been abandoned (sealed) or categorized as both bedrock and unconsolidated (Daniel Abrams, Illinois State Water Survey, written commun., March 23, 2022).

The 2020 population estimate for McHenry County was 310,229, compared to the 2010 population of 308,760 (U.S Census Bureau, 2021). The USGS has estimated water use for the United States every 5 years since 1950 at the State level, and since 1985 at the county level. The most recent

estimate of water use was for 2015 (U.S. Geological Survey, 2020a). The water-use compilation includes data from production wells, domestic-supply wells, irrigation wells, and other sources (industrial, mining, livestock, and aquaculture) and is reported at the county level. The population served and water-use estimates for 2015 indicate that total groundwater withdrawals have decreased by about 9 percent from 2010 (U.S. Geological Survey, 2020a; table 2). The USGS is transitioning from previous approaches of 5-year annual compilations to a nationally consistent modeling platform to improve

Table 2. Water use by category, volume of groundwater withdrawals, and population served, McHenry County, Illinois, 2010 and 2015. [Data from U.S. Geological Survey, 2020a. Mgal/d, million gallons per day; —, no data or not applicable]

Water use estadow	Groundwater wit	hdrawals (Mgal/d)	Population served (thousands		
Water-use category	2010	2015	2010	2015	
Public-supplied domestic	19.97	18.70	237	242	
Self-supplied domestic	5.75	5.25	72	66	
Irrigation	2.77	3.27	_	_	
Other ¹	3.22	1.58	_	_	
Total	31.71	28.80	309	308	

¹Includes industrial, mining, livestock, and aquaculture. Descriptions of estimation methodology are available at https://water.usgs.gov/watuse/data/index.html.

²Wells screened in the Shallow Bedrock Aquifer are within the upper 5 or 10 feet of the bedrock and represent the interface between the two interconnected aquifers.

temporal and spatial discretization of water use that is planned to allow forecasting capability. Water-use data were not available for 2020 at the time of this publication.

As of 2009, land use in the county was 61 percent agriculture, 11 percent open spaces, and 16 percent residential, and the remaining land use consisted of government/institutional, commercial retail, industrial, earth extraction, and some designated as "vacant" (McHenry County Planning and Development, 2020). It is estimated that by 2030, agricultural land use will be reduced to 45 percent, and residential, open spaces, and spaces designated as "environmentally sensitive" together will account for 51 percent of the land use (McHenry County Planning and Development, 2020).

Previous Water-Quality Investigations

Countywide water quality of the sand and gravel aquifers of McHenry County was assessed from sampling of municipal supply wells in 1976 (Woller and Sanderson, 1976) and nearly quarterly sampling of residential wells from April 1979 through April 1980 (Nicholas and Krohelski, 1984), which included major ions and nutrients. Results from these studies indicated concentrations of major ions and nutrients were not substantially variable and were generally consistent across seasons. Dissolved iron had the greatest amount of fluctuation but could be affected by other factors, such as age of the residential well and piping. Health and aesthetically based waterquality benchmarks were exceeded for sodium, nitrates, iron, manganese, and dissolved solids in residential wells sampled (Nicholas and Krohelski, 1984). In 2010, 41 of 42 monitoring wells in the MCGMN plus 5 NAWQA monitoring wells were sampled for field properties, major ions, metals, trace elements, and nutrients (Gahala, 2017). Health and aesthetically based water-quality benchmarks were exceeded for arsenic, sodium, nitrate (as nitrate plus nitrite), manganese, dissolved solids, chloride, and iron.

Methods of Study

This section describes methods used to collect and analyze discrete water-quality samples, including sampling procedures, data analysis, and quality assurance and quality control (QAQC). The statistical analysis applied to the water-quality results also is described.

Water-Quality Sampling

The MCGMN consists of 42 monitoring wells, of which 37 are equipped with pressure transducers that continuously monitor the groundwater levels (1 reading every 15 minutes). These data are transmitted using satellite telemetry to the USGS National Water Information System database (U.S. Geological Survey, 2020b) Water-quality samples were

collected from 41 of the 42 monitoring wells in the MCGMN and 4 NAWQA monitoring wells during June-July 2020 (fig. 1, table 1). The NAWQA monitoring wells were included to assess effects from urbanization. Two monitoring wells (2–ALD–D from MCGMN and 45N9E–7.6a from NAWQA) were planned to be sampled in 2020 but were not sampled because of site conditions and well integrity issues. All groundwater monitoring wells were sampled for field properties, major ions, trace metals, and nutrients. Field properties included dissolved oxygen, pH, dissolved solids, specific conductance, water temperature, turbidity, and alkalinity. Additionally, 12 monitoring wells were sampled for the analysis of CECs. Water-quality results can be retrieved from the USGS National Water Information System database using the station identifiers in table 1. The chloride-bromide ratio analysis from Gahala (2017) was used to help select which wells would be sampled for CECs. Based on the chloride-bromide ratio analyses, 9 of the 12 wells were identified as potentially affected by sewage, and the remaining 3 wells were considered control wells and were not identified as being affected by sewage (Gahala, 2017). Water-quality sampling procedures generally adhered to the USGS "National Field Manual for the Collection of Water-Quality Data" (U.S. Geological Survey, variously dated).

Surface-water samples also were collected in 2020 to preliminarily assess the presence of CECs downstream from wastewater-treatment plants and other industrial discharges. Five grab samples were collected along the cross-sectional transect of each stream and composited with a precleaned and prerinsed churn. This approach slightly deviated from the recommended method described in the National Field Manual (U.S. Geological Survey, variously dated), where streams greater than 5 ft in width should have a minimum of 10 sample points. Cross-sectional field properties were measured within the stream with a multiparameter device at intervals of 2–5 ft at each surface-water-quality monitoring location to determine if the stream was well mixed at that location. Water-quality properties (major ions, trace elements, and nutrients) were added at two of the three surface-water locations to assess general water quality in the two streams that contribute to downstream (beyond McHenry County) drinking water sources.

Water-quality samples were collected following the methods described in the USGS "National Field Manual for the Collection of Water-Quality Data" (U.S. Geological Survey, variously dated), and field alkalinities were completed in the field using the inflection-point method. Major ions, trace metals, and nutrients were analyzed by the methods described in Fishman (1993), Fishman and Friedman (1989), and Garbarino and others (2006). WICs are a general category for anthropogenically derived chemicals that include some pharmaceuticals, pesticides, and some volatile organic compounds all quantified in parts per billion (micrograms per liter). Pharmaceuticals are quantified in parts per trillion (nanograms per liter). Pharmaceuticals and WICs were analyzed by the methods described in Furlong and others (2008 and 2014) and

Zaugg and others (2006). Water-quality samples were shipped with appropriate sample preservation to the USGS National Water Quality Laboratory (NWQL) in Denver, Colorado, for analysis. The NWQL uses method reporting conventions for establishing the minimum concentration greater than which a quantitative measurement can be made (Oblinger Childress and others, 1999). These reporting conventions are the method reporting level and the laboratory reporting level. The method reporting level is defined by the NWQL as the smallest measured concentration of a substance that can be reliably measured using a given analytical method. The method detection level is the minimum concentration of a substance that can be measured and reported with 99-percent confidence that the concentration is greater than zero (Oblinger Childress and others, 1999; Barr and Davis, 2010). The laboratory reporting level is the interim reporting level, which is typically two times the method detection level to minimize false negative risks and is the smallest concentration of a chemical that can be reported by the laboratory (Williams and others, 2015). QAQC results are summarized in appendix 1 and are groundwater focused. No QAQC samples were collected for the preliminary CEC assessment in the surface water. QAQC samples include equipment blanks, field blanks, and replicates (replicate of the primary environmental sample) and were collected almost weekly throughout the duration of the sampling effort. For a total of 12 QAQC samples, 2 equipment blanks, 5 field blanks, and 5 replicates were collected. The full results of the QAQC samples and the full list of CECs sampled for are summarized in Gahala and Gruhn (2022).

The water-quality results of the 2020 data are compared to health-based and aesthetically based benchmarks in water-quality samples established by the U.S. Environmental Protection Agency (EPA), such as the drinking water advisories (DWAs) and established maximum contaminant level (MCL) and the secondary maximum contaminant level (SMCL) (U.S. Environmental Protection Agency, 2009a, b, 2018). Aesthetically based benchmarks are derived from EPA SMCLs, which are nonregulated, recommended criteria that establish taste and odor quality. Surface-water-quality results also were compared to aquatic ambient water-quality criteria established by the EPA to assess suitability for aquatic life (U.S. Environmental Protection Agency, 2021b).

Statistical Analysis

To identify changes in water quality between 2010 and 2020, two analyses were completed. Using paired samples, differences between 2010 and 2020 were assessed across all wells for each constituent. For constituents that did indicate a statistically significant difference across all wells, differences between 2010 and 2020 also were assessed for three groups of wells in either deep, intermediate, or shallow aquifers. A total of 20 constituents were measured in 43 wells in 2010 and 2020. The wells also were grouped according to aquifer depth; 10, 19, and 14 wells were designated as

"shallow" (13.6–58.7 ft below ground surface), "intermediate" (62.3–120.6 ft below ground surface), and "deep" (139–344 ft below ground surface), respectively. Of the constituents, nine had at least one censored (less than the reporting level) observation. An alpha level of 0.05 was used for determining statistical significance. All statistical analyses were completed using the R statistical software program (R Core Team, 2019), specifically the NADA (Lopaka and Helsel, 2020). NADA2 (Julian and Helsel, 2021) and Wilcox (R Core Team, 2021) R packages, which include statistical tests for censored data.

The applicable statistical test depended on the presence of censored values in the dataset. The WSR test (Helsel, 2012) was used to assess changes when the data were all uncensored values. The PPW test was used when data were a mix of censored and uncensored values. Both tests are nonparametric, require no assumptions about the distribution of the data, and have a null hypothesis that the difference between paired samples is zero. Rejection of the null hypothesis indicates samples collected in 2020 differ from those collected in 2010. The PPW test is a score test and allows for left-censored values to have multiple reporting levels, meaning recensoring data to the highest reporting level is not required (Helsel, 2012). These two tests also were used when testing for differences in concentration between 2010 and 2020 for each aquifer group. Because three comparisons (one for each aquifer group) were made for each parameter, probability values (p-values) from the WSR and PPW tests were adjusted using the Benjamini and Hochberg correction, which controls the false discovery rate when completing multiple comparison tests (Benjamini and Hochberg, 1995). The combination of testing for differences between 2010 and 2020 across all wells and testing for differences within each aquifer group helped to identify at what depth changes were likely.

2020 Water Quality

Results of the 2020 decadal water-quality sampling are presented in this section and include field properties, major ions, trace metals, and nutrients. Selected monitoring wells were sampled for CECs, which include pharmaceuticals, pesticides, and WICs. The results for three surface-water sites sampled also are included in tables in this report. The surface-water results did not have any detected concentrations of major ions, metals, or nutrients greater than any aquatic ambient water-quality criteria. Groundwater water-quality results, including the implications on the suitability for drinking water, are discussed in the following sections. The following sections compare constituents with the health-based standards and aesthetically based recommendations by the EPA and include the enforceable MCLs, nonenforceable SMCLs, and DWA (U.S. Environmental Protection Agency, 2009a, 2018; table 3).

Table 3. U.S. Environmental Protection Agency water-quality thresholds, number of wells exceeding water-quality thresholds in samples collected in 2020, and statistical description (maximum, minimum, median).

[MCL, maximum contaminant level; µg/L, microgram per liter; DWA, drinking water advisory; mg/L, milligram per liter; SMCL, secondary maximum contaminant level; >, greater than]

Constituent	Water-quality standard and benchmark concentration ¹	Concentration ²	Number of sites exceeding drinking water benchmark ²	Common sources of constituent	Potential effect of exceedance on health
		Health-	based benchmark		
Arsenic	MCL= 10 μg/L	Maximum = 70.29 Minimum = 0.10 Median = 0.75	11	Aquifer sediments and rocks, fertilizer	Skin damage; circula- tory system prob- lems, potentially carcinogenic.
Sodium	DWA = 20 mg/L (20,000 μ g/L)	Maximum = 180 Minimum = 4.32 Median = 14.6	18	Road salt, septic leach- ate, water softener, aquifer sediment, and rocks	May increase blood pressure in at-risk individuals.
Manganese	DWA = 0.3 mg/L (300 μ g/L)	Maximum = 740 Minimum = 0.50 Median = 25.6	1	Aquifer sediments and rocks, industrial discharges, sewage, landfill leachate	Toxicity to nervous system.
		Aesthetica	ally based benchmark		
Dissolved solids	SMCL = 500 mg/L	Maximum = 1,358 Minimum = 272 Median = 419	13	Road salt, brines, septic leachate, dissolved aquifer minerals	Objectional tastes and laxative effect at concentrations >1,000 mg/L.
Chloride	SMCL = 250 mg/L	Maximum = 517 Minimum = 0.55 Median = 16.9	2	Dissolved aquifer minerals, road salt, brines, septic leach- ate	Imparts salty taste and corrosive to metal infrastructure.
Iron	$SMCL = 300 \ \mu g/L$	Maximum = 3,478 Minimum = 10 Median = 1,640	39	Aquifer sediments and rocks	Discolors laundry and stains fixtures; imparts metallic taste.
Manganese	SMCL = $50 \mu g/L$	Maximum = 740 Minimum = 0.50 Median = 25.6	13	Aquifer sediments and rocks	Dark-brown or black staining on fixtures and laundry, black particulates at high concentrations.

¹Water-quality standards and benchmark concentrations are from U.S. Environmental Protection Agency (2018).

Field Properties

Results for field properties are presented in table 4. Dissolved oxygen concentrations at groundwater wells ranged from 0.0 to 7.0 milligrams per liter (mg/L) with a median 0.20 mg/L but were typically (80 percent) less than 1.0 mg/L. According to McCallum and others (2008), denitrification becomes more favorable at a dissolved oxygen concentration of less than 2.0 mg/L. Dissolved oxygen at surface-water sites ranged from 8.4 to 14.3 mg/L.

The pH collected at groundwater wells ranged from 6.5 to 7.6 standard pH units with a median of 7.2, and field alkalinities ranged from 268 to 429 mg/L as calcium carbonate with a

median of 340 mg/L as calcium carbonate. Surface-water sites had higher pH values than wells and ranged from 8.0 to 8.9; no field alkalinities were measured at the surface-water sites.

Specific conductance measured at groundwater sites ranged from 489 to 2,343 microsiemens per centimeter at 25 degrees Celsius (μS/cm) with a median of 719 μS/cm, and specific conductance at surface-water sites ranged from 688 to $803~\mu\text{S/cm}$. Some of the groundwater sites had elevated measurements of specific conductance and may be affected by the application of salt and salt brines on roads that can infiltrate into the groundwater. Elevated specific conductance was especially noticeable at shallow well sites that are closer to roadways.

²Data summarized from U.S. Geological Survey (2020b).

Table 4. Results for field properties for the McHenry County Groundwater Monitoring Network wells, 4 National Water-Quality Assessment Project monitoring wells, and 3 surface-water sites, McHenry County, Illinois, 2020.

[Data available from U.S. Geological Survey, 2020b. mg/L, milligram per liter; std, standard; μ S/cm, microsiemens per centimeter at 25 degrees Celsius; $^{\circ}$ C, degree Celsius; NTU, nephelometric turbidity unit; CaCO₃, calcium carbonate; IL, Illinois; —, no data]

Field identifier	Dissolved oxygen (mg/L)	pH, std units	Specific conductance (µS/cm)	Water temperature (°C)	Turbidity (NTU)	Field alkalinity (mg/L as CaCO ₃)
14-RIL-S	2.1	7.1	1,123	12.6	7.0	268
HUNT-09-03	0.1	7.4	726	11.8	0.4	377
16–GRF–I	0.1	6.6	996	11.8	5.9	325
16–GRF–D	2.5	7.3	606	13.6	0.9	294
17–ALG–S	0.4	7.1	818	22.7	10.0	391
17–ALG–D	1.1	7.5	605	18.3	2.1	276
MARS-09-01	1.9	7.2	564	13.2	0.4	353
15-COR-S	0.1	7.3	750	13.4	15.6	335
15-COR-I	0.1	7.4	743	12.7	4.3	341
15-COR-D	0.2	7.4	738	12.8	3.4	349
10-MAR-S	3.0	7.1	719	11.8	4.0	271
WAUC-02-12	0.2	7.6	534	13.3	5.8	302
11-SEN-I	0.2	7.0	679	12.3	0.6	306
11-SEN-D	0.2	7.1	592	12.0	1.9	348
MARN-09-02	0.2	7.0	678	13.3	9.6	356
WOOD-08-01	0.3	7.3	698	13.3	1.9	404
13–NUN–I	0.2	7.2	621	12.9	1.0	314
13–NUN–D	0.2	7.2	646	13.6	1.4	332
WAUC-08-13	0.1	7.1	727	12.8	0.7	355
MHEN-08-01	0.1	6.5	909	13.4	0.7	361
MARN-09-01	0.5	7.3	715	12.9	5.2	363
7–HRT–S	0.2	7.2	1,093	12.8	56.6	372
7–HRT–I	0.5	7.1	592	12.9	1.2	340
7–HRT–D	0.1	7.2	623	12.3	2.3	369
9–MCH–S	7.0	7.1	1,120	15.6	40.9	425
9–MCH–D	0.2	7.5	489	15.1	1.9	300
8-GRN-I	0.3	7.1	1,141	14.1	2.6	389
8–GRN–D	0.1	7.4	616	12.7	3.4	277
HEBR-08-02	0.1	7.3	798	12.5	4.0	322
HARV-09-01	0.4	7.1	782	12.6	2.2	331
NW-6-45-9	0.4	7.0	986	15.0	0.7	358
1-CHE-S	0.1	6.9	620	12.6	11.2	312
1–CHE–D	0.2	7.6	513	16.0	1.8	276
HEBR-09-03	0.7	7.3	753	12.6	7.1	322
I–RCH–S	3.7	7.2	1,983	25.3	13.6	373
4–RCH–I	0.1	7.3	543	12.3	5.2	311
4–RCH–D	0.5	7.4	527	12.6	12.9	302
HEBR-08-01	0.0	7.0	561	12.2	0.8	331
3–НЕВ–І	0.1	7.2	722	12.8	20.8	381
MARN-10-03	0.2	7.4	604	13.0	2.0	320
MARN-10-04	1.8	7.4	701	12.7	1.0	407

Table 4. Results for field properties for the McHenry County Groundwater Monitoring Network wells, 4 National Water-Quality Assessment Project monitoring wells, and 3 surface-water sites, McHenry County, Illinois, 2020.—Continued

[Data available from U.S. Geological Survey, 2020b. mg/L, milligram per liter; std, standard; µS/cm, microsiemens per centimeter at 25 degrees Celsius; °C, degree Celsius; NTU, nephelometric turbidity unit; CaCO3, calcium carbonate; IL, Illinois; —, no data]

Field identifier	Dissolved oxygen (mg/L)	pH, std units	Specific conductance (µS/cm)	Water temperature (°C)	Turbidity (NTU)	Field alkalinity (mg/L as CaCO ₃)
43N8E-8.2c	3.6	7.0	1,269	13.9	77.0	354
45N7E-32.4d	0.2	6.8	918	12.8	7.3	331
43N8E-3.7d	0.2	6.7	2,343	13.8	12.1	375
44N9E-20.7c	0.2	6.7	1,663	13.3	6.2	429
Nippersink Creek Near Spring Grove, IL	10.3	8.3	703	25.4	_	_
Fox River at Huntley Road at Carpentersville, IL	14.3	8.9	688	28.2	_	_
Kishwaukee River at Marengo, IL	8.4	8.0	803	21.3	_	_

Water temperature in groundwater wells ranged from 11.8 to 25.3 degrees Celsius (°C) with a median of 12.85 °C. Most wells were in the 11–16 °C range, but two wells were considerably higher and more similar to surface-water sites (17-ALG-S and 4-RCH-S), potentially from heating of the pump discharge line and flow-through chamber because of incomplete shading on sunny days or from the lower pumping rates allowing the water to warm within the discharge lines. Surface-water sites had water temperatures ranging from 21.3 to 28.2 °C, which is typical of many Midwest streams and rivers in June and July.

Turbidity at groundwater sites was generally low and ranged from 0.4 to 77.0 nephelometric turbidity units (NTU), and the median was 3.4 NTU. No turbidity measurements were made at surface-water sites.

Major lons

Major ions were analyzed in the groundwater samples and two surface-water samples to compare to applicable drinking water standards, determine changes in groundwater sources, and identify potential land-use effects. Constituents included calcium, magnesium, potassium, chloride, sodium, dissolved solids, bromide, fluoride, silica, and sulfate. Results from the analysis of these samples are presented in table 5 and further described in the following paragraphs.

Calcium in monitoring wells ranged from 41.8 to 184 mg/L with a median of 85.9 mg/L, and at the two surface-water sites, results were 44.5 mg/L and 72.5 mg/L. Magnesium in wells ranged from 27.8 to 95.0 mg/L with a median of 44.9 mg/L, and concentrations in surface-water sites were 34.8 mg/L and 37.9 mg/L. Calcium and magnesium are dissolved from soils, sediments, and rock, particularly

limestone and dolomite; these constituents can cause water hardness and scale formation on pipes and plumbing (National Ground Water Association, 2010).

Potassium in groundwater wells ranged from 0.83 to 8.00 mg/L with a median of 1.54 mg/L, and concentrations in surface-water sites were 2.39 mg/L and 2.73 mg/L. Potassium is not only commonly detected in clays, rocks, and soils, but also can be detected in fertilizers (Mullaney and others, 2009). High concentrations of potassium (greater than 500 mg/L) are usually indicative of a road salt source, especially if potassium chloride is applied as a deicer (National Ground Water Association, 2010; Gahala, 2017).

Sodium ranged from 4.3 to 180 mg/L with a median of 14.64 mg/L, and concentrations in surface-water sites were 29.3 mg/L and 52.3 mg/L. High occurrences of sodium were often incidental with high occurrences of chloride, which could indicate that sodium chloride is being used as a deicer for roadways.

Bromide in the groundwater wells ranged from an estimated 0.011 to 0.254 mg/L with a median of 0.031 mg/L, and concentrations at the two surface-water sites were 0.032 mg/L and 0.036 mg/L. Fluoride in groundwater ranged from 0.05 to 0.68 mg/L with a median of 0.29 mg/L, and concentrations at the surface-water sites were 0.14 mg/L and 0.16 mg/L. Silica in groundwater ranged from 8.8 to 25.3 mg/L with a median of 19.96 mg/L, and concentrations at the surface-water sites were 1.1 mg/L and 15.9 mg/L. Bromide is a trace element in seawater, or naturally dissolved from sedimentary rocks such as evaporites, carbonates, and shales (Salameh and others, 2016). Anthropogenic sources of bromide can come from septic tanks, deicers, agricultural chemicals, solvents, and gasoline additives (Davis and others, 1998). Fluoride in groundwater is from weathering and leaching of fluoride-bearing minerals from rocks and sediments (Jha and others, 2013).

2020 Water Quality

Table 5. Results for concentrations of major ions from samples collected from the McHenry County Groundwater Monitoring Network wells, 4 National Water-Quality Assessment wells, and 3 surface-water sites in McHenry County, Illinois, 2020.

[Data available from U.S. Geological Survey, 2020b. mg/L, milligram per liter; <, less than reporting level; E, estimated; —, no data; IL, Illinois]

Field identifier	Calcium (mg/L)	Magnesium (mg/L)	Potassium (mg/L)	Chloride (mg/L)	Sodium (mg/L)	Bromide (mg/L)	Fluoride (mg/L)	Silica (mg/L)	Sulfate (mg/L)	Dissolved solids (mg/L)	Chloride/ bromide ratio
14-RIL-S	68.3	35.6	1.53	187	114	0.084	0.13	11.7	15.3	548	2,338
HUNT-09-03	92.1	46.8	1.32	11.9	14.6	0.028	0.32	22.3	21.3	417	396
16-GRF-I	96.8	49.1	2.69	120	50.1	0.073	0.34	23.5	26.0	563	1,714
16-GRF-D	56.8	42.9	1.54	18.3	18.2	0.029	0.57	19.8	22.9	348	610
17-ALG-S	88.1	46.1	1.17	28.1	28.9	< 0.020	0.24	15.0	34.0	450	¹ E2,811
17-ALG-D	43.4	36.4	2.06	48.8	28.3	0.056	0.68	20.7	0.24	303	865
MARS-09-01	63.2	39.4	1.53	1.24	17.9	0.040	0.29	24.4	0.05	330	31
15-COR-S	96.1	47.3	1.34	14.1	7.4	0.049	0.35	19.4	86.9	466	282
15-COR-I	96.6	48.6	1.36	12.4	7.6	0.046	0.37	20.5	79.3	461	247
15-COR-D	93.2	48.5	1.43	11.7	7.4	0.048	0.40	20.6	77.6	465	235
10-MAR-S	85.8	39.0	1.92	31.3	21.1	0.033	0.11	8.8	97.7	424	1,043
WAUC-02-12	43.4	36.4	1.75	5.37	28.3	0.023	0.65	20.7	0.69	303	233
11-SEN-I	85.9	43.5	1.68	20.7	5.4	0.032	0.17	19.3	56.8	398	689
11-SEN-D	68.1	40.8	1.92	2.2	11.9	0.017	0.32	19.9	0.07	276	111
MARN-09-02	70.8	40.2	1.88	1.38	22.2	0.022	0.38	25.3	0.40	272	69
WOOD-08-01	79.6	47.3	1.60	1.53	12.0	E0.011	0.45	23.0	0.07	389	¹ E153
13-NUN-I	72.1	42.4	1.06	1.99	10.8	0.015	0.53	22.0	22.0	341	133
13-NUN-D	71.7	44.9	1.46	4.32	13.8	0.034	0.45	21.4	44.7	363	127
WAUC-08-13	96.4	48.4	1.27	3.12	5.7	0.021	0.18	23.8	60.7	415	156
MHEN-08-01	112	55.8	2.08	65.9	11.9	0.050	0.15	20.8	52.0	533	1,317
MARN-09-01	92.3	45.3	0.89	17.2	5.2	0.021	0.22	19.7	35.4	419	817
7-HRT-S	131	60.8	1.40	107	16.9	0.026	0.14	22.3	60.8	618	3,567
7-HRT-I	76.8	37.4	1.09	0.79	8.3	< 0.020	0.29	21.7	0.05	339	¹ E79
7-HRT-D	75.6	40.7	1.35	3.53	8.7	0.022	0.30	22.2	0.26	349	160
9-MCH-S	109	50.6	1.84	121	84.4	0.057	0.15	17.9	29.3	676	2,017
9-MCH-D	41.8	27.8	1.48	0.55	29.8	0.012	0.62	16.8	0.05	284	46
8-GRN-I	126	62.0	3.89	120	35.5	0.060	0.12	17.6	79.7	632	2,017
8-GRN-D	79.7	39.2	1.14	5.07	4.3	0.037	0.17	20.0	62.9	362	137
HEBR-08-02	104	49.2	1.94	24.8	9.4	0.045	0.14	20.0	96.3	505	547
HARV-09-01	96.2	46.9	2.46	37.2	12.2	0.024	0.14	14.5	51.9	455	1,537

Table 5. Results for concentrations of major ions from samples collected from the McHenry County Groundwater Monitoring Network wells, 4 National Water-Quality Assessment wells, and 3 surface-water sites in McHenry County, Illinois, 2020.—Continued

[Data available from U.S. Geological Survey, 2020b. mg/L, milligram per liter; <, less than reporting level; E, estimated; —, no data; IL, Illinois]

Field identifier	Calcium (mg/L)	Magnesium (mg/L)	Potassium (mg/L)	Chloride (mg/L)	Sodium (mg/L)	Bromide (mg/L)	Fluoride (mg/L)	Silica (mg/L)	Sulfate (mg/L)	Dissolved solids (mg/L)	Chloride/ bromide ratio
NW-6-45-9	102	49.5	3.91	83.6	41.8	0.045	0.11	15.3	44.3	562	1,672
1-CHE-S	96.9	43.7	8.00	26.6	11.8	0.024	0.16	14.7	78.9	464	1,332
1-CHE-D	50.6	29.0	1.01	0.76	21.3	0.014	0.37	11.0	0.39	274	76
HEBR-09-03	97.7	46.5	0.83	29.6	6.4	0.037	0.09	13.5	78.5	448	740
4-RCH-S	135	65.9	1.74	417	180	0.060	0.26	20.4	41.4	1,063	6,950
4–RCH–I	60.6	35.7	1.44	2.10	12.1	< 0.020	0.39	18.9	0.04	303	¹ E210
4-RCH-D	58.2	35.4	1.40	0.83	9.7	0.012	0.38	20.6	0.04	304	83
HEBR-08-01	69.0	38.4	1.00	2.31	6.6	0.025	0.33	18.0	21.1	326	621
3-НЕВ-І	84.8	46.8	1.13	16.9	12.8	0.038	0.26	22.1	16.0	404	423
MARN-10-03	61.7	35.2	1.69	1.94	24.8	0.023	0.38	21.3	0.46	283	_
MARN-10-04	76.7	43.4	1.74	1.99	17.4	0.031	0.39	23.8	0.03	392	_
43N8E-8.2c	94.6	43.2	3.08	183	118	0.035	0.10	15.3	29.8	685	5,229
45N7E-32.4d	109	48.5	1.78	69.2	25.8	< 0.020	0.15	16.6	42.3	518	¹ E6,920
43N8E-3.7d	184	95.0	7.37	517	170	0.254	0.05	18.7	73.0	1,358	2,035
44N9E-20.7c	121	56.5	2.06	227	129	0.044	0.10	15.4	35.5	802	5,675
Nippersink Creek Near Spring Grove, IL	72.5	37.9	2.39	52.8	29.3	0.036	0.16	15.9	30.7	403	_
Fox River at Huntley Road at Carpentersville, IL	44.5	34.8	2.73	89.6	52.3	0.032	0.14	1.1	24.5	385	_
Kishwaukee River at Marengo, IL	_	<u>—</u>	_	_	_	_	_		_	_	_

¹Chloride/bromide ratios are calculated by assigning one-half of the reporting level and are flagged as estimated.

Anthropogenic sources of fluoride can include coal burning, oil refining, steel productions, brick-making industries, and phosphatic fertilizer plants (Jha and others, 2011).

Sulfate from wells ranged from 0.03 to 97.7 mg/L with a median of 29.7 mg/L, and concentrations at the two surface-water sites were 24.5 mg/L and 30.7 mg/L. Hydrogen sulfides commonly are associated with a "rotten egg" odor and can impart a foul smell at higher concentrations (Agency for Toxic Substances and Disease Registry, 2011). Sniff tests at the time of sampling can often determine if there is a presence of hydrogen sulfides. Additional discussion later in this report relates these values to health-based benchmarks, implications to drinking water suitability, and aquatic life criteria for surface water.

Results for major ions did not exceed any health-based benchmarks except sodium, which exceeded the DWA of 20 mg/L in 40 percent (18 of 45 groundwater sites) of samples (fig. 2, table 5). Sources of sodium include dissolution of sediments and rocks, road salt, and septic-system leachate solutions (U.S. Environmental Protection Agency, 2002; Panno and others, 2006). Excess ingestion of sodium can increase blood pressure in at-risk individuals (Centers for Disease Control and Prevention, 2022). DWAs are not legally enforceable, and the 20 mg/L benchmark applies only to people on a 500-milligram-per-day restricted sodium diet (U.S. Environmental Protection Agency, 2018).

Dissolved solids are a measure of the dissolved ion concentrations in water. Groundwater samples ranged from 272 to 1,358 mg/L with a median of 417 mg/L, and concentrations at surface-water sites were 385 and 403 mg/L. The EPA has established an SMCL for dissolved solids because they can aesthetically affect groundwater, leave objectionable tastes, and have laxative effects when consumed at concentrations greater than 1,000 mg/L (table 3). Additionally, higher concentrations of dissolved solids can cause foaming and corrosion of metals. Dissolved solids can distinguish types of aquifers or indicate contamination from landfill leachate or road salt (Clark and Piskin, 1977). Shallower wells generally having higher dissolved solids concentrations than deeper wells (fig. 3); 29 percent of the groundwater wells (13 of 45 sites) exceeded the SMCL of 500 mg/L and 4 percent (2 of 45 sites) exceeded the 1,000-mg/L concentration (fig. 4). The two surface-water sites had similar concentrations of dissolved solids and did not exceed any thresholds of concern.

Chloride in groundwater ranged from 0.55 to 517 mg/L with a median of 16.9 mg/L, and concentrations at surface-water sites were 52.8 mg/L and 89.6 mg/L (table 5). Chloride was variable across wells, and the highest concentrations were near roadways. High concentrations of chloride are often attributed to road salt being applied as a deicer during winter weather events. Chloride has an SMCL of 250 mg/L, and greater concentrations can impart a salty taste on drinking water and can contribute to the corrosion of infrastructure (Shi and others, 2009; Pieper and others, 2018). Detections of chloride were variable across the MCGMN, and the highest concentrations were detected near roadways. A total of

4 percent of the wells (2 of 45 wells) exceeded the SMCL, and several shallow wells have elevated concentrations of chloride (fig. 5).

Trace Metals

Metals and trace elements are naturally occurring in the environment, although natural concentrations can vary depending on local geology and contributions from anthropogenic activities (U.S. Geological Survey, 2019). Some metals are essential nutrients, but all metals can be toxic at some level, and for some metals, it only takes small amounts to cause toxicity (U.S. Environmental Protection Agency, 2022a). During the 2020 sampling campaign, a variety of metals and trace elements were analyzed for, and these results are presented in table 6. These results are discussed in relation to health-based benchmarks and compared to the 2010 results later in this report to determine suitability of the groundwater and surface water for drinking and the potential implications on aquatic life in surface water. In general, most metals and trace elements were less than the reporting level or detected at low concentrations and did not exceed any health-based benchmarks (table 3); however, arsenic, iron, and manganese did exceed a benchmark and are discussed individually in this section.

Results for arsenic in groundwater and surface water were variable and ranged from less than 0.20 to 70.3 micrograms per liter (µg/L). Arsenic is a naturally occurring element that is detected in sediments and rocks, and it dissolves in anoxic groundwater (Hem, 1985; Lin and Puls, 2000). Chronic ingestion of arsenic at concentrations greater than the EPA MCL of 10 µg/L has been linked to increased risk of lung, bladder, skin, liver, kidney, and prostate cancer (U.S. Environmental Protection Agency, 2018; table 3). Arsenic concentrations exceeded the MCL in about 24 percent (11 of 45) of the monitoring wells, and most locations had concentrations similar to what was measured in 2010, with slight increases in 17–ALG–D from 16.1 µg/L in 2010 to 20.7 μ g/L in 2020 and in 1–CHE–D from 62 μ g/L in 2010 to 70.3 µg/L in 2020 (fig. 6). The screen interval at monitoring well 1-CHE-D is open to the top of the limestone bedrock aquifer and had low dissolved oxygen (0.2 mg/L). The geochemically reducing conditions may increase the dissolution of arsenic-bearing minerals within the limestone aquifer (Renard and others, 2015). MARN-10-03 (10.1 µg/L) and MARN-10-04 (16.3 μ g/L) were sampled in 2020 but were not available to sample in 2010; both wells had detections of arsenic greater than the MCL. Detections for arsenic at the two surface-water sites that were sampled did not exceed the MCL for arsenic.

Iron concentrations in groundwater ranged from less than 10 to 3,478 μ g/L with a median of 1,739 μ g/L. Groundwater in McHenry County has naturally elevated iron from the dissolution of glacial aquifer material and is consistent with previous water-quality investigations and this 2020 decadal



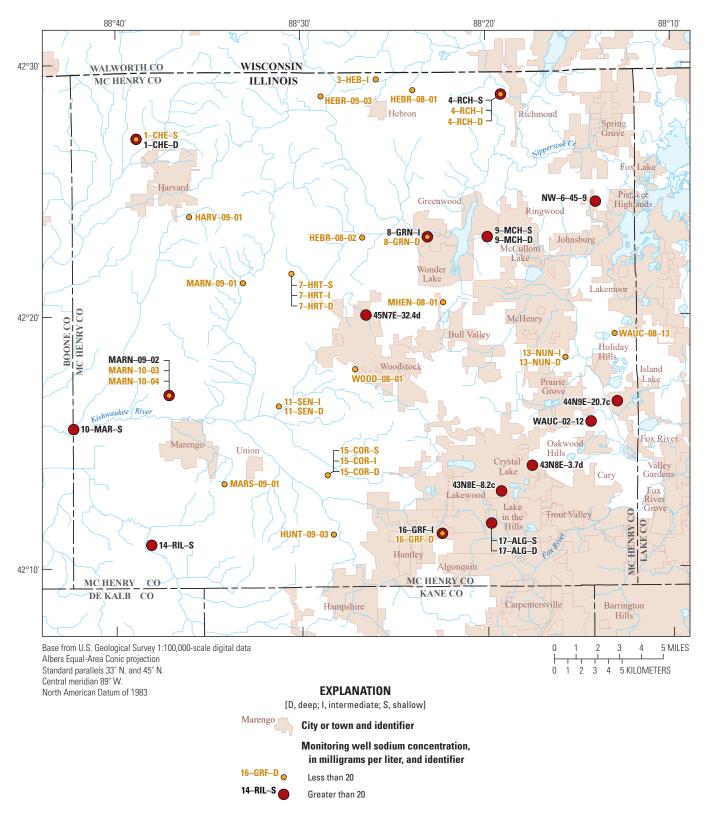


Figure 2. Distribution of sodium concentrations in samples from the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment Project monitoring wells, McHenry County, Illinois, 2020.

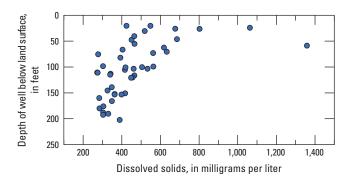


Figure 3. Dissolved solids concentrations by well depth in the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment Project monitoring wells, McHenry County, Illinois 2020.

update (Groschen and others, 2008; Gahala, 2017). Iron has an SMCL of 300 μ g/L, and concentrations exceeding the SMCL can cause staining of household fixtures and laundry (National Ground Water Association, 2010). Concentrations greater than 1,800 μ g/L can impart a metallic taste in water (Fetter, 1980, p. 355). A total of 87 percent (39 of 45 wells) exceeded the 300- μ g/L SMCL, and 40 percent (18 of 45 wells) exceeded the 1,800- μ g/L concentration (fig. 7). Iron in surface-water samples was less than the reporting level of 10 μ g/L.

Manganese ranged from 0.5 to 740 µg/L with a median of 25.6 µg/L (table 6). Manganese naturally occurs in aquifer sediments but also can have anthropogenic sources such as industrial effluent, acid-mine drainage, sewage, and landfill leachate (Nádaská and others, 2010; World Health Organization, 2011). Long-term exposure to DWA exceedances of manganese can cause toxicity to the nervous system (World Health Organization, 2011). Manganese has a DWA of 300 μ g/L (exceeded at one well) and an SMCL of 50 μ g/L. Similar to iron, manganese is naturally abundant in groundwater in McHenry County because of dissolution of manganeseoxide coatings on glacial aquifer material (Groschen and others, 2008). Concentrations greater than the SMCL can cause dark brown or black staining on household fixtures and laundry. About 29 percent (13 of 45 wells) exceeded the SMCL, and the highest concentration was nearly 15 times the SMCL (740 µg/L at 10–MAR–S; fig. 8; table 3). Manganese concentrations in surface water did not exceed the aesthetic or health benchmarks.

Nutrients

Nitrogen (nitrate plus nitrite as nitrogen, hereafter referred to as "nitrate") and phosphorus naturally occur in the environment but also can come from anthropogenic sources such as fertilizers, animal manure, and sewage (Gahala, 2017). Anthropogenic sources of nutrients, such as fertilizers, have increased the concentrations of nitrate in many streams and aquifers (U.S. Geological Survey, 1999; Dubrovsky and

others, 2010; Domagalski and Johnson, 2012). In total, 8 of 45 wells, or about 18 percent of the sampled monitoring wells, had detections for nitrate (0.08 to 5.33 mg/L) and all were shallow (less than a 50-ft depth) (table 7). Nitrite ranged from less than the reporting level of 0.001 to 0.054 mg/L with a median of less than 0.001 mg/L as nitrogen. Orthophosphate ranged from less than the reporting level of 0.004 to 0.141 mg/L as phosphorus with a median of 0.010 mg/L as phosphorus. Ammonia ranged from less than 0.01 to 2.04 mg/L as nitrogen with a median of 0.34 mg/L as nitrogen. Nippersink Creek and Fox River were the two surfacewater sites sampled for nutrients (table 7). Nitrate, nitrite, and orthophosphate were detected at Nippersink Creek, but the concentration of ammonia was less than the reporting level of 0.01 mg/L as nitrogen.

No detections of nitrate exceeded the MCL of 10 mg/L in 2020. Nitrate concentrations ranged from less than the reporting level to 5.33 mg/L as nitrogen. Although no nitrate samples exceeded the MCL of 10 mg/L, a few sites did indicate slight increases. Well 17–ALG–S increased from 0.71 mg/L as nitrogen in 2010 to 1.57 mg/L as nitrogen in 2020 and 10–MAR–S increased from 0.23 mg/L as nitrogen in 2010 to 1.52 mg/L as nitrogen in 2020. Nitrite concentrations ranged from less than the reporting level of 0.001 to 0.054 mg/L as nitrogen and did not exceed the MCL of 1 mg/L for the health-based benchmark in any samples. Nitrite is unstable in aerated water and is generally considered to be an indicator of contamination from sewage or organic waste (Hem, 1985) if present. Results for all nutrients are summarized in table 7.

A dissolved oxygen concentration of 2.0 mg/L is a common threshold for denitrification to exist; below this concentration, denitrification becomes favorable (McCallum and others, 2008). In all but one well, detections of nitrate were coincident with dissolved oxygen concentrations greater than 2.0 mg/L, and most of these wells were shallow wells (20–48 ft). The one exception is well 17–ALG–S, where the dissolved oxygen concentration was 0.4 mg/L and the nitrate concentration of 1.57 mg/L as nitrogen was relatively high. One potential explanation for the lower dissolved oxygen and elevated nitrate concentration may be that groundwater can move quickly through sand and gravel aquifer materials along a preferential flow path or that dissolved oxygen decreases at a faster rate than denitrification.

Contaminants of Emerging Concern

CECs are pharmaceuticals, detergents, natural and synthetic hormones, pesticides, and other chemicals such as WICs, plastic components, surfactant metabolites, and fire retardants (U.S. Environmental Protection Agency, 2009b; Tomasek and others, 2012) that are not typically treated for in wastewater-treatment plants. CECs that may be removed at wastewater-treatment plants are incidental to biological treatment and disinfection processes (U.S. Environmental Protection Agency, 2009b; U.S. Geological Survey, 2018).



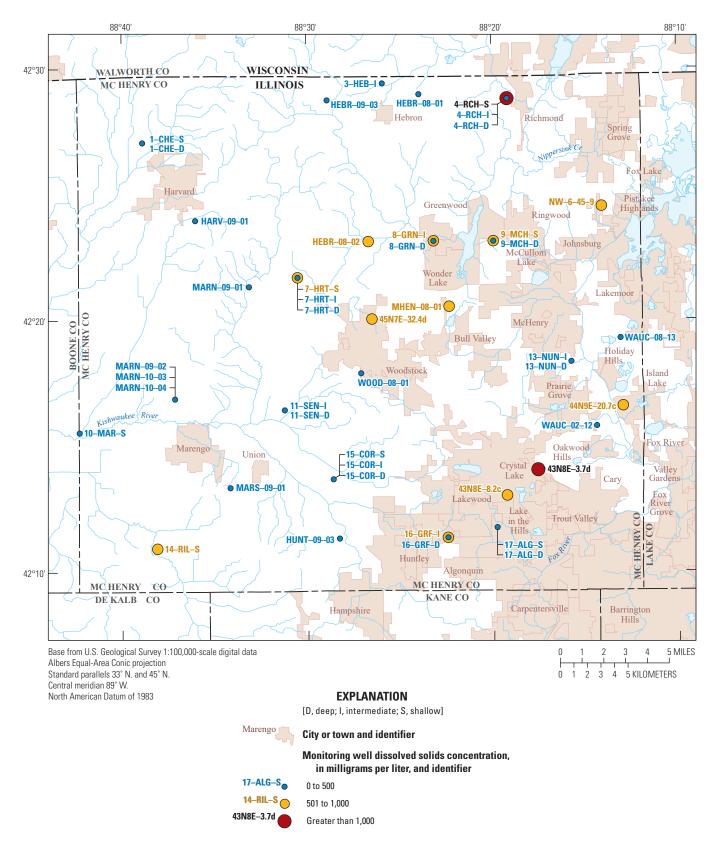


Figure 4. Distribution of dissolved solids concentrations in the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment Project monitoring wells, McHenry County, Illinois, 2020.

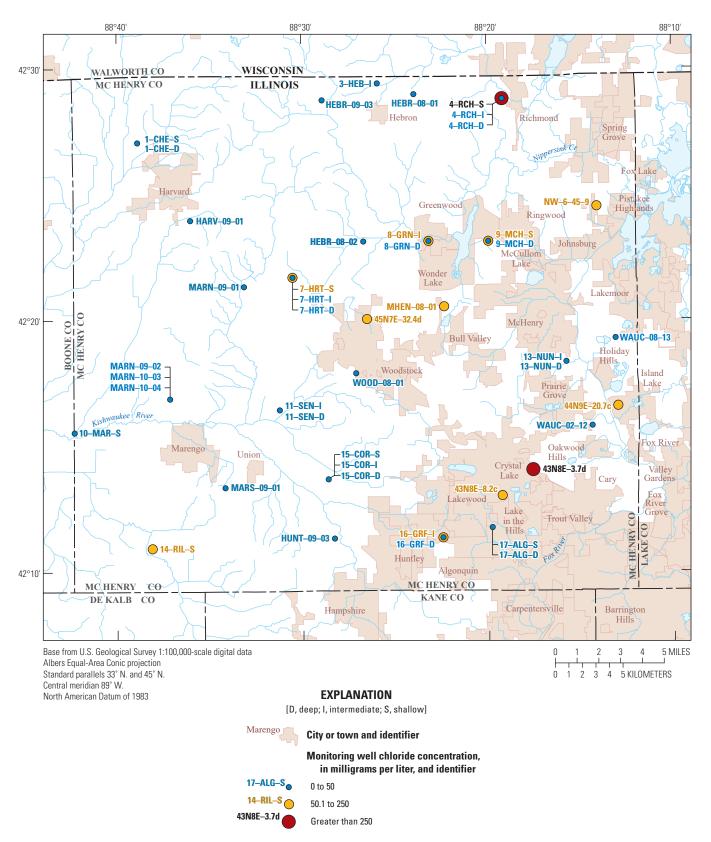


Figure 5. Distribution of chloride concentrations in the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment Project monitoring wells, McHenry County, Illinois, 2020.

Table 6. Results for concentrations of trace metals from samples collected from the McHenry County Groundwater Monitoring Network wells, 4 National Water-Quality Assessment wells, and 2 surface-water sites in McHenry County, Illinois, 2020.

 $[Data\ available\ from\ U.S.\ Geological\ Survey,\ 2020b.\ \mu g/L,\ microgram\ per\ liter;<, less\ than\ reporting\ level;\ E,\ estimated;\ IL,\ Illinois;\ ---,\ no\ data]$

Field identifier	Aluminum (μg/L)	Barium (µg/L)	Beryllium (μg/L)	Cadmium (µg/L)	Chromium (µg/L)	Iron (μg/L)	Manganese (µg/L)	Molybdenum (μg/L)
14-RIL-S	<6	75.8	< 0.02	< 0.06	<1	<10	10.2	0.61
HUNT-09-03	<3	78.4	< 0.01	< 0.03	0.6	2,667	12.2	2.33
16-GRF-I	<3	207	< 0.01	< 0.03	< 0.5	2,185	25.6	1.65
16-GRF-D	<3	115	< 0.01	< 0.03	< 0.5	526	6.4	0.70
17-ALG-S	<3	93.7	< 0.01	< 0.03	< 0.5	471	36.6	3.25
17-ALG-D	<3	88.4	< 0.01	< 0.03	< 0.5	849	11.4	4.25
MARS-09-01	<3	126	< 0.01	< 0.03	< 0.5	1,914	14.0	3.98
15-COR-S	<3	129	< 0.01	< 0.03	< 0.5	3,020	54.1	2.27
15-COR-I	<3	126	< 0.01	< 0.03	< 0.5	2,689	42.7	1.70
15-COR-D	<3	124	< 0.01	< 0.03	< 0.5	2,518	39.5	1.59
10-MAR-S	<3	52.1	< 0.01	< 0.03	< 0.5	<10	740	2.02
WAUC-02-12	<3	92.4	< 0.01	< 0.03	< 0.5	772	11.4	1.60
11-SEN-I	<3	78.1	< 0.01	< 0.03	< 0.5	946	45.4	1.07
11-SEN-D	<3	75.7	< 0.01	< 0.03	< 0.5	1,689	22.4	1.09
MARN-09-02	<3	125	< 0.01	< 0.03	< 0.5	1,927	18.5	4.05
WOOD-08-01	<3	59.0	< 0.01	< 0.03	1.1	1,277	32.5	1.78
13-NUN-I	<3	101	< 0.01	< 0.03	< 0.5	1,796	13.8	1.32
13-NUN-D	<3	64.4	< 0.01	< 0.03	< 0.5	604	8.4	1.87
WAUC-08-13	<3	78.1	< 0.01	< 0.03	0.7	1,509	29.6	0.32
MHEN-08-01	<3	80.3	< 0.01	< 0.03	< 0.5	2,434	72.8	1.37
MARN-09-01	<3	55.1	< 0.01	< 0.03	< 0.5	3,478	29.8	1.23
7-HRT-S	<6	84.2	< 0.02	< 0.06	<1	2,530	58.1	2.16
7-HRT-I	<3	35.6	< 0.01	< 0.03	< 0.5	2,525	11.2	1.58
7-HRT-D	<3	44.6	< 0.01	< 0.03	< 0.5	2,176	17.3	0.97
9-MCH-S	<6	62.2	< 0.02	< 0.06	<1	<10	0.9	0.65
9-MCH-D	<3	87.4	< 0.01	< 0.03	< 0.5	2,372	26	2.79
8-GRN-I	<6	131	< 0.02	< 0.06	<1	1,861	61.6	1.89
8–GRN–D	<3	32.9	< 0.01	< 0.03	< 0.5	775	222	2.12
HEBR-08-02	<3	109	< 0.01	< 0.03	< 0.5	1,629	53.1	1.69
HARV-09-01	<3	59.7	< 0.01	< 0.03	< 0.5	915	50.5	2.69
NW-6-45-9	<3	72.0	< 0.01	< 0.03	< 0.5	1,406	54.2	2.57
1-CHE-S	<3	148	< 0.01	< 0.03	< 0.5	1,357	141	0.89
1-CHE-D	<3	78.6	< 0.01	< 0.03	< 0.5	1,256	8.1	6.04
HEBR-09-03	<3	45.5	< 0.01	< 0.03	< 0.5	2,060	33.6	0.96
4-RCH-S	<6	173	< 0.02	< 0.06	<1	1,206	28.8	0.95
4–RCH–I	<3	51.8	< 0.01	< 0.03	< 0.5	2,044	9.9	1.34
4-RCH-D	<3	52.4	< 0.01	< 0.03	< 0.5	1,142	5.1	1.14
HEBR-08-01	<3	59.6	< 0.01	< 0.03	< 0.5	1,640	24.2	1.04
3-НЕВ-І	<3	68.0	< 0.01	< 0.03	0.8	1,764	13.5	0.30
MARN-10-03	<3	284	< 0.01	< 0.03	< 0.5	1,714	8.4	7.13
MARN-10-04	<3	103	< 0.01	< 0.03	< 0.5	2,382	22.5	3.34

Table 6. Results for concentrations of trace metals from samples collected from the McHenry County Groundwater Monitoring Network wells, 4 National Water-Quality Assessment wells, and 2 surface-water sites in McHenry County, Illinois, 2020.—Continued

 $[Data\ available\ from\ U.S.\ Geological\ Survey,\ 2020b.\ \mu g/L,\ microgram\ per\ liter;<, less\ than\ reporting\ level;\ E,\ estimated;\ IL,\ Illinois;\ ---,\ no\ data]$

Nickel (μg/L)¹	Silver (µg/L)	Antimony (μg/L)	Arsenic (μg/L)	Selenium (µg/L)	Uranium (μg/L)	Lithium (µg/L)	Strontium (µg/L)	Boron (µg/L)
0.44	<2	< 0.12	< 0.20	0.24	0.59	2.61	87	16
E1.35	<1	< 0.06	24.5	< 0.05	0.04	3.08	475	37
0.28	<1	< 0.06	0.13	< 0.05	< 0.03	8.63	874	45
< 0.20	<1	< 0.06	0.20	< 0.05	< 0.03	6.06	3,349	197
0.87	<1	0.11	0.19	< 0.05	1.32	10.98	592	76
E1.20	<1	0.06	20.7	< 0.05	< 0.03	2.74	1,680	175
0.26	<1	< 0.06	21.7	< 0.05	< 0.03	7.85	466	91
< 0.20	<1	< 0.06	0.42	< 0.05	< 0.03	5.03	265	36
< 0.20	<1	< 0.06	0.71	< 0.05	0.04	5.16	302	37
< 0.20	<1	< 0.06	0.54	< 0.05	0.05	5.44	315	39
E2.40	<1	0.18	0.24	0.18	3.80	1.61	109	21
E0.40	<1	< 0.06	0.80	< 0.05	< 0.03	10.72	1,680	175
< 0.20	<1	< 0.06	0.59	< 0.05	0.10	3.14	186	29
< 0.20	<1	< 0.06	3.82	< 0.05	< 0.03	3.01	721	90
< 0.20	<1	< 0.06	12.4	< 0.05	< 0.03	6.95	669	128
0.41	<1	< 0.06	0.62	< 0.05	< 0.03	6.70	641	73
< 0.20	<1	< 0.06	0.30	< 0.05	< 0.03	5.77	1,526	83
< 0.20	<1	< 0.06	0.10	< 0.05	0.05	6.22	1,318	107
0.29	<1	< 0.06	0.22	< 0.05	0.05	3.20	175	24
0.46	<1	< 0.06	0.36	< 0.05	0.06	4.08	159	21
< 0.20	<1	< 0.06	8.40	< 0.05	< 0.03	1.24	314	17
0.87	<2	< 0.12	0.58	< 0.1	0.15	4.98	125	93
0.54	<1	< 0.06	18.2	< 0.05	< 0.03	1.61	535	39
< 0.20	<1	< 0.06	9.72	< 0.05	< 0.03	2.44	700	41
0.67	<2	< 0.12	0.21	0.48	0.88	3.37	190	64
< 0.20	<1	< 0.06	5.38	< 0.05	< 0.03	1.44	2,341	264
E1.08	<2	< 0.12	1.04	< 0.1	0.25	11.89	148	22
< 0.20	<1	< 0.06	1.14	< 0.05	0.22	3.28	112	14
0.83	<1	0.56	3.23	< 0.05	0.26	4.76	135	19
0.27	<1	< 0.06	0.56	< 0.05	0.59	4.27	92	21
0.72	<1	0.07	0.48	0.07	0.93	6.98	76	19
0.66	<1	< 0.06	0.36	< 0.05	1.03	3.17	204	32
< 0.20	<1	< 0.06	70.3	< 0.05	< 0.03	0.45	393	48
0.31	<1	< 0.06	0.21	< 0.05	< 0.03	2.07	82	17
0.59	<2	< 0.12	0.27	< 0.1	0.14	9.37	271	26
0.59	<1	< 0.06	15.5	< 0.05	< 0.03	1.99	1,190	79
< 0.20	<1	< 0.06	16.7	< 0.05	< 0.03	3.03	1,052	58
< 0.20	<1	< 0.06	5.07	< 0.05	< 0.03	2.23	513	42
0.25	<1	< 0.06	13.3	< 0.05	< 0.03	1.90	559	55
0.31	<1	< 0.06	10.1	< 0.05	< 0.03	7.19	437	135
0.34	<1	<0.06	16.3	< 0.05	< 0.03	4.85	831	99

Table 6. Results for concentrations of trace metals from samples collected from the McHenry County Groundwater Monitoring Network wells, 4 National Water-Quality Assessment wells, and 2 surface-water sites in McHenry County, Illinois, 2020.—Continued

Data available fro	om U.S. Geological Su	rvev 2020b ug/L r	nicrogram per liter: <	less than reporting	level F	estimated: II Illinois:	no datal
TData available ill	JIII U.S. UGUIUBIGAI SU	1 v c v. 20200. u g/L. 1	microgram ber mer.	. iess man iedorung	ICVCI. L.	estimated, 1L, minors.	—, no data i

Field identifier	Aluminum (μg/L)	Barium (µg/L)	Beryllium (μg/L)	Cadmium (µg/L)	Chromium (µg/L)	Iron (μg/L)	Manganese (µg/L)	Molybdenum (µg/L)
43N8E-8.2c	<6	38.3	< 0.02	< 0.06	<1	<10	0.5	0.37
45N7E-32.4d	<3	43.4	< 0.01	< 0.03	0.6	<10	63.3	3.55
43N8E-3.7d	<6	175	< 0.02	< 0.06	<1	2,479	78.3	1.97
44N9E-20.7c	<6	85.0	< 0.02	< 0.06	<1	13	113	2.01
Nippersink Creek Near Spring Grove, IL	<3	55.9	<0.01	< 0.03	<0.5	<10	18.7	1.18
Fox River at Huntley Road at Carpentersville, IL	3.9	53.0	<0.01	< 0.03	<0.5	<10	1.6	1.52
Kishwaukee River at Marengo, IL	_	_	_	_	_	_	_	_

The release of CECs to the environment is a growing concern because of the potential to cause adverse effects in humans and other nontargeted organisms (Kolpin and others, 2002), and the use of pharmaceuticals has increased by 2.8 percent since 2015 (Martin and others, 2019; Centers for Disease Control and Prevention, 2021). The most commonly prescribed medications include analgesics, antihyperlipidemic agents, and dermatological agents (Centers for Disease Control and Prevention, 2021). WICs are a general category for anthropogenically derived chemicals that include some pharmaceuticals, pesticides, and some volatile organic compounds all quantified in parts per billion (micrograms per liter). Pharmaceuticals were analyzed in parts per trillion (nanograms per liter) and include the most commonly used prescription drugs for antibiotics, opioids, diabetes, anti-inflammatories, and high blood pressure and cholesterol (statins) medication. Atrazine is also used as an antihistamine and is referred to in the results as "pharmaceutical atrazine," and thiabendazole is an antiworm medication (Drugs.com, 2021; WebMD, 2021).

In 2020, a subset consisting of 12 monitoring wells and 3 surface-water sites was sampled and analyzed for CECs, including pharmaceuticals, pesticides, and WICs. Chloride-bromide ratio analyses completed in Gahala (2017) identified wells that plotted near the sewage mixing curve or indicated a mix of sewage and dilute road salt. Monitoring wells with chloride-bromide ratio results in 2010 that plotted near the sewage mixing curve or mixture of sewage and dilute road salt were HEBR-09-03, 15-COR-I, 10-MAR-S, 9-MCH-S, MARN-09-01, 8-GRN-I, HARV-09-01, 4-RCH-S, and HEBR-08-02. Designated background wells were selected based on their chloride-bromide ratio plotting near high road salt mixing curves (7-HRT-S) or within the dilute ground-water range or with no clear proximity towards a particular mixing curve (8-GRN-D and WOOD-08-01).

A total of 172 CECs were analyzed for in the groundwater monitoring wells, and only 5 percent of the CECs were detected in either low or estimated concentrations, indicating a lack of widespread presence of CECs in the shallow groundwater resource and supporting the overall good quality of the shallow groundwater resource (table 8). A total of 9 CECs were detected at low concentrations in groundwater samples from 8 of the 12 (about 66 percent) monitoring wells (fig. 9; table 8); however, 3 of the 9 CECs also were detected in at least 1 blank QAQC sample (appendix 1, table 1.2; Gahala and Gruhn, 2022). These CECs are formatted in bold italics and flagged as a potential but not verified detection (table 8). The most common uses of the pharmaceuticals are provided in the parentheses in table 8 and were obtained from searching the database available at https://www.drugs.com/ (Drugs. com, 2021). Of the 111 pharmaceuticals analyzed, 2 (2 percent) pharmaceuticals were detected in groundwater samples; caffeine was detected in 1 well and nicotine was detected in 6 wells (table 8). Of the 61 WICs analyzed, 7 (about 11 percent) were detected in five groundwater samples, 10-MAR-S and 9-MCH-S had detections of 2 WICs. A total of three WIC detections are flagged because they also were detected in at least one blank QAQC sample (appendix 1, table 1.2; Gahala and Gruhn, 2022). N,N-Diethyl-m-toluamide (commonly known as DEET) was detected in every groundwater sample and surface-water sample, including all the QAQC samples (appendix 1, table 1.2; Gahala and Gruhn, 2022). This compound was removed from the results tables and discussions because of its ubiquitous detection and questionable source. The low level and often estimated concentrations not only indicate the vulnerability of groundwater with respect to sewage and septic effluents but also highlight the potential capacity of the aquifer to degrade or mitigate some of the detected CECs.

Table 6. Results for concentrations of trace metals from samples collected from the McHenry County Groundwater Monitoring Network wells, 4 National Water-Quality Assessment wells, and 2 surface-water sites in McHenry County, Illinois, 2020.—Continued

ſDa	ata available from U.S.	Geological Survey, 2020b	o. ug/L, microgram per liter: <.	less than reporting level; E.	estimated; IL, Illinois; —, no data]

Nickel (μg/L)¹	Silver (µg/L)	Antimony (µg/L)	Arsenic (µg/L)	Selenium (µg/L)	Uranium (µg/L)	Lithium (µg/L)	Strontium (µg/L)	Boron (μg/L)
E1.06	<2	< 0.12	< 0.20	0.50	0.41	6.21	107	35
E2.08	<1	< 0.06	0.17	< 0.05	1.55	6.79	84	33
< 0.20	<2	< 0.12	0.53	< 0.1	0.36	10.87	175	161
E3.23	<2	< 0.12	< 0.20	0.18	1.50	9.72	156	28
E1.77	<1	0.12	1.52	0.32	0.85	2.70	146	38
E1.14	<1	0.23	1.63	0.19	0.60	3.11	278	43

¹Nickel concentrations near the method report limit are qualified because similar concentrations were detected in the equipment and field blanks (appendix 1). Italicized values are unrelated to the laboratory designated estimated values.

The surface-water sampling and analysis of CECs provide a general indication of the presence of CECs at locations downstream from multiple or specific wastewater-treatment discharges. The results are based on a single sample collection at three surface-water sites and do not constitute a robust analysis of the sources, range in concentrations, seasonality, and persistence. Of the 172 CECs analyzed, 35 (about 20 percent) CECs were detected in at least 1 surface-water sample and 18 were detected at all 3 surface-water sites. Of the 111 pharmaceuticals analyzed, 20 were detected in the surface-water samples, including the pharmaceutical atrazine. Of the 61 WICs analyzed, 15 were detected in surface-water samples, including the pesticide atrazine (table 8). Surface waters indicated the highest concentrations and most frequent detection of CECs. These locations were selected based on proximity to wastewater-treatment plant effluent, and the detections highlight the limits of the treatment systems.

The analytical methods for pharmaceutical compounds, in nanograms per liter, and WICs, in micrograms per liter, naturally create high variability in the quantified results of detections. For detailed discussion of the groundwater QAQC results, refer to appendix 1. The groundwater results discussed consider only the constituents that were detected at low or estimated quantities in the sample or replicate and omit consideration or discussion of constituents detected in the blank samples. No equipment blanks were collected for the surfacewater sample collection. Surface-water results are considered acceptable for discussion of general presence and frequency of detections among the three samples.

Pharmaceuticals are used by individuals for health reasons and include over-the-counter medication such as aspirin and pseudoephedrine, and medications prescribed by a physician such as Lipitor, albuterol, amoxicillin, and

many others (U.S. Environmental Protection Agency, 2009b). Most ingested pharmaceuticals are not fully metabolized, so parts are excreted in human waste. Metabolized and unmetabolized pharmaceuticals are discharged in domestic sewage (U.S. Environmental Protection Agency, 2009b; U.S. Geological Survey, 2021a, b). Carbamazepine is an anticonvulsant used to prevent and control seizures (Drugs.com, 2021) and was detected (2.9 nanograms per liter [ng/L]) in the replicate sample from 9-MCH-S, but was not detected in the primary environmental sample (appendix 1, table 1.2; Gahala and Gruhn, 2022). Caffeine was detected in HEBR-09-03, estimated at 5.5 ng/L. Pharmaceutical atrazine was detected in the replicate of 10–MAR–S at 5.03 ng/L (appendix 1, table 1.2; Gahala and Gruhn, 2022) but was not detected in the environmental sample at 10-MAR-S. Nicotine was detected at low and estimated concentrations in 50 percent (6 out of 12) of the groundwater samples. The number of CEC detections in the MCGMN are shown in figure 9.

Fox River and Kishwaukee River had the greatest number of pharmaceuticals detected (18 in each), whereas Nippersink Creek had 16 pharmaceutical detections. The highest concentrations detected in the three surface-water sites were for methylbenzotriazole, hexamethylenetetramine, fexofenadine, metformin, sulfamethoxazole, carbamazepine, and caffeine. The Kishwaukee River had the higher detections of lidocaine, methocarbamol, tramadol, metoprolol, venlafaxine, famotidine, and desvenlafaxine than the other two surface water sites. Nippersink Creek had higher concentrations of atrazine, cotinine, and acetaminophen than the Fox River or Kishwaukee River. Of the 13 pharmaceuticals detected in all 3 surface-water samples, concentrations were generally highest in Fox River and Kishwaukee River.

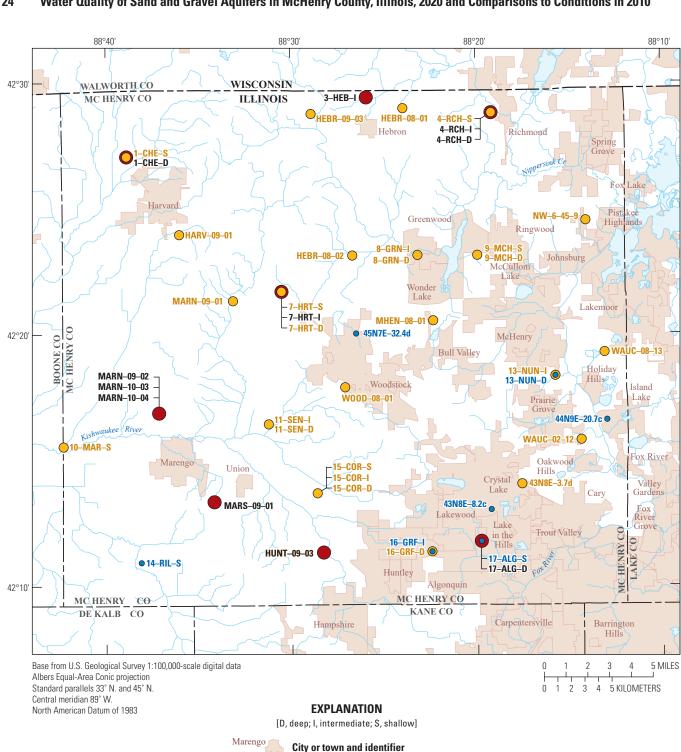


Figure 6. Distribution of arsenic concentrations in the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment monitoring wells, McHenry County, Illinois, 2020.

Less than 0.2

Greater than 10

0.2 to 10

14-RIL-S

1-CHE-S

1-CHE-D

Monitoring well arsenic concentration, in micrograms per liter, and identifier

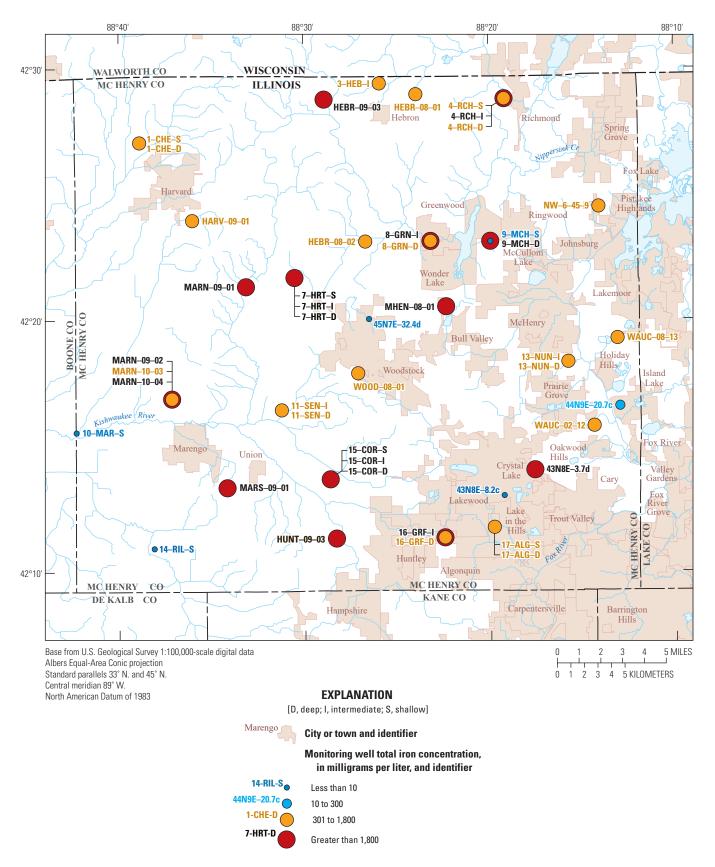


Figure 7. Distribution of iron concentrations in the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment monitoring wells, McHenry County, Illinois, 2020.

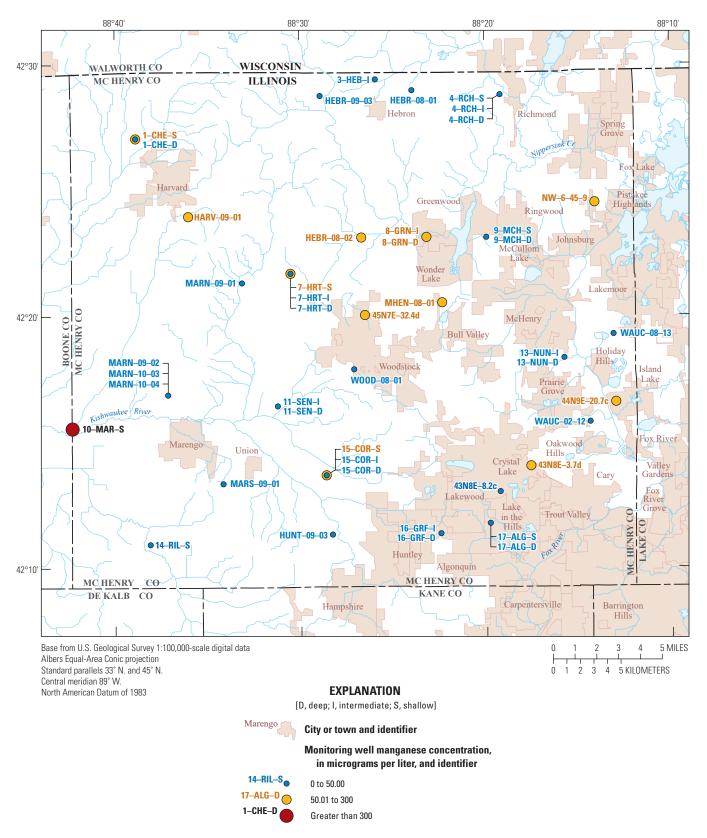


Figure 8. Distribution of manganese concentrations in the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment monitoring wells, McHenry County, Illinois, 2020.

Table 7. Results for concentrations of nutrients from samples collected from the McHenry County Groundwater Monitoring Network wells, 4 National Water-Quality Assessment wells, and 2 surface-water sites in McHenry County, Illinois, 2020.

[Data available from U.S. Geological Survey, 2020b. mg/L, milligram per liter; N, nitrogen; NO₃, nitrate; NO₂, nitrite; P, phosphorus; <, less than; U, nondetect; M, detect; IL, Illinois; —, no data]

Field identifier	Nitrate (NO ₃ + NO ₂) (mg/L as N)	Nitrite (mg/L as N)	Orthophosphate (mg/L as P)	Ammonia (mg/L as N)	Hydrogen sulfide (sniff test)
14-RIL-S	4.74	0.005	< 0.004	< 0.05	U
HUNT-09-03	< 0.04	0.001	0.096	0.80	U
16-GRF-I	< 0.04	< 0.001	0.017	0.77	U
16-GRF-D	< 0.04	0.006	0.011	1.01	M
17-ALG-S	1.57	0.027	0.009	0.37	U
17-ALG-D	< 0.04	< 0.001	0.007	0.57	U
MARS-09-01	< 0.04	< 0.001	0.008	1.49	U
15-COR-S	< 0.04	< 0.001	0.012	0.22	M
15-COR-I	< 0.04	< 0.001	0.012	0.24	U
15-COR-D	< 0.04	< 0.001	0.025	0.24	M
10-MAR-S	1.52	0.054	< 0.004	0.02	U
WAUC-02-12	< 0.04	< 0.001	0.012	0.58	M
11-SEN-I	< 0.04	< 0.001	0.005	0.17	U
11-SEN-D	< 0.04	< 0.001	0.008	1.11	U
MARN-09-02	< 0.04	< 0.001	0.023	1.54	U
WOOD-08-01	< 0.04	0.001	0.027	1.04	M
13–NUN–I	< 0.04	< 0.001	0.011	0.32	U
13–NUN–D	< 0.04	< 0.001	0.015	0.26	M
WAUC-08-13	< 0.04	< 0.001	0.014	0.16	M
MHEN-08-01	< 0.04	< 0.001	0.004	0.11	U
MARN-09-01	< 0.04	< 0.001	0.035	0.76	U
7–HRT–S	< 0.04	0.001	0.008	0.09	U
7–HRT–I	< 0.04	< 0.001	0.009	0.78	U
7–HRT–D	< 0.04	< 0.001	0.011	0.98	M
9–MCH–S	5.33	< 0.001	< 0.004	< 0.01	U
9–MCH–D	< 0.04	< 0.001	0.141	1.55	M
8-GRN-I	< 0.04	< 0.001	0.005	0.09	U
8-GRN-D	< 0.04	< 0.001	0.005	0.09	M
HEBR-08-02	< 0.04	< 0.001	0.004	0.07	U
HARV-09-01	< 0.04	< 0.001	0.016	0.02	M
NW-6-45-9	0.16	0.007	0.005	0.03	U
1-CHE-S	< 0.04	0.002	0.005	0.11	U
1-CHE-D	< 0.04	< 0.001	0.008	1.84	M
HEBR-09-03	< 0.04	< 0.001	0.004	0.04	M
4–RCH–S	0.08	0.003	0.014	0.4	U
4–RCH–I	< 0.04	< 0.001	0.073	0.95	U
4–RCH–D	< 0.04	< 0.001	0.026	0.71	U
HEBR-08-01	< 0.04	0.001	0.059	0.53	U
3–HEB–I	< 0.04	< 0.001	0.017	1.1	M
MARN-10-03	< 0.04	< 0.001	0.056	0.89	M
MARN-10-04	< 0.04	0.001	0.047	2.04	U

Table 7. Results for concentrations of nutrients from samples collected from the McHenry County Groundwater Monitoring Network wells, 4 National Water-Quality Assessment wells, and 2 surface-water sites in McHenry County, Illinois, 2020.—Continued

[Data available from U.S. Geological Survey, 2020b. mg/L, milligram per liter; N, nitrogen; NO₃, nitrate; NO₂, nitrite; P, phosphorus; <, less than; U, nondetect; M, detect; IL, Illinois; —, no data]

Field identifier	Nitrate (NO ₃ + NO ₂) (mg/L as N)	Nitrite (mg/L as N)	Orthophosphate (mg/L as P)	Ammonia (mg/L as N)	Hydrogen sulfide (sniff test)
43N8E-8.2c	2.47	0.001	0.010	0.01	U
45N7E-32.4d	< 0.04	< 0.001	0.004	< 0.05	U
43N8E-3.7d	< 0.04	< 0.001	0.007	0.04	U
44N9E-20.7c	0.13	0.004	0.008	< 0.01	U
Nippersink Creek Near Spring Grove, IL	1.96	0.036	1.992	<0.01	_
Fox River at Huntley Road at Carpentersville, IL	< 0.04	< 0.001	< 0.004	< 0.01	_
Kishwaukee River at Marengo, IL	_	_	_	_	_

Wastewater indicators are chemical compounds that indicate the presence of domestic and industrial waste in water sources such as streams and aquifers (U.S. Geological Survey, 2021b). Domestic WICs include caffeine and cotinine (a byproduct of nicotine metabolism in the liver; Bradley and others, 2007). Industrial wastewater indicators include compounds used in plastic components, surfactant metabolites, fire retardants, and many others (Tomasek and others, 2012). Several WICs were detected in the groundwater samples from the MCGMN, and many more were detected in surface-water samples from Nippersink Creek, Fox River, and the Kishwaukee River. The following paragraphs discuss the results of WICs that are classified as "organic other" and do not fall under the pharmaceutical or pesticide/herbicide categories discussed in the previous sections.

FYROL FR 2 (0.05 μ g/L) was detected in the groundwater at 10–MAR–S and is a common additive flame retardant in commercially used flexible and rigid polyurethane foams (Chemical Book, 2021). Indole (0.08 μ g/L) was detected in the groundwater at 8–GRN–D and has many uses (cleaning agent, food additive, veterinary drug, fragrance, and more; PubChem, 2021a). A common food additive and fragrance (PubChem, 2021b), 2,6-Dimethylnapththalene (0.01 μ g/L), was detected in the groundwater at HEBR–09–03.

Detections of WICs in surface water were most frequent and at greater concentrations in the Fox River but were generally ubiquitous in all three surface-water locations (table 8). WICs (organics/other, not including pesticides) detected in samples from all three surface-water locations include β -sitosterol (cosmetics, detergent, veterinary drug; PubChem, 2021c) and diethyl phthalate (plasticizer, solvent, household products, and more; PubChem, 2021d). No endocrine disrupting compounds such as 4-nonylphenols were detected at concentrations greater than reporting level in any of the surface-water samples. Endocrine disrupting compounds are known to cause various hormone and reproductive effects on aquatic life (U.S. Environmental Protection Agency, 2022b).

Pesticides are chemicals used to prevent, destroy, or repel living organisms from where they are not wanted (U.S. Environmental Protection Agency, 2009b). Pesticides are frequently used to increase crop yields and improve the quality of the crops; however, they can be transported to groundwater and surface water (U.S. Geological Survey, 1999, 2021a). Results for pesticide analyses are presented in table 8.

Four pesticides (prometon, atrazine, metolachlor, and pentachlorophenol) were detected in the water samples collected in 2020. Prometon is an herbicide used to control the emergence of most annual and many perennial broadleaf weeds and grasses in nonagricultural areas (U.S. Environmental Protection Agency, 1990). Prometon was detected in 9–MCH–S (estimated 0.04 ng/L; table 8) and also was detected in the associated replicate sample (0.04 ng/L; table 1.2). No detections of prometon were in the surfacewater samples.

Atrazine is a common and widely used herbicide that can be applied before and after planting to control broadleaf and grassy weeds (U.S. Environmental Protection Agency, 2021a). Atrazine is used primarily in agriculture, but to a lesser extent, it is sometimes used on residential lawns and golf courses (U.S. Environmental Protection Agency, 2021a). As indicated in table 8, atrazine also can be used as antianxiety medication; therefore, atrazine has two results, the pharmaceutical method analyzed concentrations in parts per trillion (nanograms per liter), and the WIC analysis measured concentrations in parts per billion (micrograms per liter). Atrazine was not detected in any environmental samples from groundwater wells, although it was detected in the replicate sample for 9-MCH-S (3.29 μg/L) and the replicate sample for 10–MAR-S at the parts per trillion concentration of 5.03 ng/L (table 1.2). This inconsistency in detection between the parts per billion and parts per trillion concentrations and between the environmental and replicate samples highlights the variability of

detections at these low levels with the laboratory methods applied as described in the "Methods of Study" section and detailed in the cited references.

All three surface-water sites had detections for atrazine (table 8). Atrazine was highest in Nippersink Creek with a concentration of 299 ng/L, as detected with the pharmaceutical method, and also was detected at 0.31 μg/L (315 ng/L) with the WIC analytical method. Consistent concentrations of atrazine between the pharmaceutical analysis and WIC analytical method also were noted for the Fox River and Kishwaukee River. Metolachlor is a broad-spectrum herbicide used for general weed control in numerous agricultural crops, lawns and turf, ornamental plants, trees, shrubs, right of ways, fence and hedge rows, and in forestry (U.S. Environmental Protection Agency, 1995). Metolachlor was not detected in any of the groundwater wells, although it was detected at all surface-water sites. One surface-water site (Fox River) had a detection for pentachlorophenol (estimated 0.13 µg/L), but no other detections of this pesticide were in surface water or groundwater.

Suitability of Water for Drinking

The largest concern for causing negative effects to the quality of groundwater as a source of drinking water is for arsenic, iron, manganese, and dissolved solids, which are naturally occurring in the sand and gravel aquifers in McHenry County (table 3). The MCL of 10 μg/L for arsenic was exceeded in about 24 percent of the wells, and chronic ingestion of arsenic in exceedance of the MCL has been linked to increased risk of lung, bladder, skin, liver, kidney, and prostate cancer (U.S. Environmental Protection Agency, 2018; table 3). Manganese, which can cause toxicity to the nervous system, has a health-based drinking water advisory of 300 µg/L, and that was exceeded at one location (U.S. Environmental Protection Agency, 2018; table 3). Aesthetically, the SMCL of 500 mg/L for dissolved solids was exceeded at 13 locations (about 29 percent). Dissolved solids in exceedance of the SMCL can cause objectionable taste and, at concentrations greater than 1,000 mg/L, can have laxative effects (U.S. Environmental Protection Agency, 2018; table 3). Two sites had dissolved solids concentrations that exceeded 1,000 mg/L (4-RCH-S and 43N8E-3.7d) (table 5). Aside from health concerns, the main concern with most of these constituents is that treatment would be needed to prevent damage to infrastructure and staining of household appliances and laundry. Another major concern for drinking water is nutrient concentrations, especially nitrate and nitrite, which have an MCL of 10 mg/L and 1 mg/L, respectively. No exceedances of the MCL were measured for either nitrate or nitrite. In general, the groundwater in McHenry County is suitable for drinking water, and some of the aesthetic issues could be mitigated with home water-treatment systems such as water softeners. The concentrations of many of these constituents could also

be reduced with an at-home water-treatment system such as a carbon filtration, ion-exchange systems, and reverse-osmosis systems (Thomas and Eckberg, 2015).

The implications of low concentrations of CECs in surface water on drinking water and aquatic life are unknown, and it is important to reiterate that these results are a single sample at three sites and do not constitute a robust understanding of the distribution and persistence of these CECs. These pharmaceuticals, pesticides, and low concentrations of WICs are potentially ingested by consumers of stream-sourced drinking water and aquatic biota within the streams (Wilson and others, 2011). In a study by Benson and others (2017), the potential long-term toxicity of CECs to humans for the most frequently detected and highest concentrations of WICs and pharmaceuticals were quantified with a risk-assessment analysis. The study concluded that the exposure to these CECs is not likely to pose a public health concern; however, additional toxicity data would be helpful to improve the calculated risks (Benson and others, 2017). Also, studies have determined that many of these CECs may degrade naturally in the streams (Bradley and others, 2007) whereas others persist miles downstream (Barber and others, 2013), including some of the contaminants detected in this study (sulfamethoxazole and carbamazepine). The persistence, seasonality, sources, and range in concentrations and effects of CECs in surface water were not investigated as part of this study. Frequent testing by water-treatment facilities provides the data on exposure levels and baseline understanding for potential effects.

Comparison to 2010 Water-Quality Results

The following sections compare the 2020 water-quality results to the baseline results of 2010. Differences in water quality on the decadal time scale were evaluated with the WSR test (Helsel, 2012) to identify statistically significant differences. Shifts in chloride to bromide ratio plots also are evaluated to identify potential changes in water quality.

Statistical Comparison

The results of the WSR to assess changes in the water-quality results between 2010 and 2020 indicated a few statistically significant differences between the two sampling events (table 9). Differences were statistically significant among all wells for concentrations of aluminum, ammonia, arsenic, barium, bromide, calcium, magnesium, molybdenum, orthophosphate, silica, specific conductance, sulfate, and dissolved solids. Statistically significant differences (*p*-value of less than [<] 0.05) were calculated among shallow (20.3–58.7 ft below ground surface), intermediate (62.3–120.6 ft below ground surface), or deep (139–344.4 ft below ground surface) aquifer

Table 8. Concentrations of pharmaceuticals and wastewater indicator compounds detected in samples collected from selected wells within the McHenry County Groundwater Monitoring Network and three surface-water sites in McHenry County, Illinois, 2020.

[Data available from U.S. Geological Survey, 2020b. Bold values are detections greater than the detection or reporting level. Bold italic values had a detection in one or more blank quality assurance and quality control samples. CEC, contaminant of emerging concern; Pharm, pharmaceuticals; ng/L, nanogram per liter; <, less than the reporting level; E, estimated; WIC, wastewater indicator compound; µg/L, microgram per liter; HHCB, Hexahydrohexamethyl cyclopentabenzopyran; TBEP, Tris(2-butoxyethyl) phosphate]

CEC type	Constituent	Units	15-COR-I	10-MAR-S	W00D-08-01	MARN-09-01	7-HRT-S
Pharm	Hexamethylenetetramine (urinary antiseptic)	ng/L	<110	<110	<110	<110	<110
Pharm	Guanylurea (metformin transformation product)	ng/L	<140	<140	<140	<140	<140
Pharm	Gabapentin (anticonvulsant)	ng/L	<160	<160	<160	<160	<160
Pharm	Acetaminophen (pain relief)	ng/L	<20	<20	<20	<20	<20
Pharm	Caffeine (stimulant)	ng/L	<91	<91	<91	<91	<91
Pharm	Carbamazepine (anticonvulsant)	ng/L	<11	<11	<11	<11	<11
Pharm	Cotinine (nicotine metabolite)	ng/L	<6.4	< 6.4	< 6.4	<6.4	< 6.4
Pharm	Sulfamethoxazole (antibiotic)	ng/L	<20	<20	<20	<20	<20
Pharm	Lidocaine (skin conditions)	ng/L	<8	<8	<8	<8	<8
Pharm	Triamterene (diuretic)	ng/L	<5.2	<5.2	< 5.2	<5.2	<5.2
Pharm	Metformin (antidiabetic)	ng/L	<20	<20	<20	<20	<20
Pharm	Nicotine (stimulant)	ng/L	<80	<80	E35.3	E35.2	E36.5
Pharm	Methocarbamol (muscle relaxant)	ng/L	<20	<20	<20	<20	<20
Pharm	Fexofenadine (antihistamine)	ng/L	<44	<44	<44	<44	<44
Pharm	Methylbenzotriazole (corrosion inhibitors)	ng/L	<80	<80	<80	<80	<80
Pharm	Tramadol (pain relief)	ng/L	<7.4	<7.4	<7.4	<7.4	<7.4
Pharm	Metoprolol (cardiac)	ng/L	<10	<10	<10	<10	<10
Pharm	Venlafaxine (antidepressant)	ng/L	<5.2	<5.2	< 5.2	< 5.2	< 5.2
Pharm	Famotidine (gastric)	ng/L	<34	<34	<34	<34	<34
Pharm	Desvenlafaxine (antidepressant)	ng/L	<84	<84	<84	<84	<84
Pharm (pesticide)	Atrazine (antianxiety) ¹	ng/L	<20	<20	<20	<20	<20
WIC (pesticide)	Atrazine (pesticide) ¹	$\mu g/L$	< 0.16	< 0.16	< 0.16	< 0.16	< 0.16
WIC	Diethyl phthalate	$\mu g/L$	< 0.4	0.11	< 0.4	< 0.4	< 0.4
WIC	Fluoranthene	$\mu g/L$	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
WIC	Isophorone	$\mu g/L$	< 0.05	< 0.05	< 0.05	< 0.05	< 0.05
WIC	Phenanthrene	$\mu g/L$	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
WIC	FYROL FR 2	$\mu g/L$	< 0.32	0.05	<1.92	< 0.32	<1.92
WIC	Triclosan	$\mu g/L$	< 0.32	< 0.32	< 0.32	< 0.32	< 0.32
WIC	β-Stigmastanol	$\mu g/L$	<3.4	<3.4	<3.4	<3.4	<3.4
WIC	2,6-Dimethylnaphthalene	$\mu g/L$	< 0.04	< 0.04	< 0.04	< 0.04	< 0.04
WIC	9,10-Anthraquinone	$\mu g/L$	< 0.04	< 0.04	< 0.04	< 0.04	< 0.04
WIC	β-Sitosterol	$\mu g/L$	<4.8	<4.8	<4.8	<4.8	<4.8
WIC	Cholesterol	$\mu g/L$	<1.6	<1.6	<1.6	<1.6	<1.6
WIC	ННСВ	$\mu g/L$	< 0.04	< 0.04	< 0.04	< 0.04	< 0.04
WIC	Indole	$\mu g/L$	< 0.04	< 0.04	< 0.04	< 0.04	< 0.04
WIC	TBEP	$\mu g/L$	< 0.64	< 0.64	<1.92	< 0.64	<1.92
WIC	Triethyl citrate	$\mu g/L$	< 0.16	< 0.16	< 0.16	E0.05	< 0.16
WIC	Triphenyl phosphate	$\mu g/L$	< 0.08	< 0.08	< 0.08	< 0.08	< 0.08
WIC (pesticide)	Pentachlorophenol	$\mu g/L$	<1.6	<1.6	<1.6	<1.6	<1.6
WIC (pesticide)	Prometon	$\mu g/L$	< 0.16	< 0.16	< 0.16	< 0.16	< 0.16
WIC (pesticide)	Metolachlor	μg/L	< 0.04	< 0.04	< 0.04	< 0.04	< 0.04

Table 8. Concentrations of pharmaceuticals and wastewater indicator compounds detected in samples collected from selected wells within the McHenry County Groundwater Monitoring Network and three surface-water sites in McHenry County, Illinois, 2020.—Continued

[Data available from U.S. Geological Survey, 2020b. Bold values are detections greater than the detection or reporting level. Bold italic values had a detection in one or more blank quality assurance and quality control samples. CEC, contaminant of emerging concern; Pharm, pharmaceuticals; ng/L, nanogram per liter; <, less than the reporting level; E, estimated; WIC, wastewater indicator compound; $\mu g/L$, microgram per liter; HHCB, Hexahydrohexamethyl cyclopentabenzopyran; TBEP, Tris(2-butoxyethyl) phosphate]

9-MCH-S	8-GRN-I	8–GRN–D	HEBR-08-02	HARV-09-01	HEBR-09-03	4-RCH-S	Nippersink Creek	Fox River	Kishwaukee River
<110	<110	<110	<110	<110	<110	<110	E14	E212	E35.1
<140	<140	<140	<140	<140	<140	<140	29	32	171
<160	<160	<160	<160	<160	<160	<160	<160	53	36.5
<20	<20	<20	<20	<20	<20	<20	E22.8	<20	<20
<91	<91	<91	<91	<91	E5.5	<91	E24.5	E24	E12.4
<11	<11	<11	<11	<11	<11	<11	4.85	9	6.5
<6.4	< 6.4	<6.4	<6.4	< 6.4	<6.4	< 6.4	4.36	<6.4	< 6.4
<20	<20	<20	<20	<20	<20	<20	E13.10	E32	E25.3
<8	<8	<8	<8	<8	<8	<8	6.24	17	30.4
<5.2	<5.2	<5.2	<5.2	<5.2	<5.2	<5.2	0.98	3.6	3.4
<20	<20	<20	<20	<20	<20	<20	33.8	64.1	59.6
23.2	E38.5	E38.6	<80	<80	<80	<80	<80	<80	<80
<20	<20	<20	<20	<20	<20	<20	17.8	9.9	19.6
<44	<44	<44	<44	<44	<44	<44	E20.31	E70.8	E65.3
<80	<80	<80	<80	<80	<80	<80	E38.1	E113	E28.3
<7.4	<7.4	<7.4	<7.4	<7.4	<7.4	<7.4	<7.4	6.2	10.7
<10	<10	<10	<10	<10	<10	<10	<10	6.4	18.1
<5.2	<5.2	<5.2	<5.2	< 5.2	<5.2	<5.2	3.76	3.8	13.2
<34	<34	<34	<34	<34	<34	<34	<34	E9.8	E12.5
<84	<84	<84	<84	<84	<84	<84	E8.01	E9.7	E30.7
<20	<20	<20	<20	<20	<20	<20	299	159	84.9
< 0.16	< 0.16	< 0.16	< 0.16	< 0.16	< 0.16	< 0.16	0.31	0.16	0.13
< 0.4	< 0.4	< 0.4	< 0.4	< 0.4	< 0.4	< 0.4	0.17	0.12	0.29
< 0.02	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02	0.05	0.01
< 0.05	< 0.05	< 0.05	< 0.05	< 0.05	< 0.05	< 0.05	< 0.05	0.02	< 0.05
< 0.02	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02	0.02	0.02
< 0.32	< 0.32	< 0.32	< 0.32	<1.92	< 0.32	< 0.32	<1.92	<1.92	<1.92
< 0.32	< 0.32	< 0.32	< 0.32	< 0.32	< 0.32	< 0.32	< 0.32	E0.06	< 0.32
<3.4	<3.4	<3.4	<3.4	<3.4	<3.4	<3.4	<3.4	E0.47	E0.45
< 0.04	< 0.04	< 0.04	< 0.04	< 0.04	0.01	< 0.04	< 0.04	< 0.04	< 0.04
< 0.04	< 0.04	< 0.04	< 0.04	< 0.04	< 0.04	< 0.04	< 0.04	E0.03	< 0.04
<4.8	<4.8	<4.8	<4.8	<4.8	<4.8	<4.8	E0.66	E0.84	E0.58
E0.09	<1.6	<1.6	<1.6	<1.92	<1.6	<1.6	<1.92	<1.92	<1.92
< 0.04	< 0.04	< 0.04	< 0.04	< 0.04	< 0.04	< 0.04	0.02	< 0.04	0.07
< 0.04	< 0.04	0.08	< 0.04	< 0.04	< 0.04	< 0.04	< 0.04	< 0.04	< 0.04
< 0.64	< 0.64	< 0.64	< 0.64	<1.92	< 0.64	< 0.64	E0.95	E0.71	<1.92
< 0.16	< 0.16	< 0.16	< 0.16	< 0.16	< 0.16	< 0.16	< 0.16	< 0.16	E0.13
< 0.08	< 0.08	< 0.08	< 0.08	< 0.08	< 0.08	< 0.08	< 0.08	< 0.08	0.04
<2.56	<1.6	<1.6	<1.6	<1.6	<1.6	<1.6	<1.6	E0.13	<1.6
E0.04	< 0.16	< 0.16	< 0.16	< 0.16	< 0.16	< 0.16	< 0.16	< 0.16	< 0.16
< 0.04	< 0.04	< 0.04	< 0.04	< 0.04	< 0.04	< 0.04	0.36	0.12	0.09

 $^{\rm l} Common$ use of pharmaceutical from Drugs.com, 2021.

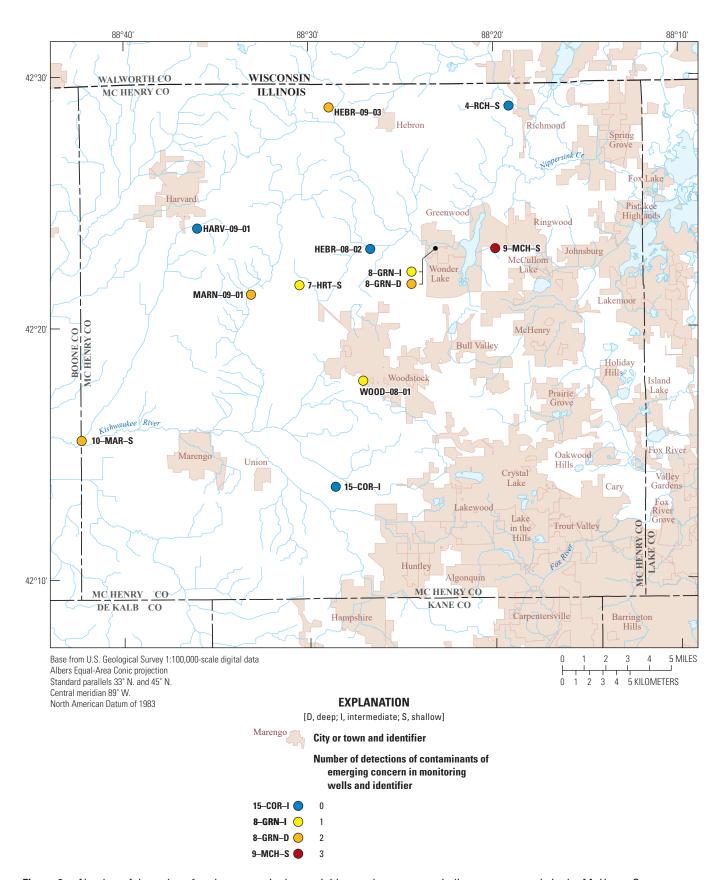


Figure 9. Number of detections for pharmaceuticals, pesticides, and wastewater indicator compounds in the McHenry County Groundwater Monitoring Network wells and four National Water-Quality Assessment monitoring wells, McHenry County, Illinois, 2020.

groups between the 2010 and 2020 water-quality results. Statistically significant differences were noted for aluminum (deep), arsenic (shallow), barium (deep and intermediate), calcium (deep and intermediate), magnesium (deep and intermediate), manganese (intermediate), molybdenum (deep and intermediate), orthophosphate (deep and shallow), and silica (deep and intermediate) (table 9). In figure 10, a change plot shows pairs of linked points that are the two samples collected in 2010 and 2020 for a given well. If both samples for a given well were censored, the lower value was recensored to the higher value. If only one of the samples was uncensored but less than the other sample's reporting level, the uncensored value was censored to match the censored sample. These "within pair" recensored values were only computed and used for plotting; they were not used in the statistical analyses. Censored values are indicated with open circles. Whether the change between 2010 and 2020 was downward, upward, or the same (common when both values were censored) is indicated by the color of the points and line.

Increases generally were detected in the intermediate and deep parts of the sand and gravel aquifer, and decreases were generally detected in the shallow part of the sand and gravel aquifer (fig. 10). The mixed distribution of increases and decreases among the various constituents and aquiferdepth groups could be reflecting dissolution and mobility of some of the redox sensitive constituents and dilution of some constituents in the shallow aquifer depths. These differences may be attributed to a combination of stable population of the past decade (2010–20), to land-use management practices, or potentially to the recent wet years of 2017 through 2019 causing a dilution of the major ions in the shallow parts of the aquifer.

The descriptive statistics and a trilinear Piper diagram (Piper, 1944; fig. 11*A*, *B*) also were reviewed to identify relative changes in the range and average of selected constituents of interest. A slight shift was seen in the anions in the Haeger-Beverly Unit (shallow aquifer) between 2010 and 2020 with a decrease in sulfate and chloride (fig. 11*A*). The trilinear Piper diagram for all other wells in the network indicates overall consistency in concentrations of major ions from 2010 to 2020 (fig. 11*B*). The minimum, median, and maximum for selected constituents for 2010 and 2020 are listed in table 9. Only calcium, magnesium, and silica indicated a slight increase in minimum, median, and maximum concentrations.

Table 9. Summary statistics (minimum, maximum, and median) for selected constituents, results of the Wilcoxon signed-rank test for groundwater-quality changes between 2010 and 2020 for uncensored data and the paired Prentice-Wilcoxon test for uncensored data, and evaluation of differences between the 2010 and 2020 aquifer-depth groups of shallow, intermediate, and deep parts of the sand and gravel aquifers in McHenry County, Illinois.

[Bold denotes statistically significant; *p*-value, probability value; μ g/L, microgram per liter; <, less than reporting level; PPW, paired Prentice-Wilcoxon test; —, no data or not applicable; mg/L, milligram per liter; N, nitrogen; WSR, Wilcoxon signed-rank test; μ S/cm, microsiemens per centimeter at 25 degrees Celsius]

Constituent	Year	Number of censored	Percent censored	Minimum	Median	Maximum	Test type	<i>p</i> -value for all wells	Deep wells <i>p</i> -value	Intermediate wells <i>p</i> -value	Shallow wells, p-value
Aluminum, μg/L	2010	27	63	<1.7	<1.7	38.9	PPW	0.0456	0.0462	_	_
Aluminum, $\mu g/L$	2020	43	100	<3	<3	<6	_	_	_	_	_
Ammonia and ammonium, mg/L as N	2010	4	9	< 0.01	0.28	1.88	PPW	0.0432	0.3054	0.7778	0.4971
Ammonia and ammonium, mg/L as N	2020	5	12	< 0.01	0.26	1.84	_	_	_	_	_
Arsenic, µg/L	2010	0	0	0.15	0.88	62	PPW	0.0013	0.1637	0.5869	0.0471
Arsenic, µg/L	2020	3	7	< 0.12	0.58	70.3	_	_	_	_	_
Barium, $\mu g/L$	2010	0	0	40.1	89.6	239	WSR	< 0.0001	0.003	0.0424	0.0745
Barium, µg/L	2020	0	0	32.9	78.4	207	_	_	_	_	_
Bromide, mg/L	2010	0	0	0.013	0.039	0.119	PPW	0.0014	0.1796	0.1035	0.1796
Bromide, mg/L	2020	4	9	< 0.011	0.032	0.254	_	_	_	_	_
Calcium, mg/L	2010	0	0	38.4	85.1	176	WSR	0.0247	0.0036	0.0447	0.1685
Calcium, mg/L	2020	0	0	41.8	88.1	184	_	_	_	_	_
Chloride, mg/L	2010	0	0	0.49	15.2	521	WSR	0.6814	0.1641	0.1712	0.1641
Chloride, mg/L	2020	0	0	0.55	17.2	517	_	_	_	_	_
Inorganic nitrogen (nitrate and nitrite), mg/L	2010	36	84	< 0.02	<0.02	10	PPW	0.2646	_	0.3359	0.3359
Inorganic nitrogen (nitrate and nitrite), mg/L	2020	35	81	< 0.04	< 0.04	5.33	_	_	_	_	_
Iron, μg/L	2010	0	0	4.9	1,480	3,450	PPW	0.0551	0.2967	0.1755	0.3017
Iron, μg/L	2020	5	12	<10	1,630	3,480	_	_	_	_	_
Magnesium, mg/L	2010	0	0	25.9	43.3	83.9	WSR	0.0026	0.003	0.0105	0.2207
Magnesium, mg/L	2020	0	0	27.8	45.2	95	_	_	_	_	
Manganese, μg/L	2010	0	0	0.45	30.2	667	WSR	0.0504	0.1264	0.0423	0.5076
Manganese, μg/L	2020	0	0	0.5	28.8	740	_	_	_	_	_
Molybdenum, μg/L	2010	0	0	0.388	2	33.8	WSR	0.0001	0.003	0.0345	0.2026
Molybdenum, µg/L	2020	0	0	0.297	1.65	6.04	_		_	_	_

Statistical Comparison

Table 9. Summary statistics (minimum, maximum, and median) for selected constituents, results of the Wilcoxon signed-rank test for groundwater-quality changes between 2010 and 2020 for uncensored data and the paired Prentice-Wilcoxon test for uncensored data, and evaluation of differences between the 2010 and 2020 aquifer-depth groups of shallow, intermediate, and deep parts of the sand and gravel aquifers in McHenry County, Illinois.—Continued

[Bold denotes statistically significant; *p*-value, probability value; µg/L, microgram per liter; <, less than reporting level; PPW, paired Prentice-Wilcoxon test; —, no data or not applicable; mg/L, milligram per liter; N, nitrogen; WSR, Wilcoxon signed-rank test; µS/cm, microsiemens per centimeter at 25 degrees Celsius]

Constituent	Year	Number of censored	Percent censored	Minimum	Median	Maximum	Test type	<i>p</i> -value for all wells	Deep wells <i>p</i> -value	Intermediate wells <i>p</i> -value	Shallow wells, p-value
Orthophosphate, mg/L	2010	0	0	0.02	0.067	0.468	PPW	<0.0001	0.003	0.099	0.0071
Orthophosphate, mg/L	2020	3	7	< 0.012	0.028	0.433	_	_	_	_	_
Potassium, mg/L	2010	0	0	0.78	1.63	7.15	WSR	0.1801	0.6477	0.6873	0.2778
Potassium, mg/L	2020	0	0	0.83	1.53	8	_	_	_	_	_
Silica, mg/L	2010	0	0	8	18.2	23.6	WSR	< 0.0001	0.0015	0.0012	0.7595
Silica, mg/L	2020	0	0	8.8	19.9	25.3	_	_	_	_	_
Sodium, mg/L	2010	0	0	3.87	11.6	181	WSR	0.9182	0.6669	1	0.6669
Sodium, mg/L	2020	0	0	4.33	13.8	180	_	_	_	_	_
Specific conductance, μS/cm	2010	0	0	512	754	2,260	WSR	0.0179	0.297	0.3341	0.2235
Specific conductance, μS/cm	2020	0	0	488	724	2,340	_	_	_	_	_
Sulfate, mg/L	2010	4	9	< 0.09	37.2	117	PPW	0.0071	0.1394	0.1341	0.1394
Sulfate, mg/L	2020	0	0	0.03	34	97.7	_	_	_	_	_
Dissolved solids, mg/L	2010	0	0	285	465	1,220	WSR	0.0015	0.1117	0.1418	0.1117
Dissolved solids, mg/L	2020	0	0	272	419	1,360	_	_	_	_	_

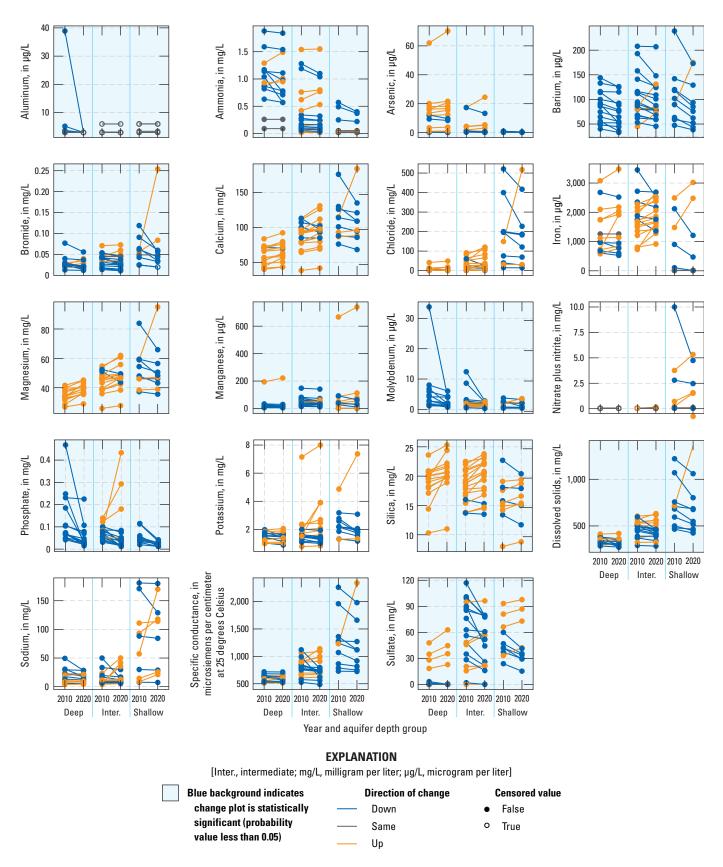


Figure 10. Differences in the shallow, intermediate, and deep parts of the sand and gravel aquifers for concentrations of selected constituents between 2010 and 2020 for water-quality samples in McHenry County, Illinois.

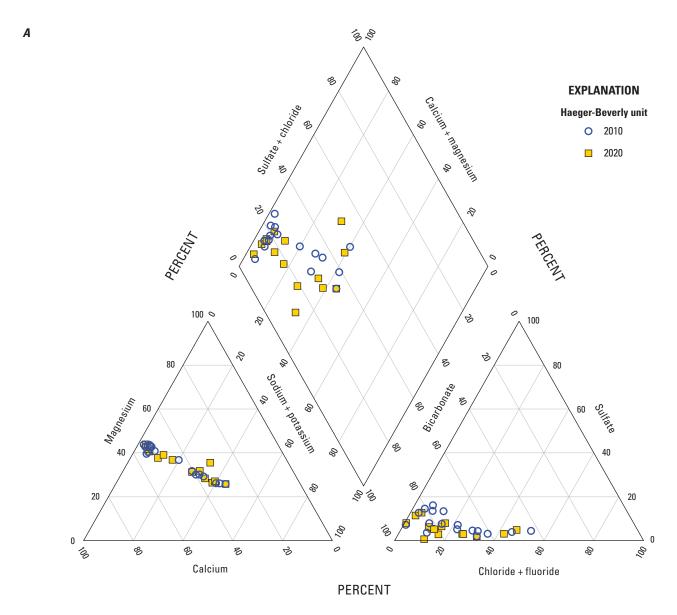


Figure 11. Major ions in wells from 2010 and 2020 in McHenry County, Illinois. *A*, wells completed in the Haeger-Beverly Unit; *B*, all other wells in the network.

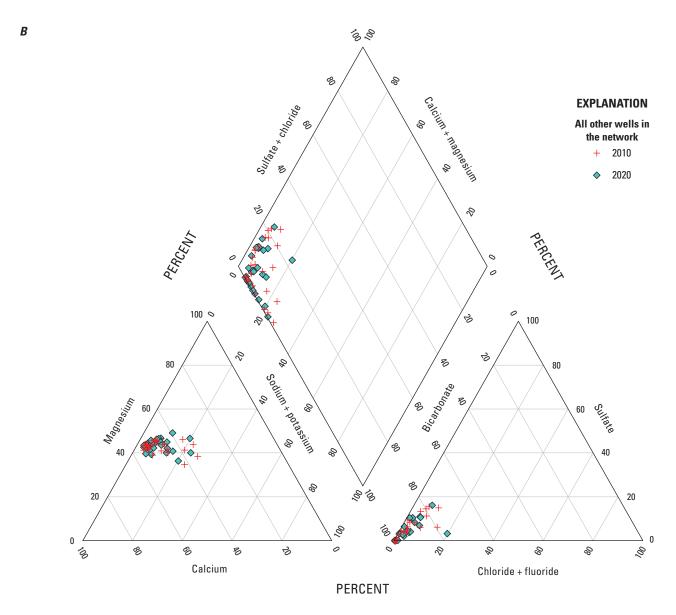


Figure 11. Major ions in wells from 2010 and 2020 in McHenry County, Illinois. *A*, wells completed in the Haeger-Beverly Unit; *B*, all other wells in the network.—Continued

Comparison of Chloride to Bromide Ratios

The methods for chloride-bromide ratio analyses used in Gahala (2017) were repeated for the 2020 samples. Bromide is a conservative ion (similar to chloride) that can be deposited from precipitation and infiltrate into groundwater (Davis and others, 1998) or dissolved from sedimentary rock such as carbonates, shale, and evaporites (Salameh and others,

2016). Concentrations of bromide in groundwater are usually 40–8,000 times less than concentrations of chloride, so small changes in the mass of bromide produce large variations in the ratios of chloride to bromide, which can help trace the source of the chloride (Davis and others, 1998; Gahala, 2017). The 2020 sample results were used, and the chloride-bromide ratios were plotted on the same graph as the source binary mixing curves to evaluate sources affecting water quality and assess any changes in water quality at each monitoring well since 2010 (figs. 12 and 13). The proximity of the

plotted chloride-bromide ratio to the source-mixing curve indicates the potential source for the chloride in the surrounding groundwater. Chloride-bromide ratios were estimated for monitoring wells with bromide concentrations less than the reporting level (<0.020 mg/L) by halving the reporting level (0.01) to estimate a point on the curve. Plot points for 17– ALG–S, 7–HRT–I, 4–RCH–I, and 45N7E–32.4d are estimated and do not plot on any curves but plot to the left of the curves.

In general, the 2020 chloride-bromide ratios plotted similarly to where they plotted in 2010, indicating the source of the chloride had not changed. CECs identified in 10–MAR–S, MARN–09–01, 9–MCH–S, 8–GRN–D, and HEBR–09–03 are consistent with the source water identified in the chloride-bromide ratio analysis. A few locations did indicate shifts toward a specific binary-mixing curve, and these locations are discussed in this section (fig. 13). The chloride-bromide

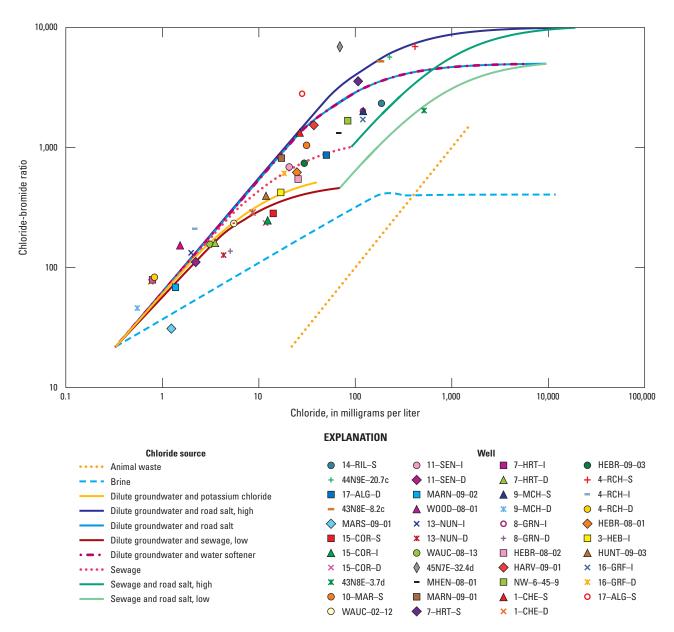


Figure 12. Chloride-bromide ratio plotted against chloride concentrations in the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment monitoring wells, McHenry County, Illinois, 2020.

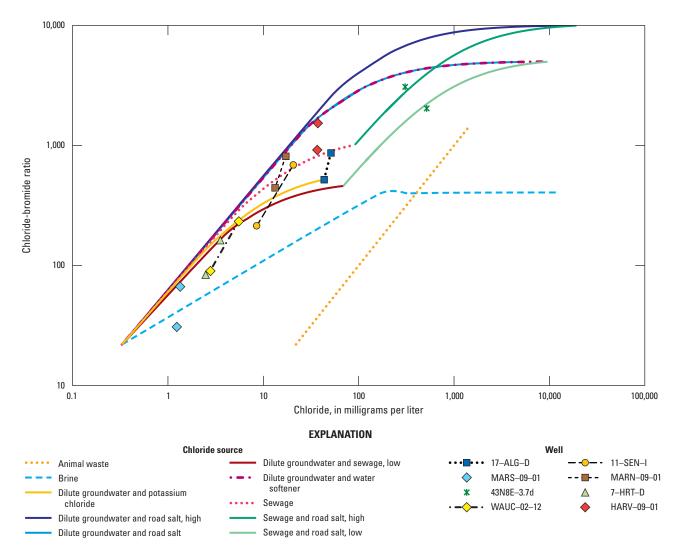


Figure 13. Chloride-bromide ratio plotted against chloride concentration for wells that show shifts between 2010 and 2020 in the McHenry County Groundwater Monitoring Network wells and the National Water-Quality Assessment monitoring wells, McHenry County, Illinois, 2020.

ratio analysis for monitoring well 11–SEN–I indicated a shift from the dilute groundwater and sewage (low) mixing curve to plotting close to the sewage mixing curve (fig. 13). Monitoring wells 11–SEN–I and 11–SEN–D are near a septic field for a municipal building. Chloride and sulfate concentrations slightly increased since 2010 in 11–SEN–I. Increased sulfate concentrations could be indicative of increased microbial activity, possibly from sewage (table 5). Chloride concentrations increased at 11–SEN–I from 8.53 to 20.7 mg/L, and sulfate concentrations increased from 45.9 to 56.8 mg/L (Gahala, 2017; table 5). The shift more towards the sewage mixing curve may be indicative of ongoing and potentially increasing septic-field effect at this location.

Monitoring well HARV–09–01 shifted from near the sewage source curve in 2010 towards the dilute groundwater and water-softener curve in 2020. Other constituents, such as alkalinity, pH, nutrients, major ions, and metals, were reviewed to observe if any changes were consistent with this

change. In general, no differences were statistically significant despite there being slight decreases in nutrient concentrations and slight increases in iron concentrations. This well also was sampled and analyzed for CECs in 2020 because it previously plotted near the sewage mixing curve. The results of the CEC analyses were all less than the reporting level, indicating sewage is not likely affecting the groundwater quality at HARV–09–01. The shift to another source curve could be highlighting the uncertainty of the chloride-bromide ratio analysis, or the source of the groundwater at this location could be from a mixture of sources; future sampling events may clarify this result.

The chloride-bromide ratios for 17–ALG–D shifted from plotting on the dilute groundwater and potassium chloride curve to plotting on the sewage curve. The surroundings of this well are split between agriculture and urban land (Gahala, 2017). A few other constituents that either increased or decreased (not significantly) could explain the shift from the

dilute groundwater and potassium chloride curve to the sewage curve. Manganese nearly doubled in concentrations from 6 to 11.4 mg/L, magnesium increased from 28.6 to 36.4 mg/L, sulfate decreased from 3.28 to 0.24 mg/L, and sodium decreased from 49.4 to 28.3 mg/L (Gahala, 2017; tables 5 and 6). Nutrients such as ammonia and orthophosphate also decreased. This well is a candidate for future sampling for CECs to monitor for any continued changes or effects.

NAWQA monitoring well 43N8E-3.7d shifted from between the dilute groundwater and water softener and the sewage and road salt (high) curves in 2010 to nearly on the curve for the sewage and road salt (low) mixing curve in 2020. Chloride concentrations increased from 149 mg/L in 2010 to 517 mg/L in 2020, sodium concentrations increased from 57.3 mg/L in 2010 to 170 mg/L in 2020, and bromide concentrations increased from 0.061 mg/L in 2010 to 0.254 mg/L in 2020 (Gahala, 2017; table 5). Other constituents also have changed with increases in sodium, iron, manganese, magnesium, potassium, barium, and dissolved solids, all of which indicate a potential effect on water quality and subsequent shift in geochemistry. This well is in an urban area and could receive anthropogenic effects from leaky sewers and road salt infiltration. Previous investigations that analyzed for volatile organic compounds did not indicate detectable levels of contaminants.

MARN-09-01 shifted from between the curves of sewage and dilute groundwater and potassium chloride in 2010 to nearly on the curve for the dilute groundwater and watersoftener curve in 2020. Few constituents indicated increases in concentrations except for chloride, calcium, sulfate, and iron, which were slight increases. The land use surrounding MARN-09-01 is agriculture and urban (Gahala, 2017). This well also was selected for analysis of CECs because the orthophosphate detected in this well in 2010 was the highest of any of the other wells sampled that year. The 2020 orthophosphate results indicated a reduction from 0.152 mg/L in 2010 to 0.035 mg/L in 2020 (table 7). It was previously concluded that this well may have been under the effect of a previous septic field from a former house near the monitoring well (Gahala, 2017). The CEC analysis detected nicotine (estimated 35.2 ng/L; table 8). The compound triethyl citrate was detected, but this compound was detected in the equipment blank (associated with 10–MAR–S); therefore, it is a suspected detection but not confirmed (appendix 1, table 1.2; Gahala and Gruhn, 2022). Nicotine is a likely detection because this constituent was not detected in field blanks or equipment blanks. The source of the chloride in 2010 plotted near the sewage mixing curve, but in 2020, the source is indicated as being affected by water softeners. In either case, an anthropogenic source of the chloride is implicated.

Summary

McHenry County, Illinois, obtains most its drinking water from the shallow sand and gravel aquifers (groundwater). To evaluate this groundwater resource, the U.S. Geological Survey collected water-quality samples from 41 of the 42 monitoring wells in the McHenry County Groundwater Monitoring Network, 4 monitoring wells from the National Water-Quality Assessment program, and 2 surface-water sites for analysis of nutrients, major ions, and trace metals. Additionally, a subset of 12 monitoring wells and 3 surface-water sites was sampled and analyzed for contaminants of emerging concern (CECs). The 2020 results were summarized and compared to the 2010 water-quality data to identify any differences in the water quality of the sand and gravel aquifers in McHenry County, Illinois, which is the primary source of drinking water.

Health-based benchmarks were exceeded for arsenic (about 24 percent; 11 of 45 monitoring wells), sodium (40 percent, 18 of 45), and manganese (about 2 percent, 1 of 45). Aesthetically based benchmarks were exceeded for dissolved solids (about 29 percent, 13 of 45), chloride (about 4 percent, 2 of 45), iron (about 87 percent, 39 of 45), and manganese (about 29 percent, 13 of 45). CECs were detected at low or estimated concentrations in 8 of the 12 (about 67 percent) monitoring wells analyzed.

In addition to sampling the groundwater monitoring wells, three surface-water-quality monitoring sites also were sampled and analyzed for pharmaceuticals and wastewater indicator compounds to provide a preliminary assessment of the presence of CECs in the surface waters. CECs were detected in all three of the surface-water-quality monitoring samples collected, and wastewater indicator compounds were more prevalent and more frequently detected than pharmaceutical compounds. These results provided a cursory understanding of the presence of CECs in surface waters and do not constitute a robust analysis of sources, seasonality, range of concentrations, persistence, or effects.

The 2020 groundwater-quality results had measurements of field properties and concentrations of major ions, trace metals, and nutrients that were consistent with 2010 results with statistically significant increases for calcium, magnesium, and silica and decreases for aluminum, ammonia, arsenic, barium, bromide, calcium, molybdenum, phosphate, specific conductance, sulfate, and dissolved solids. Increases generally were detected in the intermediate and deep parts of the sand and gravel aquifer, and decreases were detected in the shallow part of the sand and gravel aquifers. The mixed distribution of increases and decreases among the various constituents and aquifer-depth groups could be reflecting dissolution and

mobility of some of the redox sensitive constituents and dilution of some constituents in the shallow aquifer depths. These changes may be attributed to a combination of stable population of the past decade (2010–20), to land-use management practices, and to the recent wet years of 2017 through 2019 causing a dilution of the major ions in the shallow parts of the aquifer.

In addition to comparing the 2020 water-quality data to the 2010 data, an analysis for the source of chloride ions in the groundwater was completed by applying the 2020 chloride and bromide concentrations to the chloride-bromide ratio analysis. The results indicate the sources of chloride remained similar to 2010, where many of the chloride concentrations in the groundwater are from road salt applications. A few of the wells indicated some shift from road salt towards sewage or from sewage towards road salt. The change in potential sources may indicate a change in the mixture of sources within that part of the aquifer.

A total of 172 CECs were analyzed for in the groundwater monitoring wells, and only 5 percent of the CECs were detected in low or estimated concentrations, indicating a lack of widespread presence of CECs and supporting the overall good quality of the shallow groundwater resource. CECs were identified in about 66 percent (8 of 12) of the monitoring wells sampled and in all 3 surface-water samples. CECs identified in 10-MAR-S, MARN-09-01, 9-MCH-S, 8-GRN-D, and HEBR-09-03 are consistent with the source water identified in the chloride-bromide ratio analysis. The low level and often estimated concentrations not only indicate the vulnerability of groundwater with respect to sewage and septic effluents but also highlight the potential capacity of the aquifer to degrade or mitigate some the detected CECs. Surface waters indicated the highest concentrations and most frequent detection of CECs. These locations were selected based on proximity to wastewater-treatment plant effluent, and the detections highlight the limits of the treatment systems.

The suitability of the groundwater for drinking water is generally good; however, some naturally occurring minerals, such as arsenic, manganese, and iron, would likely need some filtration or treatment system to reduce the potential toxicity (arsenic) and taste objections (iron and manganese). The suitability of surface water for drinking water is generally good. However, the longer-term effects of chronically low concentrations of CECs are unknown and the sources, range of concentration, and persistence of the CECs would benefit from additional studies.

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Appendix 1. Quality Assurance and Quality Control of Water-Quality Results

Quality Assurance and Quality Control for Major Ions, Nutrients, Metals, and Trace Elements

Quality assurance and quality control (QAQC) of the water-quality results were completed by collecting blank (field and equipment) and replicate (duplication of the environmental sample for analysis of variability) samples about once per week during the sampling period between June and July of 2020 to evaluate variability and control bias in analytical results (table 1.1; Gahala and Gruhn, 2022). One equipment blank was collected for each sampling pump type before any field sampling. A total of four field equipment blanks were collected following the decontamination procedures, three for the analysis of nutrients, major ions, and metals and one for the analysis of pharmaceuticals and wastewater indicator compounds (WICs). A total of five replicate samples were collected to verify the laboratory results: two for nutrients, major ions, and metals/trace elements; one for pharmaceuticals and WICs; one for pharmaceuticals; and one for WICs.

Pre-field equipment blanks were collected before any field use or mobilization. Pumps and associated tubing were washed with liquid Alconox solution, rinsed with tap water, and rinsed a third time with de-ionized water. Inorganic blank water was then passed through the pumps and tubing and the filter, and blank samples were collected in their respective prelabeled sample bottles, preserved with nitric acid, and placed in an ice-bath cooler and sent overnight to the U.S. Geological Survey National Water Quality Laboratory in Denver, Colorado, for analysis. Field blanks were collected to test for the quality of decontamination procedures. Field blank samples were collected in a similar manner after the collection of the environmental and replicate samples, following the decontamination procedures and addition of a new filter for filtered samples.

The pre-field equipment blank from the Fultz pump detected low concentrations of several constituents. Cobalt, copper, lead, nickel, zinc, antimony, and bromide were detected at low concentrations but were within the range of concentrations detected in the environmental samples. Calcium, fluoride, silica, barium, and strontium also had some low-level detections but were below the range of concentrations detected in the environmental samples. Subsequent field rinse equipment blanks from the Fultz pump did not have detections greater than the reporting level for lead, and antimony, and therefore, the environmental sample results are considered acceptable. One field-rinse equipment blank sample (7–HRT–D) detected bromide just above the reporting limits (less than [<] 0.010 and <0.020) of 0.011 mg/L and

is within the range of some of the environmental samples. This detection could be attributed to the fluctuating reporting limit for many of the results and may increase the uncertainty of concentrations less than 0.020 mg/L. It is possible that the initial pre-field equipment blank had remnant tap water and (or) de-ionized water inadvertently collected in the samples. Equipment blanks from the Grundfos pump detected many of the same constituents as detected in the Fultz pump, and therefore, detections of some of these constituents with similar low concentrations in the environmental samples for 17–ALG–D and WAUC-02-12 collected with the Grundfos are estimated. These detections within the blanks, regardless of the pump, indicate the source is potentially from tubing or laboratory interference or tap/de-ionized rinse water. Of the three field blanks, low-level detections were noted for several constituents, including many of the same constituents detected in the blank samples. Except for copper, cobalt, and zinc, no other detections were within the range of the concentrations detected in the environmental samples. Results with detections in the pre-field equipment blanks and field blanks are estimated. Statistical analyses for cobalt, copper, nickel, and zinc were not evaluated (table 1.1).

Replicate samples were collected to assess the reproducibility of the sample-collection procedures and laboratory analysis to provide a measure of variability. Replicate samples were collected immediately after the environmental sample. Of the 34 comparisons between major ions, trace metals, and nutrients analyzed, 4 constituents (orthophosphate, cobalt, copper, and bromide) had differences greater than 10 percent. Cobalt and copper were routinely detected in blank samples and are excluded from further analysis or considerations. The differences in orthophosphate greater than 10 percent were because of low concentrations near the method detection level, but the environmental results are reliable enough to determine water types and would not alter the determination of whether a constituent exceeded water-quality standards.

Bromide had larger variability in detected concentration results (23- and 37-percent difference) between replicates and environmental samples. The replicate value for bromide in 9-MCH–S was 0.036 microgram per liter ($\mu g/L$), and the replicate chloride concentration was 123 milligrams per liter (mg/L). The environmental sample for 9–MCH–S had a detected bromide concentration of 0.057 $\mu g/L$ and the chloride concentration of 121 mg/L was consistent with the replicate concentration. The differences in detection of bromide concentration affected where the environmental and replicate results plotted on the chloride-bromide source curves. The 9–MCH–S environmental sample chloride-bromide ratio plotted ambiguously between the "sewage and road salt high curve" and "dilute groundwater and softener" curves. Previous samples from 2010 through 2015 showed 9–MCH–S chloride-bromide

Table 1.1. Chemical characteristics of selected quality-control samples collected in McHenry County, Illinois, June and July 2020.

 $[mg/L, milligram \ per \ liter; N, nitrogen; <, less than reporting level; NR, not reported; P, phosphorous; <math>\mu g/L, microgram \ per \ liter; 180C, laboratory \ dissolved solids \ dried at 180 \ degrees \ Celsius; NA, not available]$

Field identifier													
	10-MAR-S	WAUC-02-12	43N8E-8.2c	45N8E-17.7h1	45N6E-23.7d3	45N8E-17.7h1	45N6E-23.7d3	45N8E-17.7h1	45N6E-23.7d3				
				(9-MCH-S)	(7–HRT–D)	(9-MCH-S)	(7–HRT–D)	(9-MCH-S)	(7–HRT–D)				
					Sample type								
Constituent	Fultz pump equipment blank	Grundfos pump equipment blank	Field blank	Field blank	Field blank	Replicate	Replicate	Environmental sample	Environmental sample				
		Sample date (month/day/year) and time											
	6/8/2020 13:30	6/29/2020 16:00	6/22/2020 16:00	7/20/2020 12:00	7/13/2020 11:30	7/20/2020 14:00	7/13/2020 11:30	7/20/2020 14:00	7/13/2020 11:30				
Ammonia (mg/L as N)	< 0.01	0.01	< 0.01	< 0.01	< 0.01	< 0.01	1.01	< 0.01	0.97				
Nitrite (mg/L as N)	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	0.002	< 0.001	< 0.001	< 0.001				
Nitrate + nitrite (mg/L as N)	< 0.04	< 0.04	0.12	< 0.040	< 0.04	5.42	< 0.04	5.33	< 0.04				
Orthophosphate (mg/L as P)	< 0.004	< 0.004	0.005	< 0.004	< 0.004	0.011	0.006	< 0.004	0.011				
Calcium (mg/L)	0.48	0.17	0.07	0.09	0.09	108	77.6	109	75.6				
Magnesium (mg/L)	< 0.01	0.07	0.02	0.01	0.02	51.4	40.5	50.6	40.7				
Sodium (mg/L)	< 0.4	0.45	< 0.4	< 0.4	< 0.4	84.3	8.60	84.4	8.65				
Potassium (mg/L)	< 0.30	< 0.30	< 0.30	< 0.30	< 0.30	2.02	1.27	1.84	1.35				
Chloride (mg/L)	< 0.02	0.02	< 0.02	< 0.02	< 0.02	123	3.52	121	3.53				
Sulfate (mg/L)	< 0.02	0.045	< 0.02	< 0.02	< 0.02	29.75	0.28	29.3	0.26				
Fluoride	0.05	< 0.01	< 0.01	0.02	< 0.01	0.14	0.32	0.15	0.30				
Silica	0.09	< 0.05	< 0.05	0.05	< 0.05	18.0	22.1	17.9	22.1				
Arsenic (µg/L)	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.2	9.96	0.21	9.72				
Barium ($\mu g/L$)	2.06	2.25	3.59	2.8	3.15	62.4	44.5	62.2	44.6				
Beryllium ($\mu g/L$)	< 0.01	< 0.01	< 0.01	< 0.01	< 0.01	< 0.02	< 0.01	< 0.02	< 0.01				
Boron ($\mu g/L$)	2.0	<2.0	<2	<2	<2	64.6	39.9	64.1	40.5				
Cadmium (µg/L)	< 0.03	< 0.03	< 0.03	< 0.03	< 0.03	< 0.06	< 0.03	< 0.06	< 0.03				
Chromium (µg/L)	< 0.5	0.5	< 0.5	< 0.5	< 0.5	<1	< 0.5	<1	< 0.5				
Cobalt (μ g/L)	0.32	0.074	< 0.03	0.08	< 0.03	0.09	0.1	0.081	0.09				
Copper (µg/L)	1.9	6.1	10.4	3	3.8	4.2	0.7	5.3	0.4				
Iron (mg/L)	<10	<10	<10	<10	<10	<10	2,175	<10	2,176				
Lead (µg/L)	0.06	< 0.02	< 0.02	0.03	< 0.02	< 0.04	< 0.02	< 0.04	< 0.02				
Manganese ($\mu g/L$)	0.93	0.28	< 0.2	0.24	< 0.2	0.93	17.5	0.91	17.3				
$Molybdenum (\mu g/L)$	< 0.05	0.08	< 0.05	< 0.05	< 0.05	0.63	0.97	0.65	0.97				
Nickel (µg/L)	0.30	0.40	0.27	< 0.20	0.20	0.60	< 0.20	0.67	< 0.20				
Silver (µg/L)	<1	<1	<1	<1	<1	<2	<1	<2	<1				
Strontium (µg/L)	0.6	5.5	2.2	2.9	3.6	189	705	190	700				

Table 1.1. Chemical characteristics of selected quality-control samples collected in McHenry County, Illinois, June and July 2020.—Continued [mg/L, milligram per liter; N, nitrogen; <, less than reporting level; NR, not reported; P, phosphorous; μg/L, microgram per liter; 180C, laboratory dissolved solids dried at 180 degrees Celsius; NA, not available]

	Field identifier										
	10-MAR-S	WAUC-02-12	43N8E-8.2c	45N8E-17.7h1	45N6E-23.7d3	45N8E-17.7h1	45N6E-23.7d3	45N8E-17.7h1	45N6E-23.7d3		
	IU-IVIAN-3	VVAUG-02-12	431V0E-0.2C	(9-MCH-S)	(7–HRT–D)	(9-MCH-S)	(7-HRT-D)	(9-MCH-S)	(7–HRT–D)		
					Sample type						
Constituent	Fultz pump equipment blank	Grundfos pump equipment blank	Field blank	Field blank	Field blank	Replicate	Replicate	Environmental sample	Environmental sample		
	Sample date (month/day/year) and time										
	6/8/2020 13:30	6/29/2020 16:00	6/22/2020 16:00	7/20/2020 12:00	7/13/2020 11:30	7/20/2020 14:00	7/13/2020 11:30	7/20/2020 14:00	7/13/2020 11:30		
Zinc (µg/L)	9.1	<2	4.3	3.9	2.1	<4	<2	<4	<2		
Antimony (µg/L)	0.09	< 0.06	< 0.06	< 0.06	< 0.06	< 0.12	< 0.06	< 0.12	< 0.06		
Aluminum (µg/L)	<3	<3	<3	<3	<3	<6	<3	<6	<3		
Lithium (µg/L)	< 0.15	< 0.15	< 0.15	< 0.15	< 0.15	3.34	2.46	3.37	2.44		
Selenium (µg/L)	< 0.05	< 0.05	< 0.05	< 0.05	< 0.05	0.5	< 0.05	0.48	< 0.05		
Uranium (µg/L)	< 0.03	< 0.03	< 0.03	< 0.03	< 0.03	0.89	< 0.03	0.89	< 0.03		
Bromide mg/L	0.010	< 0.010	< 0.020	< 0.010	0.011	0.036	0.027	0.057	0.022		
Laboratory dissolved solids, dry, 180C, (mg/L)	424	<26.8	<27.8	NA	NA	684	360	676	NA		

ratios plotting near and along the "sewage and road salt, high" curve (Gahala, 2017). The 2020 environmental and replicate chloride-bromide ratio results shifted the datapoints away from the "sewage and road salt, high" curve towards and near the "dilute groundwater and softener" curve. The environmental sample for 7–HRT–D plotted near the dilute groundwater and sewage (low) curve, whereas the replicate was slightly farther away from that curve and slightly closer to the 2010 data point where it was between the curves of dilute groundwater and sewage (low) and the curve for brine within the dilute range. The shift was not substantially different enough to change the interpretation of the source.

Quality Assurance and Quality Control for Contaminants of Emerging Concern

For the contaminants of emerging concern (CECs), five QAQC samples were collected (one field blank, three replicates [one full, one for only pharmaceuticals, and one for only WICs], and one equipment blank). The Fultz pump was the only pump used for the collection of samples for the analysis of organic compounds (pharmaceuticals and WICs). After the equipment decontamination steps of wash, tap water rinse, and de-ionized water rinse, laboratory grade methanol was run through the Fultz pump and tubing and followed by certified organic blank water. After the methanol wash passed through the tubing, the certified organic blank water was then collected in the appropriate sample containers for the analysis of pharmaceuticals and WICs. Samples collected for pharmaceutical analysis were filtered with a 25-millimeter disposable filter. Field blank samples were collected in a similar manner after the collection of the environmental and replicate samples, following the decontamination procedures, and a new 25-millimeter filter was used. Samples were placed in an icebath cooler and sent overnight to the National Water Quality Laboratory for analysis. All detections, reported and estimated, are presented in table 1.2.

Two replicate samples were collected for the analysis of pharmaceuticals and WICs. Of the 155 comparisons for pharmaceuticals and 74 comparisons for WICs, 8 differences were detected. Two constituents, prometon and nicotine, had detections in the environmental sample and the replicate and no detections in the equipment blank or field blank. Other constituents (atrazine, less than 0.16 nanogram per liter [ng/L]; carbamazepine, 2.9 ng/L) had detections in either the sample or the replicate, but not both; and the remaining constituents had detections in either the equipment blank or field blank. Based on these results, prometon, nicotine, atrazine, and carbamazepine are considered as likely detections in the groundwater monitoring well samples.

Results from groundwater wells were variable, and some detections in the QAQC samples were not detected in the primary environmental samples. This variability was represented in the results of the analysis of the environmental sample and the associated replicate QAQC samples where a detection in one sample was not detected in the other in every instance. In other words, a detection occurred in either the environmental sample or the replicate sample, but not in both. This pattern was detected in 9-MCH-S sample results for carbamazepine and atrazine. Other types of QAQC samples (equipment and field blanks) also unexpectedly resulted in detections of several CECs. This pattern could potentially be caused by ineffective de-ionization of water at the local laboratory where equipment for sampling is cleaned and stored. If water was not being completely de-ionized, some of these CECs were potentially being detected in tap water that is not fully processed through de-ionization. Nicotine was detected in six monitoring wells, including the three hypothesized background monitoring wells. Detections were less than the reporting level of 80 ng/L and ranged from 23.2 to 38.6 ng/L. Replicate results for nicotine were less than 10-percent different (6 percent). N,N-Diethyl-m-toluamide (DEET) also was detected in every environmental sample and every QAQC sample despite personnel adhering to field sampling protocol of avoiding the use of any repellent (or pharmaceutical and personal care products). The cross contamination is suspected to be from the laboratory machines and the ubiquitous and persistent nature of DEET. The results for DEET are therefore not considered in the tally of detections or any further discussion.

 Table 1.2.
 Detections of contaminants of emerging concern in quality assurance and quality control samples.

[CEC, contaminant of emerging concern; pharm, pharmaceuticals; WIC, wastewater indicator compound; OOT, organics other; ng/L, nanogram per liter; E, estimated; <, less than reporting level; —, no data; µg/L, microgram per liter; OPE, organics pesticides]

CEC type	Group	Constituent	Units	10–MAR–S equipment blank	10–MAR–S replicate (pharm only)	7–HRT–S replicate (WIC only)	9–MCH–S replicate	9–MCH–S field blank
Pharm	OOT	Hexamethylenetetramine	ng/L	E21.1	<110	_	<110	<110
Pharm	OOT	Carbamazepine	ng/L	<11	<11		2.9	<20
Pharm	OOT	Nicotine	ng/L	<80	<80	_	24.6	<80
WIC	OOT	Diethyl phthalate	μg/L	< 0.4	_	< 0.4	0.05	0.17
WIC	OOT	Phenanthrene	μg/L	0.01	_	< 0.02	< 0.02	0.02
WIC	OOT	Pyrene	μg/L	< 0.02	_	< 0.02	< 0.02	0.01
WIC	OOT	Naphthalene	μg/L	< 0.02	_	< 0.02	< 0.02	0.06
WIC	OOT	Bis(2-ethylhexyl) phthalate	μg/L	1.9	_	<3	<2	1.39
WIC	OOT	Cholesterol	μg/L	E0.13	_	<1.9	<1.6	E0.08
WIC	OOT	4-Nonylphenol (all Isomers)	$\mu g/L$	E0.15	_	<1.6	<1.6	<1.6
WIC	OOT	Triethyl citrate	$\mu g/L$	E0.05	_	< 0.16	< 0.16	< 0.16
WIC	OOT	Isopropylbenzene	$\mu g/L$	E0.03	_	< 0.04	< 0.04	E0.08
WIC	OOT	1-Methylnaphthalene	$\mu g/L$	< 0.04	_	< 0.04	< 0.04	0.01
WIC (pesticide)	OPE	N,N-Diethyl-m-toluamide (DEET)	$\mu g/L$	E0.14	_	0.04	0.06	0.17
WIC (pesticide)	OPE	Prometon	ng/L	< 0.16	_	< 0.16	0.04	< 0.16
WIC (pesticide)	OOT	Atrazine	μg/L	< 0.16	_	< 0.16	3.29	< 0.16
Pharm (pesticide)	OPE	Atrazine	ng/L	< 0.16	5.03	< 0.16	< 0.16	< 0.16

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